

MCRA Documentation

Release 9

Biometris, Wageningen University and Research

Apr 26, 2021

CONTENTS

1	Abou	it the to	oolbox	3
	1.1	Data a	nd calculation model	3
		1.1.1		3
		1.1.2	•	4
		1.1.3		5
		1.1.4	11	5
				6
				7
				, 7
	1.2	Data r		, 7
	1.2	1.2.1		, 7
		1.2.1		, 8
		1.2.2	1 2	9
	1.3			9
	1.5	1.3.1	1	
		1.3.1	Workspace browser	
			Workspace overview page 12	
		1.3.3	Action area	
		1.2.4	Scoping: entity selection	
		1.3.4	Standard action area	
			Creating a standard action	
			Standard action reports	
		1 2 5	Converting a standard action to a normal action 19	
		1.3.5	Action zip files	9
2	Mod	ules	2	1
	2.1		ry entity modules	
		2.1.1	Effects	
			Effects data formats	
			Effects calculation	
			Effects settings	
			Effects as data	
		2.1.2	Foods	
		2.1.2	Foods data formats	
			Foods as data	
			Food coding systems 28	
			FoodEx2	
			Food hierarchies 30	
			Food unit weights 3'	
		2.1.3		
		2.1.3	1	
			1	
		014	Populations as data	
		2.1.4	Responses	
			Responses data formats	J

		Responses settings	0
		Responses as data	1
	2.1.5		1
			1
		Substances settings	12
		Substances as data	13
	2.1.6	Test systems	13
		Test systems data formats	13
		Test systems as data	4
2.2	Consu	mption modules	4
	2.2.1	Consumptions	15
		Consumptions data formats	15
		Consumptions calculation	19
		Consumptions settings	50
			50
			51
	2.2.2		51
			51
		1	52
	2.2.3	1	52
			52
			53
			53
	2.2.4		53
	2.2.1		53
			54
			55
			56
			56
2.3	Occur		56
2.5			0
	231	Concentration distributions 5	6
	2.3.1		56 56
	2.3.1	Concentration distributions data formats	56
		Concentration distributions data formats5Concentration distributions as data5	56 57
	2.3.12.3.2	Concentration distributions data formats 5 Concentration distributions as data 5 Concentration limits 5	56 57 57
		Concentration distributions data formats 5 Concentration distributions as data 5 Concentration limits 5	56 57 57 58
	2.3.2	Concentration distributions data formats 5 Concentration distributions as data 5 Concentration limits 5 Concentration limits data formats 5 Concentration limits data formats 5 Concentration limits as data 5 Concentration limits as data 5	56 57 57 58 59
		Concentration distributions data formats5Concentration distributions as data5Concentration limits5Concentration limits data formats5Concentration limits as data5Concentration limits as data5	56 57 57 58 59 59
	2.3.2	Concentration distributions data formats5Concentration distributions as data5Concentration limits5Concentration limits data formats5Concentration limits as data5Concentration limits as data5Concentration limits as data5Concentration limits as data5Concentration models5Concentration models calculation5	56 57 57 58 59 59 59
	2.3.2	Concentration distributions data formats5Concentration distributions as data5Concentration limits5Concentration limits data formats5Concentration limits as data5Concentration limits as data5Concentration limits as data5Concentration models5Concentration models calculation5Concentration models settings6	56 57 58 59 59 59 59
	2.3.2	Concentration distributions data formats5Concentration distributions as data5Concentration limits5Concentration limits data formats5Concentration limits as data5Concentration limits as data5Concentration limits as data5Concentration models5Concentration models calculation5Concentration models settings6Concentration models tiers6	56 57 57 58 59 59 59 59 57 59
	2.3.2	Concentration distributions data formats5Concentration distributions as data5Concentration limits5Concentration limits data formats5Concentration limits as data5Concentration limits as data5Concentration limits as data5Concentration models5Concentration models calculation5Concentration models settings6Concentration models tiers6Concentration models uncertainty7	56 57 58 59 59 59 57 59 57
	2.3.2	Concentration distributions data formats5Concentration distributions as data5Concentration limits5Concentration limits data formats5Concentration limits as data5Concentration limits as data5Concentration models5Concentration models calculation5Concentration models settings6Concentration models uncertainty7Calculation of concentration models7	56 57 57 58 59 59 59 59 59 57 59 75
	2.3.2	Concentration distributions data formats5Concentration distributions as data5Concentration limits5Concentration limits data formats5Concentration limits as data5Concentration limits as data5Concentration models5Concentration models calculation5Concentration models settings6Concentration models uncertainty7Calculation of concentration models7Concentration models7	56 57 57 58 59 59 59 57 59 57 59 57 57 57 6
	2.3.2	Concentration distributions data formats5Concentration distributions as data5Concentration limits5Concentration limits data formats5Concentration limits as data5Concentration limits as data5Concentration models5Concentration models calculation5Concentration models settings6Concentration models uncertainty7Calculation of concentration models7Concentrations7Concentrations7	567789999759757676
	2.3.2	Concentration distributions data formats5Concentration distributions as data5Concentration limits5Concentration limits data formats5Concentration limits as data5Concentration limits as data5Concentration models5Concentration models calculation5Concentration models settings6Concentration models uncertainty7Calculation of concentration models7Concentrations7Conc	56778999575975767633
	2.3.2	Concentration distributions data formats5Concentration distributions as data5Concentration limits5Concentration limits data formats5Concentration limits as data5Concentration models5Concentration models calculation5Concentration models settings6Concentration models uncertainty7Calculation of concentration models7Concentrations7Concentrations data formats7Concentrations data formats7Concentrations settings8Concentrations settings8	56 57 58 59 59 59 59 57 57 6 75 76 76 33 88
	2.3.2	Concentration distributions data formats5Concentration distributions as data5Concentration limits5Concentration limits data formats5Concentration limits as data5Concentration models5Concentration models calculation5Concentration models settings6Concentration models uncertainty7Calculation of concentration models7Concentrations7Concentrations data formats7Concentrations data formats7Concentrations data formats7Concentrations data formats7Concentrations calculation8Concentrations settings8Concentrations tiers9	5677899997597576783801
	2.3.2	Concentration distributions data formats5Concentration distributions as data5Concentration limits5Concentration limits data formats5Concentration limits as data5Concentration models5Concentration models calculation5Concentration models settings6Concentration models tiers6Concentration models uncertainty7Calculation of concentration models7Concentrations data formats7Concentrations data formats7Concentrations data formats7Concentrations data formats7Concentrations data formats7Concentrations settings8Concentrations tiers9Concentrations uncertainty9	56778999979757676389192
	2.3.2 2.3.3 2.3.4	Concentration distributions data formats5Concentration distributions as data5Concentration limits5Concentration limits data formats5Concentration limits as data5Concentration models5Concentration models5Concentration models calculation5Concentration models settings6Concentration models uncertainty7Calculation of concentration models7Concentrations7Concentrations7Concentrations data formats7Concentrations data formats8Concentrations settings8Concentrations uncertainty9Concentrations as data9	56778999575767638919292
	2.3.2	Concentration distributions data formats5Concentration distributions as data5Concentration limits5Concentration limits data formats5Concentration limits as data5Concentration models5Concentration models calculation5Concentration models settings6Concentration models uncertainty7Calculation of concentration models7Concentrations data formats7Concentrations data formats7Concentrations data formats7Concentrations data formats7Concentrations data formats7Concentrations data formats9Concentrations uncertainty9Concentrations as data9Deterministic substance conversion factors9	5677899979757678380 12222
	2.3.2 2.3.3 2.3.4	Concentration distributions data formats5Concentration distributions as data5Concentration limits5Concentration limits adata formats5Concentration limits as data5Concentration models5Concentration models calculation5Concentration models settings6Concentration models settings6Concentration models uncertainty7Calculation of concentration models7Concentrations data formats7Concentrations data formats7Concentrations data formats7Concentrations data formats7Concentrations adata formats9Concentrations uncertainty9Concentrations as data9Deterministic substance conversion factors9Deterministic substance conversion factors9	567789997975766380122222
	 2.3.2 2.3.3 2.3.4 2.3.5 	Concentration distributions data formats5Concentration distributions as data5Concentration limits5Concentration limits data formats5Concentration limits as data5Concentration models5Concentration models5Concentration models calculation5Concentration models settings6Concentration models uncertainty7Calculation of concentration models7Concentrations adata formats7Concentrations data formats7Concentrations data formats7Concentrations settings8Concentrations settings8Concentrations uncertainty9Concentrations settings9Concentrations uncertainty9Concentrations uncertainty9Concentrations uncertainty9Concentrations uncertainty9Concentrations uncertainty9Deterministic substance conversion factors9Deterministic substance conversion factors as data9Deterministic substance conversion factors as data9	5677899975975767838122223
	2.3.2 2.3.3 2.3.4	Concentration distributions data formats5Concentration distributions as data5Concentration limits5Concentration limits adata formats5Concentration limits as data5Concentration models5Concentration models5Concentration models calculation5Concentration models settings6Concentration models uncertainty7Calculation of concentration models7Concentrations7Concentrations data formats7Concentrations data formats7Concentrations settings8Concentrations uncertainty8Concentrations settings9Concentrations as data9Deterministic substance conversion factors9Deterministic substance conversion factors as data9Focal food concentrations9	56778999757576783812222233
	 2.3.2 2.3.3 2.3.4 2.3.5 	Concentration distributions data formats5Concentration distributions as data5Concentration limits5Concentration limits adata formats5Concentration limits as data5Concentration models5Concentration models5Concentration models calculation5Concentration models settings6Concentration models tiers6Concentration models uncertainty7Calculation of concentration models7Concentrations7Concentrations data formats7Concentrations settings8Concentrations uncertainty9Concentrations as data9Deterministic substance conversion factors9Deterministic substance conversion factors as data9Focal food concentrations9Focal food concentrations9	567789997597576738912222334
	 2.3.2 2.3.3 2.3.4 2.3.5 	Concentration distributions data formats5Concentration distributions as data5Concentration limits5Concentration limits data formats5Concentration limits as data5Concentration models5Concentration models calculation5Concentration models settings6Concentration models uncertainty7Calculation of concentration models7Concentrations7Concentrations7Concentrations data formats7Concentrations as data9Deterministic substance conversion factors as data9Focal food concentrations9Focal food concentrations9Focal food concentrations settings9Focal food concentrations settings9	56778999757566380122223344
	 2.3.2 2.3.3 2.3.4 2.3.5 2.3.6 	Concentration distributions data formats5Concentration distributions as data5Concentration limits5Concentration limits data formats5Concentration limits as data5Concentration models5Concentration models5Concentration models calculation5Concentration models settings6Concentration models liters6Concentration models uncertainty7Calculation of concentration models7Concentrations7Concentrations data formats7Concentrations data formats7Concentrations data formats7Concentrations as data9Concentrations settings9Deterministic substance conversion factors as data9Focal food concentrations settings9Focal food concentrations as data9Focal food concentrations settings9Focal food concentrations as data9Focal food concentrations as data<	56 57 57 58 59 59 59 59 57 56 57 56 57 56 75 75 75 75 75 75 75 75 75 75 75 75 75
	 2.3.2 2.3.3 2.3.4 2.3.5 	Concentration distributions data formats5Concentration distributions as data5Concentration limits5Concentration limits data formats5Concentration limits as data5Concentration models5Concentration models5Concentration models settings6Concentration models users6Concentration models users6Concentration models users6Concentration models users7Concentration models users7Concentrations7Concentrations7Concentrations data formats7Concentrations data formats7Concentrations as data9Concentrations uncertainty9Concentrations settings9Concentrations as data9Peterministic substance conversion factors data formats9Pocal food concentrations settings9Focal food concentrations as data9Focal food concentrations as data </td <td>56 57 57 58 59 59 59 57 55 75 75 75 75 75 75 75 75 75 75 75</td>	56 57 57 58 59 59 59 57 55 75 75 75 75 75 75 75 75 75 75 75

	Food extrapolations as data
2.3.8	Modelled foods
	Modelled foods calculation
	Modelled foods settings
	Calculation of modelled foods
2.3.9	Occurrence frequencies
	Occurrence frequencies data formats
	Occurrence frequencies calculation
	Occurrence frequencies Settings
	Occurrence frequencies as data
	Calculation of occurrence frequencies
2.3.10	Occurrence patterns
	Occurrence patterns data formats
	Occurrence patterns calculation
	Occurrence patterns settings
	Occurrence patterns tiers
	Occurrence patterns as data
	Calculation of occurrence patterns
2.3.11	Processing factors
	Processing factors data formats
	Processing factors calculation
	Processing factors settings
	Processing factors uncertainty
	Processing factors as data
2.3.12	Single value concentrations
	Single value concentrations data formats
	Single value concentrations calculation
	Single value concentrations settings
	Single value concentrations as data
	Calculation of single value concentrations
2.3.13	Substance authorisations
	Substance authorisations data formats
	Substance authorisations as data
2.3.14	Substance conversions
	Substance conversions data formats
	Substance conversions as data
2.3.15	Total diet study sample compositions
	Total diet study sample compositions data formats
	Total diet study sample compositions as data
2.3.16	Unit variability factors
	Unit variability factors data formats
	Unit variability factors as data
Expos	ure modules
2.4.1	Consumptions by modelled food
	Consumptions by modelled food calculation
	Consumptions by modelled food
	Calculation of consumptions by modelled food
2.4.2	Dietary exposures
	Dietary exposures calculation
	Dietary exposures settings
	Dietary exposures tiers
	Calculation of dietary exposures
2.4.3	High exposure food substance combinations
	High exposure food substance combinations calculation
	High exposure food substance combinations settings
	Calculation of high exposure food-substance combinations
2.4.4	Exposures
	Exposures calculation

2.4

		F
		Exposures settings
		Calculation of exposures
	2.4.5	Exposure mixtures
		Exposure mixtures calculation
		Exposure mixtures settings
		Calculation of exposure mixtures
	2.4.6	Food conversions
	2.4.0	
		Food conversions calculation
		Food conversion settings
		Calculation of food conversions
	2.4.7	Human monitoring analysis
		Human monitoring analysis calculation
		Human monitoring analysis settings
		Calculation of human monitoring analysis
	2 4 9	
	2.4.8	Human monitoring data
		Human monitoring data data formats
		Human monitoring data settings 198
		Human monitoring data as data
	2.4.9	Non-dietary exposures
		Non-dietary exposures data formats
		Non-dietary exposures settings
		Non-dietary exposures uncertainty
		Non-dietary exposures as data
	2.4.10	
		Single value dietary exposures data formats
		Single value dietary exposures calculation
		Single value dietary exposures settings
		Calculation of single value dietary exposures
2.5	Hozor	1 modules
2.5		
	2.5.1	Active substances
		Active substances data formats
		Active substances calculation
		Active substances settings
		Active substances as data
		Calculation of active substances
	2.5.2	AOP networks
	2.3.2	AOP networks data formats
		AOP networks settings
		AOP networks as data
	2.5.3	Dose response data
		Dose response data data formats
		Dose response data settings
		Dose response data as data
	2.5.4	Dose response models
	2.3.7	Dose response models data formats
		±
		Dose response models calculation
		Dose response models as data
		Calculation of dose response models
	2.5.5	Effect representations
		Effect representations data formats
		Effect representations as data
	2.5.6	Hazard characterisations
	2.3.0	
		Hazard characterisations calculation
		Hazard characterisations settings
		Hazard characterisations as data
		Calculation of hazard characterisations
	2.5.7	Inter-species conversions

		Inter-species conversions data formats
		Inter-species conversions settings
		Inter-species conversions as data 242
	2.5.8	Intra species factors
		Intra-species factors data formats
		Intra species factors settings
		Intra species factors as data
	2.5.9	Points of departure
		Points of departure data formats 243
		Points of departure settings
		Points of departure as data
	2.5.10	Relative potency factors
		Relative potency factors data formats
		Relative potency factors calculation
		Relative potency factors settings
		Relative potency factors as data
		Calculation of relative potency factors
2.6	In-silic	248 co modules
	2.6.1	Molecular docking models
		Molecular docking models data formats
		Molecular docking models as data
	2.6.2	QSAR membership models
		QSAR membership models data formats
		QSAR membership models as data
2.7	Kineti	c modules
2.7	2.7.1	Kinetic models
	2.7.1	Kinetic models data formats
		Kinetic models settings
		Kinetic models as data
		Available kinetic models 255
2.8	Rick n	available kinetie models
2.0	2.8.1	Risks
	2.0.1	Risks calculation
		Risks calculation 2008 Risks settings
		Calculation of risks
	2.8.2	Single value risks
	2.0.2	Single value risks calculation
		Single value risks settings
		Calculation of single value risks
Stand	lard ac	297
3.1		ic cumulative exposure assessment PFAS
3.2		ic cumulative exposure assessment PA
3.3		acute cumulative copositie assessment 1 A
3.4		ute cumulative exposure assessment (2018) Tier I and Tier II
3.5		ronic cumulative exposure assessment (2018) Tier I and Tier II
3.6		Acute Cumulative Risk Assessment No Background
3.7		teatosis from imazalil
3.8		ng prospective risk assessment acute Tier II
3.9		ng prospective risk assessment chronic Tier II
3.10		ng substance prioritisation acute neuro
5.10	1141111	
Exan	nples	307
4.1		lative dietary exposure assessment
	4.1.1	Introduction
	4.1.2	Preparation
	4.1.3	Example 1
	4.1.4	Example 2
		r =

		4.1.5	Example 3	310
	4.2	Aggrega	ate exposure assessment	311
		4.2.1	Introduction	311
		4.2.2	Preparation	311
		4.2.3	Example 1	311
		4.2.4	Example 2	313
	4.3	Hazard	characterisations from PoDs	313
		4.3.1	Introduction	313
		4.3.2	Preparation	313
		4.3.3	Example 1	313
	4.4		mpact estimates	
		4.4.1	Introduction	
		4.4.2	Preparation	
		4.4.3	Example 1	315
	4.5			
		4.5.1	Introduction	
		4.5.2	Preparation	
		4.5.3	Example 1	
		4.5.4	Example 2	318
_	D 11			310
5	Publi	ications i	using MCRA	319
6	I ist (of Symbo	ale	331
U	LISU	JI Symot	7.5	551
7	Appe	ndices		333
	7.1		cumentation	333
	7.2	Munro o	collection	356
	7.3	Unit def	initions	356
		7.3.1	Benchmark response types	356
		7.3.2	Body weight units	358
		7.3.3	Concentration units	358
		7.3.4	Concentration value types	359
		7.3.5	Consumption intake units	359
		7.3.6	Consumption units	359
		7.3.7	Consumption value types	
		7.3.8	Dose response model types	360
		7.3.9	Dose units	360
		7.3.10	Exposure route types	362
		7.3.11	Exposure types	362
		7.3.12	Exposure units	362
		7.3.13	Harvest application types	363
		7.3.14	Hazard characterisation types	364
		7.3.15	Point of departure types	364
		7.3.16	1 1	
		7.3.17	Target dose level types	365
		7.3.18	Test system types	365
	7.4	Transfo	rmations	366
		7.4.1	Box Cox power transformation	366
	7.5	Gauss-H	Iermite	366
		7.5.1		
		7.5.2	One-dimensional Gauss-Hermite integration	
		7.5.3	Two-dimensional Gauss-Hermite integration	
		7.5.4	Maximum likelihood for the LNN model with two-dimensional Gauss-Hermite integration	367
0	Color	ahor		270
8				369 369
	8.1	Contrib	utors to MCRA	309
Bi	bliogra	aphy		371

Reference and user manual for MCRA 9 (version 9.1.32).

ABOUT THE TOOLBOX

Humans are exposed to a mixture of multiple chemicals via food intake, inhalation and dermal contact. The risk to health that may result from this depends on the effects of different chemicals in the mixture and how they combine.

MCRA 9 is the model and data toolbox developed in the EuroMix project (http://www.euromixproject.eu). It implements methods for exposure, hazard and risk assessment, following guidelines from a.o. the Joint Research Centre (JRC) (https://ec.europa.eu/jrc/en) of the European Commission and the European Food Safety Authority (EFSA, https://www.efsa.europa.eu/en). The toolbox should provide computational tools for future risk management decisions on the safety of chemicals in mixtures to be taken by the European Commission and the Codex Alimentarius.

MCRA 9 is a collection of data and models. The system consists of modules that are arranged in eight categories according to a *modular design*. See *Modules overview*.

Each module represents a certain type of data, which can be computed from data provided by other (sub)modules, or the data may be obtained from a dataset selected from the *data repository*. Likewise, each module may be of interest by its own merit, or may just be required as a sub-part of larger calculations. The modular design of the toolbox reveals a network of data and models, and shows how data of types and from various sources can be combined in overarching modules. The most overarching module is *health impact estimates*. The toolbox allows the user to start in any of the modular design for performing calculations.

For each module, an *action* can be created to configure and run the module. For data modules, such as the concentrations module, such an action comprises specifying the dataset, specifying the scope (i.e., foods of interest, substances of interest, etc.), and perhaps specifying specific selections or model settings for data manipulations (e.g., imputation of water concentrations in the concentrations module). For calculation modules, when calculating the data of the module based on other data, configuration of an action comprises specification of the model settings and selection of the calculation inputs, which is data provided by other (sub-)modules. While running an action in the toolbox, the module produces output of its associated data type (which can be used as input for other modules), and a report will be generated of the selected data, the selection and model settings, and the module and all intermediate (i.e., sub-modules) results.

1.1 Data and calculation model

1.1.1 Modular design

The modular design distinguishes between three types of modules: primary entity modules, data modules, and calculation modules. For an overview see *Modules*.

- The primary entity modules are data modules determining the scope of the assessments in the toolbox. That is, in each assessment, the scope specifies the *foods*, *substances*, *effects*, *populations*, *responses*, and/or *test systems* that are of interest.
- The data modules give summaries of the available data which depend on (some of) the primary entities. For example *consumptions* data.
- The calculation modules perform calculations on input data to produce data on another type, as specified by the module name. E.g. the *dietary-exposures* calculation module calculates dietary exposures from consumption and occurrence data. Some calculation modules can also act as a data module, in which case the data are directly specified rather than calculated. Examples are,

the *relative potency factors* module: relative potency factors can be supplied as such (*Data*) or computed based on hazard characterizations (*Compute*); the *single value consumptions* module: Large Portions can be supplied as such (*Data*) or computed based on consumption distribution data of a population (*Compute*).

Risks / Relative potency factors	ac188a89 🕨 🕨	¢
Relative potency factors		0
Use data Compute		

Figure 1.1: Relative potency factors supplied as data or computed based on hazard characterizations.

1.1.2 Nominal run and uncertainty analysis

Within the toolbox two types of simulation runs are distinguished: the nominal run and the uncertainty analysis loop.

The nominal run represents a single simulation which is aimed to compute the most likely, unbiased estimates for the specified model. E.g., when a *dietary exposure assessment* is requested, in the nominal run a single exposure distribution is estimated using nominal values for all data and parameters.

In the *uncertainty analysis* loop, each simulation run is repeated a large number of times. Each run starts with a different scenario using data obtained with bootstrapping, parametric resampling and/or re-calculation of uncertain values. As a result, a large number of uncertain dietary exposure distributions is estimated which are used to estimate uncertainty limits (p5, p95).

Running a nominal run first has the advantage that the user may evaluate these modelling results before doing the final analysis. The model specification of rather complex simulation models and the corresponding output results are evaluated to detect any errors or misspecifications. Possible errors in the data and/or model settings are identified and corrected. In the final exposure assessment, the uncertainty analysis is included and the uncertainty of estimates is assessed.

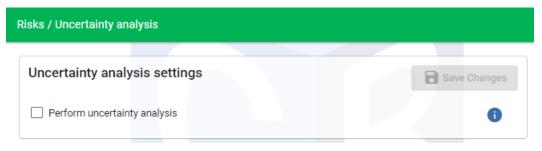


Figure 1.2: Uncertainty analysis settings.

1.1.3 Retain & Refine and tiered approaches

A basic idea of Retain & Refine is that entities (e.g., substances) can be handled in different ways (more or less refined) while still being considered together in the same risk assessment (retain). We refer to such different approaches as tiers.

In the modular design, a tier is defined as a specific set of settings for a module or a group of modules. Tiers differ in many aspects, and there is no single dimension to rank tiers as low vs. high. In risk assessment, typical tiers contrast deterministic to probabilistic approaches, conservative to realistic approaches, approaches using restricted data to approaches using more extensive data, and approaches using different degrees of model complexity. For each of the modules of the toolbox, as many tiers are implemented as considered useful for the practice of risk assessment.

Each calculation in the modular design may involve multiple, nested, calculations of sub-modules. A *risk* (or health impact) assessment builds on an *exposure assessment* and a *hazard assessment*, the exposure assessment builds on a *dietary* and a *non-dietary exposure* assessment, the dietary exposure assessment builds on a *consumption assessment* and an *occurrence assessment*, etc. Tiers can be defined at each node of the assessment network. An example consists of the tiers *TESTI'*, *'EFSA basic optimistic'* and *'EFSA basic pessimistic'* which are defined at the level of a dietary exposure assessment, but include the settings for the corresponding tiers at the level of the concentration model calculator.

Each calculator has as a main output entities that can be specified to have different tiers (tiered entities). For example, in a *hazard assessment*, some substances may be assessed using a tier 'Hazard Dose from dose-response data', other substances may be assessed using a tier 'TTCx100' or 'sample from general NOAEL distribution x100' (which only requires knowledge of the Cramer class of the substance). As another example, in dietary exposure assessment some food-substance combinations may be recognised as risk drivers for which a more complex approach (e.g. probabilistic modelling) is required, whereas a simpler approach (e.g. *deterministic modelling*) may be sufficient for all other food-substance combinations. So in this case the tiered entity is 'food-substance'. A typical risk assessment will start at a tier that is simple to perform for all tiered entities (potential risk drivers). Note that, based on data availability and ease of application, the initial assessment can already include more complex elements, such as probabilistic modelling. If the initial calculations produce risk estimates that do not exclude concern, refinement of the modelling for the perceived risk drivers is useful for checking whether this concern is real.

1.1.4 Uncertainty

Uncertainties arise in different forms in many of the models and data of the toolbox.

- uncertainty in the data values (e.g., uncertain NOAELs, uncertain RPFs, or uncertain processing factors),
- uncertainty due to limited data (e.g., a limited number of food samples),
- uncertainty due to a lack of data (e.g., missing concentration data for some foods/substances or missing processing factors),
- uncertainty of the models, (e.g., due to a lack of detail).

The toolbox offers the following options to handle uncertainty:

- for many types of data, the possibility to provide data including quantifications of uncertainty,
- · imputation methods for filling in missing data in various types of models, and
- a generic uncertainty analysis method that providing uncertainty estimates of the modelling results for many of the modules, which are based on bootstrapping, parametric resampling, and/or re-calculation on all sub-modules for which this is possible.

Uncertainty due to limited sampled data

For some type of data, e.g., processing factors, it is possible to not only provide nominal estimates of the data values, but also to provide quantified estimates of the uncertainties of these values. Occasionally, quantifications of the uncertainties of these estimates are not available. The toolbox provides the possibility to work with both quantified and unquantified uncertainties: include uncertainties in a quantitative uncertainty analysis when available, or, when not available, use nominal estimates, followed by an offline qualitative uncertainty analysis.

Uncertainties of data values are available in different forms. For some data values, uncertainty may be quantified by means of parametric distribution parameters (e.g., *processing factor uncertainties*, or kinetic model instance parameter uncertainties). Alternatively, uncertainty values may be provided in the form of an empirical set of uncertainty values (e.g., *relative potency factor uncertainties*, or *points of departure uncertainties*).

For each data sub-module that has quantified uncertainties, it is optional to include the source of uncertainty in the uncertainty analysis of the main module. Then, when specified, data values are resampled in each *uncertainty analysis cycle*.

The basic *acute exposure* distribution is estimated in a Monte Carlo simulation by combining dietary consumption records (person-days) with sampled residue values. The resulting distribution represents a combination of variability in consumption within the population and between residues in a food lot. Percentiles may be used for further quantification e.g. the median or 99th percentile. Due to the limited size of the underlying data, these outcomes are uncertain. Confidence (or uncertainty) intervals reflect the uncertainty of these estimates, where MCRA uses bootstrap methodology and/or, depending on the available data, parametric methods to estimate the uncertainty.

Empirical method, resampling

The empirical bootstrap is an approach to estimate the accuracy of an outcome. In its most simple, non-parametric form, the bootstrap algorithm resamples a dataset of n observations to obtain a *bootstrap sample* or *resampled set* of again n observations (sampling with replacement, that is: each observation has a probability of 1/n to be selected at any position in the new resampled set). By repeating this process *B* times, one can obtain *B* resampled sets, which may be considered as alternative data sets that might have been obtained during sampling from the population of interest. Any statistic that can be calculated from the original dataset (e.g. the median, the standard deviation, the 99th percentile, etc.) can also be calculated from each of the *B* resampled sets. This generates a *uncertainty distribution* for the statistic under consideration. The uncertainty distribution characterises the uncertainty of the inference due to the sampling uncertainty of the original dataset: it shows which statistics could have been obtained if random sampling from the population would have generated another sample than the one actually observed [[Efron, 1979]], [[Efron et al., 1993]].

Parametric methods

Instead of bootstrapping the observed data, inference about parameters is based on parametric methods. For processing, where factors are specified through a nominal and/or upper value this is the natural choice. For concentration data, where the lognormal model is used to represent less conservative scenario's (EFSA, 2012) [[EFSA, 2012]], the *parametric bootstrap* may be an alternative, especially when data are limited and the empirical bootstrap fails.

According to Cochran's theorem, sample variance $\hat{\sigma}_y^2$ follows a scaled chi-square distribution. In the parametric bootstrap for the *lognormal* distribution, the sample variance $\hat{\sigma}_y^2$ is replaced by a random draw from a chi-square distribution with $n_1 - 1$ degrees of freedom; the sample mean $\hat{\mu}_y$ is replaced by a random draw from a normal distribution with parameters $\hat{\mu}_y$ and $\hat{\sigma}_y^{*2}/n_1$, giving a new set of parameters $\hat{\mu}_y$ and $\hat{\sigma}_y^{*2}$. This is repeated *B* times.

For the *truncated lognormal* and *censored lognormal*, large sample maximum likelihood theory is used to derive new parameters $\hat{\mu}_y$ and $\hat{\sigma}_y^{*2}$. This is repeated B times.

The binomial fraction of non-detects for the *mixture lognormal* and *mixture truncated* distribution is sampled using the beta distribution with uniform priors a = b = 1 (with the *beta* distribution as the empirical Bayes estimator for the binomial distribution). This is repeated B times.

Uncertainty due to missing data

In some cases, data are only available for specific (primary) entities and missing for others. E.g., points of departure (such as NOAELs or BMDs) may only be available for some of the substances of interest.

Uncertainty due to modelling approach

Model uncertainty or uncertainty of model outcomes arise by applying different modelling approaches or applying alternative model assumptions.

Note: TODO

1.2 Data repository

The data used for the *modelling actions* of the toolbox is organised in the data repository. All users have their own (personal) repository folder in which they can *upload* their own data files and organise these in folders and sub-folders to their own preference. In addition, there are shared repository folders that are specifically created for sharing data sources among user groups. Users may be granted access to one or more shared repositories: shared, maintained, and used by multiple users. Shared repositories and their contents are free to use by granted users in their own calculations.

Each data source in the data repository contains data of one or more data groups, indicating the module(s) for which the data set can be used as data. Each primary entity module and data module of the *modular design* is linked to a data group. If a data source is recognized as a data source belonging to a module, then it can be used as a data source of this module in a modelling action. The data groups of a data source are automatically recognized when the data is uploaded, based on whether the data matches with the specified data format(s) of the data group.

Figure 1.3 shows the toolbox data repository browser. The repository browsers allows users to browse through the data repository, upload and organise their own datasets and share these with other users. The central panel of the repository browser shows the data sources and sub-folders of the currently opened folder/repository. The top bar of the repository browser shows the path of the currently opened repository, buttons to collapse/expand the repository folder tree-view sidebar on the left \equiv and the info-sidebar on the right \bullet , and a button to open the action menu \ddagger . The tree-view sidebar shows the hierarchical structure of the repositories and sub-repositories to which the user has access. The info panel shows the details of the selected data source or folder. If the selected item is a data source, then the info panel shows the types of data available in the data source and the different data source versions of the data source. If the selected item is a folder, then the info panel shows info about the other users and user groups that have access to this repository.

Users with read-write access (or higher) may upload new data source files by pressing the add button + on the bottom right and selecting the *upload new file(s)* item. A new sub-repository can be created by pressing the same add button and selecting the *create new folder* item. A third option is to create an external Proast link, which can be seen as a data source repository folder in which the data sources link to datasets (outputs) available on Proast web.

1.2.1 Uploading data files

Users with an access level of read-write (or higher) are allowed to upload new data source files to a repository folder. This can be done by pressing the add button + on the bottom right and selecting the *upload new file(s)* item.

Accepted upload file types are:

• **Microsoft Excel files (.xlsx):** An Excel file contains one or more sheets, each sheet containing tabular data. The sheet names and the fields of the data tables must comply with the conventions as specified in the data format section of the module(s) for which the data is presented.

MCRA 9 - EuroMix tool Exposure, Hazard & Risk Assessment	box					💄 kruisselbrink 🗉 🗮 🏭 🕐
≡ Data / EuroMix / Dose	-resp	onse data				0 :
Folders		Name 👻	Versi	on Date	Uploader	BfR-HepG2-RGA-Mixtures-2.xlsx
kruisselbrink	<	↑ (EuroMix)				
Acropolis	<	BfR-HepaRG-AdipoRed-Mixtures.xlsx	2	25-06-2019 10:20	kruisselbrink	Data groups
EuroMix Combined datasets	~	BfR-HepaRG-AdipoRed-Single.xlsx	2	25-06-2019 10:20	kruisselbrink	Dose response data
Concentrations		BfR-HepG2-RGA-Mixtures-2.xlsx	3	25-06-2019 10:19	kruisselbrink	Versions
Consumptions		HepaRG-AdipoRed-one-expfive-subst-for training no summary.xlsx	: 1	18-03-2019 16:23	kruisselbrink	BfR-HepG2-RGA-Mixtures-2.xl v1 (28/02/2019 04:02 kruisselbrink)
Dose-response data Effects and AOP networks		HepaRG-AdipoRed-one-exp-five-subst-for training.xlsx	1	18-03-2019 16:23	kruisselbrink	BfR-HepG2-RGA-Mixtures-2.xl v2 (25/06/2019 10:06 kruisselbrink)
Foods and food translations		HepaRG-AdipoRed-one-exptwo-subst-for training no summary.xlsx	(1	18-03-2019 16:23	kruisselbrink	BfR-HepG2-RGA-Mixtures-2.xl v3 (25/06/2019 10:06 <u>kruisselbrink</u>)
Hazard data		HepaRG-AdipoRed-one-exptwo-subst-for training.xlsx	1	18-03-2019 16:23	kruisselbrink	
In-silico data		RIVM-EST-CardioDiff-Mixtures.xlsx	2	25-06-2019 10:19	kruisselbrink	
Kinetic models		UGent-HepaRG-Mitochondria-Mixtures.xlsx	2	25-06-2019 10:20	kruisselbrink	
Non-dietary exposures						
Processing						
Substances						
Test-systems and responses						•

Figure 1.3: The toolbox data repository browser.

- Microsoft Access files (.mdb): An Access file contains one or more tables. The table names and the fields of the data tables must comply with the conventions as specified in the data format section of the module(s) for which the data is presented.
- Zip archives with CSV files (.zip): A CSV file (.csv) is a comma-separated values file containing data in tabular format. One or more CVS files are archived in a zipped file format (.zip) to facilitate the upload of collections of multiple CSV files. The names of the CSV files in the zip archive must follow the accepted table names of the module(s) for which the data is presented and the tables in the CSV files must follow the data format of that/those module(s). Note, that it is not allowed to upload single CSV files.

1.2.2 Repository access levels

Shares and access rights can be granted on the level of repositories and sub-folders. Data sources inherit the access rights of the repository/folder in which these are located. The following access rights are available:

- **visible:** the user can only see that the repository exists, but cannot see its contents, except for sub-folders that may also visible to the user.
- **use:** the user is only allowed to use the data sources in this repository, but is **not** allowed to download the original data of the data sources of the repository.
- **read:** the user can use data sources in this repository **and** is allowed to download the original data files of the data sources of the repository.
- **read/write:** the user can use and download data sources in this repository and is allowed to add/remove files and folders to/from this repository.
- **admin:** the is considered as an administrator of this repository and has full control over it, including the rights to add/remove files and folders to/from this repository and to add/remove user and group shares.
- owner: the user is considered to be the owner of this repository and therefore has full control over it.

Users with administrator or owner rights on a repository/folder are allowed to add/remove user and group access using the *edit shares dialog* (Figure 1.4) that can be opened by pressing the *edit shares* button \leq .

MCRA 9 - EuroMix Exposure, Hazard & Risk Asse	toolbox ^{sment}	👤 kruisselbri	^{nk} 🗊 🚍 🏭 💡
≡ Data / EuroMix / D	ose-response data		0 :
Folders	Nama - Narcian D		onse data
💶 kruisselbrink	Dose-response data - Access rights	×	
Acropolis	User access rights Group access rights		fo
EuroMix			rink
Combined datasets	Add user share		
Concentrations			min
Consumptions	User		
Dose-response data	Access level	•	iers
Effects and AOP networks			k (owner)
Foods and food translation		Add share	(read/write)
🔚 Hazard data	Members		Edit shares
HumanMonitoring	& kruisselbrink	Quiner	Luit Sinarco
🖿 In-silico data		Owner	
Kinetic models		Close	
Non-dietary exposures			
Processing			
Substances			
Test-systems and respons	25		+

Figure 1.4: The edit-shares dialog of the toolbox data repository browser: user and group access rights are added and removed by repository owners and administrators.

1.2.3 Linking remote data repositories

The toolbox also offers to link external data repositories \bigoplus . These are remote websites not part of the toolbox, but containing data sources that can be used for calculations. Currently, only one remote source can be linked as external repository in the toolbox, the PROASTweb (https://proastweb.rivm.nl/). PROASTweb users may link directly the outputs of their PROAST analyses (i.e., dose response models) as an external repository to the toolbox.

Figure 1.5 shows how PROAST outputs of a PROAST web user are linked to an external repository in the toolbox. Data sources of remote repositories have to be explicitly imported in the toolbox before they can be used in analyses. Initially, all data sources in a remote repository have a status of not-imported \triangle . Pressing the import button Φ , the toolbox will attempt to import the data source and once that is finished, the data source is ready to be used in analyses.

A new PROAST remote repository link is created by pressing the add button + on the bottom right and selecting the *Create Proast link* option. A dialog (Figure 1.6) opens asking for the local name of the external repository/folder, the PROASTweb username of the user of which the outputs should be linked, and the PROASTweb access key of the user, which is required as authentication token to access the analyses of the specified user.

1.3 Workspaces and actions

In MCRA, user work is organised in workspaces. A workspace is collection of work items (actions) and data that are logically grouped together. Each workspace has a name and can be given a description and tags. The *workspace browser* page provides users with an overview of all available workspaces and allows for creation of new workspaces. The *workspace overview page* page is shown when opening a workspace and provides an overview of the actions, data and tasks of the workspace.

The modelling tasks of MCRA are specified through actions. Each action is of a certain action type, which is the *main module* for which this action specifies the modelling task. Depending on the type of action, the user can configure the settings and the data sources of the main module and all relevant/linked sub-modules in the *action area* or, if it is a so-called standard action, in the *standard action page*. When all required settings and data sources of an action are configured and the action is in a valid state, then the modelling task described by the action can be started by pressing the run button. The status of this modelling task (which, depending on the complexity, may take some time to run)

· · · ·	A DAS11 Image: Constraint of the second				
Folders			Version	Date	
xruisselbrink	~	↑ (kruisselbrink)			
Imports		🛆 DAS11			C
Issues	<	🛆 DAS11-NoCovar			G
Misc		Foetal	1	04-07-2019 11:46	
Proast Kruisselbrink		Furan	1	01-07-2019 13:55	
Projects	<	Furan-NoCovar	1	01-07-2019 12:36	
Topics	<	△ MijnErod			6
Validation					
Acropolis	<				
EuroMix	<				

Figure 1.5: The remote (PROASTweb) repository in the toolbox data repository browser.

MCRA 9 - EuroMix t Exposure, Hazard & Risk Assess		よ kruisselbrink 🗉 🚍 🏭 🕐
Folders	Nama – Varojan Data Uklandar	kruisselbrink
 Imports Issues Misc 	Create Proast link Name *	X ruisselbrink evel: Admin
Projects Topics Velidentee	Proast user name * Proast access key *	ory users
Acropolis EuroMix	Create Can	

Figure 1.6: The dialog for creating a new PROASTweb remote repository link.

is shown in the results page and when the task has completed, output is available in the form of reports or in the form of data that can be used as input in other actions. Note that users are free to change the settings of an action at any time. Actions can therefore have multiple outputs from multiple tasks created with different settings.

A special kind of action is a so-called standard action. In a custom action, the user has to link up all data and configure all settings to run the action. To get an action up and running can be quite cumbersome and the result is an output report which can be quite detailed. Although output sections follow the modular design, specific sections are not instantly available. To encourage users to explore MCRA, *standard actions* are implemented to facilitate an easy introduction to the use of MCRA. The actions can be specified with only a few settings that are important within the scope of the standard action, and the output report contains a limited selection of main sections only.

A new action can be added to a workspace by clicking the + button at the bottom right of the *workspace overview* page. This will present an action menu with three options (see Figure 1.7):

- **Create a new action:** will open a wizard to create (normal) *action* by selecting the desired action type, specifying name, description and tags, and, depending on the action type, some main action settings.
- Create a new standard action: will open a dialog to create a *standard action* by selecting one of the available standard actions.
- **Import an action from zip file:** will open a local file browser that let's the user *import* a specifically formatted zip-file containing an action definition.

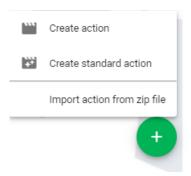


Figure 1.7: Three options to add an action to the workspace: create new action, create new standard action, and import action from zip file.

1.3.1 Workspace browser

Figure 1.8 shows the workspace browser. Users scroll through their workspaces and select the workspace which they want to work with. Detailed information about the selected item in the browser is shown in the info panel, which can be expanded/collapsed using the info button \bullet on the right of the toolbar. The *filter text box* \triangleleft is used to quickly find/filter workspace by name or tag. A workspace is opened by clicking on the workspace name or selecting the *open workspace* \bullet option of the *action menu* \vdots of the workspace. Opening a workspace will redirect you to the *workspace overview page*.

A new workspace is added by pressing the add button + on the bottom right of the screen. Delete workspaces by opening the *action menu* : of the workspace item in the browser and selecting the delete \blacksquare option.

MCRA 9 - EuroMix toolbox / Euro Exposure, Hazard & Risk Assessment workspace	Mix - Exposure			💄 kruisselbrink	▣≡	₩ (
Workspaces	Order	by	▼ Q euromix		×	
ame 👻	Created	Last modified	Tags			
EuroMix - CAG memberships calculations	01-10-2018 15:51	08-07-2019 14:42	EuroMix			:
EuroMix - Calculations case study R&R	15-12-2018 10:31	08-07-2019 14:37	EuroMix			:
EuroMix - Dose response models	14-03-2019 09:11	08-07-2019 14:40	EuroMix			:
EuroMix - Effect representations	01-05-2018 12:11	08-07-2019 14:40	EuroMix			:
EuroMix - Examples hazard characterisations	23-08-2018 10:01	08-07-2019 16:07	EuroMix			:
EuroMix - Examples PROAST	03-04-2018 10:41	08-07-2019 16:41	EuroMix	PROAST		:
EuroMix - Exposure mixtures calculations	23-04-2019 15:21	08-07-2019 16:42	EuroMix	Mixtures		:
EuroMix - Hazard characterisation calculations	18-09-2018 14:51	08-07-2019 14:36	EuroMix			:
EuroMix - Hazard characterisation tests	13-07-2018 14:41	08-07-2019 14:41	EuroMix			:
EuroMix - Human monitoring example	11-11-2018 14:41	08-07-2019 15:54	EuroMix			:
EuroMix - Ivive	28-04-2019 11:41	08-07-2019 16:14	EuroMix	IVIVE Risk		:
EuroMix - Kinetic model calculations	07-01-2019 09:41	08-07-2019 16:40	EuroMix	PBPK Kinetic-model	s	-
EuroMix - RPF calculation scenarios	17-07-2018 15:51	08-07-2019 16:43	EuroMix	RPF		+
EuroMix - Target exposure assessments	30-04-2019 16:31	08-07-2019 15:54	EuroMix			

Figure 1.8: The workspace browser.

1.3.2 Workspace overview page

Figure 1.9 shows the workspace overview page. This page provides an overview of the actions, data, tasks, and results of a workspace, shown as four tabs at the top of the page. The actions tab shows all actions of the workspace, and from this tab, actions are opened. The data tab shows all data sources used in this workspace. I.e., all data sources that are used by the actions of the workspace. The results tab shows all tasks and results of simulation jobs that have been submitted by the actions of the workspace. The properties tab shows the general information of the workspace (i.e., name, descriptions, and tags) and edit functionality.

In the actions tab, all actions of the workspace are listed. The list of actions can be filtered by action type or by filter text using the controls on the toolbar. An action is opened by clicking on the action name or by selecting the *open action* option of the action menu \odot of the selected action item. Opening a workspace will redirect you to the *action area*. A new action is added to the workspace by pressing the *add button* + at the bottom right of the page.

1.3.3 Action area

After opening an action, the user is directed to the main panel of the action. Each action has its own specific panel. In the main action page and sub-action pages, an action is configured, simulation jobs started, and output results are evaluated. The panel in Figure 1.10 shows the following sections:

- Scope: Links to the scope-panels in which the scope entities of the action are set (e.g., foods or substances).
- **Inputs:** Links are shown for panels in which the calculation inputs or selection inputs are set (e.g., concentration models that are inputs for computing dietary exposures).
- **Data source:** If the action is a data action, then a form is shown in which the data source should be specified (e.g., selection of the concentration data source in a concentrations action).
- Settings: A form is shown in which the calculation and/or selection settings of the action are set/changed (e.g., specify the exposure type, chronic/acute, of an exposure assessment).

All modules of the toolbox have equally structured panels. In each panel, data sources and settings for the action are specified and the scope and input sub-module links that are relevant are shown. This presentation reflects the modular

Actions Data	II. Results	Properties			
Workspace actions			(44 selected)	✓ Q, Type filter text here	
lame 👻	Туре	Created	Last modified	Tags	
Aggregate exposure assessment	Exposures	07-05-2018 12:45	22-06-2018 13:57	aggregate target-exposures	
Dietary exposure assessment	Dietary exposures	07-05-2018 16:56	08-05-2018 08:58	dietary	
Example hazard characterisation calculation	Hazard characterisations	07-05-2018 14:08	08-06-2018 14:43	target euromix example	
Example target exposures calculation	Exposures	07-05-2018 17:06	08-06-2018 16:56	target-exposures	
Relative potency factors	Relative potency factors	07-05-2018 16:04	07-05-2018 16:41	rpf	
Relative potency factors from data	Relative potency factors	07-05-2018 16:45	07-05-2018 16:53	rpf	
Risk assessment example	Risks	08-05-2018 09:19	08-05-2018 09:44	Risk	



design and allows the user to select the data and settings required for running the action. In the summary panel \blacksquare the main settings and data of the action are summarized. The output settings panel is used to specify general output settings. In the uncertainty settings panel $\stackrel{=}{\Rightarrow}$ the number of uncertainty runs and uncertainty sources is specified. In the results panel \bigoplus running tasks and output results of the actions are shown. An alternative form of navigating from action to sub-action is provided by the navigation menu in the left sidebar that can be expanded/collapsed by clicking the menu button on the top left in the Action bar. In this menu, all required modules for the action are shown in one list, allowing a linear way of navigation.

An action is valid and ready to run when all scopes and inputs are valid and all required data and settings are configured. For each sub-action, the check symbol \checkmark indicates that it has been configured correctly and is ready to run. In case a sub-action has a warning symbol \blacktriangle , some user action is required. When the main action is ready to run, a simulation job is started by clicking the run button \blacktriangleright in the grey action bar on the top right. Optionally, sub-actions can be started by clicking the run button \blacktriangleright in the green (sub)action bar on the top right. Clicking the run button will send the simulation task of this (sub)action to the job-scheduler, and the progress of the task is shown in the results panel \bigoplus . After completing the task, output is available in the form of a screen report, download as pdf, or download of tables in csv format.

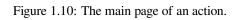
Scoping: entity selection

Each action starts with the selection of the relevant primary entities. In this context, entity selection or scoping plays an important role. Scoping of the action is defining the members for its primary entities, and, occasionally, also for other entities.

As an example, Figure 1.11 shows the substances module panel. At the top, the data source file with substances is selected containing the primary entity data of substance codes. In the selection card, a selection is made of the entities in the dataset that are relevant for the current action (3 in scope). Note that if no explicit selection is made, the scope is set to all entities by default. In the settings form, additional (selection) settings are shown, e.g., selection of the index substance (relevant for a cumulative assessment). In this way, the scope of the action is specified by selection of the primary entities.

The panels for the data modules have a similar structure and selection is essentially the same. The only difference is that data actions always have a scope. I.e., data modules always relate to one or more primary entities.

Exposure, Hazard & Risk Assessment	x / = EuroMix - User gro / Aggregate exposure workspace	
■ Aggregate exposure as Exposures action	essment	■ ≎ ። ▶ (
≣ Summary	Exposures	24dbf8c7 ►
Effects		
Foods	✓ Scope	
Populations (optional)	Populations (optional)	
Substances	Foods (2277 in scope) Processing types (0 in scope)	×
Exposures	Substances (30 in scope)	
Concentration models	Effects (25 in scope)	✓
Consumptions by food as measured		
Dietary exposures	A Inputs	
Food conversions	Dietary exposures	
Foods as measured	Non-dietary exposures (8 non-dietary surveys selected)	
Active substances (optional)	Kinetic models (defaults) (3 kinetic model instances selected	ed)
AOP networks	Relative potency factors	
Concentrations		
Consumptions	Target exposures settings	Save Changes



MCRA 9 - EuroMix to Exposure, Hazard & Risk Assessme	olbox / ^{ent}	The Acties / W Hazard characteris	K 🗖 🗮 🏼 🕐
Hazard characterisa Hazard characterisations action Hazard characterisations action Hazard characterisations A characterisation A characterisation A characterisation A characterisation A characterisation A characteris A charact	ations fi	om dose response models	: ः ▶ •
Summary	Â	Hazard characterisations / Substances	8df476ec 🕨 🕒
Effects	~		
Responses	~	Substances data source	
Substances	~	✓ EuroMix Substances Inventory (v6).zip	11
Test systems	~	Substances: 3 in scope (clear filter) 1626 only in table (add to scope)	~ ₹
🗱 Hazard characterisations	~		
Active substances	~	Substances selection	
🔯 Dose response models	~	Substances: 3 in scope	-
AOP networks	~		
Dose response data	~	Substance settings	Save Changes
Effect representations	~	Index substance Imazalil (aka enilconazole) (RF-0246-001-PPP)	- 0
Inter-species conversions (defa	aults)		
Intra species factors (defaults)			
<	+		

Figure 1.11: The substances module panel as an example of a primary entity module panel.

Implicit versus explicit scoping

MCRA distinguishes between implicit and explicit selection of entities (scoping). By default, the selection is defined implicitly as 'all entities' found in all data are linked to the action. For instance, the substance scope will contain all substance codes found. That is, not only substances as specified in the substance data source, but also all other substances found in data sources that link to substances like concentration sample data or points of departure data. These are implicit selections. Explicit selections are made in the specific module panel of this data type (e.g., by selecting the substances in the substances panel). Once made explicit, selections are no longer automatically expanded when new data sources are linked to the action.

For example, the substances scope shown in Figure 1.11 is defined explicitly, having three substances in the scope, and excluding 1626 substances also present provided through substances data source and/or other linked data sources like concentration samples. By pressing the *clear filter* button, the explicit scope is cleared and is made implicit again. Then, the scope contains all substances found as primary entities and found in all linked data sources, in total 1629 (1626 + 3) substances.

Comparing new data to set scopes

After linking a data source to an action, MCRA performs a check whether the new data links well to the current scope (selected entities) of the action and reports the results. For instance, after linking new substance concentration data to an action which already has an implicit or explicit substance scope, it should be checked whether the substance codes used in the concentration data match with the current substances in scope. Note that this check is also performed after linking a primary entity substances data source to an action which already has a set of substances in scope, i.c. substances already specified in other selected data sources.

After linking a data table from a new data source to an action which already has a defined scope for one of the entities in the table, there are three possible states for entity codes:

- codes included in both the scope and the data source
- codes included in the scope, but not present in the data source
- codes included in the data source, but not present in the scope

The first case represents a successful link, no further action is required. For the second and third type of mismatch, it depends on the type of data link whether this is considered a serious problem (red flag \blacktriangle) or merely a point of attention (green flag \bigstar). For instance, in the case of concentration data, for some substances no concentrations are available, and therefore MCRA allows missing concentration data for part of the substances in the scope: a green warning symbol is shown. The concentration data source may equally well contain codes that are not in the scope (e.g., concentrations for substances that are not specified in the primary entity data for substances). It may be desirable to extend the scope with these substances found in the concentration data. Also this situation is flagged with a green warning symbol.

Figure 1.12 shows an example of a point of departure action. The substances scope has already been defined by other data in the action (in this case points of departure data), and subsequently a substances data source is selected. Here, there are 140 substances in the current scope (explicitly defined). However, 132 of these 140 substances are not present in the substances data source (*not in table*). Hence, we are missing the definitions of these substances. This is considered a critical linking issue that should be solved by updating the substances data source to include these substances, therefore a red warning symbol is shown. On the other hand, the substances data source also contains 3 substances that are not part of the current scope (*only in table*). This is a non-critical error, normally leading to a green warning symbol, but in this case, it is overruled by the red warning symbol.

Another example is shown in Figure 1.13. The primary entities effects and substances are selected and in the scope. Then, a points of departure data source is selected containing effect and substance codes. For effects, no linking errors are observed, hence the new data source matches perfectly with the effects already in scope. For substances, we see that there are 7 substances that are in the points of departure data source but not in the substances scope (*new*) and for 3 substances in the scope no points of departure are available (*not in table*). The former is fine, but it might be needed to extend the scope with these 7 substances (*add to scope*). The latter, in general, is not a problem but just a point of consideration. These substances might be removed from the scope (*remove from scope*) or not.

Points of departure / Substances	cfb31e19	►	¢
Substances data source			
✓ NetherlandsTriazoles2007-10.mdb		A	/
Substances: 140 in scope (<u>clear filter</u>) 3 only in table (<u>add to se</u> 132 not in table (<u>remove from scope</u>)	cope)	A	Ŧ
Substances selection			
Substances: 140 in scope			-

Figure 1.12: Checking substances data in a substances data source against an already set substances scope.

Points of departure be168c0b	► ¢
Scope	
Effects (1 in scope)	A
Substances (140 in scope)	A
 ✓ CAG_steatose_PESTICIDES_april 2017.mdb 	~ <i>i</i>
✓ Hazard doses:	~
Effects: no linking errors	Ŧ
Substances: 7 new (add to scope) 3 not in table (remove from scope)	▲ =

Figure 1.13: Checking substances data in a POD data source against an already set substances scope.

1.3.4 Standard action area

Standard actions facilitates MCRA users in organizing the data and configuring all setting to run an action. Some standard actions are for demonstration purposes only, other standard actions are realistic examples of risk assessment or dietary exposure assessment using real data.

Creating a standard action

A new standard action is added to the workspace by pressing the *add button* + at the bottom right of the workspace page, see *workspace overview page*. Select the \bowtie *Create standard action* option.

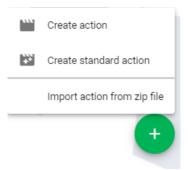
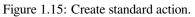


Figure 1.14: Add standard action.

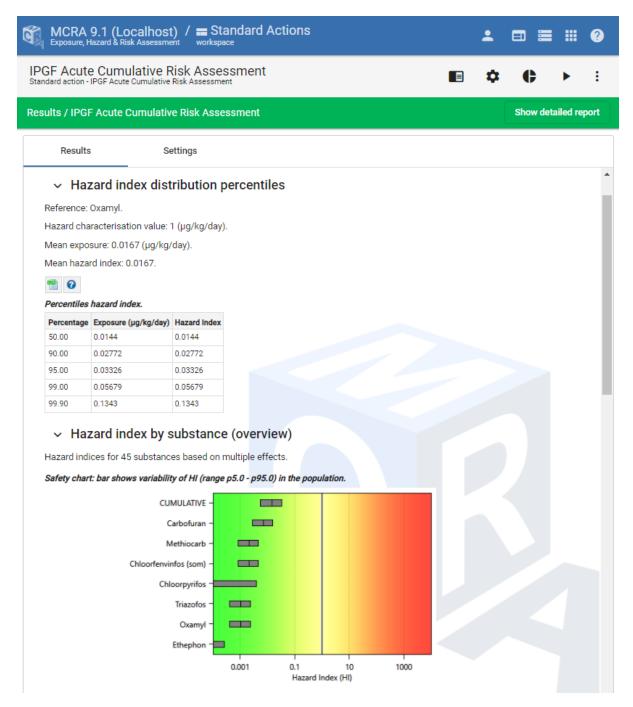
Then a pop-up appears, see Figure 1.15 and after clicking one of the available panes the standard action is created. Currently, a limited number of standard actions is available for authorized users. Please contact the MCRA system administrator at https://rivm.nl for more information. New standard actions will be developed in the near future.

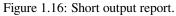
elect standard ac	tion type 2
🗘 Chronic cumula	ative exposure assessment PA
This standard action ca	n be used to calculate a chronic cumulative exposure for 18 PA. Assuming equipotency, or using proposed RPFs.
🗘 Chronic cumula	ative exposure assessment PFAS
his standard action car	n be used to calculate a chronic cumulative exposure for four PFAS. Assuming equipotency, and using proposed RPFs.
🗘 Demo acute cu	mulative risk assessment
	us data, acute cumulative risk assessments can be performed following various calculation methods (EFSA 2012 Optimistic and Pessimistic, EC 2018 Tier 1 ne effect of applying processing factors can be assessed.
🔉 EU acute cumu	lative exposure assessment (2018) Tier I and Tier II
	enable you to reproduce the exposure assessment of acute cumulative effects of pesticide residues in food affecting the nervous system. These are
	assessments of the cumulative exposure for the nervous system using monitoring data from 2014, 2015 and 2016. In this standard action Dutch monitoring and sed. The results, data used and methodology are reported in a scientific report following published on the EFSA website in September 2019. The methodology
	set. The results, data used and mentodology are reported in a scientific report following published on the EPSA website in September 2019. The methodology set by the European Commission.
🛊 EU chronic cum	nulative exposure assessment (2018) Tier I and Tier II
exposure assessments	I enable you to reproduce the exposure assessment of chronic cumulative effects of pesticide residues in food affecting the thyroid. These are retrospective of the cumulative exposure for the thyroid using monitoring data from 2014, 2015 and 2016. In this standard action Dutch monitoring and consumption data are sed and methodology are reported in a scientific report following published on the EFSA website in September 2019. The methodology fulfils the requirements
	nulative Risk Assessment
PGF base action	



Standard action reports

Although a standard action produces a short output report, see Figure 1.16, by clicking the *Show detailed report* the full report becomes available. By clicking *Show short report*, the short output report is returned.





Converting a standard action to a normal action

A standard action is easily converted to a custom action by opening the *action menu* \vdots in the white bar of the standard workspace in your browser and selecting the \bowtie *Convert to custom action* option.

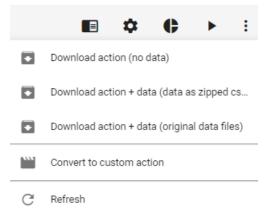


Figure 1.17: Convert to custom action.

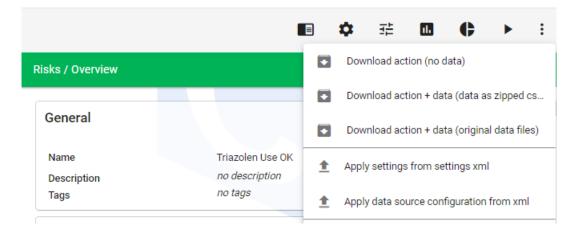
1.3.5 Action zip files

MCRA also provides the functionality to import and export actions (with or without data) as specially formatted zip files (.zip) containing 'ready to run' actions. This special file archive contains two XML files (.xml) named _*ActionSettings.xml* and _*ActionData.xml* that respectively describe the action's settings and linked data sources. In addition, this zip archive may include the action data itself, either in their original form, located in a sub-folder (data) or in the form of (.csv) files.

These action zip files can be imported into and workspace by pressing the *add button* + at the bottom right of the workspace page, see *workspace overview page* and then selecting the *Import action from zip file* option. Once uploaded, the action doesn't require further user interaction and is ready to run.

An action can be exported as a zip archive by by clicking the *action menu* : located in the white bar on top of the panel. Currently, this allows export in three different formats (see Figure 1.18):

- The **download action** (**no data**) option will create a zip archive containing only the two xml files defining the action, but will not include the data.
- The **download action + data (data as zipped cvs)** option will create a zip archive containing the action definition xml files and also the action data in the form of csv files of the internally used data table formats.
- The **download action + data (data as zipped cvs)** option will create a zip archive containing the action definition xml files and also the original action data files (i.e., the files that were originally uploaded in the repository).



CHAPTER TWO

MODULES

MCRA is a modular system. The diagram of Figure 2.1 shows the modules and their relations. Each module is associated with its own type of data, and is linked to one or more other modules. Note that not all details can be fully shown in the scheme, for details consult the table below, which specifies all relations between the modules in MCRA.

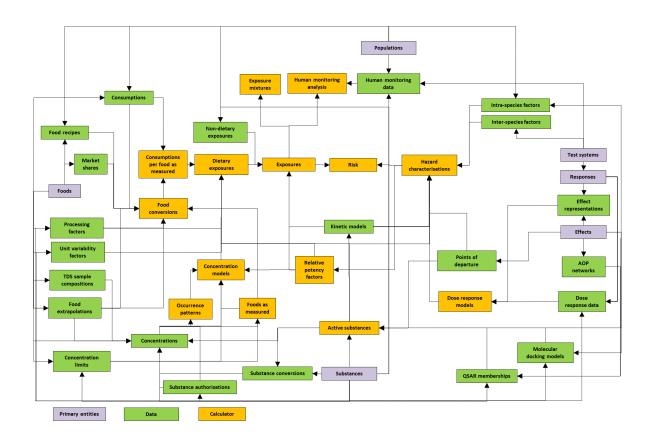


Figure 2.1: Diagram of the modular design of MCRA.

2.1 Primary entity modules

The MCRA modular system is based on six primary entities, defining who (*Populations*) is to be protected against what impact (*Effects*) caused by what agent (*Substances*) originating from where (*Foods*), with an indication how the effects are quantified (*Responses* in *Test systems*).

2.1.1 Effects

Effects are biological or toxicological consequences for human health, that may result from chemical exposure and are the focus of hazard or risk assessment.

Output of this module is used by: Concentration models High exposure food-substance combinations Dietary exposures Exposure mixtures QSAR membership models Molecular docking models Active substances Relative potency factors Hazard characterisations Points of departure Effect representations Inter-species conversions Intra species factors AOP networks Risks Single value risks

Effects data formats

Effects

Effects are primary entities of the data model. Health effects are defined as (critical) changes relative to a treatment or exposure.

Effects

Effects are uniquely identified by a code (idEffect). Optionally, a name and description can be added. Health effects are commonly distinguished in two types, acute and chronic. Further properties may be specified, e.g. in relation to decision schemes such as the use of thresholds of toxicological concern (TTCs).

Name	Туре	Description	Aliases	Required
idEffect	AlphaNumeric(50)	Unique identification code of the effect.	idEffect, EffectId, Code- FocalEffect, Id, Code, KeyEvent, idKeyEvent	Yes
CodeSystem	AlphaNumeric(100)	Identifier of the coding system of the effect code.	CodeSystem	No
Name	AlphaNumeric(100)	Name of the effect.	Name	No
Description	AlphaNumeric(200)	Additional description or label of the effect.	Description	No
Biological- Organisation	AlphaNumeric(100)	Biological organisation of the effect: Molecular, Cellular, Tissue, Organ, Individual. This is in line with AOP wiki terminology and can be used for grouping.	Biological- Organisation	No
KeyEvent- Process	AlphaNumeric(100)	Description of AOP Key event component process. E.g., receptor signalling.	Process	No
KeyEvent- Object	AlphaNumeric(100)	Description of AOP Key event component object. E.g., PPAR-alpha.	Object	No
KeyEvent- Action	AlphaNumeric(100)	Description of AOP Key event component action. E.g., decreased.	Action	No
KeyEventOrgan	AlphaNumeric(100)	Description of AOP Key event organ. E.g., liver.	Organ	No
KeyEventCell	AlphaNumeric(100)	Description of AOP Key event organ. E.g., hepatocyte.	Cell	No
AOPwikiKE	AlphaNumeric(200)	Key event ID number in AOP wiki https://aopwiki.org/events Several ID possible Some effects might not be in the wiki, and this field will be empty.	AOPWikiIds, AOPwikiKE	No
Reference	AlphaNumeric(200)	External reference(s) to sources containing more information about the AOP key event. E.g., the AOP wiki, and the associated AOP wiki Ids.	References	No

Table aliases: Effects, Effect, KeyEvents, KeyEvent.

Effects calculation

Option *Multiple effects analysis* selects multiple selects. Effects are selected using **Effects selection**. Press **Clear filter** and/or **change selection** and select multiple effects in the scroll down box. If both this option and *Include related effects of AOP network* is unchecked, it is obligatory to select one and only one effect. If *Include related effects of AOP network* is checked, a **Focal effect** is specified and all related effects in the *AOP network* are selected as well.

Effects settings

Selection settings

Name	Description
Multiple effects analysis	Specifies whether the analysis should consider multiple effects.
	Otherwise, a single focal effect should be selected.
Focal effect	The main (health) effect of interest.
Include related effects of AOP	Include all related key events of the AOP network.
network	

Effects as data

Effect definitions are provided as lists/catalogues of effect definitions.

- Effects data formats
- Effects calculation

2.1.2 Foods

Foods are uniquely defined sources of dietary exposure to chemical substances. Foods may refer to 1) foods as eaten, foods as coded in food consumption data (e.g. pizza); 2) modelled foods, foods as coded in concentration data (e.g. wheat, tomato); 3) any other type of food (e.g. ingredients like flour, tomato sauce).

Output of this module is used by: Consumptions Single value consumptions Market shares Food recipes Concentrations Concentration distributions Single value concentrations Processing factors Unit variability factors Occurrence patterns Occurrence frequencies Substance authorisations Deterministic substance conversion factors Concentration limits Concentration models Modelled foods Focal food concentrations Total diet study sample compositions Food extrapolations Food conversions Consumptions by modelled food High exposure food-substance combinations Dietary exposures Single value dietary exposures Exposures Exposure mixtures

Foods data formats

Foods

Foods are of interest in (dietary) consumption assessments and the sources of exposure within exposure assessments. The foods table is the main table of the food definitions. Relevant food related data, such as processing types, additional properties (e.g., unit weight and brand loyalty), facets, and hierarchies, can be described in the food properties, food hierarchies, and faces and facet descriptors tables.

Foods

Each food is identified by a unique code (idFood) in a code system of choice, a name, and a description. In the EuroMix data collection, FoodEx1 codes are used for both foods in consumption surveys (foods as eaten) and for raw agricultural commodities (foods-as-measured). Example: 'A.19.01.002.002' is pizza and pizza-like pies, cheese, and vegetables and 'A.01.02.001' is wheat grain. Food codes can have a hierarchical structure (as in the FoodEx1 and FoodEx2 coding systems), using '.' or '\$' as separator between adjacent hierarchical levels, e.g. 'A.05' is fruits and fruit products, 'A.05.01' is citrus fruits, and 'A.05.01.001' is grapefruit (citrus paradisi). Additional forms of foods, such as foods in processed form, can be specified via food facets according to the FoodEx2 system of EFSA.

Name	Туре	Description	Aliases	Required
idFood	AlphaNumeric(50)	The unique identification code	idFood, Code,	Yes
		of the food.	FoodId,	
			FoodCode,	
			Food, Id	
Name	AlphaNumeric(100)	The name of the food.	Name,	No
			FoodName,	
			Name1,	
			FoodName1	
Description	AlphaNumeric(200)	Food description.	Description	No

Table 2.3: Table definition for Foods.

Table aliases: Foods, Food.

Food properties

Additional food properties, such as portion sizes can be attached using the food properties table.

Name	Туре	Description	Aliases	Required
idFood	AlphaNumeric(50)	The code of the food to which	idFood, FoodId,	Yes
		the property is attached. The	Food,	
		provided food code should	FoodCode,	
		match with a code of the	Code	
		foods table.		

Table 2.4: Table definition for FoodProperties.

Table aliases: FoodProperties, FoodProperty.

Food unit weights

Food unit weights as specified for a food, and possibly a location.

Name	Туре	Description	Aliases	Required
idFood	AlphaNumeric(50)	The unique identification code of the food.	idFood, Code, FoodId, FoodCode, Food, Id	Yes
Location	AlphaNumeric(50)	The location for which this food unit weight is defined. If not specified, then the value is considered a default unit weight that can be used when there is no location specific unit weight.	Location	No
ValueType	UnitWeightValue- Type	The value type of the unit weight value (i.e., raw agricultural commodity or edible portion).	ValueType, UnitWeight- ValueType	No
Qualifier	QualifierType	Qualifier of the unit weight value, e.g. equal-to (=) or smaller-than (<). If omitted, = is assumed.	Qualifier, QualifierType	No
Value	Numeric	Unit weight value in grams.	Value, Unit- WeightValue, UnitWeight	Yes
Reference	AlphaNumeric(200)	External reference(s) to source of the unit weight value.	Reference, References	No

Table 2.5: Table definition for FoodUnitWeight	s.
--	----

Table aliases: FoodUnitWeights, UnitWeights.

Food hierarchies

Food items are commonly categorised in hierarchies, e.g. oranges and mandarins are citrus fruits. For example FoodEx is a food description and food classification (FDFC) system consisting of a large number of individual food items aggregated into food groups and broader food categories in a hierarchical structure of parent-child relationships.

Table 2.6:	Table	definition	for	FoodHierarchies.
10010 2.0.	ruoie	acimition	101	i oournerurenes.

Name	Туре	Description	Aliases	Required
idFood	AlphaNumeric(50)	Food node.	idFood, FoodId,	Yes
			Food, Code	
idParent	AlphaNumeric(50)	Parent node of the food.	idParent,	Yes
			ParentId, Parent,	
			ParentCode	

Table aliases: FoodHierarchies, FoodHierarchy, FoodsHierarchy.

Facets

Food codes can be linked to facets, as e.g. in FoodEx.

Name	Туре	Description	Aliases	Required
idFacet	AlphaNumeric(5)	The food code of the food to	idFacet, Code,	Yes
		which the facet is attached.	Id, FacetCode,	
			FacetId	
Name	AlphaNumeric(200)	Facet name	Name,	Yes
			FacetName	

Table aliases: Facets, Facet, FoodFacets, FoodFacet.

Facet descriptors

Name	Туре	Description	Aliases	Required
idFacet-	AlphaNumeric(5)	The identification code of the	idFacet-	Yes
Descriptor		facet descriptor.	Descriptor,	
			Code, Id,	
			FacetCode,	
			FacetId	
Name	AlphaNumeric(200)	The name of the facet	Name, Facet-	Yes
		descriptor.	DescriptorName	

Table 2.8	Table	definition	for	FacetDescriptors.
1 abic 2.0.	raute	ucinnuon	IUI	r accubescriptors.

 $Table\ aliases:\ FacetDescriptors,\ FacetDescriptor,\ FoodFacetDescriptors,\ FoodFacetDescriptor.$

Processing types

Name	Туре	Description	Aliases	Required
idProcessing- Type	AlphaNumeric(50)	The unique identification code of the processing type.	idProcessing- Type, ProcessingType- Id, ProcType, Id	Yes
Name	AlphaNumeric(100)	The processing name.	ProcName, Name	No
Description	AlphaNumeric(200)	The processing type description.	Description	No
Distribution- Type	AlphaNumeric	The distribution type. Simulated processing factors are restricted to the interval (0,1) using a logistic-normal distribution (= 1) or simulated processing factors are restricted to positive values using a log-normal distribution (= 2).	Distribution- Type, DistType	Yes
Bulking- Blending	AlphaNumeric(10)	For types of processing applied on large batches e.g. juicing, sauce/puree. No bulking/blending = 0, bulking blending = 1.	Bulking- Blending, BulkBlending	Yes

Table 2.9: Table definition for ProcessingTypes.

Table aliases: ProcessingTypes, ProcessingType.

Foods as data

Food definitions are provided as lists/catalogues of food definitions, optionally with encompassing processing type definitions, facet definitions, hierarchy definitions, and additional food property information.

• Foods data formats

Food coding systems

MCRA is intended to retain complete transparence of the results of risk assessment in terms of the foods that were actually consumed (foods-as-eaten). In many cases measurements of substances have not been made on the **food-as-eaten**, e.g. pizza, but on a raw agricultural commodity (RAC), e.g. tomato, onion etc. The food on which the concentration measurements have been made is termed the **modelled food**. MCRA implements a *recursive search algorithm* to link foods-as-eaten to modelled foods. This means that there can be intermediate steps, e.g. if unpeeled *apple* and *grapes* are the modelled foods, the food-as-eaten *apple pie* contains *peeled apple* and *raisins*, *peeled apple* is linked to unpeeled *apple*, and *raisins* are dried *grapes*. *Peeled* and *dried* are the *processing types*.

Food classification: FoodEx1

Food code definition

In MCRA, a food code is a string consisting of symbols. Some special symbols (., \$, -, #) are reserved for special use (see below), and can not be used freely in own codes.

Codes can be hierarchical. Any code can be followed by \$ or . plus a subtype code. This can be repeated any number of times, e.g. A\$B\$C\$D, or A.B.C.D.

Codes can specify the food processing type (e.g. peeling). Any code can be followed by a hyphen ('-') plus a processing type code (e.g. FP0226-2). More than one level of processing code is allowed (e.g. FP0226-2-13). Subtype codes should precede processing codes (e.g. NL005\$123\$456-2).

Food codes in consumption surveys

Any coding system for foods-as-eaten can be used in MCRA. For example, in Europe EFSA develops a Food Classification and Description System for exposure assessment named FoodEx 2 ([[EFSA, 2011a]], [[EFSA, 2011b]]), featuring a hierarchical system of a core list of foods, an extended list, and domain-specific hierarchies.

Food codes in concentration data

Any coding system for modelled foods can be used in MCRA.

Food processing

Concentrations of substances in foods may change when foods are processed. Examples of *processing types* are peeling (e.g. of apples), cooking (e.g. of spinach), drying (e.g. of grapes), juicing (e.g. of oranges). In MCRA a processing factor can be specified for any food. Processing factors specify the ratio of concentrations in the processed and unprocessed food. The food code of the processed food (e.g. FP0226-2) will be converted to the food code of the unprocessed food will then be multiplied by the processing factor. Special attention is needed when food processing also includes changes of the weight of the food. Traditionally, processing factors combine the effects of chemical alteration and weight change, so the weight change should not be double-counted. The *processing correction factor* is introduced to correct processing factors that combine both effects, e.g. when 100g *raisins* (dried grapes) are translated to 300g *grape* (modelled food) and the processing factor for drying combines both effects, the processing correction factor is 3.

Recipes and food translation

Recipes specify the composition of composite foods, e.g. *pizza*, in terms of relevant ingredients, e.g. 100g pizza contains 10g *tomato*, 5g *cheese* and 50g *flour*. Recipes are also used to specify weight changes, e.g. to obtain 100g *raisins* (dried grapes) 300g of the modelled food *grape* is needed, see also *processing correction*.

A special use of recipes and food translation is found in *Total Diet Studies*. Here, the composition of a Total Diet Study food is specified, e.g. TDS-food *FruitMix* is composed of *apple, orange* and *pear* with a default translation proportion of 100%. So in MCRA, the food-as-eaten *apple* is converted to *FruitMix* (100%) and *FruitMix* is considered as the modelled food (TDS-food). A conversion from *apple-pie* (food-as-eaten) to *FruitMix* (modelled food) is based on a recipe for apple-pie and a TDS composition for FruitMix.

Another use of converting foods (as-eaten or as an intermediate step), is through the specification of so-called food extrapolations (read across translations), e.g. for *pineapple* no measurements are found but by specifying that *pineapple* is converted to *FruitMix* (with a default proportion of 100%), the TDS sample concentration value of *FruitMix* will be used for *pineapple* (as-eaten or as ingredient).

Market shares and brand loyalty

Sometimes measurements of substances in food are available at a more detailed food coding level than consumption data. For example, measurements may have been made for specific brands of a food whereas the consumption survey did not record the brand. MCRA allows to specify market share data for subtypes of a food (e.g. A\$1, A\$2, A\$3 are three brands of food A), and to calculate acute exposure based on such *market shares*.

Supertypes

Sometimes measurements of substances on food are available at a less detailed food coding level than consumption data. MCRA allows to use the concentration data of a supertype for all underlying food codes. However, this is not the default, and an explicit permission should be given to allow this feature.

Maximum Residue Levels

Maximum residue levels are the upper legal levels of a concentration for substance residues in a food, e.g. pesticide, or feed based on good agricultural practices and to ensure the lowest possible consumer exposure.

MCRA food code conversion algorithm

The conversion algorithm links food as eaten codes to modelled food codes using a 7-step procedure.

Food classification: FoodEx2

'The collection and evaluation of data on levels of chemical occurrence or presence of biological agents in food and feed are important tasks of EFSA. By combining the data with information on food consumption allows for detailed intake and exposure estimates crucial to any food and feed safety risk assessment or nutrient adequacy analysis. The EU Member States provide an increasing volume of data to EFSA and other European bodies. To provide a common link to all the diverse food and feed databases, a system for the unique and universal identification and characterisation of food and feed items is essential. EFSA has developed a preliminary standardised food classification and description system called FoodEx2 (version 2 of the EFSA Food Classification and Description System [FCDC] for exposure assessment). The system consists of descriptions of a large number of individual food items aggregated into food groups and broader food categories in a hierarchical parent-child relationship. Central to the system is a common 'core list' of food items or generic food descriptions that represent the minimum level of detail needed for intake or exposure assessments. More detailed terms may exist in addition to the core list and these are identified as the 'extended list'. A parent-child relationship exists between a core list food item and its related extended list food items. The terms of the core and extended list may be aggregated in different ways according to the needs of the different food safety domains. In the present version four hierarchies are proposed: three domain-specific and a general purpose one. Facets are used to add further detail to the information provided by the food list term. Facets are collections of additional terms describing properties and aspects of foods from various perspectives'. For more information visit: http://www.efsa.europa.eu/en/datex/datexfoodclass.htm.

For MCRA, having a different set of food codes is in itself not a problem. That is, for MCRA, it does not matter how foods are coded, as long as they can be linked to consumptions and concentrations within an exposure assessment. What makes FoodEx2 different from other food coding systems is that it provides additional food hierarchies, food facets, and a combined food/facet coding system. Below follows a brief summary of these main features of the FoodEx 2 coding system from the perspective of exposure assessment using MCRA.

Foods and food hierarchies

FoodEx 2 contains different food hierarchy definitions and allows for creation of custom food hierarchy definitions. These hierarchies could, for exposure assessment, allow to assess intake or consumption data based on the groups defined by these hierarchies.

Code	Level	Name	ParentCode	Scopenotes
A000J	1	Grains and grain-based products	ROOT	The category covers all
A000K	2	Cereals and similar	A000J	
A0001	3	Cereal and cereal-like grains	A000K	
A000M	4	Amaranth grain	A000L	
A000N	5	Buckwheat grain	A000L	
A000P	6	Barley grain	A000L	

Table 2.10: Food hierarchy export from FOODEX 2.0 Browser version 0.1.3

Facets and facet descriptors

FoodEx 2 allows to provide supplementary details on specific aspects of foods by means of so-called facets and facet descriptors. Facets are collections of terms defining specific characteristics of food from particular points of view and facet descriptors describe specific characteristics foods. For example, *processing technology* is a facet, and *baking* is a facet descriptor belonging to this facet. Currently, 26 facets are defined, containing in total 2172 descriptors (EFSA 2011b) [[EFSA, 2011b]]. Facets are also defined in a hierarchical system. For instance, *cooking in fat (A07GR)* and *baking (A07GX)* are sub-items of the descriptor *cooking and similar thermal preparation processes (A0BA1)*. Facets are coded as small strings that consist of a facet code and a facet descriptor code separated by a '.'-character. For example, the facet code *F28.A07GX* holds

- 1. the facet code F28, which is the facet code for process technology, and
- 2. *A07GX* , which is the descriptor code for *baking*.

	(F()1).		
Code	Level	Name	ParentCode	Scopenotes
A04SF	1	Animals	ROOT	
A056H	2	Mammals (food source animal)	A04SF	
A056Z	3	Farmed / non-game mammals (food source animal)	A056H	
A057A	4	African buffalo (food source animal)	A056Z	
A057B	4	American buffalo (food source animal)	A056Z	
A057C	4	Buffalo (food source animal)	A056Z	
A057D	4	Cape buffalo (food source animal)	A056Z	
A057E	4	Cattle (food source animal)	A056Z	

Table 2.11: Part of the FoodEx 2 facet descriptor codes of the source facet

Implicit facets

Implicit facets are facets of a product that are already implied by the food product itself. Consider, for example, *potato boiled (A011P)*, where *boiling (A011P)* is an implicit facet, because boiling is already implied by the product. According to EFSA [[EFSA, 2011a]] 'inclusion of implicit facets in the string recorded for each food database record is not encouraged' and it is suggested to identify and record the implicit facet descriptors in a separate table.

Foods as facets

Foods and facet descriptors share the same unique alphanumerical coding system; in some cases, like *characterising ingredient or sweetening agent* food list elements may be used as facet descriptors.

The FoodEx 2 coding system

In the coding system, facets can be added to the primary food codes to provide supplementary detailed information of particular data records. The structure of the FoodEx 2 codes is:

idFood#idFacet.idFacetDescriptor\$idFacet.idFacetDescriptor\$....

The code starts with the primary FoodEx2 food code. Then, when there are supplementary facets, the food code is followed by a '#'-character and the facets string. The facets string is constructed as a concatenation of the individual facets strings, separated by means of the '\$' character. As an example, consider the string *A011P#F28.A07GL\$F28.A07KQ* which is composed of:

- Food: A011P Potato boiled
- Facet 1: F28.A07GL Process technology Boiling
- Facet 2: F28.A07KQ Process technology Freezing

FoodEx2

For MCRA, FoodEx 2 introduces the following points of attention:

- Reading and dealing with FoodEx 2 coded data sets
- Reading and dealing with food facets
- Reading and exploiting food hierarchy data

Reading and dealing with FoodEx 2 codes

All data entities that contain foods data are potentially affected by the introduction of FoodEx 2. In MCRA, the following data tables are adapted to allow for input of full FoodEx 2 food codes:

- Foods
- Consumptions
- Concentrations

For these tables, the food code is allowed to be the complete FoodEx 2 food code and automatically recognized as such. As an example, Table 2.12 shows how the FoodEx 2 coded consumptions should be provided to the system. On important note: the maximum field length of the food code is 50. This means that there is a maximum of five facets that can be specified for a food.

Individual	DayOfSurvey	Food	Amount	FoodSurvey
14233701	1	A011R# F28.A07GX	153.43	FS01
18843004	1	A011R# F28.A07GX	125.23	FS01
34025701	1	A011R# F28.A07GX	153.60	FS01
14720005	2	A011R# F28.A07GX	105.00	FS01
49174010	1	A011R# F28.A07GX	140.00	FS01
62794010	1	A011R# F28.A07GX	67.00	FS01
61392002	1	A011P# F28.A07GL\$F28.A07KQ	104.72	FS01
61281231	1	A011P# F28.A07GL\$F28.A07KQ	109.72	FS01

Table 2.12: Integrated coding of the facets in the consumed foods field of food consumptions. Implementation.

Reading and dealing with facets data

Within MCRA, the following facets related aspects are accounted for:

- Reading facets data
- · Dealing with facets
- Facets in concentration data
- · Facets in food conversion
- Using facets as processing factors
- Using hierarchy data in the output

Reading facets data

To incorporate input of facets data in MCRA, two tables Facets and FacetDescriptors are introduced as optional tables of the Foods data group. The *table for Facets* and *table for FacetDescriptors*.

Within MCRA, the facets of FoodEx 2 coded foods, consumptions, and concentrations are automatically linked to the provided facets and facet descriptors. Also, the facet descriptor names are added automatically to the foods containing these facets.

Dealing with facets

The introduction of food facets allows for much more detailed specifications of consumption and concentration data. However, it introduces the problem of deciding on which level of detail the exposure assessment should be performed. That is, should concentration models be generated on the level of foods-without-facets or on the level of foods-with-facets? E.g., should the concentrations of *clementine peeled (A01CE#F28.A07LC)* and *clementine unprocessed (A01CE#F28.A07CS)* be modelled separately or should one model be constructed for *clementine (A01CE)*? Treating all clementine's as equal may yield over-simplified conversions, whereas treating all separately may lead to many concentration models based on only few measurements. In MCRA, no implicit grouping of concentrations of equal foods with different facets is applied. If concentrations are provided for both *clementine peeled (A01CE#F28.A07LC)* and *clementine unprocessed (A01CE#F28.A07LC)* and *clementine unprocessed (A01CE#F28.A07LC)* and *clementine unprocessed (A01CE#F28.A0C0S)*, then these are modelled separately. Another question is whether the order of the facets is relevant or not. E.g., is *A0BYV#F02.A06GF\$F03.A06HY* the same as *A0BYV#F03.A06HY\$F02.A06GF*? Regarding this matter, MCRA considers the facet order to be important. I.e., *A0BYV#F02.A06GF\$F03.A06HY* is not the same as *A0BYV#F02.A06GF*.

Facets in food conversion

For conversion of foods-as-eaten to modelled foods, MCRA considers foods with different facet strings as different foods. I.e., there is no implicit conversion of foods-with-facets to foods-without-facets and also the order of the facets is important. However, as it is realistic to convert food-with-facets to the base food without facets, an additional (explicit) conversion step remove-all-facets is added that converts foods with facets to the base foods. I.e., the action is "remove all". There is no conversion step for "stripping off one facet at a time". The reason for this is that there is no good way of deciding which facet to strip off first. This new conversion step is somewhat equivalent to the already existing default processing conversion step (step 6), and is therefore implemented as step 6b of the conversion algorithm. Particular rules followed by this step:

· Conversion of food-with-facets to food-without-facets.

Using facets that reveal processing data

Facets containing processing information, such as *part-consumed-analysed* (F20) and *processing technology* (F28) could be integrated with processing data. As an example, consider *clementine peeled* (A01CE#F28.A07LC). This could be linked to *clementine* (A01CE), with processing type *removal of external layer* (A07LC). Linking to processing data could be achieved by entering processing data using the facet codes. As an alternative to the current processing factor tables, a facet-based processing factors table is defined for processing facets. That is, the codes for food processed and unprocessed are implicitly defined for FoodEx 2.

Table 2.13: Example of a MCRA processing factors table using FoodEx 2 foods and facets codes.

FacetCode	Substance	FoodCode	ProcNom	ProcUpp	Proc- NomUnc- Upp	Proc- UppUnc- Upp
A07LC	SubstanceX	A01CE	0.5	0.6	0.05	0.06
F28.A07GV	SubstanceX	A0BY	0.2	0.1	0.03	0.04

Note that in the example, the facet code could be specified as the full facet code, or just the code of the facet descriptor. As a more elaborate example consider

French fries from cut potato (A0BYV#F02.A06GF\$F03.A06HY\$F04.A00ZT\$F28.A07GR)

For this food code, the substring of the processing facet is extracted from the list of facets.

- A0BYV#F02.A06GF\$F03.A06HY\$F28.A07GR\$F04.A00ZT with processing facet link A07GR
- *A0BYV#F02.A06GF\$F03.A06HY\$F04.A00ZT*

In MCRA, a table FacetProcessingFactors is introduced that allows for specification of processing factors by means of facets. This table has the following structure:

Column name	Key	Required	Туре	Size	Description
idProcessingType	Yes	Yes	String	5	The facet code of this processing
					factor definition. May be specified
					as full facet code, i.e., facet code
					plus facet descriptor code, or as the
					facet descriptor code.
idFood	Yes	Yes	String	200	The food code
idCompound	Yes	No	String	50	The substance for which this
					processing factor is defined.
Nominal	No	Yes	Double		Nominal value (best estimate of
					50th percentile) of processing factor
					(defines median processing factor)
Upper	No	Yes	Double		Upper value (estimate of 95th
					percentile or "worst case" estimate)
					of processing factor due to
					variability
NominalUncertaintyUpper	No	Yes	Double		Upper 95th percentile of nominal
					value (Nominal) due to uncertainty.
					A standard deviation for uncertainty
					of the nominal value (Nominal) is
					derived using the nominal value
					(Nominal) and upper 95th
					percentile
					(NominalUncertaintyUpper)
UpperUncertaintyUpper	No	Yes	Double		Upper 95th percentile of upper
					value (Upper) due to uncertainty.
					From the nominal value (Nominal),
					upper value (Upper) and the
					specified uncertainties of these
					values (NominalUncertaintyUpper
					and UpperUncertaintyUpper,
					respectively) the degrees of
					freedom of a chi-square distribution
					describing the uncertainty of the
					standard

The integration with the food conversion algorithm is as follows: Conversion step 2 (*processing*) is extended with a step 2c (*processing facet*) that attempts to match facets of a food code to processing data provided in the processing facets table. The following important rules are followed:

- Processing factors can be defined for base-food-code/facet-code combinations and translate as food-withprocessing-facet to food-without-processing-facet.
- If multiple processing facets are present in the food-as-eaten code, then the last processing facet is used first for conversion.
- Facet processing factors can be specified using the full facet code (i.e., facet-code plus facet-descriptor-code) or just the facet descriptor code. If both are specified for the same food, the full facet code is used.
- Facet processing factors can be defined substance-specific, and non-substance-specific. Processing factors that are defined substance-specific always precede non-substance specific processing factors.
- Processing factors defined by a food-processed/food-unprocessed combination precede processing factors defined through facets.

Weight reduction factors for processing factors defined for facets should be included in the food translation table and should match exactly.

Food hierarchies

Reading and dealing with food hierarchy data

Within MCRA, the following hierarchy related aspects are accounted for:

- Reading food hierarchy data
- Using hierarchical data for conversion of foods
- Using hierarchy data in the output

Reading food hierarchy data

A new data group named *Foods* is added. In this group, a new *table for FoodHierarchies* is used for input of food hierarchies. This table contains food hierarchy node-definition records that reflect a hierarchical structure. For foods that are not in this list as idFood, it is implicitly assumed that these foods are root items.

Note: It is common practice to describe hierarchies using tree structures. Here, the elements of the tree are named *nodes*, the lines connecting the nodes are named *branches*, and nodes without children are *leaf nodes/end-nodes*. This terminology is also used throughout the remainder of this document.

Using food hierarchies for food conversion

The introduction of the hierarchy structure allows for integration with step 4 and step 5 of the food conversion algorithm; the *subtype* and *supertype* linking steps. That is, when no concentration data is found for a certain product, the concentration data of a (according to the hierarchy) related product could be used. In MCRA, the *supertype* conversion step also contains a *hierarchy-supertype* step based on the food hierarchy.

Supertype link (step 5):

- a) **Supertype**: Try to find supertypes base on '\$'-coded strings, e.g., 'xxx\$yyy' is converted to 'xxx'
- b) **Hierarchy-supertype**: try to find the supertype of the current food based on the food hierarchy (i.e., convert the current food to its parent).

Note 1: the supertype conversion step is optional and should be specified in the conversion settings panel.

Note 2: the *hierarchy-supertype* step only applies for foods-without-facets. The reason for this is that for the conversion, the base type of a food-with-facets can be considered as a better conversion candidate than the parent food with the same facets.

Using hierarchy data in the output

Food hierarchy information could be used in presentation of various tables of the output of MCRA. That is, in the tables in which foods data is presented, these records could be grouped based on the hierarchy and/or a tree-like display can be built for the presentation of this data. Tables that are candidate for being extended are, for example, the input data tables foods-as-eaten/modelled foods and the exposure by food-as-eaten/modelled food output tables.

Summarizing over the food hierarchy is many cases not a straightforward task. Consider, for instance, the statistic *number of consumption days* given the artificial hierarchy of *Citrus Fruits* containing two child-nodes *Mandarin* and *King Mandarin*: the number of consumption of *Citrus Fruits* is not "just" the sum of the consumption day of *Mandarin* and *King Mandarin*. A difficulty for summarizing based on a hierarchy arises when a node contains both data and child-nodes with data. E.g., concentrations are defined on the level of *Citrus Fruits* and on the level of *Mandarin*. In this case, the hierarchy view should ideally summarize for both *Citrus Fruits* as data record and *Citrus Fruits* as summary node. An additional complication is the status of facet-coded foods within the hierarchy. In a hierarchical view, foods-with-facets should ideally be added to their base-foods for visualization.

In MCRA, an alternative view (treetable) is added that can display hierarchical data. This alternative view is used to present a hierarchical view based on the foods hierarchy for the consumption input summary tables food as eaten and

modelled food. The data summary methods for these tables are updated such that the data is also summarized per hierarchy-node.

XI 🛱 🎬 🕢							
Food name	Food code	Mean consumption (g)	Mean consumption days (g)	Consumption days	Percentage consumption days	Total weights consumption days	Percentage total weights consumption days
 Fruit and fruit products 	A01BS	167	200	5	83.3 %	5.0	83.3 %
+ Fresh fruit	A04RK	167	200	5	83.3 %	5.0	83.3 %
 Starchy roots or tubers and products thereof, sugar plants 	A00ZR	100	600	1	16.7 %	1.0	16.7 %
Starchy root and tuber products	A011B	66.7	400	1	16.7 %	1.0	16.7 %
 Processed root and tuber products 	A04MJ	66.7	400	1	16.7 %	1.0	16.7 %
Potato boiled	A011P	66.7	400	1	16.7 %	1.0	16.7 %
 Potato boiled Tuber (as part-nature) 	A011P#F02.A067V	16.7	100	1	16.7 %	1.0	16.7 %
 Potato boiled Tuber (as part-nature), Potatoes, Boiling 	A011P#F02.A067V\$F27.A00ZT\$F28.A07GL	16.7	100	1	16.7 %	1.0	16.7 %
 Potato boiled Tuber (as part-nature), Potatoes, Boiling 	A011P#F02.A067V\$F28.A07GL\$F27.A00ZT	16.7	100	1	16.7 %	1.0	16.7 %
 Potato boiled Tuber (as part-nature), Potatoes, Boiling, Baking 	A011P#F02.A067V\$F27.A00ZT\$F28.A07GL\$F28.A07GX	16.7	100	1	16.7 %	1.0	16.7 %
 Starchy roots and tubers 	A00ZS	33.3	200	1	16.7 %	1.0	16.7 %
 Tubers 	A04MC	33.3	200	1	16.7 %	1.0	16.7 %
 Potatoes 	A00ZT	33.3	200	1	16.7 %	1.0	16.7 %
 Potatoes Potatoes (food source plant), Tuber (as part-nature) 	A00ZT#F01.A05KG\$F02.A067V	16.7	100	1	16.7 %	1.0	16.7 %
 Potatoes Potatoes (food source plant), Tuber (as part-nature), Baking 	A00ZT#F01.A05KG\$F02.A067V\$F28.A07GX	16.7	100	1	16.7 %	1.0	16.7 %

Figure 2.2: Hierarchy view for the foods as eaten input summary table.

If a node contains both data and a child record, then this node is split-up in two nodes: a summary node that summarizes the data of the node and all of its child nodes, and a data record with the string "(unspecified)" added as a child of this summary node. See Figure 2.2 for an example (*Citrus Fruits versus Citrus Fruits (unspecified)*). In MCRA, foods-with-facets are added as child nodes of the foods-without-facets.

Food unit weights

Food unit weights specify the standard weights of food units. E.g., the standard weight of an apple. This unit weight may be specified as the weight of the whole food (raw agricultural commodity/RAC) or the weight of the edible portion (EP), e.g., without peel. Unit weights are specified in the table *table for FoodUnitWeights* and used in combination with *unit variability factors* to account for unit-to-unit variation in concentrations between single units of the same food in *single value dietary exposures assessments* and (*individual*) *dietary exposures assessments*.

Food unit weights can be location specific or specified as overall (default) unit weights. For some models, e.g., the *IESTI model*, location specific unit weights are preferred over overall unit weights. The overall unit weights are then used when no location specific uses are available. For other methods, only overall unit weights are used. If, for a food, an overall unit weight is not available, but there are location specific unit weights available, then the overall unit weight is computed as the average weight of the location specific unit weights (similar to EFSA PRIMo revision 3 [[EFSA, 2018]]).

Note: Note that in earlier versions of the software, food unit weights were specified in the *table for FoodProperties*. Although this is still possible, the recommended way of specifying unit weights is in the *table for FoodUnitWeights*. If, for a food, unit weights are specified in both tables, then the unit weights specified in the *table for FoodUnitWeights* have priority. The unit weights specified in the *table for FoodProperties* are then only used as fallbacks for the overall unit weight when no overall unit weight is specified in the *table for FoodUnitWeights*.

2.1.3 Populations

Populations are groups of human individuals that are the scope of exposure or risk assessments. Optional descriptors of populations are location (e.g. a country), time period (start date, end date), age range and gender. Example: the French population in 2005-2007 of women of child-bearing age (18-45 yr).

Output of this module is used by: Consumptions Single value consumptions Consumptions by modelled food Dietary exposures Single value dietary exposures Non-dietary exposures Exposures Human monitoring analysis Risks Single value risks

Populations data formats

Populations

Populations are primary entities of the data model.

Populations

Populations identify human groups, and e.g. dietary, nondietary and human monitoring surveys. Optionally, a name and description can be added. Population can be restricted to a certain time period. AgeMin, AgeMax and Gender are optional properties of a population.

Name	Туре	Description	Aliases	Required
idPopulation	AlphaNumeric(50)	Unique identification code of	IdPopulation,	Yes
		the population.	PopulationId,	
			Code, Id	
Name	AlphaNumeric(100)	The name of the population.	Name,	No
			PopulationName	
Description	AlphaNumeric(200)	Description of of the	Description	No
		population.		
Location	AlphaNumeric(50)	Location.		No
StartDate	DateTime	Starting date of the specific	StartDate	No
		time window marking this		
		population.		
EndDate	DateTime	End date of the specific time	EndDate	No
		window marking this		
		population.		
AgeMin	Integer	Inclusive minimum bound (in	AgeMinimum	No
		years) of the specific age		
		group of this population.		
AgeMax	Integer	Inclusive maximum bound (in	AgeMaximum	No
		years) of the specific age		
		group of this population.		
Gender	AlphaNumeric(50)	Gender levels of this	Sex	No
		population.		
NominalBody-	Numeric	Nominal body weight (in kg)	NominalBody-	No
Weight		of the individuals of this	Weight,	
		population.	BodyWeight	
		·		

Table 2.15:	Table definition	for Populations.
-------------	------------------	------------------

Table aliases: Populations, Population.

Populations settings

Selection settings

Table 2.16: Selection settings for module Populations.
--

Name	Description
Population	Specifies which population is selected.
Filter individual days	Specifies a subset selection on specific dates.

Populations as data

Populations are provided as data.

• Populations data formats

2.1.4 Responses

Responses are measurable entities in test systems. Responses are used to represent effects (see effect representations) and their measured values are collected in dose response data.

This module has as primary entities: Test systems

Output of this module is used by: Dose response models Dose response data Effect representations

Responses data formats

Responses

A response is a measurable endpoint on in a test system. E.g., in a rat test system a response may be the percentage of fatty hepatocytes observed after 90 days. Responses are defined in the responses table.

Responses

Each response is identified by a unique code (idResponse) in a code system of choice, a name, and a description. Also, each response should be linked to a test system (idTestSystem) on which the response is measured. Responses can be of various types (ResponseType), e.g., ContinuousMultiplicative (= non-negative real values using a ratio scale), ContinuousAdditive (= real values using an interval scale), Ordinal, Quantal, or Binary. For continuous variables, the response unit (ResponseUnit) is also relevant. Additionally, also a reference to the test method guideline, e.g., standardised assay kit may also be specified (GuidelineMethod).

Name	Туре	Description	Aliases	Required
idResponse	AlphaNumeric(50)	Unique identification code of	idResponse,	Yes
		the response. In the EuroMix	ResponseId,	
		data collection, a EuroMix	Response, Id	
		coding system has been set up		
		in which the id of the test		
		system prefixes the id of the		
		response. E.g.,		
		'HepaRG-PCR-PPARA',		
		'RatWEC-PCR-CYP26a1'		
		and 'MouseDevelopmental-		
		FacialPrimordia-malformed-		
		E9'.		
CodeSystem	AlphaNumeric(100)	Identifier of the coding	CodeSystem	No
		system of the response code.		
Name	AlphaNumeric(100)	Name of the response.	Name	No
Description	AlphaNumeric(200)	Additional description or label	Description	No
		of the response.		
idTestSystem	AlphaNumeric(50)	Unique identification code of	idTestSystem,	Yes
		the test system.	idSystem,	
			SystemId,	
			TestSystem	
Guideline-	AlphaNumeric(200)	Reference to the test method	Guideline-	No
Method		guideline, e.g., standardised	Method	
		assay kit.		
ResponseType	ResponseTypes	The data type of the response	ResponseType	Yes
		measurements (e.g.,		
		continuous multiplicative,		
		ordinal, categorical).		
ResponseUnit	AlphaNumeric(100)	If the response type is	ResponseUnit	No
		Continuous, then this should		
		be the unit of the response,		
		e.g., kg.		

Table 2.17: Table definition for Responses.

Table aliases: Responses, Response.

Responses settings

Selection settings

Table 2.18: Selection settings for module Responses.

Name	Description
Response(s)	The response(s) of interest.

Responses as data

A response is a measurable endpoint defined in a test system. It has a unit and a measurement type (e.g., continuous non-negative, quantal).

• Responses data formats

2.1.5 Substances

Substances are chemical entities that can refer to: 1) active substances such as investigated in toxicology; 2) measured substances such as defined in specific analytical methods. MCRA assessments can have one or more substances as the scope. When more than one substance is specified, there is an option to perform a cumulative assessment. In that case one of the substances has to be indicated as the index/reference substance, and results will be expressed in equivalents of the index substance.

Output of this module is used by: Concentrations Concentration distributions Single value concentrations Processing factors Unit variability factors Occurrence patterns Occurrence frequencies Substance authorisations Substance conversions Deterministic substance conversion factors Concentration limits Concentration models Modelled foods Focal food concentrations Food conversions Consumptions by modelled food High exposure food-substance combinations Dietary exposures Single value dietary exposures Non-dietary exposures Exposure fixtures Human monitoring data Human monitoring analysis QSAR membership models Molecular docking models Kinetic models Active substances Relative potency factors Hazard characterisations Points of departure Dose response models Dose response data Inter-species conversions Intra species factors Risks Single value risks

Substances data formats

Substances

Substances are primary entities of the data model. Substance intakes are of main interest in exposure assessments and the effect of intake on human health is of interest in risk assessments. In the substances table, the substance entities and other relevant substance properties that are relevant for the assessment at hand should be defined.

Substances

Each substance should have a unique identification code (idSubstance), and optionally, a name and description may be used for a more detailed description of the entity. Additional properties, such as the molecular mass (MolecularMass) and Cramer class (CramerClass) may also be specified. Example: Captan (idSubstance RF-0061-001-PPP) has MolecularMass 300.5922 and CramerClass 3.

Name	Туре	Description	Aliases	Required
idSubstance	AlphaNumeric(50)	The unique identification code of the substance. This code may be from an existing coding system, such as CAS-codes or Param codes of EFSA, or it may be a used-defined code.	idSubstance, SubstanceId, Substance, Code, Id	Yes
Name	AlphaNumeric(100)	The substance name.	Name, SubstanceName, PesticideName	No
Description	AlphaNumeric(200)	Substance description.	Description	No
Concentration- Unit	ConcentrationUnits	Contains a coding to determine the default unit in which concentrations for this substance are expressed.	Concentration- Unit, Unit, Reference- Concentration- Unit	No
CramerClass	Integer	The Cramer class of the substance.	CramerClass	No
MolecularMass	Numeric	The molecular (molar) mass.	MolecularMass, Mass, MolarMass, Molecular- Weight, MolarWeight	No

Table 2.19: Table de	finition for	Compounds.
----------------------	--------------	------------

Table aliases: Substances, Substance.

Substances settings

Selection settings

Name	Description
Multiple substances analysis	Specifies whether the assessment involves multiple substances.
Index substance	The substance of interest or index substance.

Substances as data

Substances are provided as data (code, name).

• Substances data formats

2.1.6 Test systems

Test systems are biological or artificial systems used for assessing hazard in relation to chemical exposure from substances in varying doses. Test systems may refer to 1) in-vivo test systems (e.g. a rat 90-day study, a human biomonitoring study); 2) in-vitro test systems (e.g. HepaRG cells).

Output of this module is used by: Responses Dose response models Dose response data

Test systems data formats

Test Systems

Test systems are the biological systems (e.g., animals) or in-vitro systems on which responses related to health effects can be measured.

Test Systems

Each test system should have a unique identification code (idSystem), and (optionally) a name and a description. The test system's type (TestSystemType) indicates the type whether the test system is an in-vivo test system (in which case it is a model for external exposure) or any of a range of other, in-vitro, options (cell-line, etc., which all will be interpreted as models for internal exposure). Additionally, if applicable, the organ (e.g., liver) of the test system and the route of exposure (RouteExposure) for in-vivo test systems (oral, dermal or inhalation) may be specified.

Name	Туре	Description	Aliases	Required
idTestSystem	AlphaNumeric(50)	Unique identification code of the test system.	idTestSystem, idSystem, Id, Code	Yes
CodeSystem	AlphaNumeric(50)	Identifier of the code system of the test systems.	CodeSystem	No
Name	AlphaNumeric(100)	Name of the test system.	Name	No
Description	AlphaNumeric(200)	Additional description or label of the test system.	Description	No
TestSystem- Type	TestSystemTypes	The type of the test system, i.e., in-vivo, cell-line, etc.	TestSystem- Type, SystemType	No
Organ	AlphaNumeric(100)	If applicable, the organ that the cells originate from associated with the in vitro test-system.	Organ	No
Species	AlphaNumeric(100)	If applicable, the species associated with the test-system.	Species	No
Strain	AlphaNumeric(100)	If applicable, the strain of the species associated with the test-system.	Strain	No
RouteExposure	ExposureRouteTypes	If applicable, the route of exposure associated with the in vivo test-system, oral, dermal, inhalation, s.c., i.v.	ExposureRoute- Type, ExposureRoute, RouteExposure	No
Guideline- Method	AlphaNumeric(200)	Reference to test guideline.	GuidelineStudy	No
Reference	AlphaNumeric(200)	External reference(s) to other sources containing more information about the test system. E.g., publications, website, documents.	Reference	No

Table 2.21: Table definition for TestSystems.

Table aliases: TestSystems, TestSystem, Systems, System.

Test systems as data

Test systems are provided as data.

• Test systems data formats

2.2 Consumption modules

Consumption modules specify the *consumptions* or *single value consumptions* of *foods* by surveyed individuals in *populations*. Foods can be related to each other using *food recipes*.

2.2.1 Consumptions

Consumptions data are the amounts of foods consumed on specific days by individuals in a food consumption survey. For acute exposure assessments, the interest is in a population of person-days, so one day per individual may be sufficient. For chronic exposure assessments, the interest is in a population of persons, so preferably two or more days per individual are needed.

This module has as primary entities: Populations Foods

Output of this module is used by: Food conversions Consumptions by modelled food

Consumptions data formats

Consumption data is often collected in 24-hour dietary recall studies and contains the food consumptions and consumption amounts for a number of individuals on a number of days. For each of the individuals, the bodyweight should be specified, and optionally also age, sex, and other properties may be recorded. If applicable, sampling weights may also be specified that can be used to correct the sample of individuals in the survey to a more representative sample of the targeted population. The consumption amounts are usually expressed in grams, but may also be expressed in alternative units of plates, cups, or spoons. Optionally, the uncertainty of food consumption quantifications can be specified, see [[Souverein et al., 2011]].

Consumptions

Consumption surveys are described using three tables: FoodSurveys, Individuals, and Consumptions. Individuals are linked to food surveys using the survey code (idFoodSurvey), and consumptions are linked to individuals using the individual codes (idIndividual). The food codes used to identify the consumed foods should match with the codes provided by the foods entity definitions.

Food consumption surveys

The records of the food consumption surveys table contain the ids, names, descriptions, and other relevant metadata of consumption surveys.

Name	Туре	Description	Aliases	Required
idSurvey	AlphaNumeric(50)	Unique identification code of	idSurvey,	Yes
		the food consumption survey.	idFoodSurvey,	
			Survey,	
			FoodSurvey,	
			SurveyId,	
			FoodSurveyId,	
			Code, Id	
Name	AlphaNumeric(100)	The name of the food	Name,	No
		consumption survey.	SurveyName	
Description	AlphaNumeric(200)	Description of the food	Description	No
		consumption survey.		
Location	AlphaNumeric(50)	The location or country where	Location,	No
		survey is held. It is	Country	
		recommended to use ISO		
		Alpha-2 country codes.		
BodyWeight-	Body Weight Units	The unit of bodyweight of the	BodyWeight-	No
Unit		individuals of the survey: kg	Unit,	
		(default) or g.	UnitBody-	
			Weight,	
			WeightIn	
AgeUnit	AgeUnit	The unit of age, i.e., year or	UnitAge, agein,	No
		month.	AgeUnit	
Consumption-	ConsumptionUnits	The unit of the	AmountUnit,	No
Unit		use/consumption amounts of	UnitAmount,	
		the consumptions of the	AmountUnit,	
		survey: g (default) or kg or	Consumption-	
		CustomUnit (see table food	Unit	
		consumption quantifications		
		table).		
StartDate	DateTime	The start date of the survey.	StartDate	No
EndDate	DateTime	The end date of the survey.	EndDate	No
NumberOf-	Integer	The number of days each	NumberOf-	Yes
SurveyDays	_	individual participated in the	SurveyDays,	
		survey.	NDaysInSurvey	
idPopulation	AlphaNumeric(50)	Unique identification code of	IdPopulation,	No
÷	· · · /	the population.	PopulationId	

Table 2.22: Table definition for FoodSurveys.

 $Table\ aliases:\ FoodConsumptionSurveys,\ ConsumptionSurveys,\ FoodSurveys,\ Surveys.$

Individuals

The individuals of a survey are recorded in the individuals table.

Name	Туре	Description	Aliases	Required
idIndividual	AlphaNumeric(50)	Unique identification code of	idIndividual,	Yes
		the individual.	IndividualId,	
			Individual, Id	
idFoodSurvey	AlphaNumeric(50)	The identification code / short	idSurvey,	Yes
		name of survey.	idFoodSurvey,	
			Survey,	
			FoodSurvey,	
			SurveyId,	
			FoodSurveyId,	
			SurveyCode	
BodyWeight	Numeric	The body weight of the	BodyWeight,	Yes
		individual.	Weight	
Sampling-	Numeric	The sampling weight for an	SamplingWeight	No
Weight		individual (default = 1).		
NumberOf-	Integer	The number of days the	NumberOf-	No
SurveyDays		individual participated in the	SurveyDays,	
		survey.	NumberOfDays-	
			InSurvey,	
			DaysInSurvey,	
			NDaysInSurvey	
Age	Numeric	The age of the individual.	Age	No
Gender	AlphaNumeric(50)	The gender of the individual.	Gender	No
		It is recommended to use the		
		codes Male/Female for coding		
		the gender.		
Other individual		Other individual properties		No
properties		can be added just like the		
		fields age and gender. These		
		properties are automatically		
		parsed as co-factors or		
		co-variables.		

 $Table\ aliases:\ Individuals,\ Survey Individuals,\ Consumption Survey Individuals,\ Food Consumption Survey Individuals.$

IndividualDays

The individuals and days in the survey are recorded in the individualDays table.

Name	Туре	Description	Aliases	Required
idIndividual	AlphaNumeric(50)	The identification code of the	idIndividual,	Yes
		individual.	IndividualId,	
			Individual, Id	
idDay	AlphaNumeric(50)	Identification code of the day	idDay, DayId,	Yes
		of consumption, sequential	Day,	
		number	DayOfSurvey	
SamplingDate	DateTime	The date of the consumption.	Date,	No
			SamplingDate,	
			SurveyDate	

Table 2.24: Table definition for IndividualDays.

 $Table\ aliases:\ Individual Days,\ Survey Individual Days,\ Consumption Survey Individual Days,\ Food Consumption Survey Individual Days.$

Individual properties

Individual properties, additional columns that can also be specified as additional columns in the Individuals table

Name	Туре	Description	Aliases	Required
Name	AlphaNumeric(50)	The name of the property.	Id	Yes

Table 2 25.	Table	definition	for	IndividualProperties.
	Table	deminition	101	multiluair toperties.

Table aliases: IndividualProperties, IndividualProperty.

Individual property values

Individual property values, additional columns that can also be specified as additional columns in the Individuals table

Tuble 2.20. Tuble deministration reporty values.					
Name	Туре	Description	Aliases	Required	
idIndividual	AlphaNumeric(50)	The identification number of	Id	Yes	
		the Individual.			
PropertyName	AlphaNumeric(50)	The name of the property.	Name	Yes	
TextValue	AlphaNumeric(50)	The value of the property as		No	
		text value.			
DoubleValue	Numeric	The value of the property as		No	
		number.			

 Table 2.26: Table definition for IndividualPropertyValues.

Table aliases: IndividualPropertyValues, IndividualPropertyValue.

Consumptions

The individual consumptions are recorded in the consumptions table.

Name	Туре	Description	Aliases	Required
idIndividual	AlphaNumeric(50)	The unique identification code	idIndividual,	Yes
		of the consumer (individual).	IndividualId,	
			Individual	
idFood	AlphaNumeric(50)	The food code (food as eaten	idFood, Food,	Yes
		code).	FoodId,	
			FoodConsumed,	
			FoodAsEaten	
idUnit	AlphaNumeric(50)	Identification code of the unit	idUnit, Unit,	No
		in which the food is consumed	UnitId	
		(e.g. plate, cup, spoon).		
idDay	AlphaNumeric(50)	Identification code of the day	idDay, DayId,	Yes
		of consumption, sequential	Day,	
		number	DayOfSurvey	
idMeal	AlphaNumeric(50)	Identification code of the meal	idMeal, MealId,	No
		(eating occasion within a day).	Meal	
Amount	Numeric	The consumed portion of food	Amount,	Yes
		in g (default) or kg or quantity	Amount-	
		of a plate, cup, spoon. Days	Consumed	
		without consumptions are not		
		recorded.		

Table 2.27: Table definition for Consumptions.

 $Table \ a liases: \ FoodConsumptions, \ FoodConsumption, \ Consumptions, \ Consumption.$

Food consumption quantifications

Food consumption quantifications record information about food consumption quantities that are associated with unit-consumptions of foods.

Name	Туре	Description	Aliases	Required
idFood	AlphaNumeric(50)	The food code of the	idFood, FoodId,	Yes
		quantification.	Food	
idUnit	AlphaNumeric(50)	The code of the unit of	idUnit, UnitId,	Yes
		consumption. E.g spoon,	Unit	
		plate, cup. Units may depend		
		on food.		
UnitWeight	Numeric	The unit weight/portion size	UnitWeight	Yes
		of the food, specified in		
		grams.		
UnitWeight-	Numeric	The uncertainty in unit	UnitWeight-	No
Uncertainty		weight/portion size (%).	Uncertainty,	
			UnitWeight%	
Amount-	Numeric	The uncertainty in amount	Amount-	No
Uncertainty		consumed (%). The label	Uncertainty,	
		'general' specifies a default	Amount%	
		value for the uncertainty when		
		specific information for		
		combinations of food and unit		
		in food consumptions table is		
		not available.		

Table 2.28: Table definition for FoodConsumptionQuantifications.	
--	--

Table aliases: FoodConsumptionQuantifications, FoodConsumptionQuantification.

Consumptions calculation

The consumptions module offers a number of options and filters that specify how the concentration data should be included in the assessment.

- When individual sampling weights have been specified in the *data*, the option *ignore sampling weights* can be used to ignore these sampling weights so that they won't be used in the calculations.
- It is possible to restrict to consumptions of specific foods.
- Depending on whether the *exposure type* is acute or chronic, the *ignore sampling weights*, the individuals (chronic) or individual-days (acute) can be filtered in several ways. It is possible to restrict to the *consumers or consumer days only*, which can relate to any consumption or consumptions of *specific (focal) products*.

Consumptions settings

Selection settings

Table 2.29. Selection settings for module Consumptions.			
Name	Description		
Risk type	The type of exposure considered in the assessment; acute (short		
	term) or chronic (long-term).		
Food survey	The food consumption representative for the population of		
	interest.		
Restrict population to	Specifies whether the population should be restricted to the		
consumers or consumer days	individuals (chronic) or individual days (acute) that have non-zero		
only	consumption.		
Restrict population to	Specifies whether the population should be restricted to the		
consumers or consumer days	individuals (chronic) or individual days (acute) consuming any of		
with consumptions of specific	the foods of the specified subset.		
foods			
Selected foods-as-eaten	Set of consumed foods that are of particular interest for restricting		
	the consumers / consumption days.		
Consumption subset: restrict to	If checked, then the consumptions are restricted to those of the		
consumptions of specific foods	specified food-as-eaten subset.		
Selected foods-as-eaten	Set of consumed foods that are of particular interest.		
Ignore sampling weights	If checked, individual sampling weights are not used (sampling		
	weight = 1). If unchecked, the specified sampling weights are		
	used.		

Table 2.29:	Selection	settings	for r	nodule	Consumption	S.
10010 2.27.	Selection	settings	101 1	nouure	consumption	

Uncertainty settings

Table 2.30:	Uncertainty	settings	for module	Consumptions.

Name	Description			
Resample individuals	Individual data are resampled from the original database using the			
	bootstrap methodology (Efron 1979, Efron & Tibshirani 1993).			
Resample portion sizes	Specifies whether portion sizes should be resampled based on			
	food consumption quantification data, see (Souverein et al. 2011).			

Consumptions uncertainty

In MCRA, in an *acute exposure* assessments, individual consumption day data are *resampled*, thus preserving the multivariate consumption patterns and associated weights and/or other individual characteristics. In MCRA we resample the set of individuals x number of survey days. We think that this implementation better reflects the notion of acute exposure which is expressed as the normalized intake per day. For *chronic exposure* assessments the resampling algorithm remained unchanged and the set of individuals (with corresponding days) is *resampled*.

Consumptions as data

Consumptions data are the amounts of foods consumed on specific days by individuals in a food consumption survey.

- Consumptions data formats
- Consumptions calculation

2.2.2 Food recipes

Food recipes data specify the composition of specific foods (typically: foods-as-eaten) in terms of other foods (intermediate foods or foods-as-measured) by specifying proportions in the form of a percentage.

This module has as primary entities: Foods

Output of this module is used by: Food conversions

Food recipes data formats

Food recipes

Recipe data to specify the ingredients of foods. Food recipes can be used to describe the ingredients of a composite food (e.g., of apple pie), or to specify the amount of a primary ingredient needed to obtain 100g of the food (e.g., grapes to raisins). Recipe is commonly used recursively (e.g., apple pie contains apple and flour, flour contains wheat).

Recipes

Name	Туре	Description	Aliases	Required
idFromFood	AlphaNumeric(50)	The code of the composite	idFromFood,	Yes
		food (from-code), i.e., the	FromFoodId,	
		code of the food for which the	FromFood,	
		ingredient(s) are specified.	FoodFrom,	
			Food	
idToFood	AlphaNumeric(50)	The code of the ingredient	idToFood,	Yes
		food (to-code).	ToFoodId,	
			ToFood,	
			FoodTo,	
			Ingredient	
Proportion	Numeric	Proportion of each ingredient	Proportion,	Yes
		in the food (%).	Proportion%	
idPopulation	AlphaNumeric(50)	Unique identification code of	IdPopulation,	No
		the population.	PopulationId	

Table aliases: FoodTranslations, FoodTranslation, FoodCompositions, FoodComposition.

Food recipes as data

Food recipes are provided as data in the form of simple composition tables.

• Food recipes data formats

2.2.3 Market shares

Market shares data specify for a given food, percentages of more specific foods (subfoods, e.g. brands) representing their share in a market. Market shares are used when consumption data are available at a more generalised level than concentration data.

This module has as primary entities: Foods

Output of this module is used by: Food conversions

Market shares data formats

MarketShares

Describes the shares (proportions) in a market.

Market shares

Market shares main table.

Name	Туре	Description	Aliases	Required
idFood	AlphaNumeric(50)	The subtype of the food.	idFood, FoodId, Food, FoodType	Yes
Percentage	Numeric	Market share of each subtype (%)	Percentage, Marketshare- Percentage, MarketShare, MarketShare- Percentage, MarketShare%	Yes
BrandLoyalty	Numeric	A parameter used in brand loyalty modelling, where 0 (default) specifies no brand loyalty (on each eating occasion a random selection of the next lower level in the hierarchy of food codes), and 1 specifies absolute brand loyalty (on subsequent eating occasions the same selection of the next lower level in the hierarchy of food codes).	BrandLoyalty	No

Table 2.32:	Table	definition	for	MarketShares.
-------------	-------	------------	-----	---------------

Table aliases: MarketShares, MarketShare, FoodMarketShares, FoodMarketShare.

Market shares as data

Market shares are provided as data in the form of percentages.

• Market shares data formats

Market shares and brand loyalty

Sometimes measurements of substances in food are available at a more detailed food coding level than consumption data. For example, measurements may have been made for specific brands of a food whereas the consumption survey did not record the brand. MCRA allows to specify market share data for subtypes of a food (e.g. A\$1, A\$2, A\$3 are three brands of food A), and to calculate acute exposure based on such market shares.

For chronic assessments **brand loyalty** should be specified according to a simple Dirichlet model [[Goodhardt et al., 1984]]. Technically, the Dirichlet model for brand choice needs nbrand parameters α_i (which should be positive real numbers). The average brand choice probability for each brand is

 α_i/S

where

$$S=\sum \alpha_i$$

By definition, the market shares m_i should be proportional to the brand choice probabilities, and thus to the parameters α_i . Thus means that S, the sum of the alphas, is the only additional parameter that should be specified, and indeed this is the parameter that determines brand loyalty. S = 0 corresponds to absolute brand loyalty, and brand loyalty decreases with increasing S. We define $L = (1 + S)^{-1}$ as an interpretable brand loyalty parameter, where now L = 0 and L = 1 correspond to the situations of no brand loyalty and absolute brand loyalty, respectively. Given empirical or parametric distributions of consumption and concentration values, the algorithm for chronic exposure assessment now operates as follows:

- 1. Simulate consumptions for a large number n of individuals.
- 2. Simulate n selection probabilities from the Dirichlet distribution
- 3. For each individual, simulate d brand choices from a multinomial distribution using the individual specific selection probabilities from step 2.
- 4. For all individuals and days simulate values from the appropriate concentration distribution.
- 5. Multiply consumption with concentration to obtain exposure.

2.2.4 Single value consumptions

Single value consumption data are the single value amounts (Large Portion, Mean Consumption, p97.5Consumption) of modelled foods (foods-as-measured) consumed in a population.

This module has as primary entities: Populations Foods

Output of this module is used by: Single value dietary exposures

Single value consumptions data formats

Single value consumptions data provides a single per-individual-day and per-food consumption amount for a population. Also the bodyweight should be specified, and optionally also age, sex, and other properties may be recorded. The consumption amounts are usually expressed in grams, but may also be expressed in alternative units of plates, cups, or spoons. Optionally, the uncertainty of food consumption quantifications can be specified, see [[Souverein et al., 2011]].

Single value consumptions

Single value consumptions are described using one table: PopulationConsumptionSingleValues.

Population consumption single values

Population consumption single values describe population food consumptions in the form of single value statistics.

Name Type		Description	Aliases	Required	
idPopulation	AlphaNumeric(50)	Unique identification code of the population.	IdPopulation, PopulationId	Yes	
1 17 1	$\mathbf{A} = \mathbf{b} \mathbf{A} + \mathbf{b} \mathbf{b} \mathbf{b} \mathbf{b} \mathbf{b} \mathbf{c} \mathbf{c} \mathbf{c} \mathbf{c} \mathbf{c} \mathbf{c} \mathbf{c} c$			N/	
idFood	AlphaNumeric(50)	The unique identification code	idFood,	Yes	
		of the consumed food.	FoodCode, Food		
Value type of	Consumption Value-	The value type of this	Consumption-	Yes	
the single value	Types	consumption value.	Туре,		
consumption			ValueType,		
amount.			Consumption-		
			ValueType,		
			Consumption-		
			SingleValue-		
			Туре		
Percentile	Numeric	The percentile (if	Percentile	No	
		consumption value type is a			
<u>a</u> .:	NY .	percentile).	A		
Consumption-	Numeric	The consumed amount.	Amount,	Yes	
Amount			Consumption,		
			Consumption-		
			Amount,		
			Amount-		
			Consumed		
Consumption-	ConsumptionIntake-	The unit of the consumption	AmountUnit,	No	
Unit	Units	amount.	UnitAmount,		
			Consumption-		
			Unit		
Reference	AlphaNumeric(200)	Reference to the source from	Reference,	No	
		which this value is obtained.	References,		
			Source, Sources		

Table 2.33: Table definition for PopulationConsumptionSingleValues.

Table aliases: ConsumptionSingleValues, SingleValueConsumptions, PopulationConsumptionSingleValues, PopulationConsumptionValues.

Single value consumptions calculation

Single value consumptions can be supplied *as data* or computed. When single value consumptions are computed from *consumptions by modelled food*, then the mean, median and large portion (p97.5 percentile) are computed for all modelled food consumption distributions. Besides these statistics, also the mean bodyweight of the population is computed. The following options are relevant in this calculation:

- Set the *risk type* option to *acute* if the consumptions should be based on the individual-day distributions. Otherwise, choose *chronic* to base them on the distributions aggregated by individual.
- Checking the *apply processing factors* option will compute the single value consumptions for the processed foods. When using this option, the output will also show a reverse yield factor, that is the ratio of the quantity of the raw commodity required to to obtain the processed commodity. Note, when no processing factors are available, the single-value consumption amounts of processed foods are expressed in terms of the processed

commodities. The yield factor, i.e. the factor for translating the processed amount to the unprocessed amount, is not applied. In the IESTI calculations (and also chronic single-value calculations), calculations are done using the processed amounts, this is on the level of raw processed foods.

- Check the *restrict population to consumers or consumer days only (modelled-food)* option to compute the single value consumption statistics for each food based on the food consumers only. Note that checking this option will also affect the computed bodyweight, which is then computed by food based on the food-consumers only and can be different for each food.
- There is also an option to *ignore sampling weights* in the calculation.
- Check the *standardise consumption with body weight before calculation of single values or afterwards (with mean bodyweight)* option to compute the single value consumptions from the per bodyweight distribution. If unchecked, the per-person distribution will be used for computing the statistics. Note that although the results are reported per-day, the statistics are established by multiplying the statistics obtained from the per bw distribution by the bodyweight.

Single value consumptions settings

Calculation settings

Name	Description	
Risk type	The type of exposure considered in the assessment; acute (short	
	term) or chronic (long-term).	
Restrict population to	Specifies whether the population should be restricted to the	
consumers or consumer days	individuals (chronic) or individual days (acute) that have non-zero	
only	consumption.	
Ignore sampling weights	If checked, individual sampling weights are not used (sampling	
	weight = 1). If unchecked, the specified sampling weights are	
	used.	
Use standardised consumption	Specifies whether single values are calculated on individual	
distributions before calculation	consumptions standardised with body weight and then multiplied	
of single values	by the mean body weight. Otherwise, single values are calculated	
	on the original consumptions (per day). Note that both methods	
	lead to different estimates for the single value.	
Apply processing factors	Specified in table ProcessingFactor. If checked, processing factors	
	are applied. Concentrations in the consumed food may be	
	different from concentrations in the modelled food in monitoring	
	programs (typically raw food) due to processing, such as peeling,	
	washing, cooking etc. If unchecked, no processing information is	
	used. This is in most (though not all) cases a worst-case	
	assumption	
Restrict population to	Specifies whether the population should be restricted to the	
consumers or consumer days	individuals (chronic) or individual days (acute) with consumptions	
only (food-as-measured)	containing any of the foods-as-measured.	

Table 2.34: Calculation settings for module Single value consumptions.

Single value consumptions as data

Single value consumption data are the single value amounts of modelled foods (foods-as-measured) consumed in a population.

• Single value consumptions data formats

Calculation of single value consumptions

Single value consumptions are calculated as a percentile (p97.5 or p99) or mean of the modelled food consumption distribution. For an acute single value dietary exposure assessment, this is the individual day consumption distribution, for chronic single value dietary exposure assessment, the individual consumption distribution is used.

• Single value consumptions calculation

Inputs used: Consumptions by modelled food

Settings used

• Calculation Settings

2.3 Occurrence modules

The basic occurrence data are *concentrations* for *substances* in *foods*, sometimes specified separately for a focal food as *focal food concentrations*. In some cases *concentration limits* are used as a stand-in when data are missing.

Concentration data are recalculated (if needed) as *active substance concentrations* in *modelled-foods*. If substance concentrations are not specified directly for the *active substances*, then they are converted using *substance conversions* and/or specified authorised *occurrence patterns*. The composition of mixed samples in total diet studies is described in *total diet study sample compositions*. *Food extrapolation rules* specify if insufficient data for a food can be suppleted with data from another food. From these basic data the list of *modelled-foods* is derived.

Active substance concentrations in modelled-foods are modelled in concentration models, optionally allowing for occurrence pattern models. In addition, processing factors and unit variability factors can be provided for further use in dietary exposure assessment.

2.3.1 Concentration distributions

Concentration distributions describe substance concentrations on foods in the form of summary statistics.

This module has as primary entities: Foods Substances

Output of this module is used by: Concentration models Dietary exposures

Concentration distributions data formats

Concentration distributions

Concentration distributions describe substance concentrations on foods in the form of summary statistics. These distributions can be characterised by a mean and a dispersion factor, the standard deviation or, preferably, a percentile point e.g. p95.

Concentration distributions

Substance concentrations on foods specified in the form of summary statistics.

Name	Туре	Description	Aliases	Required	
idFood	AlphaNumeric(50)	Food code, the raw	idFood	Yes	
		agricultural commodity.			
idSubstance	AlphaNumeric(50)	The code of the substance.	idSubstance,	Yes	
			SubstanceId,		
			SubstanceCode,		
			Substance		
Mean	Numeric	The mean of (monitoring)	Mean	Yes	
		samples, on the original scale			
		(in mg/kg).			
CV	Numeric	Coefficient of variation, for	CV	No	
		samples of the size of the			
		TDS pooled amount.			
Percentile	Numeric	The percentile at the point	Percentile	No	
		specified by the percentage.			
Percentage	Numeric	The percentage that belongs	Percentage No	No	
		to the given the percentile,			
		e.g., 95 (in mg/kg).			
Limit	Numeric	The specified norm value or	Limit	No	
		limit value (in mg/kg).			
Concentration-	ConcentrationUnits	The unit of the limit value	Concentration-	No	
Unit		(default mg/kg).	Unit, Unit		

Table 2.35:	Table	definition	for	ConcentrationDistribution	ıs.
-------------	-------	------------	-----	---------------------------	-----

Table aliases: ConcentrationDistributions, ConcentrationDistribution.

Concentration distributions as data

Concentration distributions describe substance concentrations on foods in the form of summary statistics.

• Concentration distributions data formats

2.3.2 Concentration limits

Concentration limits specify (legal) limit values for substance concentrations on foods and are sometimes used as conservative values for concentration data. In the framework of pesticides the legal Maximum Residue Limit (MRL) is the best known example.

This module has as primary entities: Foods Substances

Output of this module is used by: Concentrations Single value concentrations Concentration models Modelled foods

Concentration limits data formats

The concentration limits table describes limit values (e.g., MRLs) for specific food/substance combinations. This data may be used, for instance, for the food/substance combinations for which no concentration data is available. The food codes (idFood) and substance codes (idSubstance) should match the codes of the foods and substances table respectively.

Concentration limits

Concentration limits are concentration limit values for specific food and substance combinations originating from regulations (e.g., MRLs). This data may be used, for instance, for the food/substance combinations for which no concentration data is available.

Concentration limits

The food codes (idFood) and substance codes (idSubstance) should match the codes of the foods and substances table respectively.

Name Type Description		Description	Aliases	Required	
idFood	AlphaNumeric(50)	Code of the food of this residue limit definition.	idFood, FoodId, Food	Yes	
10.1				X 7	
idSubstance	AlphaNumeric(50)	Code of the substance of this	idSubstance,	Yes	
		residue limit definition.	SubstanceId,		
			SubstanceCode,		
			Substance		
Value	Numeric	Residue limit value.	Value, Limit,	Yes	
			Maximum-		
			ResidueLimit,		
			Maximum-		
			ResidueLimits,		
			MRL		
StartDate	DateTime	Start date of the period during	StartDate	No	
		which the limit applies.			
EndDate	DateTime	End date of the period during	EndDate	No	
		which the limit applies.			
Concentration-	ConcentrationUnits	The unit of the limit value	Concentration-	No	
Unit		(default mg/kg).	Unit, Unit		
ValueType	ConcentrationLimit-	Value type of the	ValueType,	No	
	ValueType	concentration value.	Concentration-		
			LimitValue-		
			Туре,		
			Concentration-		
			SingleValue-		
			Туре		
Reference	AlphaNumeric(200)	Reference to the source from	Reference,	No	
		which this concentration	References,		
		single value is obtained.	Source, Sources		

Table aliases: ResidueLimits, ResidueLimit, MaximumResidueLimits, MaximumResidueLimit, MRLs, MRL.

Concentration limits as data

Maximum Residue Limits (MRL) are provided as data.

• Concentration limits data formats

2.3.3 Concentration models

Concentration models are distributional models of substance concentrations on foods. They describe both the substance presence (yes/no, with no representing an absolute zero concentration) and the substance concentrations. Concentration models are specified per food/substance combination.

This module has as primary entities: Foods Substances Effects

Output of this module is used by: High exposure food-substance combinations Dietary exposures

Concentration models calculation

There are a number of *concentration model types* are available. A basic distinction is between using the empirical concentration data (empirical model), fitting a statistical model to the concentration data (parametric model), or to construct a model from (conservative) limit values. Settings relevant for some of these model types as well as other settings are described under *concentration model settings*.

Concentration data from *Total Diet Studies* does not include variability information. In order to include variability in the concentration models created for TDS samples, the variance of the TDS concentrations may be *estimated using variances of the foods making up the composite TDS food*.

Concentration model types

Empirical model

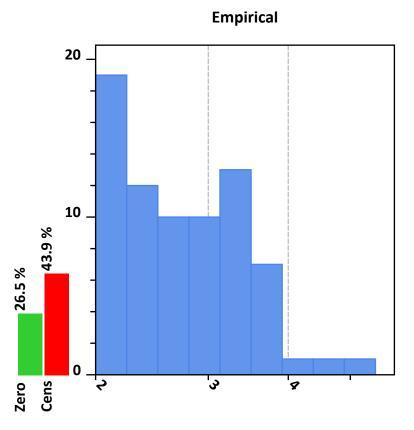
Data points are sampled at random from the available set. Non-detects are handled by imputation. If occurrence patterns are used, a proportion p_0/p_{ND} of non-detects is set as 0. See also concentration models.

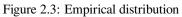
Non-detect spike lognormal model

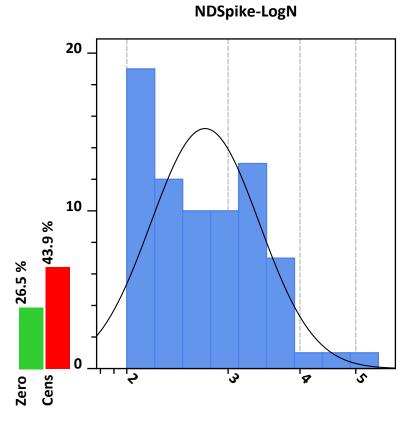
A binomial model is used to estimate the proportion p of positive values (detects). This is just the proportion observed in the data (unless *agricultural use* data have been used to set a proportion of true zeroes). A lognormal model is fitted to the positive data. This provides estimates of μ and σ , which are the mean and standard deviation of the natural logarithm of the concentration. Simulated concentrations are a non-detect with probability $p_{ND} = 1 - p$ or a value sampled from the fitted lognormal distribution with probability p. Non-detects are handled by imputation. If occurrence patterns are used, a proportion p_0/p_{ND} of non-detects is set as 0. Minimum requirements: at least two positive concentration values. See also *concentration models*.

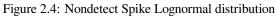
Non-Detect-Spike Truncated lognormal model

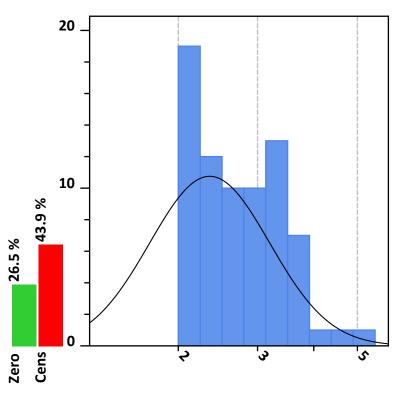
A binomial model is used to estimate the proportion p of positive values (detects). This is just the proportion observed in the data (unless agricultural use data have been used to set a proportion of true zeroes in which case p is calculated on the remaining proportion). A truncated lognormal model, with LOR as the truncation limit, is fitted to the positive data, leading to estimates of μ and σ , which are the mean and standard deviation of the natural logarithm of the concentration. Simulated concentrations are a non-detect with probability $p_{ND} = 1 - p$ or a value sampled from the fitted lognormal distribution with probability p. Non-detects are handled by imputation. If occurrence patterns are used, a proportion p_0/p_{ND} of non-detects is set as 0. Minimum requirements: at least two positive concentration values, all non-detects must have one LOR value. See also *concentration models*.











NDSpike-TruncLogN

Figure 2.5: Nondetect Spike Truncated Lognormal distribution

Censored Lognormal model

A censored lognormal model, with LOR as the censoring limit, is fitted to the data, both positives and non-detects. This provides estimates of μ and σ , which are the mean and standard deviation of the natural logarithm of the concentration. If agricultural use data are being used, then a proportion p_0/p_{ND} of non-detects will be excluded, where p_0 will be lowered to p_{ND} if it would be higher. Simulated concentrations are sampled from the fitted lognormal distribution. If agricultural use data have been used, simulated concentrations are 0 with probability p_0 or are sampled from the fitted lognormal distribution with probability $1 - p_0$. Minimum requirements: at least one positive concentration value. See also *concentration models*.

Zero-spike censored lognormal model

A mixture distribution of a spike of true zeroes and a censored lognormal model, with LOR as the censoring limit, is fitted to the data (non-detects and positives. This provides estimates of p_0 , which is the proportion of true zeroes, and of μ and σ , which are the mean and standard deviation of the natural logarithm of the concentration. Simulated concentrations are 0 with probability p_0 and are sampled from the fitted lognormal distribution with probability $1-p_0$. Minimum requirements: at least one positive concentration value, no agricultural use data for the food-substance combination (which directly specify p_0 , therefore it should not be estimated from the data). See also *concentration models*.

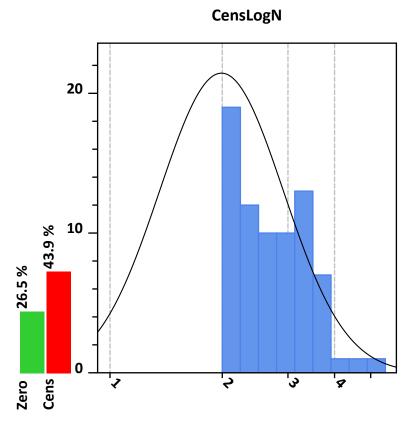
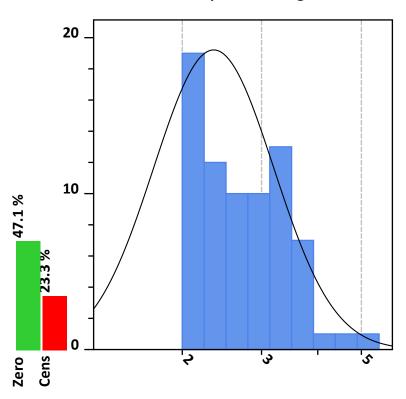
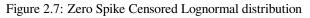


Figure 2.6: Censored Lognormal distribution





ZeroSpike-CensLogN

Non-detect spike MRL model

This model simply takes values specified in an input table as Maximum Residue Level (MRL) to be used for the proportion of positive values in the concentration dataset, and can be used to force the use of a pessimistic value.

Summary statistics model

For this model, no individual measurements on raw agricultural commodities are needed. The final estimates of μ and σ are simply provided or pooled or estimated using e.g. a coefficient of variation. Specific use of this model is found in *Total Diet Study* assessments. In general, each TDS food sample is prepared only once, yielding one measurement for a TDS food sample. The variability of the underlying distribution is unknown. However, a rough guess can be made using the e.g. coefficient of variation of the subsamples (in general raw agricultural commodities) that compose the TDS food sample. The estimated standard deviation is *calculated as a pooled estimate using the coefficient of variation of the TDS food*.

Concentration models

Let x denote a random variable from a lognormal distribution. Then, the log transformed variable y = ln(x) is normally distributed with μ and variance σ . The probability density function (p.d.f.) of y may be expressed as:

$$f_y(y, p_0, \mu_y, \sigma_y^2) = p_0 I(y; 0) + (1 - p_0)(1 - I(y; 0)) \cdot \frac{1}{\sqrt{2\pi\sigma_y}} \exp \frac{(y - \mu_y)^2}{2\sigma_y^2}$$

where $p_0 = Pr(y < log(X_{lor})), x_{lor}$ is the limit of reporting and I(y; 0) is an indicator function for $y < log(X_{lor})$. For $p_0 = 0$ the p.d.f. of y reduces to the usual lognormal density. The left truncated density for $y \ge log(X_{lor})$ may be expressed as:

$$f_y(y;\mu_y,\sigma_y^2) = \frac{1}{\sqrt{2\pi\sigma_y}} \exp{\frac{(y-\mu_y)^2}{2\sigma_y^2}} / (1-\Phi(z))$$

with $\Phi(\cdot)$ the standard normal c.d.f. and $z = (\log(x_{lor}) - \mu_y)/\sigma_z$. Model parameters are estimated using maximum likelihood estimation based on the loglikelihood functions specified below. The loglikelihood functions are evaluated in R, using the **optim** algorithm to find estimates for μ_y, σ_y^2 and p_0 .

Mixture zero spike and censored lognormal

The loglikelihood may be expressed as:

$$\log L(p_0, \mu_y, \sigma_y^2) = \sum_{i=1}^{n_0} \log(p_0 + (1 - p_0) \Phi(z_i)) + n_1 \log(\frac{1 - p_0}{\sqrt{2\pi\sigma_y}}) - \sum_{i=n_0+1}^n \frac{(y_i - \mu_y)^2}{2\sigma_y^2}$$

where $y_i = \log(x_i)$, $\Phi(\cdot)$ is the standard normal c.d.f., $z = (\log(x_{i,lor}) - \mu_y)/\sigma_y$, $z_{lor} = (\log(lor) - \mu_y)/\sigma_y$ with n_0 number of censored values ($x_i < x_{i,lor}$), n_1 number of uncensored values ($x_i \ge x_{i,lor}$) and x_i , $i = 1 \cdots n$. Multiple values for LOR are allowed.

1

Censored lognormal

When $p_0 = 0$ the loglikelihood reduces to:

$$\log L(\mu_y, \sigma_y^2) = \sum_{i=1}^{n_0} \log(\Phi(z)) + n_1 \log(\frac{1}{\sqrt{2\pi\sigma_y}}) - \sum_{i=n_0+1}^n \frac{(y_i - \mu_y)^2}{2\sigma_y^2}$$

Multiple values for LOR are allowed.

Mixture non-detect spike and truncated lognormal

Ignoring the n_0 values below x_{lor} , the loglikelihood may be expressed as:

$$\log L(\mu_y, \sigma_y^2) = -n_1 \log(1 - \Phi(z)) + n_1 \log(\frac{1}{\sqrt{2\pi\sigma_y}}) - \sum_{i=n_0+1}^n \frac{(y_i - \mu_y)^2}{2\sigma_y^2}$$

Only one value for LOR is allowed.

Mixture non-detect spike and lognormal

Ignoring the n_0 values below x_{lor} , the loglikelihood may be expressed as:

$$\log L(\mu_y, \sigma_y^2) = n_1 \log(\frac{1}{\sqrt{2\pi\sigma_y}}) - \sum_{i=n_0+1}^n \frac{(y_i - \mu_y)^2}{2\sigma_y^2}$$

Only one value for LOR is allowed.

Imputation

A complication in concentration modelling occurs if results are reported as being below a limit. Different names may be used for such a limit, e.g. limit of detection or limit of quantification. For the purpose of exposure assessment it is only relevant whether results are reported as a positive value or as a non-detect, therefore we refer to any limit as the **Limit Of Reporting** (LOR), and any result reported as '<LOR' is termed a **nondetect**. The value of LOR should always be known for the particular analytical method used.

Non-detects are a very common phenomenon for some classes of substances like pesticides. Non-detects can be handled by replacing them with a given value (**imputation**), or by incorporating them in a parametric model. In the imputation approach, non-detects (values reported less than LOR) can be replaced in simulations by any value between 0 and LOR * *constant*.

Imputation may be also dependent on the authorisation status of a substance i.c. whether the use of a substance on a agricultural crop is allowed or not.

In Figure 2.8 to Figure 2.11, the various scenarios are displayed. Two substances, Fenamidine and Hexythiazox are indicated with a brown box, these substances are authorized.

No imputation

Impute all nondetects

Impute nondetects based on authorized uses

No imputation except for authorized uses

Deriving the variance of TDS samples from monitoring

Variability of TDS food sample concentrations can be derived using *concentration distributions* for the sub-foods of the TDS food samples (defined by the *TDS compositions*). For each sub-food, e.g. *apple* (sub-food of TDS food *FruitMix*), a coefficient of variation (CV) is specified that is derived using the available monitoring samples. Note that monitoring samples may be composite samples. For *apple*, composite food samples are measured and each sample contains, for instance, 12 apples with unit weight 200 g. So monitoring concentrations, c_{mi} , are based on composite samples with a total weight $w_{mi} = 2400$ g each.

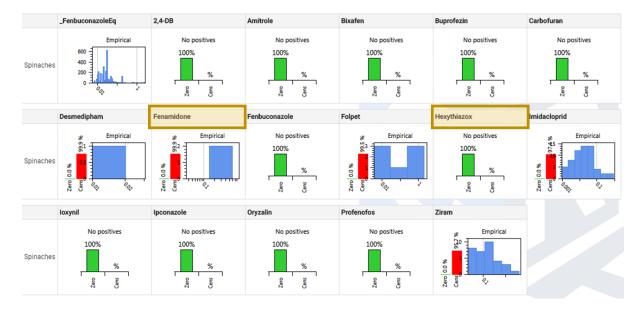


Figure 2.8: Tier 1: Non-detects are not replaced. For Fenamidine and Hexythiazox (brown boxes) authorized use is assumed.

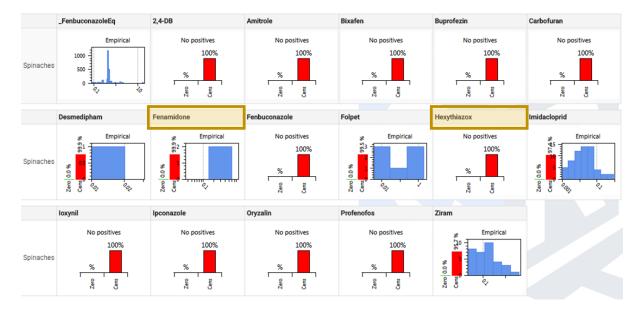


Figure 2.9: All non-detects are replaced by a constant factor x LOR. For Fenamidine and Hexythiazox (brown boxes) authorized use is assumed.

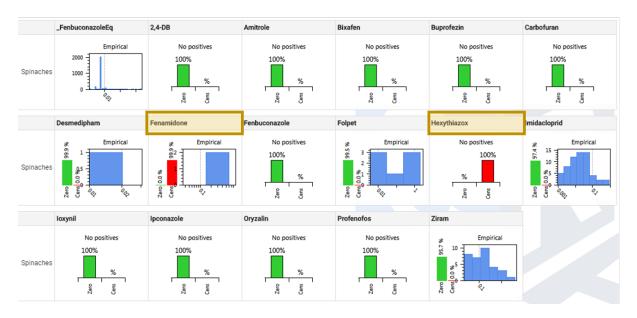


Figure 2.10: Nondetects are replaced by a constant factor x LOR for authorized uses. For Fenamidine and Hexythiazox (brown boxes) authorized use is assumed.

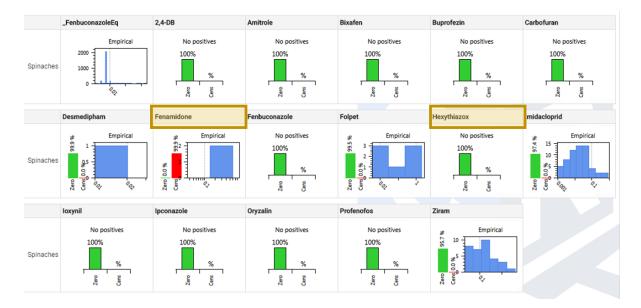


Figure 2.11: Tier 1: Nondetects are not replaced except for authorized uses (replaced by a constant factor x LOR). For Fenamidine and Hexythiazox (brown boxes) authorized use is assumed.

A TDS food sample is composed of w_i g of food i with i = 1...k, w_i represents the *PooledAmount* in *TDS food* sample compositions table. Then, the concentration of a TDS food sample may be represented as:

$$c_{\textit{TDS}} = \sum_{i=1}^k (w_i \cdot c_i) / \sum_{i=1}^k w_i$$

with variance:

$$var(c_{\textit{TDS}}) = \sum_{i=1}^k (w_i \cdot var(c_i)) / \sum_{i=1}^k w_i$$

and $var(c_i)$ is the variance of concentrations c_i of food *i* with portion sample size w_i .

It is expected that increasing the number of units in a composite sample will have a reverse effect on the variation between concentrations. Suppose TDS food *FruitMix* is composed of $2 \times 200 = 400$ g *apple*. The expected variation between portion sizes of 400 g will be larger than between portion sizes of 2400 g:

$$var(c_i) = var(c_{mi}) \cdot w_{mi}/w_i$$

The variance of the monitoring samples are corrected as follows, calculate:

1.
$$var(c_{mi}) = \log(CV_{mi}^2 + 1)$$

2.
$$var(c_i) = var(c_{mi}) \cdot w_{mi}/w_i$$

3. $CV_i = \sqrt{\exp(var(c+i)-1)}$

Concentration models settings

Calculation settings

Name	Description
Concentration model tier	Custom model, or set according to EFSA Guidance 2012. Note: you may need to set the tier separately in sub-modules.
Default concentration model	The concentration model type that will be used as default for all food/substance combinations. If this model type cannot be fitted, e.g., due to a lack of data, a simpler model will be chosen automatically as a fall-back.
Include MRL fallback model	Use the MRL as fallback model in case the occurrence data is insufficient for other concentration modelling options.
Restrict LOR imputation to authorised uses	Specifies whether imputation of factor x LOR should be limited to authorised uses only.
Non-detects replacement	How to replace non-detects (when not co-modelled, as in censored models).
Factor f (f x LOR)	Replace non-detects by Limit Of Reporting (LOR) times this factor. Constant (f), e.g. 0.5.
MRL Factor (f x MRL)	Use f x MRL as concentration estimate of the MRL models.
Sample based	Include co-occurrence of substances in samples in simulations. If checked, substance residue concentrations are sampled using the correlations between values on the same sample. If unchecked, any correlation between substances is ignored, substance residue concentrations are sampled ignoring the correlations between values on the same sample.
Impute missing values from available values (if unchecked, missing values are imputed with 0)	If checked, in procedure of EFSA Guidance 2012, Appendix 1, impute missing values using substance based concentration models. If unchecked, missing values are imputed by 0.
Correlate imputed values with sample potency	If checked, in procedure of EFSA Guidance 2012, Appendix 1, correlate high imputed values with high cumulative potency samples. If unchecked, random imputation.
Use occurrence frequencies for imputation	Use of occurrence frequencies (e.g., agricultural use frequencies) is relevant for imputation of non-detects in the concentration data. Part of the observed non-detects and missing values may be imputed with zero when the occurrence frequency is smaller than 100%. If checked, occurrence frequencies are expected as input of this action, otherwise 100% potential presence is assumed for all substances on all foods.

Table 2.37: Calculation settings for module Concentration models.

Uncertainty settings

Table 2.38: Uncertainty settings for module Concentration models.

Name	Description
Parametric uncertainty	For resample concentrations: specifies whether the uncertainty
	assessment is based on a parametric approach.

Concentration models tiers

In addition to the possibility for users to work with their own choices for all settings, MCRA implements four tiers from two documents:

- The optimistic and pessimistic basic assessments from the EFSA 2012 Guidance on the Use of Probabilistic Methodology for Modelling Dietary Exposure to Pesticide Residues [[EFSA, 2012]].
- Tier 1 and 2 from the *European Commission working document SANTE-2015-10216 rev.* 7 (2018) on risk management aspects related to the assessment of cumulative exposure [[EC, 2018]].

Overview

70

NameEFSA 2012 Op- timisticEFSA 2012 Op- timisticEFSA 2012 Op- pes- simistic - AcuteEFSA 2012 Pes- simistic - ChronicEC 2018 Tier 1EC 2018 Tier 2Default concentration modelEmpiricalNonDe- tect- SpikeLog- NormalNonDe- tect- SpikeLog- NormalEmpirical tect- SpikeLog- NormalEmpirical tect- SpikeLog- NormalEmpirical tect- SpikeLog- NormalEmpirical tect- SpikeLog- NormalEmpirical tect- SpikeLog- NormalEmpirical tect- SpikeLog- NormalReplace- ByLORReplace- 							
timistic AcutePes- simistic - AcutePes- simistic - ChronicPes- simistic - ChronicPeriod Simistic - ChronicP		EC 2018	EC 2018	EFSA	EFSA	EFSA	Name
simistic - Acutesimistic - Chronicsimistic - ChronicEmpirical Empirical Empirical Empirical Empirical tect- SpikeLog- NormalEmpirical SpikeLog- NormalEmpirical Empirical Empirical falseEmpirical Empirical Empirical falseEmpirical False <td></td> <td>Tier 2</td> <td>Tier 1</td> <td>2012</td> <td>2012</td> <td>2012 Op-</td> <td></td>		Tier 2	Tier 1	2012	2012	2012 Op-	
AcuteChronicImage: Chronic in the ct- in the ct- spike Log- NormalNonDe-tect- it ct- spike Log- NormalEmpirical it ct- spike Log- Normal				Pes-	Pes-	timistic	
Default concentration modelEmpirical kect- SpikeLog- NormalNonDe- tect- SpikeLog- NormalEmpirical tect- SpikeLog- NormalEmpirical tect- SpikeLog- NormalEmpirical tect- spikeLog- NormalEmpirical tect- spikeLog- NormalEmpirical tect- spikeLog- NormalEmpirical tect- spikeLog- NormalEmpirical tect- spikeLog- NormalEmpirical tect- spikeLog- NormalEmpirical tect- spikeLog- NormalEmpirical tect- spikeLog- NormalEmpirical tect- spikeLog- NormalEmpirical tect- spikeLog- NormalEmpirical tect- spikeLog- NormalEmpirical tect- spikeLog- NormalEmpirical tect- spikeLog- NormalEmpirical tect- spikeLog- NormalEmpirical tect- spikeLog- NormalEmpirical tect- spikeLog- NormalEmpirical tect- spikeLog- NormalEmpirical tect- spikeLog- NormalEmpirical tect- tect- mortalEmpirical tect- spikeLog- NormalEmpirical tect- spikeLog- NormalEmpirical tect- tect- methodEmpirical tect- tect- mutationEmpirical tect- trueEmpirical tect- tect- tect- mutationEmpirical tect- trueEmpirical tect- tect- tect- mutationEmpirical tect- trueEmpirical tect- tect- tect- mutationEmpirical teat- trueEmpirical teat- teat- trueEmpirical teat- teat- trueEmpirical teat- teat- teat- teat- teat- teat- teat- teat- teat- teat- teat- teat- teat- tea	1			simistic -	simistic -		
modeltect- SpikeLog- Normaltect- SpikeLog- Normaltect- SpikeLog- Normaltect- SpikeLog- NormaltruefalsefalseInclude MRL fallback modelfalsefalsetruetruefalsefalsefalseNon-detects replacementReplace- ByZeroReplace- ByLORReplace- I'ueReplace- I'ueReplace- I'ueReplace- I'ueReplace- I'ueReplace- I'ueReplace- I'ueReplace- I'ueReplace- I'ueReplace- I'ueReplace- I'ueReplace- I'ueReplace- I'ueReplace- I'ueReplace- I'ueReplace- I'ueReplace- I'ueReplace- I'ueReplace- I'ueR				Chronic	Acute		
modeltect- SpikeLog- Normaltect- SpikeLog- Normaltect- SpikeLog- Normaltect- SpikeLog- NormaltruefalsefalseInclude MRL fallback modelfalsefruetruetruefalsefalsefalseNon-detects replacementReplace- ByZeroReplace- ByLORReplace- ByLORReplace- ByLORReplace- ByLORReplace- ByLORReplace- ByLORReplace- ByLORReplace- ByLORSample basedtruetruetruetruetruetruetrueImpute missing values (fi unchecked, missing values are imputed with 0)falsetruetruetruetrueCorrelate imputed values with sample potencyfalsetruetruetruefalseValues with sample potencyfalsetruefalsefalsefalseParametric uncertainty authorised usesfalsetruefalsefalsefalseFactor (f x LOR)110.50.51MRL Factor (f x mutation to authorised uses1111Apply occurrence pattern percentagesImputentionfalsetruefalseSubstance conversion substance allocationImputentionImputentionImputentionImputentionReplace- falseFalseFalseImputentionImputentionImputentionApply occurrence pattern percentagesImputentionImputentionImputentionImputention <td>1</td> <td>Empirical</td> <td>Empirical</td> <td>NonDe-</td> <td>NonDe-</td> <td>Empirical</td> <td>Default concentration</td>	1	Empirical	Empirical	NonDe-	NonDe-	Empirical	Default concentration
NormalNormalNormalInclude MRL fallback modelfalsetruetruefalsefalseNon-detects replacementReplace- ByZeroByLORByLORByLORByLORSample basedtruetruetruetruetruetrueImpute missing values from available values (if unchecked, missing values are imputed with 0)falsetruetruetruetrueCorrelate imputed values with sample potencyfalsetruetruetruetruefalseValues occurrence imputationfalsetruetruefalsetruetruetrueParametric Uncertainty authorised usesfalsetruefalsefalsefalsefalseFactor f (f x LOR)110.50.5111Apply occurrence pattern percentages111111Apply occurrence pattern percentages111111Apply occurrence pattern percentages111111Apply occurrence pattern percentages11 <td< td=""><td></td><td>1</td><td>1</td><td>tect-</td><td>tect-</td><td>1</td><td>model</td></td<>		1	1	tect-	tect-	1	model
Include MRL fallback modelfalsetruefalsefalseNon-detects replacementReplace- ByZeroReplace- ByLORReplace- ByLORReplace- ByLORReplace- ByLORReplace- ByLORReplace- ByLORReplace- ByLORSample basedtruetruetruetruetruetruetrueImpute missing values from available values (if unchecked, missing values are imputed with 0)falsetruetruetruetrueCorrelate imputed values with sample potencyfalsetruetruetruetruefalseVuse occurrence imputationfalsetruetruefalsetruetruetrueParametric uncertainty authorised usesfalsetruefalsefalsefalsefalseFactor (f x LOR)110.50.5111Apply occurrence pattern percentages11111Apply occurrence pattern percentages11111Apply occurrence pattern percentages11111Apply occurrence pattern percentages11111Account for substance authorisations in11111Account for substance authorisations in11111Account for substance authorisations in11111Account for substance authorisations in<					SpikeLog-		
Include MRL fallback modelfalsetruetruefalsefalsefalseNon-detects replacementReplace- ByZeroReplace- ByLORReplace- IncentionReplace- IncentionReplace- IncentionReplace- IncentionReplace- IncentionReplace- IncentionReplace- IncentionReplace- IncentionReplace- IncentionReplace- IncentionReplace- IncentionReplace- IncentionReplace- IncentionReplace- IncentionReplace- IncentionReplace- IncentionReplace- IncentionReplace- 							
modelImage: scalar	1	false	false			false	Include MRL fallback
Non-detects replacementReplace- ByZeroReplace- ByLORReplace- Content trueReplace- 		1000	i ulo e			10100	
replacementByZeroByLORByLORByLORByLORByLORSample basedtruetruetruetruetruetruetrueImpute missing values from available values (if unchecked, missing values are imputed with 0)falsetruetruetruetrueCorrelate imputed values with sample potencyfalsetruetruetruefalseUse occurrence frequencies for imputationfalsetruefalsetruefalseParametric uncertainty substance dlocationfalsetruefalsefalseFactor (f x LOR)110.50.5MRL Factor (f x bylo ccurrence false110.50.5MRL Factor (f x substance allocated1110.50.5Max Mather authorised uses1110.50.5Max Mather authorised uses1110.50.5Max Mather authorised uses1110.50.5Max Max Max Max Max Mather authorised uses1110.50.5Max Mather authorised uses1110.50.5Max Max Max Max Max Max Mather authorised uses1110.50.5Max Max Max Max Max Max Max Max Max Max	-	Replace-	Replace-	Replace-	Replace-	Replace-	
Sample basedtruetruetruetruetruetrueImpute missing values from available values (if unchecked, missing values are imputed with 0)falsetruetruetruetruetrueCorrelate imputed values with sample potencyfalsetruetruetruefalseUse occurrence frequencies for imputationfalsetruefalsetruetrueParametric uncertainty falsefalsetruefalsefalsefalseFactor f (f x LOR)110.50.50.5MRL Factor (f x pattern percentages111UseMost- ToxicDrawRan- domSubstance conversion methodImpute and impute actionfalsetruefalsetrueAccount for substance authorisations inImpute actionfalsetruetrueSample basedImpute actionImpute actionImpute actionImpute actionAccount for substance authorisations inImpute actionImpute actionImpute actionImpute authorisations inImpute actionImpute actionImpute actionImpute actionImpute authorisations inImpute actionImpute actionImput		-	-	-	-	-	
Impute missing values from available values (if unchecked, missing values are imputed with 0)falsetruetruetruetrueCorrelate imputed values with sample potencyfalsetruetruetruefalseUse occurrence frequencies for imputationfalsetruefalsetruefalseParametric uncertainty authorised usesfalsetruefalsefalsefalseFactor f (f x LOR)110.50.5MRL Factor (f x pattern percentages1110.5Substance conversion method1110.50.7Account for substance authorisations in11111Account for substance authorisations in111111Account for substance authorisations in1111111Account for substance authorisations in11111111Account for substance authorisations in1111111111111111 <td>-</td> <td></td> <td></td> <td></td> <td></td> <td></td> <td>-</td>	-						-
from available values (if unchecked, missing values are imputed with 0)falsetruetruetrueCorrelate imputed values with sample potencyfalsetruetruetruefalseUse occurrence frequencies for imputationfalsetruefalsetruefalseParametric uncertainty authorised usesfalsetruefalsefalsefalseFactor f (f x LOR)110.50.5MRL Factor (f x pattern percentages1110.5DrawRan- ToxicSubstance conversion methodImputationImputationTruefalsetrueAccount for substance authorisations inImputationImputationImputationImputationAccount for substance authorisations inImputationImputationImputationImputationTowic1110.50.5MRL Factor (f x conversion1ImputationImputationImputationSubstance allocationImputationImputationImputationImputationAccount for substance authorisations inImputationImputationImputationImputationImputation to authorisations inImputationImputationImputationImputationImputation to authorisations inImputationImputationImputationImputationImputation to authorisations inImputationImputationImputationImputationImputation to authorisations in <td>-</td> <td></td> <td></td> <td></td> <td></td> <td></td> <td>÷</td>	-						÷
(if unchecked, missing values are imputed with 0)Image: second seco		true	true	true	true	Talse	
values are imputed with 0)falsetruetruetruetrueCorrelate imputed values with sample potencyfalsetruetruefalsevalues with sample potencyfalsetruetruefalseUse occurrence frequencies for imputationfalsetruefalsetrueParametric uncertainty authorised usesfalsetruefalsefalseFactor f (x LOR)110.50.5MRL Factor (f x pattern percentages1111Apply occurrence pattern percentagesfalsefalsetrueSubstance allocationIIIIIAccount for substance authorisations inIIIIIAccount for substance authorisations inIIIIIAccount for substanceIIIIIIAccount for substance authorisations inIIIIIImplicition Implicition ImplicitionIIIIIImplicition Implicition Impli							
with 0)Image: constraint of the second s							
Correlate imputed values with sample potencyfalsetruetruetruefalseJuse occurrence frequencies for imputationfalsetruefalsetruetrueParametric uncertainty authorised usesfalsetruefalsefalsefalseFactor f (f x LOR)110.50.5MRL Factor (f x pattern percentages1111Apply occurrence pattern percentagesImage: second sec							
values with sample potencyImage: sample potencyImage: sample potencyImage: sample potencyImage: sample parametric uncertaintyfalseImage: sample parametric uncertaintyImage: sample							
potencyImage: space of the space		false	true	true	true	false	
Use occurrence frequencies for imputationfalsetruefalsetruetrueParametric uncertainty method substance authorised usesfalsetruefalsefalsefalsefalseRestrict LOR imputation to authorised usesfalsefalsefalsefalsefalsefalseFactor f (f x LOR)110.50.50.5MRL Factor (f x pattern percentages1111Substance conversion method1110.5DrawRan- ToxicRetain all allocated substance allocation11falsetrueAccount for substance authorisations in11falsetrue							values with sample
frequencies for imputationImage: second sec							potency
imputationImputation<]	true	true			false	Use occurrence
imputationImputation<							frequencies for
Parametric uncertaintyfalsetruefalsefalsefalsefalseRestrict LORfalsefalsefalsefalsefalsefalsefalseimputation to authorised uses110.50.5Factor f (f x LOR)110.50.5MRL Factor (f x MRL)1111Apply occurrence pattern percentages-falsefalseSubstance conversion method-UseMost- ToxicDrawRan- domRetain all allocated substance allocation-truetrueAccount for substance authorisations in-falsetrue							-
Restrict LOR imputation to authorised usesfalsefalsefalsefalsefalseFactor f (f x LOR)110.50.5MRL Factor (f x MRL)1111Apply occurrence pattern percentages-falsetrueSubstance conversion method-UseMost- ToxicDrawRan- domRetain all allocated substance allocation-truetrueAccount for substance authorisations in-falsetrue	1	false	false	false	true	false	-
imputation to authorised usesII0.50.5Factor f (f x LOR)110.50.5MRL Factor (f x MRL)11IIApply occurrence pattern percentagesIIISubstance conversion methodIIUseMost- ToxicDrawRan- domRetain all allocated substance allocationIIIrueAccount for substance authorisations inIIrueIrue	-						
authorised usesII0.50.5Factor f (f x LOR)110.50.5MRL Factor (f x111IIMRL)IIIIIApply occurrenceIIIIIIIIIIIIIIIIIIIIIIIIIIIIIIIIIII		10100	i ulo e	14100	10100		
Factor f (f x LOR)110.50.5MRL Factor (f x MRL)1111Apply occurrence pattern percentagesfalsetrueSubstance conversion methodUseMost- ToxicDrawRan- domRetain all allocated substance allocationtrueAccount for substance authorisations infalsetrue							-
MRL Factor (f x MRL)111Apply occurrence pattern percentagesfalsetrueSubstance conversion methodUseMost- ToxicDrawRan- domRetain all allocated substance allocationtruetrueAccount for substance authorisations infalsetrue	-	0.5	0.5	1	1		
MRL)Image: MRL)Image: MRL)Image: MRL)Apply occurrence pattern percentagesfalsetrueSubstance conversion methodUseMost- ToxicDrawRan- domRetain all allocated substance allocationtruetrueSubstance allocationImage: MRL)Image: MRL)Account for substance authorisations inImage: MRL)Image: MRL)	-	0.5	0.5				
Apply occurrence pattern percentagesfalsetrueSubstance conversion methodUseMost- ToxicDrawRan- domRetain all allocated substances after active substance allocationtruetrueAccount for substance authorisations infalsetrue				T	1		,
norm Use Most- Substance conversion Use Most- method Toxic Retain all allocated true substance allocation true Account for substance false authorisations in false	-	4.00-0	falas				,
Substance conversion methodUseMost- ToxicDrawRan- domRetain all allocated substances after active substance allocationtruetrueAccount for substance authorisations infalsetrue		true	Taise				
methodToxicdomRetain all allocatedtruetruesubstances after activetruetruesubstance allocationfalsetrueAccount for substancefalsetrue	-						
Retain all allocated true true substances after active true true substance allocation false true Account for substance false true							
substance allocation Image: substance allocation Image: substance allocation Image: substance allocation Account for substance authorisations in Image: substance allocation Image: substance allocation Image: substance allocation		dom	Toxic				
substance allocation false Account for substance authorisations in false		true	true				
Account for substance authorisations in false true							
authorisations in							
		true	false				Account for substance
substance conversions							authorisations in
							substance conversions
Use extrapolation rules true true	1	true	true				Use extrapolation rules
Threshold for 10 10	1						
extrapolation		-	-				
Restrict extrapolations true	1	true	true	<u> </u>			1
to equal MRLs							
Restrict extrapolations true	-	true	true				-
to authorised uses		auc	ii uc				-
	-	true	true	<u> </u>			
Impute water true true	1	urue	true				
concentrations O 1 0.05		0.05	0.1				
Water concentration 0.1 0.05	-	0.05	0.1				
value (µg/kg)	-						
Restrict water true true	-	true	true				
imputation to the five	-						imputation to the five
most toxic substances Chapter 2. M	-						
Restrict water false false	Modul	Chapter 2.	(most toxic substances
	Modul						
imputation to	Modul						Restrict water

Table 2.39: Tier overview for module Concentration models.

The sections below describe the settings specified by each tier in detail.

EFSA 2012 Optimistic

Use the optimistic model settings according to the EFSA Guidance 2012. Non-detects and missing values are replaced by zero.

Name	Setting
Default concentration model	Empirical
Include MRL fallback model	false
Non-detects replacement	ReplaceByZero
Sample based	true
Impute missing values from available values (if unchecked,	false
missing values are imputed with 0)	
Correlate imputed values with sample potency	false
Use occurrence frequencies for imputation	false
Parametric uncertainty	false

EFSA 2012 Pessimistic - Acute

Concentration model settings for acute pessimistic dietary exposure assessments according to the EFSA Guidance 2012. A non-detect spike lognormal model is fitted to the positive residue values and non-detects are replaced by the LOR. When the number of positives is smaller than 2, the maximum residue limit (if available) is used instead. Missing values are imputed.

Name	Setting
Default concentration model	NonDetectSpikeLogNormal
Include MRL fallback model	true
Restrict LOR imputation to authorised uses	false
Non-detects replacement	ReplaceByLOR
Factor f (f x LOR)	1
MRL Factor (f x MRL)	1
Sample based	true
Impute missing values from available values (if unchecked,	true
missing values are imputed with 0)	
Correlate imputed values with sample potency	true
Parametric uncertainty	true

Table 2.41: Tier definition for EFSA 2012 Pessimistic - Acute.

EFSA 2012 Pessimistic - Chronic

Concentration model settings for acute pessimistic dietary exposure assessments according to the EFSA Guidance 2012. A non-detect spike lognormal model is fitted to the positive residue values and non-detects are replaced by the LOR. When the number of positives is smaller than 2, the maximum residue limit (if available) is used instead. Missing values are imputed.

Name	Setting
Default concentration model	NonDetectSpikeLogNormal
Include MRL fallback model	true
Restrict LOR imputation to authorised uses	false
Non-detects replacement	ReplaceByLOR
Factor f (f x LOR)	1
MRL Factor (f x MRL)	1
Sample based	true
Impute missing values from available values (if unchecked,	true
missing values are imputed with 0)	
Correlate imputed values with sample potency	true
Parametric uncertainty	false

Table 2.42: Tier definition for EFSA 2012 Pessimistic - C	hronic.

EC 2018 Tier 1

Name	Setting	From input tier	In module
Default concentration model	Empirical		
Include MRL fallback model	false		
Restrict LOR imputation to	false		
authorised uses			
Non-detects replacement	Replace-		
	ByLOR		
Factor f (f x LOR)	0.5		
Sample based	true		
Impute missing values from available	true		
values (if unchecked, missing values			
are imputed with 0)			
Correlate imputed values with sample	true		
potency			
Use occurrence frequencies for	true		
imputation			
Parametric uncertainty	false		
Apply occurrence pattern percentages	false	EC 2018	Occur-
		Tier 1	rence
			patterns
Substance conversion method	UseMost-	EC 2018	Concen-
	Toxic	Tier 1	trations
Retain all allocated substances after	true	EC 2018	Concen-
active substance allocation		Tier 1	trations
Account for substance authorisations	false	EC 2018	Concen-
in substance conversions		Tier 1	trations
Use extrapolation rules	true	EC 2018	Concen-
		Tier 1	trations
Threshold for extrapolation	10	EC 2018	Concen-
		Tier 1	trations
Restrict extrapolations to equal MRLs	true	EC 2018	Concen-
		Tier 1	trations
Restrict extrapolations to authorised	true	EC 2018	Concen-
uses		Tier 1	trations
Impute water concentrations	true	EC 2018	Concen-
		Tier 1	trations
Water concentration value (µg/kg)	0.1	EC 2018	Concen-
		Tier 1	trations
Restrict water imputation to the five	true	EC 2018	Concen-
most toxic substances		Tier 1	trations
Restrict water imputation to	false	EC 2018	Concen-
authorised uses		Tier 1	trations

Table 2.43: Tier definition for EC 2018 Tier 1.

EC 2018 Tier 2

Name	Setting	From	In
		input tier	module
Default concentration model	Empirical		
Include MRL fallback model	false		
Restrict LOR imputation to	false		
authorised uses			
Non-detects replacement	Replace-		
	ByLOR		
Factor f (f x LOR)	0.5		
Sample based	true		
Impute missing values from available	true		
values (if unchecked, missing values			
are imputed with 0)			
Correlate imputed values with sample	false		
potency			
Use occurrence frequencies for	true		
imputation			
Parametric uncertainty	false		
Apply occurrence pattern percentages	true	EC 2018	Occur-
		Tier 2	rence
			patterns
Scale up use frequency to 100%	true	EC 2018	Occur-
1 1 2		Tier 2	rence
			patterns
Restrict use percentage up-scaling to	true	EC 2018	Occur-
authorised uses		Tier 2	rence
			patterns
Substance conversion method	DrawRan-	EC 2018	Concen-
	dom	Tier 2	trations
Retain all allocated substances after	true	EC 2018	Concen-
active substance allocation		Tier 2	trations
Account for substance authorisations	true	EC 2018	Concen-
in substance conversions		Tier 2	trations
Use extrapolation rules	true	EC 2018	Concen-
1		Tier 2	trations
Threshold for extrapolation	10	EC 2018	Concen-
1.		Tier 2	trations
Restrict extrapolations to equal MRLs	true	EC 2018	Concen-
1 1 1 1 1		Tier 2	trations
Restrict extrapolations to authorised	true	EC 2018	Concen-
uses		Tier 2	trations
uses			
	true	EC 2018	Concen-
Impute water concentrations	true	<i>EC 2018</i> <i>Tier 2</i>	Concen- trations
Impute water concentrations	true 0.05		trations
		<i>Tier 2</i> <i>EC 2018</i>	
Impute water concentrations Water concentration value (µg/kg)	0.05	<i>Tier 2</i> <i>EC 2018</i> <i>Tier 2</i>	trations Concen- trations
Impute water concentrations Water concentration value (µg/kg) Restrict water imputation to the five		Tier 2 EC 2018 Tier 2 EC 2018	trations Concen- trations Concen-
Impute water concentrations Water concentration value (µg/kg)	0.05	<i>Tier 2</i> <i>EC 2018</i> <i>Tier 2</i>	trations Concen- trations

Table 2.44: Tier definition for EC 2018 Tier 2.

EFSA 2012 Pessimistic

Note: This tier is deprecated and has been replaced by separate acute/chronic tiers.

Concentration model settings for pessimistic dietary exposure assessments according to the EFSA Guidance 2012. A non-detect spike lognormal model is fitted to the positive residue values and non-detects are replaced by the LOR. When the number of positives is smaller than 2, the maximum residue limit (if available) is used instead. Missing values are imputed.

Table 2.45: Tier definition for EFSA 2012 Pessimistic	Table 2.45:	Tier definition	for EFSA 201	2 Pessimistic
---	-------------	-----------------	--------------	---------------

Name	Setting
Default concentration model	NonDetectSpikeLogNormal
Include MRL fallback model	true
Restrict LOR imputation to authorised uses	false
Non-detects replacement	ReplaceByLOR
Factor f (f x LOR)	1
MRL Factor (f x MRL)	1
Sample based	true
Impute missing values from available values (if unchecked,	true
missing values are imputed with 0)	
Correlate imputed values with sample potency	true
Parametric uncertainty	true

Concentration models uncertainty

When using empirical distributions, concentration model uncertainty is covered by the the inputs. I.e., concentration models can be recomputed from *resampled/bootstrapped* concentration data. This happens for both the univariate concentration models, being recomputed from the bootstrapped residue collections for each food and substance, and also for the samples of the sample-based approach that are re-generated from the bootstrapped samples (including the necessary steps of missing value imputation and imputation of non-detects).

When parametric uncertainty is preferred over empirical bootstrapping, the parameters of the univariate concentration models fitted as a parametric distributions can be *resampled parametrically*.

Let x denote a random variable from the specified distribution. The log transformed variable y = ln(x) is normally distributed with mean μ_y and variance σ_y . The maximum likelihood estimates are $\hat{\mu}_y$ and $\hat{\sigma}_y$. In each bootstrap sample, values are drawn from a normal distribution where the maximum likelihood estimates are replaced by ($\hat{\mu}_y^*$, $\hat{\sigma}_y^*$).

Calculation of concentration models

Concentration models can be computed from concentration data.

• Concentration models calculation

Inputs used: Concentrations Concentration limits Modelled foods Substance authorisations Occurrence frequencies Relative potency factors Concentration distributions Total diet study sample compositions

Settings used

Calculation Settings

2.3.4 Concentrations

Concentrations data are analytical measurements of chemical substances occurring in food samples. In their simplest form, concentration data can just be used as provided by datasets. Optionally, concentrations data can be manipulated for active substances, extrapolated to other foods, and/or default values can be added for water.

This module has as primary entities: Foods Substances

Output of this module is used by: Single value concentrations Occurrence patterns Concentration models Modelled foods

Concentrations data formats

Three schemes for data are implemented:

- 1. MCRA scheme: relational tables that can hold all information about Food samples (e.g. sampling date and location), Analytical methods, Analytical method properties for substances (e.g. LOR), Analysis samples (e.g. analysis date) and Concentrations;
- 2. SSD scheme: data according to the EFSA Standard Sample Description (SSD) guideline; SSD data are converted automatically to the MCRA scheme;
- 3. Tabulated data scheme: simplified data format, where samples and analytical methods are not explicitly specified. Tabulated concentration data are converted automatically to the MCRA scheme.

Concentration data

In this group all tables are collected that store information related to concentration or concentration related entities.

Sample-based concentration data

This sub-group contains five tables to specify food samples, analytical methods, their properties for given substances, analyses and concentrations.

Analytical methods

The analytical methods used for analyzing the samples are recorded in the analytical methods table. Each analytical method should have a unique identification code (idAnalyticalMethod). The description field may be used for a more detailed description of the analytical method. The records of this table should be linked to one or more analytical-method-substance records, which record the substances that are measured by this method (and their limits of reporting).

Name	Туре	Description	Aliases	Required
idAnalytical- Method	AlphaNumeric(50)	The code for the method of analysis.	idAnalytical- Method, Analytical- MethodId, Analytical- MethodName, Id	Yes
Description	AlphaNumeric(200)	Additional description of method of analysis.	Description	No

Table aliases: AnalyticalMethod, AnalyticalMethods.

Analytical method properties for substances

Name	Туре	Description	Aliases	Required
idAnalytical- Method	AlphaNumeric(50)	The code of method of analysis.	idAnalytical- Method,	Yes
			Analytical-	
			MethodName,	
			Analytical-	
			MethodId	
idSubstance	AlphaNumeric(50)	The substance code.	idSubstance,	Yes
			SubstanceId,	
			Substance	
LOR	Numeric	The limit of reporting (LOR). In MCRA, LOR just means	LOR	Yes
		the limit below which no		
		quantitative result has been		
		reported. Depending on a		
		laboratory's format of		
		reporting, LOR may be a		
		limit of detection (LOD), a		
		limit of quantification (LOQ)		
		or another limit.		
Concentration-	<i>ConcentrationUnits</i>	The unit of the	Concentration-	No
Unit		concentrations/LORs	Unit, Units, Unit	
		reported by the analytical		
		method for this substance		
		(default mg/kg).		

Table 2.47: Table definition for AnalyticalMethodCompounds.

 $Table\ aliases:\ Analytical Method Substances,\ Analytical Method Substance.$

Food samples

Food sample for analysis of concentrations. May be characterised by location and/or date of sampling. A sample can be analysed multiple times, the results per analysis are stored as analysis samples.

Name	Туре	Description	Aliases	Required
idFoodSample	AlphaNumeric(50)	The identification number of	idFoodSample,	Yes
		the food sample.	idSample,	
			SampleId, Id	
idFood	AlphaNumeric(50)	The food code.	idFood, FoodId,	Yes
			Food, FoodCode	
Location	AlphaNumeric(50)	The location or country code,	Location,	No
		sampling location.	Location-	
			Sampling,	
			Sampling-	
			Location,	
			Country	
Region	AlphaNumeric(50)	The area or region within the	Region, Area,	No
		sampling location.	Sampling-	
			Region,	
			SamplingArea	
DateSampling	DateTime	The date of sampling.	DateSampling,	No
			SamplingDate	
Production-	AlphaNumeric(50)	Additional information on the	Production-	No
Method		type/method of production of	Method,	
		the sampled food.	ProductionType	

Table 2.48: Table definition for FoodSamples.	
---	--

Table aliases: FoodSamples, FoodSample, Samples, Sample, PrimarySample, PrimarySamples.

Sample properties

Food sample properties, additional columns that can also be specified as additional columns in the food samples table

Table 2.49: Table definition for SampleProperties.
--

Name	Туре	Description	Aliases	Required
Name	AlphaNumeric(50)	The name of the property.	Id	Yes

Table aliases: SampleProperties, SampleProperty.

Sample property values

Food sample property values, additional columns that can also be specified as additional columns in the food samples table

		1 1 2		
Name	Туре	Description	Aliases	Required
idSample	AlphaNumeric(50)	The identification number of	Id,	Yes
		the food sample.	IdFoodSample	
PropertyName	AlphaNumeric(50)	The name of the property.	Name	Yes
TextValue	AlphaNumeric(50)	The value of the property as		No
		text value.		
DoubleValue	Numeric	The value of the property as		No
		number.		

Table 2.50: Table definition for SamplePropertyValues.

Table aliases: SamplePropertyValues, SamplePropertyValue.

Sample Analyses

An analysis sample specifies the analysis of a sample by an analytical method. A sample can be analysed multiple times, the results per analysis are stored as analysis samples.

Name	Туре	Description	Aliases	Required
idSample-	AlphaNumeric(50)	The identification number of	id, idSample-	Yes
Analysis		the analysed sample.	Analysis,	
			SampleAnalysis,	
			idAnalysis-	
			Sample,	
			AnalysisSample-	
			Id	
idFoodSample	AlphaNumeric(50)	The identification number of	idFoodSample,	Yes
		the food sample.	idSample,	
			SampleId,	
			Sample	
idAnalytical-	AlphaNumeric(50)	The code of method of	idAnalytical-	Yes
Method		analysis.	Method,	
			Analytical-	
			MethodId	
DateAnalysis	DateTime	The date of the analysis.	DateAnalysis,	No
			AnalysisDate,	
			Date	

Table aliases: AnalysisSamples, AnalysisSample, SampleAnalysis, SampleAnalyses.

Sample concentrations

The positive concentration values for substances from analysis in the unit specified in table AnalysisSamples. Nondetects (i.e. results 'less than LOR') are not included, their existence can be inferred from the tables AnalysisSamples and AnalyticalMethodSubstances, and the LOR itself from the table AnalyticalMethods.

Name	Туре	Description	Aliases	Required
idSample-	AlphaNumeric(50)	The identification number of	idSample-	Yes
Analysis		the analysed sample.	Analysis,	
			SampleAnalysis,	
			idAnalysis-	
			Sample,	
			AnalysisSample-	
			Id	
idSubstance	AlphaNumeric(50)	The substance code.	idSubstance,	Yes
			SubstanceId,	
			Substance	
Concentration	Numeric	The measured concentration.	Concentration	Yes

Table 2.52: Table definition for ConcentrationsPerSample.

Table aliases: SampleConcentrations, ConcentrationsPerSample, ConcentrationPerSample.

Tabulated concentration data

Tabulated concentration data provide a simplified concentration data format, where samples and analytical methods are not explicitly specified and analysis results can be tabulated for repeats of the same outcome. This is a convenient data format for single-substance analyses, but it should be noted that it is not possible to use this data in sample-based methods of multiple substances, because it does not record co-occurrence information of substances in samples. Tabulated concentrations data is converted to the internal, relational data format of MCRA.

Tabulated concentrations

In the tabulated concentration data table, each record represents one or multiple samples, and each sample contains a concentration value for a food/substance combination. Non-detects (i.e. concentrations less than LOR) are specified as negative values, i.e. 'less than LOR' should be specified as minus the LOR value. MCRA uses the concept of samples analysed by analytical methods, where the analytical method is characterised by the substances analysed and the LORs for these substances. However, the tabulated data do not provide this information explicitly. Samples are reconstructed from the tabulated records using the NumberOfSamples field to create that number of single substance samples. Analytical methods are reconstructed from the data, with each analytical method having only one analysed substance with a LOR and concentration unit. When a negative concentration value is given (i.e., it is a non-detect measurement), this value is recorded as the LOR (negated). All non-detect measurements of the same substance with the same LOR and concentration unit are linked to the same analytical method. When a positive concentration value is given, this value is recorded as the measured concentration of the sample. All positive measurements of the same substance are linked to the same analytical method that has an artificial LOR that is smaller than the lowest positive concentration. When a concentration of 0 (zero) is given, the measurement is considered to be a non-detect measurement and the LOR is set as default to the value 1E-08.

Table 2 53	Table	definition	for	ConcentrationTabulated.
1 abic 2.55.	rabic	uchintion	101	concentration rabulated.

Table aliases: ConcentrationTabulated, ConcentrationValues, TabulatedConcentrations, TabulatedConcentration.

EFSA SSD concentration data

MCRA provides an option to upload concentration data that is formatted according to the EFSA Standard Sample Description (SSD) guideline. SSD formatted concentrations data is converted to the internal, relational data format of MCRA.

SSD concentrations

MCRA uses the concept of samples analysed by analytical methods, where the analytical method is characterised by the substances analysed and the LORs for these substances. However, the SSD data do not provide information on the analytical methods at this level of detail. Therefore, from the provided SSD records, analytical methods are reconstructed and samples are linked to these analytical methods. All SSD records with the same labSampCode and labSubSampCode are considered to be from the same sample. All SSD samples that have records for the same substances, with the same LOQ/LOD values and resUnit are considered to originate from the same reconstructed analytical method. If both LOQ and LOD are provided, LOQ is used as LOR of the reconstructed analytical method. It is highly recommended to supply LOQ/LOD values, even for positive measurement, because this reduces the number of reconstructed analytical methods.

Name	Туре	Description	Aliases	Required
labSampCode	AlphaNumeric(30)	Code of the laboratory sample. MCRA will use the combination of labSampCode and labSubSampCode as unique code for a sample.	labSampCode	Yes
labSubSamp- Code	AlphaNumeric(4)	Code of the laboratory sub-sample. MCRA will use the combination of labSampCode and labSubSampCode as unique code for a sample.	labSubSamp- Code	No
sampCountry	AlphaNumeric(2)	Two-letter code to identify the country of sampling.	sampCountry	Yes
sampArea	AlphaNumeric(5)	Area where the sample was collected.	sampArea	Yes
prodCode	AlphaNumeric(20)	Code identifying the modelled food. Should be equal to a code idFood in the Foods table.	prodCode	Yes
prodProdMeth	AlphaNumeric(5)	Code providing additional information on the type of production for the food under analysis.	prodProdMeth	No
sampY	Integer(4)	Year of sampling.	sampY	Yes
sampM	Integer(2)	Month of sampling.	sampM	No
sampD	Integer(2)	Day of sampling.	sampD	No
analysisY	Integer(4)	Year of analysis.	analysisY	Yes
analysisM	Integer(2)	Month of analysis.	analysisM	No
analysisD	Integer(2)	Day of analysis.	analysisD	No
paramCode	AlphaNumeric(20)	Code identifying the substance.	paramCode	Yes
resUnit	AlphaNumeric(5)	Unit of residue measurement.	resUnit	Yes
resLOD	Numeric	Residue Limit Of Detection. Required if resType is LOD. MCRA will use resLOD as LOR if resLOQ is not provided.	resLOD	No
resLOQ	Numeric	Residue Limit Of Quantification. Required if resType is LOQ MCRA will use resLOQ as LOR if provided.	resLOQ	No
resVal	Numeric	Required if resType is VAL.	resVal	No
resType	AlphaNumeric(3)	Type of residue data. Should be VAL, LOQ or LOD.	resType	Yes

Table 2.54: Table definition for ConcentrationsSSD.

Table aliases: ConcentrationsSSD, SSDConcentrations.

Concentrations calculation

Occasionally, concentrations of substances measured in food samples are exceeding a specified concentration limit e.g. the *Maximum Residue Limits* (MRL). An MRL is the highest level of a substance that is legally tolerated in or on food or feed when substances are applied correctly. *Filter samples* exceeding the concentration limits filter out all samples where one of the substances measured is exceeding the *MRL*.

Substance conversions rules may be used to *convert* concentration data at the level of measured substances to concentration data at the level of potentially active substances. These rules (provided as data) may be applicable, for example, when a measured substance represents multiple substances and these measurements should be converted into measurement values for these substances. This conversion may depend on *substance authorisations* which provides information on the likelihood of certain translations to occur. *points of departure* or *relative potency factors* might be needed when the substance conversion should select the most toxic candidate in case a measured substance translates to multiple active substances.

If there are only a few measurements in the concentration data, then *extrapolation of concentration data* may be desired. In that case, *food extrapolation rules* may be provided to specify per food the alternative foods from which extrapolation is allowed. The extrapolation of concentrations will then be performed within this module and the results are included in the resulting active substance concentrations data. *Substance authorisations* and/or *concentration limits* may be used to further restrict the to-food/from-food combinations per substance for which extrapolation is possible.

Concentration data for water are often not available in the concentration data, but it may be desirable to include them in the assessments. For this, *imputation* of low-tier, deterministic estimates of water concentrations of the most toxic substances may be used to include (typically conservative) estimates in the calculations.

In some scenarios it may be desired to perform a prospective analysis in which anticipated (or foreground) *focal commodity concentration data* for a particular focal commodity food (and substance) is added to, or replaces part of the background concentration data that is used for the null-scenario. The concentrations module offers various options to perform such *focal commodity scenario analyses*.

It is also possible to *filter (or subset) samples by specific sample properties* (e.g., year, location). This can be done by checking the option to *filter samples by specific property values (subset selection)*.

Filter samples exceeding the concentration limits

If the option **Filter samples exceeding the concentration limits** is checked, all samples with one or more substance concentrations exceeding the *MRL* are filtered out. Then a **concentration limit filter exceedance factor** (*factor*) is specified, which filters out samples with at least one substance concentration higher than $factor \cdot MRL$.

If the option **Filter samples exceeding the concentration limits** remains unchecked all samples are retained in the analysis.

Substance conversion

When concentration data at the level of measured substances have to be converted to concentration data at the level of *active substances* (or perhaps also inactive substances), then *substance conversion rules* can be specified to provide the rules. This section first describes the basic substance conversion, and then the refinements using available *substance authorisations*.

For each measured substance in the concentration data, there may be zero or more conversion rules (records in the substance conversion rules data source), each linking to an active or inactive substance. Substance conversion rules may specify a link to an exclusive substance or not. For an exclusive conversion it is assumed that only one substance is present in the sample, therefore the measured substance is considered to be just one of the linked substances. It can also be that measured substances link to one or more exclusive substances plus one (non-exclusive) substance that is considered a metabolite of the other exclusive substances. The metabolite can occur together with any of the exclusive substances. It is assumed that either all conversion rules linked to a measured substance are marked as exclusive (case 1), or precisely one rule is marked as exclusive and the other rules are marked as not exclusive (case 2). If this is not the case for any set of rules linked to a measured substance, then this is regarded as erroneous data.

Four methods are implemented for substance conversion:

1. Allocate most potent (EC 2018 Tier 1): For each measured substance, the linked substances are restricted to the active substances of interest. The concentration of the measured substance is assigned to the most potent active substance in this set. Potency is specified by the *relative potency factors*. All other candidate active substances are assigned a zero concentration. I.e., the measured substance concentration is allocated for 100% to the most potent substance specified by the conversion rules and for this allocation, the concentration or LOR is multiplied by the molecular weight correction factor. See *EC2018 Tier 1*.

2. Random allocation (EC 2018 Tier 2): One of the conversion rules is drawn randomly (with equal probability), including the rules of both active and other substances. This drawn rule is used as follows to generate active substance concentrations:

- If the drawn conversion rule is marked as exclusive, the concentration or LOR is allocated to the linked substance.
- If the drawn conversion rule is marked as not exclusive, a proportion *p*, specified by the drawn conversion rule, of the concentration or LOR is allocated to the linked substance. The remaining proportion (*1-p*) is allocated to one other substance, which is the substance that is linked to the measured substance in a conversion rule marked as exclusive (in this case it is assumed that precisely one record per measured substance is marked as exclusive).

All assigned concentrations are multiplied by the molecular weight correction factor. All unselected candidate substances are assigned a zero concentration. See *EC2018 Tier 2*.

3. Nominal estimate: The substances specified through the conversion rules are allocated with a nominal value based on all possible conversion rules. This may be regarded as the nominal or average allocation value of the random sampling method.

- All conversion rules are marked as exclusive: The measured substance concentration is divided over all *n* active substances specified with equal proportions *1/n*, accounting for the molecular weight correction factor for all substances.
- Precisely one conversion rule is marked exclusive and n conversion rules are marked as not exclusive: The measured substance concentration is divided over all active substances specified, with a proportion 1/2 + 1/n for the substance belonging to the exclusive conversion rule, and equal proportions 1/n for the other substances, accounting for the molecular weight correction factor for all substances.

4. Allocate all: The concentration of a measured substance is allocated to each active substance associated with the measured substance as if it were the most potent substance. I.e., the same measured substance is allocated to all associated active substances simultaneously. This method is not sensible when using it in a cumulative assessment, but it is of use in substance screening assessments, where in a combined analysis of multiple substances all active substances are considered independently.

Use of substance authorisations in substance conversion

When *substance authorisations* are available, then these can be used to exclude conversions of measured substances to unauthorised substances on a given food. The information is used as follows in the substance conversion procedures:

1. Allocate most potent: The set of candidate active substances from which the most potent active substance is to be drawn is reduced to only the substances with authorised uses. However, if none of the candidate active substances is authorised, then the most potent of the unauthorised substances is selected for active substance allocation.

2. Random allocation: The set of conversion rules from which to draw is reduced to the rules linking to authorised substances or the non-exclusive substance (thus allowing the selection of a possibly unauthorised metabolite of an authorised substance). If none of the conversion rules links to an authorised substance, then one rule is drawn from the full set of all (unauthorised) conversion rules.

3. Nominal estimate: The set of conversion rules is reduced in the same way as in *Tier 2*. Nominal calculation is performed on the resulting set of conversion rules.

4. Allocate all: For this method, the same rules apply as for *allocate most potent*. The set of candidate active substances that are to be allocated is reduced to only the substances with authorised uses. Hence, a substance is not allocated when it is not authorised and there is at least one other candidate active

substance that is authorised. However, if none of the candidate active substances is authorised, then the most potent of the unauthorised substances is selected for active substance allocation.

Food extrapolation

If the *food extrapolation* setting has been checked, extrapolation of concentrations is performed for all food/*active substance* combinations for which:

- 1. the number of measurements in the analytical scope is smaller than a given threshold for extrapolation (default 10), and
- 2. there is an *extrapolation rule* allowing extrapolation of concentrations from one or more other foods (the from-food(s)) to the given food (the to-food), and
- 3. (optional criterion:) the substance is associated with *authorised use* for both foods, and
- 4. (optional criterion:) *concentration limits (e.g. MRLs)* on the from-food and to-food exist and are equal. Note: if the **active substance** is not a **measured substance**, then the MRL check has to be made per measurement at the level of the measured substance which provided the concentrations assigned to the active substance.

Food extrapolation is performed by one of the following procedures: 1) Substance-specific imputation of missing values by extrapolated measurements, or 2) Extrapolation of complete samples for multiple substances.

1. Substance-specific imputation of missing values by extrapolated measurements

The missing values in the active substance concentrations of the to-food are *imputed* in a random order by active substance concentrations (positive, nondetect or zero) from a randomised list obtained from the fromfood(s). By matching the randomised lists, each from-food measurement is assigned at most once, so after extrapolation there may still be missing values left, or not all measurements of the from-food(s) may have been used for extrapolation.

Note: In this method, it is assumed that the to-food has a sufficient number of samples. No extrapolation is applied for foods with no samples at all, and data gaps will also remain for foods with fewer than n samples, because no new samples are added.

Note: the resulting *occurrence patterns* will be random with respect to the extrapolated substances, i.e., observed occurrence patterns for the from-food are not extrapolated to the to-food.

2. Extrapolation of complete samples for multiple substances

(not yet implemented)

All samples of the from-food(s), i.e., complete samples with data for all active substances, are copied as samples for the to-food and added to the existing to-food samples. For example, extrapolate all apple sample records to the available pear sample records. However, measurements for substances that do not fulfil the (optional) criteria 3 and 4 above are non-valid extrapolations and are replaced by missing values. The status of the extrapolated samples is stored to distinguish between extrapolated and non-extrapolated sample records. Note that this method maintains correlations in the occurrence patterns and postpones imputation of MVs until the concentration models step.

Water imputation

If water has been selected as an additional source of exposure, but concentration data is missing, then, fixed concentration values can be assigned to water for the five most toxic *active substances*, with the toxicity ranking being based on the *relative potency factors*. For all other substances, zero concentrations are *imputed*. The default imputation value is 0.05 µg/L, but this value can be chosen as a setting. If specified, *substance authorisations* may be used to restrict to the set of active substances for which water concentrations are imputed to only those for which concentrations may be expected from *authorised use*.

Focal commodity scenario analysis

There are different methods for modifying the (background) concentration data for specific (prospective) focal commodity scenario analyses. In the front end, these focal commodity scenario analysis method are accessible through the option *include focal commodity concentrations*. Checking this option will open the focal commodity scenario analysis form (see Figure 2.12) where the method and the focal commodity food/substance can be selected, and accompanying other settings can be configured.

Focal commodity concentrations replacement method	
Replace measurements of focal food/substance combinations with measurements from focal com	nodity samples 🔹
Include focal commodity concentrations	
Beans (with pods)	-
Focal commodity substances	
Emamectin	•
Focal commodity substance occurrence percentage	
25	
Adjustment factor for the focal food/substance concentration	
1	

Figure 2.12: Focal commodity scenario analysis form of the front end. This form is a sub-form of the concentrations module panel.

Replace samples with focal commodity samples

This method will replace all samples for the selected focal commodity food by samples from the *focal commodity concentration data*. It works substance independent, and will therefore replace all substance concentrations of the focal commodity food in the background concentration data.

Append focal commodity samples

This method adds the *focal commodity samples* of the selected focal commodity food to the background concentration data. This method is also substance-independent and may be a useful approach when the substances measured in the field trial do not overlap with the substances of the (background) concentration data. In this case, the focal commodity substance concentrations will be missing for the background concentration data and (also the other way around) the substance concentrations of all other substances will be considered missing for the focal commodity samples. These missing values may be imputed at a later stage following the "normal procedures".

Replace measurements of focal food/substance combinations

This method replaces, for the selected (focal) combination of food and substance, all substance concentrations with focal concentrations. This method knows two variants:

- **Replace by focal commodity samples:** The focal food/substance measurements are obtained from *focal commodity samples*. Here, substance measurements of the focal commodity food in the background concentration data set are replaced by randomly assigned substance measurements of the focal commodity samples.
- **Replace by concentration limits:** The focal food/substance measurements are obtained from *focal commodity samples*. Here, substance measurements of the focal commodity food in the background concentration data set are replaced by the concentration limit value (e.g., an MRL) obtained from the provided *concentration limits data*.

Using the *focal commodity substance occurrence percentage*, it is possible to specify an occurrence percentage for the combination of focal food and substance. When this percentage is less than 100%, this will partly (i.e., for the selected percentage) replace the concentrations of the focal commodity food and substance with the focal concentrations, and for the other part replace the concentrations with zero concentrations. E.g., when aiming to replace background concentrations of the substance fluopyram on potatoes with an MRL value, then specifying a focal commodity substance occurrence percentage of 40% will replace 40% of the measurements with the MRL, and 60% of the measurements with zero concentrations. Note that, because the allocation is random (i.e., each substance measurement has a probability of being assigned a focal concentration or a zero defined by the percentage), the realized replacement percentage may differ from the specified percentage. This option can, for example, be used to simulate a percentage of agricultural use.

Using the *adjustment factor for the focal food/substance concentration*, it is possible to adjust the (positive) concentrations of the focal food and substance measurements. This factor can be used when the focal commodity concentrations (e.g., from field trials) are assumed to be higher than what may be reasonably expected in practice. In this case, this factor could be set for instance, to the expected ratio of mean monitoring concentration and mean field trial concentration. Note that for replacement by focal commodity measurements, this factor will only adjust the positive concentrations and not the LORs.

By default, the focal commodity substance measurements are replaced before the optional step of *converting the concentrations from measured to active substance concentrations*. This also means that for these replaced measurements, the same rules apply, and the measurements may be converted to active substance measurements after replacement. Alternatively, it is possible to replace substance measurements after having done the allocation, and to use *deterministic substance conversions factors* for the focal commodity food and substance to convert these measurements to the level of *active substances*.

Note that when also using using *substance authorisations*, the focal food and substance combination will be treated as authorised, even if there is no authorisation supplied for the combination. The approved authorisation status is considered to be part of this scenario analysis.

Remove measurements of focal food/substance combinations

This method will simply remove all background concentrations for the selected focal commodity food and substance combination, and will not replace them with other values. This method may be useful when a separate analysis is desired for the background and foreground concentrations.

Filter samples by specific sample properties

When the option to *filter samples by specific property values (subset selection)* is checked in the *main panel of the concentrations module*, a new form will appear in which various sample property subsets can be specified (see Figure 2.13). The visibility/availability of the sample property filters depends on the availability of these properties in the data. The following filters are available:

- The options to *filter samples by year* and *filter samples by month* allow the user to filter the samples by sampling date. The additional option to *include samples with missing sampling date* determines whether samples for which the sampling date is unknown/missing should be included or not.
- The options to *filter samples by location* and *filter samples by region* allow for subsets by location (country) and/or region within a location. Also for these options there is the possibility to include/exclude samples for which the location/region is unknown/missing.
- The option to *filter samples by production method* allows for subset selection on production method of the sampled product (e.g., organic or conventional).
- The option to filter samples by additional sample properties available in the data (see *data format*).

Concentrations settings

 Filter samples by years 2018 Filter samples by month Include samples with unspecified sampling date Filter samples by location Filter samples by region Filter samples by region Filter samples by production method Filter samples by additional sample properties Filter samples by additional sample properties 	Sample subset settings	Save Changes
2018 Image: Constraint of the symples of the symple	Filter samples by year	0
 Include samples with unspecified sampling date Filter samples by location Filter samples by region Filter samples by production method 		- 0
 Filter samples by location Filter samples by region Filter samples by production method 	Filter samples by month	0
Filter samples by region i Filter samples by production method i	Include samples with unspecified sampling date	0
Filter samples by production method	Filter samples by location	0
	Filter samples by region	0
✓ Filter samples by additional sample properties (i	Filter samples by production method	0
	Filter samples by additional sample properties	0
Filter samples by importseason		
Filter samples by season		
Selected values * Winter		•

Figure 2.13: Sample subset selection form of the front end. This form is a sub-form of the concentrations module panel.

Selection settings

90

	Selection settings for module Concentrations.
Name	Description
Concentrations tier	Specifies the concentration data should be treated according to a
	pre-defined tier or custom.
Filter samples exceeding the	If checked, samples with at least one substance concentration
concentration limits	higher than some factor (concentration limit filter exceedance
	factor) times the MRL are filtered out.
Concentration limit filter	The multiplication factor for the concentration limit exceedance
exceedance factor	filter.
Use substance conversion rules	If checked, concentrations are modelled in terms of active
	substances (using substance conversion).
Substance conversion method	Allocation method for assigning active substance concentrations
	from measured substance concentrations based on substance
Retain all allocated substances	translations.
	If checked, all allocated substances kept after substance
after active substance allocation	conversion. Otherwise, the concentration data is restricted to the active substances of the assessment group.
Account for substance	Account for substance authorisations when allocating measured
authorisations in substance	substances to active substance using substance conversions.
conversions	substances to active substance using substance conversions.
Use extrapolation rules	Use extrapolation rules to extrapolate food samples for foods with
ose extrapolation rules	a limited amount of samples (data poor foods) from other foods
	(data rich foods).
Threshold for extrapolation	Threshold for extrapolation.
Restrict extrapolations to equal	Restrict extrapolations to equal MRLs.
MRLs	restrict extrupolations to equal mixtus.
Restrict extrapolations to	Only extrapolate if substance use is authorised.
authorised uses	
Impute water concentrations	Impute constant concentration values on the selected (water)
	commodity.
Water commodity	The commodity for which constant concentration values should be
2	added.
Water concentration value	Constant concentration value that should be used for water (in
(µg/kg)	μg/kg).
Restrict water imputation to	Restrict water imputation to the five most toxic substances.
the five most toxic substances	
Restrict water imputation to	Restrict water imputation to authorised uses.
authorised uses	
Include focal commodity	Specifies whether there is monitoring data that should replace part
concentrations	of the consumption data for the specified focal commodities.
Focal commodity foods	The foods for which background concentration data are to be
-	replaced by focal commodity concentrations.
Focal commodity substances	The substances for which background concentration data are to be
	replaced by focal commodity concentrations.
Focal commodity	Replacement method to be used for replacing base concentration
concentrations replacement	data with concentration data of the focal commodity/commodities
method	concentrations.
Focal commodity substance	Anticipated occurrence percentage / agricultural use percentage
occurrence percentage	of the focal commodity.
Adjustment factor for the focal	Optional adjustment factor for the focal food/substance
food/substance concentration	concentration. E.g., the expected ratio of mean monitoring
	concentration and mean field trial concentration.
Use deterministic substance	Convert measured substance concentrations of focal commodity
conversions for focal	to active substance concentrations using deterministic substance
commodity	conversion factors.
Filter samples by specific	Specifies whether a subset selection on specific sample properties
property values (subset	should be made (e.g., by country or by year). Chapter 2. Modu
selection)	
Sample locations	The locations for which samples are filtered.The years for which samples are filtered.
Sample years	

Table 2.55: Selection settings for module Concentrations.

Uncertainty settings

Table 2.56:	Uncertainty	settings for	or module	Concentrations.
14010 2.00.	Oncortainty	bettings io	n moaaie	concentrations.

Name	Description
Resample concentrations	Specifies whether concentrations are resampled by empirical
	bootstrap or using a parametric uncertainty model.

Concentrations tiers

In addition to the possibility for users to work with their own choices for all settings, MCRA implements Tier 1 and 2 from the European Commission working document SANTE-2015-10216 rev. 7 (2018) on risk management aspects related to the assessment of cumulative exposure.

Overview

Name	EC 2018	EC 2018
	Tier 1	Tier 2
Substance conversion method	UseMost-	DrawRan-
	Toxic	dom
Retain all allocated substances after active substance allocation	true	true
Account for substance authorisations in substance conversions	false	true
Use extrapolation rules	true	true
Threshold for extrapolation	10	10
Restrict extrapolations to equal MRLs	true	true
Restrict extrapolations to authorised uses	true	true
Impute water concentrations	true	true
Water concentration value (µg/kg)	0.1	0.05
Restrict water imputation to the five most toxic substances	true	true
Restrict water imputation to authorised uses	false	false

Table 2.57: Tier overview for module Concentrations.

EC 2018 Tier 1

Table 2.58: Tier definition for EC 2018 Tier 1.

Name	Setting
Substance conversion method	UseMostToxic
Retain all allocated substances after active substance allocation	true
Account for substance authorisations in substance conversions	false
Use extrapolation rules	true
Threshold for extrapolation	10
Restrict extrapolations to equal MRLs	true
Restrict extrapolations to authorised uses	true
Impute water concentrations	true
Water concentration value (µg/kg)	0.1
Restrict water imputation to the five most toxic substances	true
Restrict water imputation to authorised uses	false

EC 2018 Tier 2

Name	Setting
Substance conversion method	DrawRandom
Retain all allocated substances after active substance allocation	true
Account for substance authorisations in substance conversions	true
Use extrapolation rules	true
Threshold for extrapolation	10
Restrict extrapolations to equal MRLs	true
Restrict extrapolations to authorised uses	true
Impute water concentrations	true
Water concentration value (µg/kg)	0.05
Restrict water imputation to the five most toxic substances	true
Restrict water imputation to authorised uses	false

Table 2.59: Tier definition for EC 2018 Tier 2.

Concentrations uncertainty

Uncertainty due to a limited number of samples can be accounted for by resampling/bootstrapping. Resampling is done on a sample-based basis preserving co-occurrence of substance residue values on the same sample for multiple-substance analyses.

Concentrations as data

Concentration data can be entered using the internal, relational data format or using the EFSA SSD format. Depending on the settings, the entered concentration data can be pre-processed for conversion to active substances, extrapolation to other foods, and/or default values can be added for water.

- Concentrations data formats
- Concentrations calculation

Inputs used: Focal food concentrations Food extrapolations Substance conversions Deterministic substance conversion factors Relative potency factors Substance authorisations Active substances Concentration limits

2.3.5 Deterministic substance conversion factors

Deterministic substance conversion factors.

This module has as primary entities: Substances Foods

Output of this module is used by: Concentrations Single value concentrations

Deterministic substance conversion factors data formats

Deterministic substance conversion factors

Deterministic substance conversion factors. Foods are optional.

Deterministic substance conversion factors

Deterministic substance conversion factors for translating measured substance concentrations to active substance concentrations.

Name	Туре	Description	Aliases	Required
idMeasured-	AlphaNumeric(50)	Substance code of the	idMeasured-	Yes
Substance		measured substance.	Substance,	
			idResidue-	
			Definition,	
			Residue-	
			Definition,	
			Measured-	
			Substance	
idActive-	AlphaNumeric(50)	Substance code of the active	idActive-	Yes
Substance		substance.	Substance,	
			idSubstance,	
			Active-	
			Substance,	
			Substance	
idFood	AlphaNumeric(50)	The unique identification code	idFood, Code,	No
		of the food.	FoodId,	
			FoodCode, Food	
Conversion-	Numeric	Specifies the conversion factor	Factor,	Yes
Factor		to translate concentrations of	Conversion-	
		the measured substance to	Factor	
		(equivalent) concentrations of		
		the active substance according		
		to e.g. the system used in		
		PRIMo.		
Reference	AlphaNumeric(200)	Reference to the source from	Reference,	No
		which this value is obtained.	References,	
			Source, Sources	

Table 2.60: Table definition for DeterministicSubstanceConversionFactors.

Table aliases: SingleValueSubstanceConversionFactors, SingleValueConversionFactors, SingleValueConversions, SubstanceConversionsFixed, DeterministicSubstanceConversionFactors.

Deterministic substance conversion factors as data

Deterministic substance conversion factors.

• Deterministic substance conversion factors data formats

2.3.6 Focal food concentrations

In some cases the attention in an assessment is to evaluate concentrations (e.g., from specific field trials) for a specific food (and substance), in combination with a background of concentration data for other foods. Focal food concentrations can be included to provide these separate (foreground) concentration data for one or more focal food commodities that should replace measurements in the (background) *concentration data* in *focal commodity scenario analyses*.

This module has as primary entities: Foods Substances

Output of this module is used by: Concentrations

Focal food concentrations data formats

See concentration data formats.

Focal food concentrations settings

Selection settings

Table 2.61 Selection	settings for module	e Focal food concentrations.
	sounds for mouuld	

Name	Description	
Focal commodity foods	The foods for which background concentration data are to be	
	replaced by focal commodity concentrations.	
Focal commodity substances	The substances for which background concentration data are to be	
	replaced by focal commodity concentrations.	

Calculation settings

Table 2.62: Calculation settings for module Focal food concentrations.

Name	Description
Focal commodity	Replacement method to be used for replacing base concentration
concentrations replacement	data with concentration data of the focal commodity/commodities
method	concentrations.

Focal food concentrations as data

Focal food concentrations are concentration data and specified in the exact same manner. The difference is that this data will be used to replace part of the concentration data in order to combine specific concentration data with a background of ordinary concentration data.

• Focal food concentrations data formats

2.3.7 Food extrapolations

Food extrapolations data specify which foods (data rich foods) can be used to impute concentration data for other foods with insufficient data (data poor foods).

This module has as primary entities: *Foods*

Output of this module is used by: Concentrations Food conversions

Food extrapolations data formats

Food extrapolation rules

Food extrapolations (or read-across food translations) can be used to specify whether data (e.g, occurrence data) on a food for which this is missing (a data poor food) may be extrapolated from another food for which data is available (read-across food).

Food extrapolations

Food extrapolations are simply specified as combinations of two food codes. One code for the food for the data poor food, and one for the data rich food (or read-across food).

Name	Туре	Description	Aliases	Required
DataPoorFood	AlphaNumeric(50)	The code of the data poor food. I.e., the food for which missing data is allowed to be extrapolated.	IdFoodData- Poor, FoodDataPoor, idFromFood, FromFoodId, FromFood, FoodFrom, Food, IdFood	Yes
CodeDataRich- Food	AlphaNumeric(50)	The code of the read-across food (or data rich food). I.e., the food from which data is used for extrapolation.	IdFoodData- Rich, FoodDataRich, IdFoodRead- Across, FoodRead- Across, IdReadAcross- Food, ReadAcross- Food, idToFood, ToFoodId, ToFood, FoodTo	Yes

Table aliases: ReadAcrossFoodTranslations, ReadAcrossFoodTranslation, ReadAcrossTranslations, ReadAcrossTranslation, FoodExtrapolations, FoodExtrapolation.

Food extrapolations as data

Food extrapolations are specified as data in the form of simple tuples of data rich food and data poor food for which extrapolation is allowed/reasonable.

• Food extrapolations data formats

2.3.8 Modelled foods

Modelled foods are foods within the foods scope for which concentration data or MRLs of substances are available (or expected).

This module has as primary entities: Foods Substances

Output of this module is used by: Concentration models Food conversions

Modelled foods calculation

Modelled foods are the foods within the foods scope for which concentration data or MRLs of substances are available (or expected). Modelled foods are derived primarily from *concentration data*. That is, all foods for which food samples are available in the concentration data or MRL data are considered to be modelled foods. In addition, this set may be extended when *concentration limits* such as MRLs are available (see *calculation settings*) and/or when *food extrapolation rules* are used. Foods for which such data is available are considered to be modelled foods. The set of foods can also be restricted by omitting foods with only non-detect measurements (see *calculation settings*).

Modelled foods settings

Calculation settings

6
Description
If checked, then the assessment is restricted to the specified
modelled foods.
Set of modelled foods that are of particular interest.
Derive modelled foods from sample based concentration data.
Derive modelled foods from single value concentrations.
Derive modelled foods from concentration limits.
Specifies whether foods with only non-detect measurements are
part of the exposure assessment (default yes).
Specifies whether substances with only non-detect measurements
are part of the exposure assessment (default yes).
Specifies whether substances without any measurements should be
included.

Table 2.64: Calculation settings for module Modelled foods.

Calculation of modelled foods

Modelled foods are computed from concentration data (which may also be in the form of single-value concentrations) and/or derived from available maximum residue limits.

• Modelled foods calculation

Inputs used: Concentrations Single value concentrations Concentration limits

Settings used

• Calculation Settings

2.3.9 Occurrence frequencies

Occurrence frequencies specify how often substances occur on foods. Frequencies are expressed as percentages.

This module has as primary entities: Foods Substances

Output of this module is used by: Concentration models Single value dietary exposures

Occurrence frequencies data formats

Occurrence frequencies

Occurrence frequencies are described by one simple table, specifying for pairs of food and substance, the associated occurrence frequencies as percentages.

Occurrence frequencies

Occurrence frequencies are specified as percentages for pairs of food and substance. Optionally, a reference can be included in each record to specify the source (e.g., from literature) from which the percentage was obtained.

Name	Туре	Description	Aliases	Required
idFood	AlphaNumeric(50)	The food code.	idFood,	Yes
			CodeFood,	
			FoodId,	
			FoodCode, Food	
idSubstance	AlphaNumeric(50)	Code of the substance.	idSubstance,	Yes
			CodeSubstance,	
			SubstanceId,	
			SubstanceCode,	
			Substance	
Percentage	Numeric	The occurrence frequency	Percentage,	Yes
		percentage.	Frequency-	
			Percentage	
Reference	AlphaNumeric(200)	Reference to the source from	Reference,	No
		which this use frequency	References,	
		value is obtained.	Source, Sources	

Table 2.65: Table definition for OccurrenceFrequencies.

Table aliases: OccurrenceFrequencies.

Occurrence frequencies calculation

Occurrence frequencies can be provided as data or computed from *occurrence patterns*. For a food and substance, they are computed by collecting all occurrence patterns of this food and summing up the frequencies of the occurrence patterns containing the substance. In the unlikely case that the total frequency of the occurrence patterns of a food exceeds 100%, then a rescaling is applied first. If the sum of the frequencies does not sum up to 100%, the interpretation of the remaining unspecified percentage can be designated as either "no use" or "all use". In the the first case it is assumed that none of the substances occur on this remaining percentage. In the latter it is assumed that all of the substances occur on this remaining percentage. This choice is available as the setting *associate the unspecified percentage with no-occurrence for foods with at least one specified occurrence pattern*.

Depending on the setting *apply occurrence pattern percentages*, occurrence frequencies can be computed in a crisp form in which the occurrence frequency is either 0% or 100% or as percentages ranging from 0% to 100%.

Occurrence frequencies Settings

Selection settings

Table 2.66: Selection settings for module Occurrence frequencies.			
Name	Description		
Associate the unspecified	If checked, for foods with at least one specified occurrence		
percentage with no-occurrence	pattern, unspecified occurrence patterns for the same food are		
for foods with at least one	assumed to be associated with no use. If unchecked, all		
specified occurrence pattern	substances are considered to be authorised (potentially present in		
	samples). Note that this setting cannot be used for foods that have		
	no specified AUs. These foods have 100% potential presence of		
	all substances. To declare all AUs on such a food un-authorised,		
	include an empty AU with percentage 100% in the AU data table		
	(i.e., use an AU for this food, without specifying substances in the		
	AU Substances table)		
Apply occurrence pattern	If checked, use the percentages of potential presence as specified		
percentages	by the occurrence patterns. If unchecked, 100% potential		
	presence in samples is assumed for all substances identified by the		

T 11 2 (C 0 1 . .

Occurrence frequencies as data

Occurrence frequencies are described by one table, specifying for a food and substance the associated occurrence frequency as percentage.

occurrence patterns.

• Occurrence frequencies data formats

Inputs used: Active substances

Calculation of occurrence frequencies

Occurrence frequencies for a food and substance are computed according to the model that is part of the EC 2018 Tier II definition (see van Klaveren et al. 2019)

• Occurrence frequencies calculation

Inputs used: Occurrence patterns

2.3.10 Occurrence patterns

Occurrence patterns (OPs) are the combinations (or mixtures) of substances that occur together on foods and the frequencies of these mixtures occurring per food, expressed in percentages. In the context of pesticides, occurrence patterns are associated with agricultural use percentages. Occurrence patterns are relevant to account for co-occurrence of active substances in exposed individuals. Occurrence patterns may be specified as data or modelled based on observed patterns of positive concentrations.

This module has as primary entities: Foods Substances

Output of this module is used by: Occurrence frequencies Dietary exposures

Occurrence patterns data formats

Agricultural uses

Agricultural use percentages for plant protection products (PPPs) may be of use for concentration modelling, as they provide information about what substance mixtures are expected to be present simultaneously on food samples. Especially for non-detect concentration measurements, this information may aid to determine whether the non-detect measurement originated from a true zero or may be a positive measurement below the limit of detection. Agricultural use percentages are specified using the agricultural uses and agricultural use substances table. This data format expects agricultural use percentages to be specified for mixtures of substances. Each mixture has an id (idAgriculturalUse) and a list of substances that are part of this mixture (agricultural use substances). These agricultural uses are assumed to be exclusive (i.e., only one mixture or PPP is used per sample). Hence, the sum of the agricultural uses for one food should not exceed 100%.

Agricultural uses

The AgriculturalUses contains the definitions of the agricultural use mixtures, or PPPs and the specification of the percentage of the products treated with this mixture. Optionally also the time period of the use percentage may be specified.

Name	Туре	Description	Aliases	Required
idAgricultural-	AlphaNumeric(50)	The unique identification code	idAgricultural-	Yes
Use		of the agricultural use group /	Use,	
		plant protection product	AgriculturalUse-	
		(PPP).	Id, Id	
idFood	AlphaNumeric(50)	The food code.	idFood, FoodId,	Yes
			Food	
Location	AlphaNumeric(50)	The location or country code,	Country,	No
		agricultural use location.	Location	
StartDate	DateTime		StartDate	No
EndDate	DateTime		EndDate	No
Percentage-	Numeric	The percentage agricultural	PercentageCrop-	Yes
CropTreated		use (%).	Treated,	
			Percentage,	
			PercCrop-	
			Treated,	
			PercentageUse	

Table 2.67: Table definition for AgriculturalUses.

Table aliases: AgriculturalUses, AgriculturalUse.

Agricultural use substances

The agricultural use substances table records the substances that are part of the agricultural use mixtures (PPPs).

	Tuble 2.00. Tuble u	children for Agriculture sestiu	seompounds.	
Name	Туре	Description	Aliases	Required
idAgricultural-	AlphaNumeric(50)	The agricultural use code,	idAgricultural-	Yes
Use		normally a code for a	Use,	
		combination of authorised	AgriculturalUse-	
		substances.	Id	
idSubstance	AlphaNumeric(50)	The code of the substance.	idSubstance,	Yes
			SubstanceId,	
			SubstanceCode,	
			Substance	

Table 2.68: Table definition for AgriculturalUsesHasCompounds.

Table aliases: AgriculturalUseHasSubstances, AgriculturalUsesHasSubstances, AgriculturalUseSubstances, AgriculturalUseGroups, AgriculturalUseGroup.

Occurrence patterns calculation

Assumptions can be made for each food on the basis of findings in concentration data.

Tier 1: 0% occurrence is assumed for all substances with no positive concentrations at all; 100% occurrence is assumed for all substances with at least one positive concentration;

Tier 2: 0% occurrence is assumed for all substances with no positive concentrations at all; for substancefood combinations with at least one positive (finding), use findings patterns to implement a specific interpretation of Option 5 in the SANTE document, as described below.

Therefore in both tiers, substance-food combinations without any positive finding are handled in the optimistic way by assuming absolute zeroes for any non-detect observation.

If Tier 2 is selected, then for each of the modelled foods a tabulation is made of the observed frequencies of positives for all substance combinations (including the empty set), based on the *active substance concentrations*. For an OP consisting of just one substance, the basic frequency is the number of samples with a positive concentration divided by the number of samples where the substance has been measured (i.e., is not a MV). For an OP consisting of multiple substances, the basic frequency is the number of samples with all concentrations positive for the members divided by the number of samples where all members of the set have been measured.

After calculation of the basic frequencies for all occurrence patterns, these frequencies are rescaled such that the overall sum of frequencies is 100%. When *substance authorisations* are available, then patterns involving unauthorised substances are not rescaled and only those patterns for which all substances are authorised are rescaled such that the sum of all frequencies is 100%.

Note: the Tier 2 procedure is not what is literally written in the SANTE document, but is an interpretation agreed upon by EFSA and RIVM. An alternative model, not yet implemented, but perhaps more in line with the text of the SANTE document, would be to double the basic frequencies to modelled occurrence pattern frequencies. Only if the sum of all frequencies becomes larger than 100%, the set of frequencies would be normalised to 100% sum.

Occurrence patterns settings

Selection settings

Name	Description
Associate the unspecified	If checked, for foods with at least one specified occurrence
percentage with no-occurrence	pattern, unspecified occurrence patterns for the same food are
for foods with at least one	assumed to be associated with no use. If unchecked, all
specified occurrence pattern	substances are considered to be authorised (potentially present in
	samples). Note that this setting cannot be used for foods that have
	no specified AUs. These foods have 100% potential presence of
	all substances. To declare all AUs on such a food un-authorised,
	include an empty AU with percentage 100% in the AU data table
	(i.e., use an AU for this food, without specifying substances in the
	AU Substances table)
Apply occurrence pattern	If checked, use the percentages of potential presence as specified
percentages	by the occurrence patterns. If unchecked, 100% potential
	presence in samples is assumed for all substances identified by the
	occurrence patterns.
Scale up use frequency to	Scale up use frequency to 100%.
100%	
Restrict use percentage	Restrict use percentage up-scaling to authorised uses.
up-scaling to authorised uses	

Table 2.69: Selection settings for module Occurrence patterns.

Uncertainty settings

Name	Description
Recompute occurrence	Specifies whether occurrence patterns should be recomputed in
patterns	the uncertainty runs.

Table 2.70: Uncertainty settings for module Occurrence patterns.

Occurrence patterns tiers

Overview

Table 2.71: Ther overview for module Occurrence patterns.				
Name	EC 2018	EC 2018		
	Tier 1	Tier 2		
Apply occurrence pattern percentages	false	true		
Substance conversion method	UseMost-	DrawRan-		
	Toxic	dom		
Retain all allocated substances after active substance allocation	true	true		
Account for substance authorisations in substance conversions	false	true		
Use extrapolation rules	true	true		
Threshold for extrapolation	10	10		
Restrict extrapolations to equal MRLs	true	true		
Restrict extrapolations to authorised uses	true	true		
Impute water concentrations	true	true		
Water concentration value (µg/kg)	0.1	0.05		
Restrict water imputation to the five most toxic substances	true	true		
Restrict water imputation to authorised uses	false	false		
Scale up use frequency to 100%		true		
Restrict use percentage up-scaling to authorised uses		true		

Table 2.71: Tier overview for module Occurrence patterns.

EC 2018 Tier 1

Name	Setting	From	In
		input tier	module
Apply occurrence pattern percentages	false		
Substance conversion method	UseMost-	EC 2018	Concen-
	Toxic	Tier 1	trations
Retain all allocated substances after	true	EC 2018	Concen-
active substance allocation		Tier 1	trations
Account for substance authorisations	false	EC 2018	Concen-
in substance conversions		Tier 1	trations
Use extrapolation rules	true	EC 2018	Concen-
		Tier 1	trations
Threshold for extrapolation	10	EC 2018	Concen-
		Tier 1	trations
Restrict extrapolations to equal MRLs	true	EC 2018	Concen-
		Tier 1	trations
Restrict extrapolations to authorised	true	EC 2018	Concen-
uses		Tier 1	trations
Impute water concentrations	true	EC 2018	Concen-
		Tier 1	trations
Water concentration value (µg/kg)	0.1	EC 2018	Concen-
		Tier 1	trations
Restrict water imputation to the five	true	EC 2018	Concen-
most toxic substances		Tier 1	trations
Restrict water imputation to	false	EC 2018	Concen-
authorised uses		Tier 1	trations

Table 2.72: Tier definition for EC 2018 Tier 1.

EC 2018 Tier 2

Name	Setting	From input tier	In module
Apply occurrence pattern percentages	true		
Scale up use frequency to 100%	true		
Restrict use percentage up-scaling to authorised uses	true		
Substance conversion method	DrawRan-	EC 2018	Concen-
	dom	Tier 2	trations
Retain all allocated substances after	true	EC 2018	Concen-
active substance allocation		Tier 2	trations
Account for substance authorisations	true	EC 2018	Concen-
in substance conversions		Tier 2	trations
Use extrapolation rules	true	EC 2018	Concen-
		Tier 2	trations
Threshold for extrapolation	10	EC 2018	Concen-
		Tier 2	trations
Restrict extrapolations to equal MRLs	true	EC 2018	Concen-
		Tier 2	trations
Restrict extrapolations to authorised	true	EC 2018	Concen-
uses		Tier 2	trations
Impute water concentrations	true	EC 2018	Concen-
		Tier 2	trations
Water concentration value (µg/kg)	0.05	EC 2018	Concen-
		Tier 2	trations
Restrict water imputation to the five	true	EC 2018	Concen-
most toxic substances		Tier 2	trations
Restrict water imputation to	false	EC 2018	Concen-
authorised uses		Tier 2	trations

Table 2.73: Tier definition for EC 2018 Tier 2.

Occurrence patterns as data

Occurrence patterns are provided as data by specification of the occurrence mixtures and their associated occurrence/agricultural use percentages.

• Occurrence patterns data formats

Inputs used: Substance authorisations Active substances

Calculation of occurrence patterns

Occurrence patterns are computed from the observed patterns of positive concentrations in the concentration data.

Occurrence patterns calculation

Inputs used: Concentrations

2.3.11 Processing factors

Processing factors are multiplication factors to derive the concentration in a processed food from the concentration in an unprocessed food and can be specified for identified processing types (e.g., cooking, washing, drying). Processing factors are primarily used in dietary exposure assessments to correct for the effect of processing on substance concentrations in dietary exposure calculations.

This module has as primary entities: Foods Substances

Output of this module is used by: Food conversions Dietary exposures Single value dietary exposures

Processing factors data formats

Processing factors connect two food codes, one for the processed food and one for the unprocessed food. There are two schemes to make this connection:

- 1) specify the two food codes and the processing type, or
- 2) use food facets, i.e. specify only the code of the unprocessed food and the processing type (facet), the code of the processed food is defined by the other two.

Processing factors

Processing factors are defined for triplets of processing type, food, and substance. The processing types are defined in the processing types table and the processing factors are defined in the processing factors table.

Processing factors

Processing factor records should be linked to processing types using the processing type code (idProcessingType) and for the foods and substances. The codes of the processing factor records should match the codes of the foods, substances, and processing type definitions.

Name	Туре	Description	Aliases	Required
idProcessing- Type	AlphaNumeric(50)	The code of the processing type.	idProcessing- Type,	Yes
			ProcessingType- Id, ProcessingType,	
idSubstance	AlphaNumeric(50)	The code of the substance.	ProcType idSubstance, SubstanceId, SubstanceCode, Substance	No
idFood- Processed	AlphaNumeric(50)	The code of the processed food.	idFood- Processed, FoodProcessed- Id, FoodProcessed	Yes
idFood- Unprocessed	AlphaNumeric(50)	The code of the unprocessed food.	idFood- Unprocessed, Food- UnprocessedId, idFood, FoodId, Food- Unprocessed	Yes
Nominal	Numeric	The nominal value (best estimate of 50th percentile) of processing factor (defines median processing factor).	Nominal, ProcNom	No
Upper	Numeric	The upper value (estimate of 95th percentile or "worst case" estimate) of processing factor due to variability.	Upper, ProcUpp	No
Nominal- Uncertainty- Upper	Numeric	The upper 95th percentile of nominal value (Nominal) due to uncertainty. A standard deviation for uncertainty of the nominal value (Nominal) is derived using the nominal value (Nominal) and upper 95th percentile (NominalUncertaintyUpper).	Nominal- Uncertainty- Upper, ProcNomUnc- Upp	No
Upper- Uncertainty- Upper	Numeric	The upper 95th percentile of upper value (Upper) due to uncertainty. From the nominal value (Nominal), upper value (Upper) and the specified uncertainties of these values (NominalUncertaintyUpper and UpperUncertaintyUpper, respectively) the degrees of freedom of a chi-square distribution describing the uncertainty of the standard deviation for variability is derived.	Upper- Uncertainty- Upper, ProcUppUnc- Upp	No

Table aliases: ProcessingFactors, ProcessingFactor, Processing.

Food facet processing factors

This table can be used to define processing factors for (FoodEx2) food/food-facet combinations.

Name	Туре	Description	Aliases	Required
idProcessing- Type	AlphaNumeric(50)	The code of the processing type.	idProcessing- Type, ProcessingType- Id, ProcessingType, ProcType, facet, idFacet, codeFacet	Yes
idFood	AlphaNumeric(50)	The food to which this facet should be linked.	idFood, FoodId, Food	Yes
idSubstance	AlphaNumeric(50)	The code of the substance.	idSubstance, SubstanceId, SubstanceCode, Substance	No
Nominal	Numeric	The nominal value (best estimate of 50th percentile) of processing factor (defines median processing factor).	Nominal, ProcNom	No
Upper	Numeric	The upper value (estimate of 95th percentile or "worst case" estimate) of processing factor due to variability.	Upper, ProcUpp	No
Nominal- Uncertainty- Upper	Numeric	The upper 95th percentile of nominal value (Nominal) due to uncertainty. A standard deviation for uncertainty of the nominal value (Nominal) is derived using the nominal value (Nominal) and upper 95th percentile (NominalUncertaintyUpper).	Nominal- Uncertainty- Upper, ProcNomUnc- Upp	No
Upper- Uncertainty- Upper	Numeric	The upper 95th percentile of upper value (Upper) due to uncertainty. From the nominal value (Nominal), upper value (Upper) and the specified uncertainties of these values (NominalUncertaintyUpper and UpperUncertaintyUpper, respectively) the degrees of freedom of a chi-square distribution describing the uncertainty of the standard deviation for variability is derived.	Upper- Uncertainty- Upper, ProcUppUnc- Upp	No

Table 275.	Table	definition	for	FoodFacetProcessingFactors.
Table 2.75 .	rable	deminition	101	roouraceir focessingraciors.

 $Table\ aliases:\ FoodFacetProcessingFactors,\ FoodFacetProcessingFactor,\ FacetProcessingFactors,\ FoodFacetProcessingFactors,\ FacetProcessingFactors,\ FacetProcessingFacetProcessingFactors,\ FacetProcessingFacetProces$ FacetProcessingFacetProcessingFacetProcessingFacetProcessi

FacetProcessingFactor, FacetProcessing.

Processing factors calculation

Processing factors fixed or distribution based

Processing factors can be specified as fixed factors (nominal) or as statistical distributions for the variability across samples.

- The distribution is either *the logistic-normal distribution* for processing types with factors restricted between 0 and 1 (e.g. washing),
- or the lognormal distribution *the lognormal distribution* for processing types with non-negative factors (e.g. drying).

Variability distribution parameters are specified indirectly via the 50th and 95th percentile. Uncertainty for processing factors can be specified using uncertainty distributions of the same form as for variability. Uncertainty distribution parameters are specified indirectly via the 95th uncertainty percentiles on the 50th and 95th variability distribution percentiles.

For distribution based processing factors specify $f_{k,nominal}$ and $f_{k,upper}$ (Nominal and Upper in table **Processing-Factors**). Two situations are distinguished depending on the type of transformation.

Nonnegative processing factors

Equate the logarithms of $f_{k,nominal}$ and $f_{k,upper}$ to the mean and the 95% one-sided upper confidence limit of a normal distribution. This normal distribution is specified by a mean

$$ln(f_{k,nominal})$$

and a standard deviation

$$ln(f_{k,upper}) – ln(f_{k,nominal}) / 1.645$$

Processing factors between 0 and 1

Equate the logits of $f_{k,nominal}$ and $f_{k,upper}$ to the mean and the 95% one-sided upper confidence limit of a normal distribution. This normal distribution is specified by a mean

$$logit(f_{k,nominal})$$

and a standard deviation

$$logit(f_{k,upper}) - logit(f_{k,nominal})/1.645.$$

See also processing correction

Processing factors settings

Uncertainty settings

Table 2.76:	Uncertainty	settings f	or module	Processing	factors.

Name	Description
Resample processing factors	Specifies whether processing factors are resampled from a
	parametric uncertainty distribution.

Processing factors uncertainty

Processing effects are modelled either by a fixed processing factor, or by a lognormal or logistic-normal distribution (depending on the distribution type of the *processing type*). In case of a fixed factor, the uncertainty distribution is lognormal or logistic-normal with the same mean μ as the fixed value, and with a standard deviation σ_{unc} which is calculated from the specified central value μ (or nominal) and an estimate of the p95 of the *uncertainty distribution* (set *NominalUncertaintyUpper* in the *table for ProcessingFactors*).

The calculation is:

$$\sigma_{unc} = \frac{f(\textit{NominalUncertaintyUpper}) - f(\mu)}{1.645}$$

with f() = logit for the logistic-normal distribution (distribution type 1) and f() = ln for the lognormal distribution (distribution type 2). Values lower than 0.01 or higher than 0.99 (distribution type 1 only) are replaced by default values (0.01 and 0.99); this is useful computationally to avoid problems. In each iteration of the uncertainty analysis a new value is drawn from this distribution to be used as a fixed factor in the Monte Carlo calculation. In case of distribution based processing factors (describing the variability of processing factors) two uncertainties can be specified. For σ_{unc} , specification and calculation is as before (set *NominalUncertaintyUpper* in the *table for ProcessingFactors*).

The uncertainty about the variability standard deviation

$$\sigma_{var} = \frac{f(Upper) - f(\mu)}{1.645}$$

can be specified by the UpperUncertaintyUpper value. This value is specified as the p95 upper limit on Upper. The specified value is used to derive in a iterative search the number of degrees of freedom df (van der Voet et al. 2009) [[van der Voet et al., 2009]]. In the uncertainty analysis, a modified chi-square distribution with df degrees of freedom is used to generate new values of σ_{var} . A very high value of df means little uncertainty and σ_{var} will be almost equal in all iterations of the uncertainty analysis. A df close to 0 means a large uncertainty and very different values of σ_{var} will be obtained in the iterations of the uncertainty analysis. The p95 upper limit on Upper is set through parameter UpperUncertaintyUpper.

Processing factors as data

Specify for a combination of processing type, food and substance the processing factor (nominal, upper).

- Processing factors data formats
- Processing factors calculation

2.3.12 Single value concentrations

Single value concentrations data are the single value estimates (High Residue, Maximum Residue Limit, Supervised Trials Median Residue) of residue concentrations on modelled foods.

This module has as primary entities: Foods Substances

Output of this module is used by: Modelled foods Single value dietary exposures

Single value concentrations data formats

Single value concentrations data provides a single value concentration for a substance.

Single value concentration data

Concentration single values

The food codes (idFood) and substance codes (idSubstance) should match the codes of the foods and substances table respectively.

Name	Туре	Description	Aliases	Required
idFood	AlphaNumeric(50)	Code of the food of this	idFood, FoodId, Food	Yes
10 1.4	A 1.1. N	concentration single value.		Yes
idSubstance	AlphaNumeric(50)	Code of the substance of this	idSubstance,	res
		concentration single value.	SubstanceId,	
			Substance,	
			idCompound,	
			CompoundId,	
X7.1	NT		Compound	X
Value	Numeric	Concentration single value.	Value,	Yes
			Concentration,	
			Concentration-	
<u></u>			Value	
ValueType	Concentration Value-	Value type of the	Concentration-	Yes
	Types	concentration value.	SingleValue-	
			Туре,	
			Concentration-	
			ValueType,	
			SingleValue-	
			Туре,	
			Concentration-	
			Туре,	
			ValueType,	
			Туре	
Percentile	Numeric	Percentile.	Percentile	No
Concentration-	ConcentrationUnits	The unit of the concentration	Concentration-	No
Unit		single value (default mg/kg).	Unit, Unit	
Reference	AlphaNumeric(200)	Reference to the source from	Reference,	No
		which this concentration	References,	
		single value is obtained.	Source, Sources	

 $Table\ aliases:\ ConcentrationSingleValues,\ SingleValueConcentrations.$

Single value concentrations calculation

Single value concentrations as data are supplied as mean concentrations, median concentrations, highest residues, percentiles, LOQs or maximum residue limits. Specify the 'Use data' option in the interface. In a retrospective context, the single values are computed based on the concentration distributions available for the modelled food as supplied in the *Concentrations module*. Specify option 'Compute' in the Single value concentrations action.

Single value concentrations settings

Selection settings

T_11_ 0 70.	Cala atiana a attin	an fam man derl	- Cim -1 1	
-1 and $2/3$	Selection sellin	gs for moanne	- Single value	e concentrations.

Name	Description
Use substance conversion	Specifies whether to use substance conversion factors to convert
factors	measured substance concentrations to active substance
	concentrations.

Single value concentrations as data

Single value concentrations data are the single value concentrations of residues on modelled foods.

• Single value concentrations data formats

Inputs used: Active substances

Calculation of single value concentrations

Single value concentrations are calculated as a percentile (p50, p97.5 or maximum residue limit) of the modelled food concentration distribution.

• Single value concentrations calculation

Inputs used: Concentrations Concentration limits Deterministic substance conversion factors

2.3.13 Substance authorisations

Substance authorisations specify which food/substance combinations are authorised for (agricultural) use. If substance authorisations are used, then only the food/substance combinations that are specified in the data are assumed to be authorised and all other combinations are assumed to be not authorised. This information may, for instance, be used to determine whether concentration measurements below the LOR could be assumed true zeros. I.e., if a food/substance combinations is assumed to be unauthorised, then the LOR may be assumed to be a zero.

This module has as primary entities: Foods Substances

Output of this module is used by: Concentrations Occurrence patterns Concentration models

Substance authorisations data formats

Substance authorisations

Authorised uses data provides information about whether substance use is allowed for specified foods. For cumulative exposure assessments, this information is used for imputation of non-detects/missing values.

Authorised uses

The authorised uses table

Name	Туре	Description	Aliases	Required
idFood	AlphaNumeric(50)	The food code.	idFood, FoodId,	Yes
			Food	
idSubstance	AlphaNumeric(50)	The substance code.	idSubstance,	Yes
			Substance,	
			SubstanceId	
Reference	AlphaNumeric(200)	External reference(s) to	Reference,	No
		sources containing more	References	
		information about the effect		
		(key event) relationships.		

Table 2.79: Table definition for AuthorisedUses.

Table aliases: AuthorisedUses, AuthorisedUse.

Substance authorisations as data

Substance authorisations are specified as data in the form of a list of authorised food/substance combinations, with combinations not on the list associated with no authorised use.

• Substance authorisations data formats

2.3.14 Substance conversions

Substance conversions specify how measured substances are converted to active substances, which are the substances assumed to cause health effects. In the pesticide legislation such measured substances and the substance conversion rules are known as residue definitions.

This module has as primary entities: Substances

Output of this module is used by: Concentrations

Substance conversions data formats

Two types of substance conversions are implemented, with two subtypes for the first type:

1a) The measured substance is one or more of a set of possible substances (e.g. isomers or metabolites), and the toxicity of all substances in this set is assumed to be the same and is expressed in one active substance. Example: The measured substance Parathion-methyl(RD) is either Parathion-methyl or paraoxon-methyl, but both are expressed as the active substance Parathion-methyl.

1b) The measured substance is one or more of a set of possible substances (e.g. isomers or metabolites), and the toxicity of all substances in this set is assumed to relate with equal probability to one of a subset of active substances. Example: The measured substance Dithiocarbamates includes the active substances maneb, mancozeb, metiram, propineb, thiram and ziram, one of which will be assumed to be the active substance present with equal probability.

2) If :n active substances all metabolise to the same active substance (the metabolite), it is assumed that all :n + 1 substances have equal probability of being the source of the measured concentration. The measured substance then is either one active substance (the metabolite) or a mixture of two active substances, one being the metabolite and the other one of the possible parent substances. Example: The measured substance Carbofuran(RD) is either the active substance Carbufuran or a mixture of Carbofuran and one of the possible active parent substances Benfuracarb or Carbosulfan.

Substance conversion rules

Substance conversions are described by a single substance conversions table.

Substance conversion rules

The records of the substance translations definitions table specify which active substances (idActiveSubstance) link to a measured substance (idMeasuredSubstance). Each record contains a conversion factor that specifies how a concentration of the measured substance translates to a concentration of the active substance, a flag that states whether the residue definition should be assumed to translate exclusively to one of its active substances, and a proportion. The proportion specifies the proportion of the samples that should translate to this specific active substance in case the translation is exclusive, otherwise it specifies the proportion of the concentration that is assumed to be attributed to the active substance.

Name	Туре	Description	Aliases	Required
idMeasured- Substance	AlphaNumeric(50)	Substance code of the measured substance.	idResidue- Definition, Residue- Definition, Measured-	Yes
idActive- Substance	AlphaNumeric(50)	Substance code of the active substance.	Substance idActive- Substance, idSubstance, Active- Substance, Substance	Yes
Conversion- Factor	Numeric	Specifies the (molecular weight) conversion factor to translate the concentration of the residue definition to a concentration of the active substance	Conversion- Factor	Yes
IsExclusive	Boolean	Specifies whether a measurement of the residue substance should be translated exclusively to this active substance, or if the residue definition represents/breaks down to a mixture of active substances.	IsExclusive	Yes
Proportion	Numeric	In case the definition is exclusive: the proportion of measurements of the residue definition that can be assumed to translate exclusively to a concentration of the active substance. In case the residue definition is not exclusive, the proportion of the concentration that is assumed to be attributed to the active substance.	Proportion	No

Table aliases: ResidueDefinitions, ResidueDefinition.

Substance conversions as data

Substance conversions are provided as data.

• Substance conversions data formats

Inputs used: Active substances

2.3.15 Total diet study sample compositions

Total diet study sample compositions specify the composition of mixed food samples, such as used in a total diet study (TDS), in terms of their constituting foods.

This module has as primary entities: *Foods*

Output of this module is used by: Concentration models Food conversions

Total diet study sample compositions data formats

Total diet study data

Total diet studies (TDS) complement traditional monitoring of substance concentrations on raw commodities by measuring substance occurrence in main foods prepared as consumed and pooled into representative food groups. To include occurrence data from TDS for exposure assessment, the composition of the TDS samples is needed in order to link the composite samples to the consumed foods (either directly or indirectly). TDS composition data describes the composition of TDS samples by specifying the foods (and the amounts) of TDS samples.

TDS food sample compositions

The TDS food sample compositions table contains the descriptions of the TDS samples and specifications of the foods (with amounts) included in the TDS samples.

Name	Туре	Description	Aliases	Required
idTDSFood	AlphaNumeric(50)	The code of the TDS food.	idTDSFood	Yes
idFood	AlphaNumeric(50)	Sub-food of the TDS food.	idFood	Yes
PooledAmount	Numeric	Total weight (in g) or volume	PooledAmount,	Yes
		(in ml) of the food.	Weight	
Description	AlphaNumeric(200)	Additional description of the	Description	No
		TDS sample (e.g. number of		
		subsamples).		
Regionality	AlphaNumeric	Regionality information.	Regionality	No
Seasonality	AlphaNumeric	Seasonality information.	Seasonality	No

Table 2.81: Table definition for TDSFoodSampleCompositions.

 $Table\ a liases:\ TDSFoodSampleCompositions,\ TDSFoodSampleComposition,\ CompositionTDSFoodSamples,\ CompositionTDSFoodSample.$

Total diet study sample compositions as data

Total diet study sample compositions are provided as data.

• Total diet study sample compositions data formats

2.3.16 Unit variability factors

Unit variability factors specify the variation in concentrations between single units of the same food, which have been put together in a mixture sample on which the concentration measurements have been made. Unit variability factors are used for *modelling unit variability* in acute *(individual) dietary exposures calculations* to account for the fact that concentration data often relate to composite samples, whereas an acute risk may result from consumption of single food units. For the same purpose, they are also used in the *IESTI model* for *single value dietary exposures calculations*.

This module has as primary entities: Foods Substances

Output of this module is used by: Dietary exposures Single value dietary exposures

Unit variability factors data formats

Unit variability factors

Unit variability factors specify the unit-to-unit variation of substance concentrations on foods. Unit variability factors are described by a single unit variability factors table.

Unit variability factors

Unit variability factors are defined for a food, and may possibly also be specified for a specific substance and/or processing type. The unit variability factors are linked to the foods by means of the food code (idFood). Unit variability factors can be specified as unit variability factors (P97.5/mean) or as coefficients of variation of a statistical distribution.

Name	Туре	Description	Aliases	Required
idFood	AlphaNumeric(50)	The food code.	idFood, FoodId,	Yes
idSubstance	AlphaNumeric(50)	The code of the substance.	FoodidSubstance,	No
			SubstanceId,	
			SubstanceCode,	
			Substance	
idProcessing-	AlphaNumeric(50)	The processing type code.	idProcessing-	No
Туре			Type,	
			ProcessingType-	
			Id,	
			ProcessingType,	
			РгосТуре	
Factor	Numeric	The variability factor.	Factor, VarFac,	No
			VariabilityFactor	
UnitsIn-	Numeric	The number of units in the	UnitsIn-	Yes
Composite-		composite sample.	Composite-	
Sample			Sample,	
C	NT		NoUnitComp	NL.
Coefficient	Numeric	The coefficient of variation.	Coefficient,	No
			Variability-	
			Coefficient,	
			CoefVar,	
			VarCoef	

Table 2.82: Table definition for UnitVariabilityFactors.

Table aliases: UnitVariabilityFactors, UnitVariabilityFactor, VariabilityFactor, VariabilityFactors, VariabilityProcCompProd, UnitVariability.

IESTI special cases

IESTI special cases for specified combinations of food, substance. The application type (post-harvest or pre-harvest) determines whether Case 1 or Case 3 should be used.

Name	Туре	Description	Aliases	Required
idFood	AlphaNumeric(50)	The unique identification code	idFood, Code,	Yes
		of the food.	FoodId,	
			FoodCode,	
			Food, Id	
idSubstance	AlphaNumeric(50)	The unique identification code	idSubstance,	Yes
		of the substance. This code	SubstanceId,	
		may be from an existing	Substance,	
		coding system, such as	Code, Id	
		CAS-codes or Param codes of		
		EFSA, or it may be a		
		used-defined code.		
Application-	HarvestApplication-	Harvest application type	Application-	Yes
Туре	Types	(pre-harvest or post-harvest).	Туре,	
			Harvest-	
			ApplicationType	
Reference	AlphaNumeric(200)	External reference(s) to	Reference	No
		pre-harvest use.		

Table 2.83: Table definition for IestiSpecialCases.

Table aliases: IestiSpecialCases.

Unit variability factors as data

Unit variability factors are provided as data.

• Unit variability factors data formats

2.4 Exposure modules

Exposures are, in the simplest applications, *dietary exposures*, which combine consumption and occurrence data, either for single or for multiple *substances* causing the same adverse *effect*. Links between the foods-as-eaten and the *modelled foods* are made using *food conversions*, and the consumptions are expressed as *consumptions per modelled food*. For large assessment groups, the use of *dietary exposures screening* may be used to reduce the complexity of the calculations and only focus calculations on the risk drivers.

In aggregate exposure assessments, *exposures* combine *dietary exposures* with *non-dietary exposures*, which have to be entered as pre-calculated data.

Human monitoring data can be compared to exposures using human monitoring analysis.

In cumulative assessments, important mixtures of substances can be identified using exposure mixtures.

2.4.1 Consumptions by modelled food

Consumptions by modelled food are consumptions of individuals expressed on the level of the foods for which concentration data are available (i.e., the modelled-foods). These are calculated from consumptions of foods-as-eaten and food conversions that link the foods-as-eaten amounts to modelled-foods amounts.

This module has as primary entities: Populations Foods Substances

Output of this module is used by: Single value consumptions High exposure food-substance combinations Dietary exposures

Consumptions by modelled food calculation

Consumptions by modelled food are calculated from *consumptions* of *modelled foods* and *food conversions* that link the foods-as-eaten amounts to modelled-foods amounts. Given that the food conversion is already available, the procedure for computing the consumptions by modelled-food is straightforward. For each consumption of each individual, a modelled-food consumption record is created for each modelled-food that is linked to the consumed foods through the food conversion, with the amount being the total consumption amount multiplied by the proportion indicated by the food conversion. Also, if in the *food conversion algorithm* one or more *processing types* are found, then these types are recorded in the consumption by modelled food record.

Consumptions by modelled food

Calculation settings

Table 2.84: Calculation settings for module Consumptions by modelled food.

Name	Description
Restrict population to	Specifies whether the population should be restricted to the
consumers or consumer days	individuals (chronic) or individual days (acute) with consumptions
only (food-as-measured)	containing any of the foods-as-measured.
Risk type	The type of exposure considered in the assessment; acute (short
	term) or chronic (long-term).
Restrict population to	Specifies whether the population should be restricted to the
consumers or consumer days	individuals (chronic) or individual days (acute) with consumptions
with consumptions of specified	containing any of the specified food-as-measured subset.
foods-as-measured only	
Selected foods-as-measured	Set of consumed modelled foods that are of particular interest for
	restricting the consumers / consumption days.

Calculation of consumptions by modelled food

Consumptions by modelled food are calculated from consumptions of foods-as-eaten and food conversions that link the foods-as-eaten amounts to modelled-foods amounts.

• Consumptions by modelled food calculation

Inputs used: Consumptions Food conversions

Settings used

• Calculation Settings

2.4.2 Dietary exposures

Dietary exposures are the amounts of substances, expressed per kg bodyweight or per individual, to which individuals in a population are exposed from their diet per day. Depending on the exposure type, dietary exposures can be short-term/acute exposures and then contain exposures for individual-days, or they can be long-term/chronic exposures, in which case they represent the average exposure per day over an unspecified longer time period.

This module has as primary entities: Populations Foods Substances Effects

Output of this module is used by: Exposures

Dietary exposures calculation

In probabilistic exposure assessment we consider a population of individuals. Exposure assessment with MCRA can address *acute exposure* or *chronic exposure*. Acute exposure is relevant when the short-term effect on individuals is relevant, chronic exposure when the long-term effects on the individuals matter. In MCRA short-term is operationalised as one day, so effectively acute exposure assessment is concerned with a population of person-days, whereas chronic exposure assessment is concerned with a population of person-days.

The basic operation in exposure assessment is integrating consumptions and concentrations per food. With multiple foods, consumptions are typically correlated, therefore MCRA works with the multivariate distribution of a consumption vector, as represented by the consumption data of individuals in a consumption survey. In contrast, the distributions of concentration for each food are typically considered to be independent between foods. E.g., eating an apple with an accidentally high residue concentration does not predict that another food eaten on the same day will also have a high residue concentration. As a consequence of this assumption, concentrations of substances are modelled for each food independently.

For large assessment groups, the use of *dietary exposures screening* may be used to reduce the complexity of dietary exposures calculations and only focus calculations on the risk drivers. In this case, only detailed information is recorded for the risk drivers. With or without screening MCRA produces the same estimated cumulative exposure distribution summarized by percentiles and exceedance percentages, the same contributions of all substances and all modelled foods. After screening, contributions related to food-as-eaten are available for the risk drivers only.

In cumulative exposure calculations two simple approaches are used to identify and select mixtures contributing to the exposure of a target population:

1. qualitative approach: counting of co-exposure. To which combinations of substances are individuals exposed?

Co-exposure of substances is a qualitative approach where the number of combinations of substances to which an individual is exposed is recorded. There is no cut-off level, the only criterion is the presence of a substance in the simulated daily diet or not. For an *acute* or short term exposure assessment, a simulated individual day is the smallest entity to determine co-exposure. For a *chronic* or long term exposure assessment, co-exposures are summarized at the individual level, e.g. co-exposure is determined combining all consumption days of an individual. For more information see *co-exposure of substances*.

2. quantitative approach: *maximum cumulative ratio (MCR)*. To what degree are mixtures more important than single substances?

A quantitative approach is available in the exposures mixtures module.

Acute exposure assessment

In an acute exposure assessment, the short term exposure to a substance or group of substances is estimated. The interest is in the distribution of individual day exposures and derived statistics like the fraction of days that exceed an intake limit or point of departure (PoD). The PoD is calculated as the acute reference dose (ARfD) * safety factor (SF). The basic model for the exposure to a substance in an acute exposure assessment is:

$$y_{ij} = \frac{\sum_{k=1}^{p} x_{ijk} c_{ijk}}{b w_i}$$

where y_{ij} is the intake by individual *i* on day *j* (in microgram substance per kg body weight), x_{ijk} is the consumption by individual *i* on day *j* of food *k* (in g), c_{ijk} is the (*simulated*) concentration of that substance in food *k* eaten by individual *i* on day *j* (in mg/kg), and bw_i is the body weight of individual *i* (in kg). Finally, *p* is the number of foods accounted for in the model. Within parenthesis, the default unit definitions are assumed, but decimal multiples or submultiples of units are easily specified using the relevant tables.

In the exposure assessment, individual days enter the Monte Carlo sample using the inverse of the sampling weights w_i when the number of MC iterations is > 0 (see *table for Individuals*, field *SamplingWeight*).

Modelling unit-to-unit variation

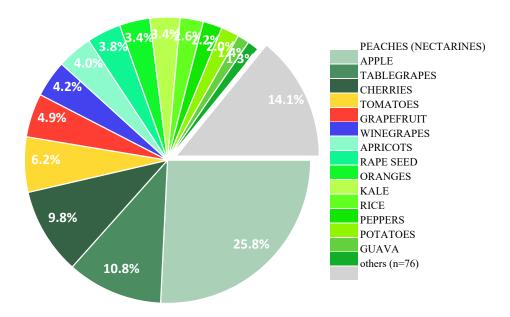
The basic model for an acute exposure assessment assumes that the concentration of the substance displays the variation of residues between units in the marketplace. In general, both monitoring data and controlled field trial data are obtained using composite samples. As a result some of the unit-to-unit variation is averaged out. The model for unit variability aims to adjust the composite sample mean such that sampled concentrations represent the originally unit-to-unit variation of the units in the composite sample.

MCRA offers three distributions to sample from:

- 1. the beta distribution,
- 2. the lognormal distribution,
- 3. and the bernoulli distribution.

The beta distribution simulates values for a unit in the composite sample. It requires knowledge of the number of units in a composite sample and of the variability between units.

The lognormal distribution simulates values for a new unit in the batch. It requires only knowledge of the variability between units.



Contribution to total exposure distribution for foods as measured

Figure 2.14: Example MCRA dietary exposure contributions foods as measured.

Contribution to total exposure distribution for foods as eaten

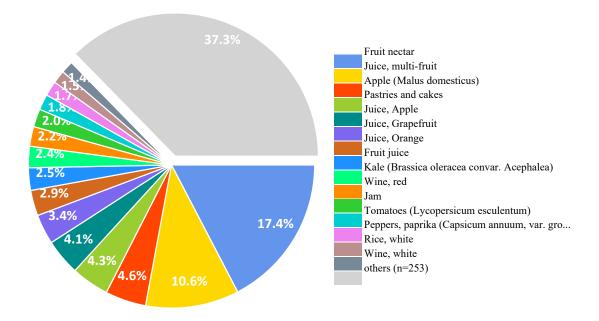
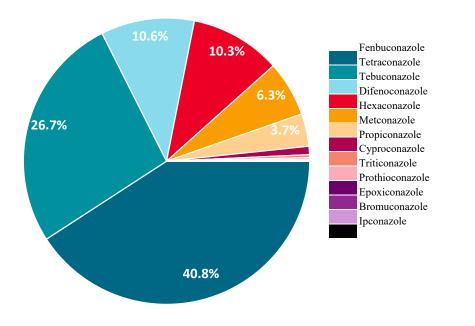


Figure 2.15: Example MCRA dietary exposure contributions foods as eaten



Contribution to total exposure distribution for substances

Figure 2.16: Example MCRA dietary exposure contributions substances

Contribution to total exposure distribution for foods as measured x substances (MSCC)

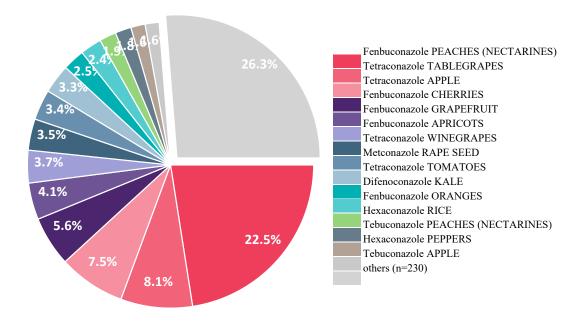


Figure 2.17: Example MCRA dietary exposure contributions foods as measured x substances

The bernoulli distribution is considered as a limiting case of the beta distribution when knowledge of the variability between units is lacking and only the number of units in the composite sample is known. For the beta and lognormal distribution, estimates of unit variability are either realistic (no censoring at the value of the monitoring residue) or conservative (unit values are left-censored at the value of the monitoring residue). For the lognormal distribution sampled concentrations have no upper limit. Whereas for the beta distribution, sampled concentration values for a unit are never higher than the monitoring residue times the number of units in the composite sample.

Variability between units is specified using a variability factor v (defined as 97.5th percentile divided by mean) or a coefficient of variation c_v (standard deviation divided by mean). Following FAO/WHO recommendations, the default variability factor v = 1 for small crops (unit weight < 25 g). For large crops (unit weight > 25 g) v = 5. For foods which are processed in large batches, e.g. *juicing, marmalade/jam, sauce/puree, bulking/blending* the variability factor v = 1 is proposed.

Estimation of intake values using the concept of unit variability

A composite sample for food k is composed of nu_k units with *nominal (whole food/RAC) unit weight* wu_k . The weight of a composite sample is $wm_k = nu_k \cdot wu_k$ with mean residue value cm_k .

- For each iteration *i* in the MC-simulation, obtain for each food k a simulated intake x_{ik} , and a simulated composite sample concentration cm_{ik} .
- Calculate the number of unit intakes nux_{ik} in x_{ik} (round upwards) and set weights w_{ikl} equal to unit weight wu_k , except for the last partial intake, which has weight $w_{ikl} = x_{ik} (nux_{ik} 1)wu_k$.
- For the beta or bernoulli distribution: draw nux_{ik} simulated values bc_{ikl} from a beta or bernoulli distribution. Calculate concentration values as $c_{ikl} = bc_{ikl} \cdot cm_{ik,max} = bc_{ikl} \cdot cm_{ik} \cdot nu_k = svf_{ikl} \cdot cm_{ik}$, where nu_k is the number of units in a composite sample of food k, and svf_{ikl} is the stochastic variability factor for this simulated unit, i.e. the ratio between simulated concentration c_{ikl} and the simulated composite sample concentration cm_{ik} . Sum to obtain the simulated concentration in the consumed portion:

$$c_{ik} = \sum_{l=1}^{nux_{ik}} w_{ikl} c_{ikl} / x_{ik}$$

• For the lognormal distribution: draw nux_{ik} simulated logconcentration values lc_{ikl} from a normal distribution with (optional) a biased mean $\mu = ln(cm_{ik})$ or (default) unbiased mean $\mu = ln(cm_{ik}) - 1/2\sigma^2$ and standard deviation σ . Calculate concentration values as

$$c_{ikl} = \exp(lc_{ikl}) = svf_{ikl} * cm_{ik}$$

where svf_{ikl} is the stochastic variability factor for this simulated unit, i.e. the ratio between simulated concentration c_{ikl} and the simulated composite sample concentration cm_{ik} . Back transform and sum to obtain the simulated concentration in the consumed portion:

$$c_{ik} = \sum_{l=1}^{nux_{ik}} w_{ikl} c_{ikl} / x_{ik}$$

For cumulative exposure assessments, a sensitivity analysis may be performed by specifying a full correlation between concentrations from different substances on the same unit. As a result, high (or low) concentrations from different substances occur together on the same unit. In MCRA, for each unit the random sequence is repeatedly used to generate concentration values for all substances.

Beta distribution

Under the beta model simulated unit values are drawn from a bounded distribution on the interval $(0, c_{max})$ with $c_{max} = nu_k \cdot cm_k$. The standard beta distribution is defined on the interval (0, 1) and is usually characterised by two parameters a and b, with a > 0, b > 0 (see e.g. Mood et al. 1974) [[Mood et al., 1974]]. Alternatively, it can be parameterised by the mean

$$\mu = a/(a+b)$$

and the variance

$$\sigma^2 = ab/(a+b+1)^{-1}(a+b)^{-2}$$

or, as applied in MCRA, by the mean μ and the squared coefficient of variation

$$c_{\rm V}^2 = ba^{-1}(a+b+1)^{-1}$$

For the simulated unit values in *each iteration of the program* we require an expected value cm_k . This scales down to a mean value $\mu = cm_k/c_{max} = 1/nu_k$ in the (standard) beta distribution. From this value for μ and an externally specified value for cv_k the parameters a and b of the beta distribution are calculated as:

$$a = b(nu_k - 1)^{-1}$$

and

$$b = \frac{(nu_k - 1)(nu_k - 1 - cv_k^2)}{nu_k cv_k^2}$$

From the second formula it can be seen that cv_k should not be larger than $\sqrt{nu_k - 1}$ in order to avoid negative values for b. When the unit variability is specified by a variability factor

$$v_k = \frac{p97.5_k}{cm_k}$$

instead of a coefficient of variation cv_k then MCRA applies a bisection algorithm to find a such that the cumulative probability

$$P[Beta(a, b)] = 0.975$$

for $b = a(nu_k - 1)$.

Sampled values from the beta distribution are rescaled by multiplication with cm_{max} to unit concentrations c_{ijk} on the interval $(0, cm_{max})$.

Lognormal distribution

The lognormal distribution is characterised by μ and σ , which are the mean and standard deviation of the logtransformed concentrations. The unit log-concentrations are drawn from a normal distribution with mean $\mu = ln(cm_{ik}) - 1/2\sigma^2$. The coefficient of variation cv is turned into the standard deviation σ on the log-transformed scale with:

$$\sigma = \sqrt{\ln(cv^2 + 1)}$$

The variability factor is defined as the 97.5th percentile of the concentration in the individual measurements divided by the corresponding mean concentration seen in the composite sample. A variability factor v is converted into the standard deviation σ as follows:

$$v = \frac{p97.5}{mean} = \frac{e^{\mu+1.96\sigma}}{e^{\mu+1/2\sigma^2}} = e^{1.96\sigma-1/2\sigma^2}$$

with μ and σ representing the mean and standard deviation of the log-transformed concentrations. So

$$ln(v) = 1.96\sigma - 1/2\sigma^2$$

Solving for σ gives:

$$\sigma^2 - 2 \cdot 1.96\sigma + 2log(v) = 0$$

with roots for σ according to:

$$\sigma=1.96\pm\sqrt{(1.96^2-2log(v))}$$

The smallest positive root is taken as an estimate for σ .

Bernoulli distribution

The bernoulli model is a limiting case of the beta model, which can be used if no information on unit variability is available, but only the number of units in a composite sample is known (see van der Voet et al. 2001). As a worst case approach we may take the coefficient of variation cv as large as possible. When cv is equal to the maximum possible value $\sqrt{nu_k - 1}$, the (unstandardised) beta distribution simplifies to a bernoulli distribution with probability

$$(nu_k-1)/nu_k$$

or

$$(v_k - 1)/v_k$$

 $1/nu_k$

 $1/v_k$

for the value 0 and probability

or

for the value $c_{max} = nu_k \cdot cm_k$.

In MCRA values 0 are actually replaced by cm_k , to keep all values on the conservative side. For example, with $nu_k = 5$, there will be 80% probability at $c_{ijk} = cm_k$ and 20% probability at $c_{ijk} = c_{max}$. When the number of units nu_k in the composite sample is missing, the nominal unit weight wu_k is used to calculate the parameter for unit variability.

Chronic exposure assessment

In a chronic exposure assessment, usual exposure is defined as the long-run average of daily exposure to a substance or group of substances by an individual. The interest is in the distribution of individual exposures and derived statistics like the fraction of individuals that exceed an intake limit or point of departure *PoD*). The PoD is calculated as the average daily intake (ADI) * safety factor (SF). Usually, for an individual, dietary recall data are available on 2 (or more) consecutive days. We assume an equal number of days for each individual, unless specified differently in *table for Individuals*.

For a chronic exposure assessment the available data are used to calculate exposures per person-day (daily exposure):

$$y_{ij} = \frac{\sum_{k=1}^{p} x_{ijk} c_{ijk}}{b w_i}$$

where y_{ij} , x_{ijk} and bw_i are defined as before but now concentrations of the substance found in food k enter the model as the *estimated mean substance concentration value* c_k . Using the person-day exposures MCRA, provides a number of *exposure models* to calculate the distribution of usual exposure at the person level.

Chronic exposure models

Using the person-day exposures MCRA uses one of the following models to calculate the distribution of usual exposure at the person level:

- 1. The observed individual means observed individual means (OIM) model;
- 2. The *logisticnormal-normal (LNN) model*, in a full version that includes the estimation of correlation between exposure frequency and amount, and in a simpler version without this estimation;
- 3. The betabinomial-normal (BBN) model;
- 4. The *discrete/semi-parametric* model known as the Iowa State University Foods (ISUF) model. For this model, an equal number of days per individual is assumed.

In modelling usual exposure, two situations can be distinguished. Foods are consumed on a *daily basis* or foods are *episodically consumed*. For the logistic normal-normal model and the betabinomial-normal model, the latter requires fitting of a two-part model,

- 1. a model for the frequency of consumption, and
- 2. a model for the exposure amount on consumption days.

In the final step, both models are integrated in order to obtain the usual exposure distribution. For daily consumed foods, fitting of the frequency of consumption is skipped and modelling resorts to fitting the model to daily exposure amounts only. Note that the distinction between BNN and LNN disappears and modelling will give equivalent results.

Observed individual means (OIM)

The usual exposure distribution for a population is estimated with the empirical distribution of individual means. Each mean is the average of all single-day exposures for an individual. The mean value for an individual still contains a considerable amount of within-individual variation. As a consequence, the distribution of within-individual means has larger variance than the true usual exposure distribution and estimates using the OIM-method are biased, leading to a too high estimate of the fraction of the population with a usual exposure above some standard. Despite its known tendency to over-estimate high-tail exposures, the OIM method is the method to be used in EFSA (2012) [[EFSA, 2012]] basic assessments.

Model based and model assisted

Following Kipnis et al. [[Kipnis et al., 2009]], some of the models available in MCRA are extended to predict individual usual exposures. This model assisted approach has been added to BBN and LNN when used without correlation) and may be a useful extension in evaluating the relationship between health outcomes and individual usual exposures of foods. In contrast, the estimation of the usual exposure distribution in the general population is called the model based approach. Summarizing, we get Table 2.85:

posure models	
Model based approach	Model assisted approach
	observed individual means (OIM)
betabinomial-normal (BBN)	betabinomial-normal (BBN)
logisticnormal-normal (LNN) without correlation	logisticnormal-normal (LNN) without correlation
logisticnormal-normal (LNN) with correlation	
Iowa State University Foods (ISUF)	

Table 2.85: Model based and assisted approach available for chronic exposure models

The model assisted approach builds on the proposal of Kipnis et al. [[Kipnis et al., 2009]], but is modified to ensure that the population mean and variance are better represented. The method is based on shrinkage of the observed individual means (modified BLUP estimates) and shrinkage of the observed exposure frequencies. The model-assisted usual exposure distribution applies to the population for which the consumption data are representative, and automatically integrates over any covariates present in the model. Model-assisted exposures are not yet available for LNN, and when a covariable is modelled by a spline function of degree higher than 1. In case of a model with covariates the usual exposure is presented in graphs and tables as a *function of the covariates* (conditional usual exposure distributions).

Betabinomial-Normal model (BBN)

The *Betabinomial-Normal (BBN)* model for chronic risk assessment is described in [[de Boer et al., 2009]], including its near-identity to the STEM-II model presented in [[Slob, 2006]]. The BBN model combines a betabinomial model for the exposure frequencies with a normal model for transformed positive exposures.

Logisticnormal-Normal model (LNN with and without correlation)

In the logistic normal-normal (LNN) model, exposure frequencies are modelled by a logistic normal distribution. In notation, for probability p:

$$logit(p) = log(p/1 - p) = \mu - i + \underline{c}_i$$

where μ_i represents the person specific fixed effect model and \underline{c}_i represent person specific random effects with estimated variance component $\sigma_{between}^2$. Similarly as in the BBN model, the positive exposure amounts are modelled, after transformation (logarithmic or Box-Cox), with a normal distribution. This model is referred to as the *LogisticNormal-Normal (LNN)* model. The full *LNN model* model includes the estimation of a correlation between exposure frequency and exposure amount. This is similar to the NCI model described in Tooze et al. [[Tooze et al., 2006]]. A simple and computationally less demanding version of the LNN method does not estimate the correlation between frequency and amount. The models are fitted by maximum likelihood, employing *Gauss-Hermite integration*.

For chronic models amounts are usually transformed before the statistical model is fit. The power transformation, given by y^p , has been replaced by the equivalent Box-Cox transformation. The Box-Cox transformation is a linear function of the power transformation, given by $(y^p - 1)/p$, and has a better numerical stability. *Gauss-Hermite integration* is used for back-transformation (see also *Box Cox power transformation*).

Discrete/semi-parametric model (ISUF)

Nusser et al. [[Nusser et al., 1996]] described how to assess chronic risks for data sets with positive exposures (a small fraction of zero exposures was allowed, but then replaced by a small positive value). The modelling allowed for heterogeneity of variance, e.g. the concept that some people are more variable than others with respect to their consumption habits. However, a disadvantage of the method was the restricted use to contaminated foods which were consumed on an almost daily basis, e.g. dioxin in fish, meat or diary products. The estimation of usual exposure from data sets with a substantial amount of zero exposures became feasible by modelling separately zero exposure on part or all of the days via the estimation of exposure probabilities as detailed in Nusser et al. [[Nusser et al., 1997]] and Dodd [[Dodd, 1996]]. In MCRA, a discrete/semi-parametric model is implemented allowing for zero exposure sure and heterogeneity of variance following the basic ideas of Nusser et al. and Dodd ([[Nusser et al., 1996]]). This implementation of the ISUF model for chronic risk assessment is fully described in de Boer et al. [[de Boer et al., 2009]].

Model-Then-Add

The traditional approach can be termed the Add-Then-Model approach, because adding over foods precedes the statistical modelling of usual exposure. MCRA offers, as an advanced option, an alternative approach termed Model-Then-Add (van der Voet et al. 2014). In this approach the statistical model is applied to subsets of the diet (single foods or food groups), and then the resulting usual exposure distributions are added to obtain an overall usual exposure distribution. The advantage of such an approach is that separate foods or food groups may show a better fit to the normal distribution model as assumed in all common models for usual exposure (including MCRA's *betabinomialnormal* (BBN) model and *logisticnormal-normal* model (LNN)). That this principle can work in practice was shown in previous work (de Boer et al. 2009 [[de Boer et al., 2009]], Slob et al. 2010 [[Slob et al., 2010]], Goedhart et

al. 2012) [[Goedhart et al., 2012]], and a simulation model was developed and implemented in MCRA 7.1 to show how multimodal distributions can arise from adding unimodal distributions of foods that are not always consumed (Slob et al. 2010 [[Slob et al., 2010]], de Boer and van der Voet 2011, [[de Boer et al., 2011]]). For specific cases involving separate modelling of dietary supplements and the rest of the diet, proposals have been made (Verkaik-Kloosterman et al. 2011) [[Verkaik-Kloosterman et al., 2011]]. However, a practical approach to apply the Model-Then-Add approach to general cases of usual exposure estimation was still missing. Therefore a module in MCRA was developed to implement such an approach based on a visual inspection of a preliminary estimate of the usual exposure distribution using the *Observed Individual Means* (OIM) method.

The Model step

At this stage of development the division of foods into a number of food groups is performed in an interactive process, where the MCRA user is presented with a visual display (see example in Figure 2.18) which shows:

- 1. The OIM distribution represented as a histogram, where each bar shows the frequency of exposures (summed over foods) of individuals in a certain exposure interval; each bar is subdivided according to the contributions of the individual foods contributing to those exposures (left panel Figure 2.18).
- 2. The contributions graph, where each of the bars in the OIM histogram is expanded to 100%. This graph allows a better view of the lower bars in the OIM histogram.

The visual display identifies the nine foods that contribute most to the total exposure; the remaining foods are grouped in a rest category to avoid identification problems because of too many colours (right panel Figure 2.18).

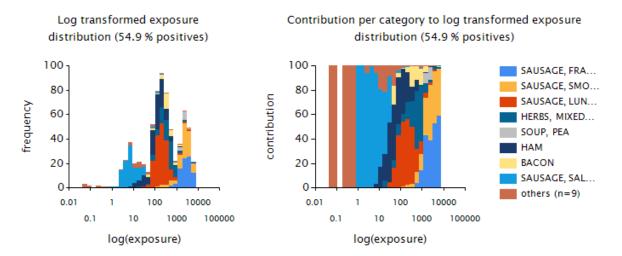


Figure 2.18: Left panel: OIM usual exposure distribution to smoke flavours via the different foods (excluding the zero exposures) in young children; right panel: Contribution of foods to exposures within each bar of the OIM distribution histogram.

The user has now the possibility to select one or more foods and to split these from the main exposure histogram. A separate graph shows the OIM distribution for the split-off food or food group. The graphs for the main group (now called the rest group) are adapted to show the OIM distribution and the contributions for the remaining foods only (see Figure 2.19 upper two panels). This splitting-off can be repeated several times for other foods or food groups. In this way the user can try to obtain foods or food groups that show unimodal OIM distributions. If the result is not what is intended, a food or food group can be added again to the rest group. Per split-off food or food group the usual exposure can be modelled using either BBN or LNN, with a logarithmic or power transformation. The rest group will always be modelled as OIM. It is possible that the rest group is empty, when the total exposure via the different split-off foods and /or food groups is modelled with BBN or LNN.

After a split-off selection has been made, the OIM distribution is summarised in terms of the defined grouping (Figure 2.20), and the usual exposure distribution per split-off food or food group is fitted according to the chosen modelling settings.

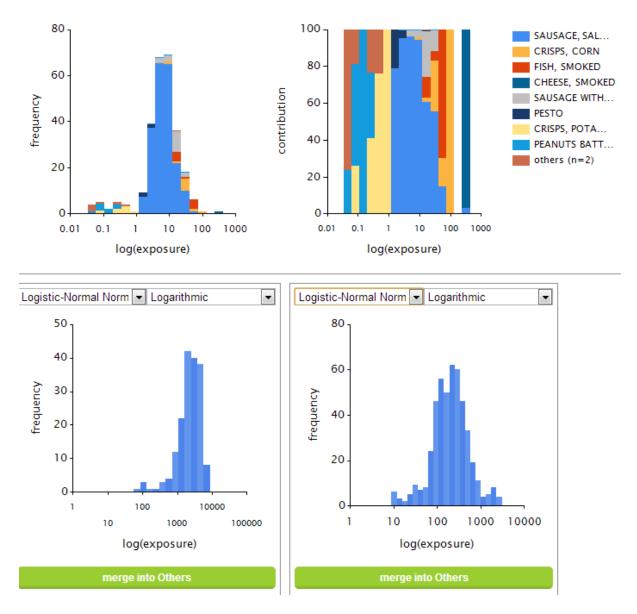
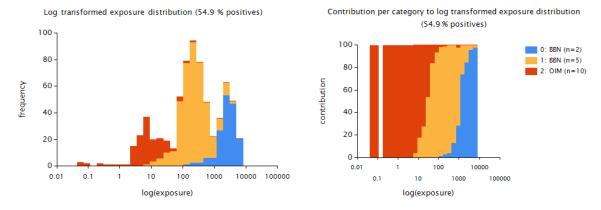


Figure 2.19: Result of a selection into two split-off groups and a rest group. The graph bottom left represents the exposure via a food group containing 'Sausage, frankfurter' and 'Sausage, smoked cooked'. The graph bottom right represents the exposure via a food group containing 'Sausage, luncheon meat', Herbs, mixed, main brands, not prepared', 'Soup, pea', 'Ham', and 'Bacon'. The top graph represents the exposure via the rest group.



Usual exposures per model

Figure 2.20: OIM usual exposure distribution showing the contributions from the three food groups as constructed in Figure 2.19.

The Add step

Consumptions of foods may be correlated. In the traditional Add-Then-Model approach the Add step automatically reflects any correlations that are apparent in the consumptions at the individual-day or individual level. In the Model-Then-Add approach the estimated usual exposure distributions for different foods or food groups have to be combined to assess the total usual exposure. Two approaches are available for this:

- 1. *Model-based approach*: adds independent samples from the usual exposure distribution per food or food group, ignoring any correlations in consumption;
- 2. *Model-assisted approach*: adds the model-assisted, person-specific usual exposure estimates per food or food group, taking correlations in consumptions into account.

See also, episodically consumed foods, model-based, model-assisted.

Before the addition is made, in the model-based approach, model-based estimates of the usual exposure amounts distribution per food or food group are back-transformed values from the normal distribution assumed for transformed amounts per food or food group, and the *model-based frequency* distribution is sampled to decide if a simulated individual has exposure via the food or food group or not. Model-assisted estimates of the usual exposure distribution are back-transformed values from a shrunken version of the transformed OIM distribution, also done per food or food group, where the shrinkage factor is based on the variance components estimated using the linear mixed model for amounts at the transformed scale (van Klaveren et al. 2012). For individuals with no observed exposure (OIM=0) no model-assisted estimate of usual exposure can be made and a model-based replacement is used.

The model-based approach was investigated in Slob et al. (2010) [[Slob et al., 2010]] and performed surprisingly well, even if correlations in consumptions of foods were present. The model-assisted approach adds exposures at the individual level, and therefore retains effects of correlations between foods in the usual exposure distribution.

MCRA calculates both the model-based and model-assisted usual intake distributions.

Chronic exposure as a function of covariates

The intake frequency and transformed intake amounts may be modelled as a function of covariates. MCRA allows one covariable and/or one cofactor.

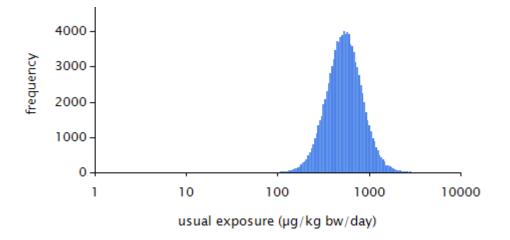


Figure 2.21: Model-assisted estimated usual exposure distributions (excluding the zero exposures).

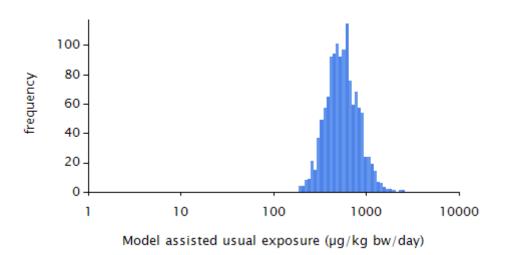


Figure 2.22: Model-based estimated usual exposure distributions (excluding the zero exposures).

001	unutes.	
	Frequencies	Amounts
cofactor	$\mathit{logit}(\pi) = \beta_{0l}$	$transf(y_{ij}) = \beta_{0l} + c_i + u_{ij}$
covariable	$\textit{logit}(\pi) = \beta_0 + \beta_1 f(x_1; df)$	$transf(y_{ij}) = \beta_0 + \beta_l f(x_1; df) + c_i + u_{ij}$
both	$\textit{logit}(\pi) = \beta_{0l} + \beta_1 f(x_1; df)$	$transf(y_{ij}) = \beta_{0l} + \beta_l f(x_1; df) + c_i + u_{ij}$
interaction	$\textit{logit}(\pi) = \beta_{0l} + \beta_{1l} f(x_1; df)$	$\mathit{transf}(y_{ij}) = \beta_{0l} + \beta_{1l}f(x_1; \mathit{df}) + c_i + u_{ij}$

Table 2.86: Intake frequencies and amounts, modelled as a function of covariates.

Here $l = 1 \cdots L$ and L is the number of levels of the cofactor, y_{ij} , the intake amount, x_1 is the covariable, f is a polynomial function with the degrees of freedom df, c_i and u_{ij} are the individual effect and interaction effect, respectively. These effects are assumed to be normally distributed $N(0, \sigma_{between}^2)$ resp. $N(0, \sigma_{within}^2)$. The degree of the function is determined by backward or forward selection. In the output, the usual intake is displayed for a specified number of values of the covariable and/or the levels of the cofactor.

Total Diet Study

In Total Diet Studies (TDS), substance occurrence data is obtained from measuring food products as consumed. TDS offers a more direct measure of substance concentrations compared to traditional monitoring and surveillance programs that are concerned with contamination of raw agricultural commodities. In a TDS, food selection is based on national consumption data in such a way that 90 to 95% of the usual diet is represented by the samples. Selected foods are collected, prepared as consumed and related foods are pooled prior to analysis. The compositions these TDS food samples are described by the *TDS food sample compositions* data module.

In MCRA, TDS concentration data can also be used in *dietary exposure assessments*, using it as an alternative type of concentration data where the modelled foods are not the raw primary commodities (RACs), but these are TDS food compositions. To link the concentration data to the consumed foods, the *TDS food sample composition information* is used in the *food conversion algorithm* in a manner analogous to the use of *food recipes* describing the composition of a composite food. The main difference is that the translation proportion is always 100% (default). Take, as an example, a TDS food *FruitMix* that is composed of *apple, orange* and *pear*, then a consumed food (food-as-eaten) *apple-pie* is converted to *apple, wheat* and *butter* (in some specific proportions) and subsequently, *apple* is converted to modelled food *FruitMix* (100%). Not necessarily all foods as consumed are represented in a TDS food sample. In addition to the TDS food sample compositions, there may be additional foods that are not officially part of a TDS food, but which can be extrapolated to a TDS food sample. Through the use of *food extrapolations* (read across translations), these foods may be directly linked to a TDS food sample, e.g., by specifying that *pineapple* is translated to *FruitMix*, *pineapple* or foods containing *pineapple* will also be matched to a *FruitMix* concentration.

Because TDS samples only contain one single, average measurement, TDS occurrence data can currently only be used for only applicable for chronic exposures assessments. However, when variability information is available for the raw primary foods in the TDS food samples (e.g., from monitoring), this information may be used *to approximate the variance of TDS samples*.

For more information about Total Diet Studies, visit the TDS-Exposure website http://www.tds-exposure.eu.

TDS reduction to limit scenario analysis

The outcome of a MCRA risk assessment may be that some foods dominate the right upper tail of the exposure distribution. A scenario analysis answers the question to what extent the risk of foods with a high exposure would have been diminished by an intervention or by taking any precautions. To be able to do so, some information is needed about the concentration distributions of the raw agricultural commodities (RACs) that make up the TDS food sample. The decision to intervene or not can be based on comparison between the p95 percentile point of the concentration distribution and a concentration limit value that associated with a high risk.

- For p95 ≤ limit, most concentration values are below the value that is considered as a potential risk, so there is no urgency to take any precautions.
- When the opposite is true, i.c. p95 > limit, there may be an argument to intervene for this specific food.

In MCRA, limits and p95's are supplied by the *concentration distributions module*. In the MCRA interface, a scenario analysis can be checked and the scroll down list allows to select the foods that should be included in the scenario analysis. For the selected foods, concentration reduction factors are computed based on the p95 percentile and the limit value:

$$f_{\rm reduction} = limit/p95$$

These reduction factors (computed for the RACs) are applied to the simulated concentrations on the level of the TDS composite foods of which the conversion paths contain the RACs included in the scenario analysis.

$$f_{\mathrm{reduction}} \cdot c_{TDS}$$

Here, c_{TDS} is the concentration value of the TDS food.

Substance concentrations generation

Both *chronic* and *acute* dietary exposure assessments rely on assigning substance concentrations to consumed modelled foods. For chronic exposure assessments, this concentration should be the mean concentration of the food and substance, as obtained from the concentration models. For acute, these concentrations are obtained through random sampling, for which there are two distinct approaches: sample-based and substance-based.

Sample-based concentrations generation

In the sample-based approach, the analytical samples from the concentration data form the basis for generating concentrations. For each identified modelled food of a consumption, substance concentrations are generated by drawing a random sample from the set of all samples available for that modelled food. Assuming that for the drawn sample, substance concentration values are known for all substances of interest (i.e., all missing values and non-detects are imputed with either a zero concentration or a positive concentration at or below LOR), the substance concentrations for all substances of the assessment group are set to the substance concentrations of the drawn samples. The rationale behind this approach is that it maintains correlations between substance concentrations on the same food.

As mentioned, the sample based approach relies on all samples being analysed for all substances of interest. Often, this is not the case and for a given sample, concentration may missing for one or more substances. Also, this approach requires non-detect values to be imputed with either positive concentration or a zero concentration.

For *imputation* of missing values there are two approaches:

- 1. Imputation by zero: all missing values are assumed zero.
- 2. **Imputation using substance-based concentration models:** all missing values are imputed by drawing a concentration value from the substance-based concentration models.

For imputation of non-detects, two approaches exist:

- 1. **Replace by zero:** Non-detect values are imputated by a zero concentration value. This is an optimistic approach.
- 2. Replace by factor times LOR: Each non-detect value is replaced by a factor (e.g., 1 or 1/2) times its LOR.

Substance-based concentrations generation

In the substance-based approach, substance concentrations for a given food are drawn independently per substance from the food/substance concentration models.

Processing correction

Concentrations in the consumed food (food as eaten) may be different from concentrations in the modelled food in monitoring programs (typically raw food) due to processing, such as peeling, washing, cooking etc. Concentrations are therefore corrected according to

$$c_{jhk}' = p f_{jhk} \cdot c_{jhk} = \left(\frac{P F_k}{c f_k}\right) \cdot c_{jhl}$$

where c_{jhk} is the concentration of substance k in the food j with processing type h, and where $pf_{jhk} = \frac{PF_{jhk}}{cf_{jhk}}$ is a factor indicating the mass change for a specific combination k of modelled food and processing. The processing correction factor cf_{jhk} is used to correct for the fact that the processing factors PF_{jhk} as commonly available from the input data describe both the effects of chemical alteration and weight change. E.g. for a dried food with a consumption of 100 gram which is translated to 300 gram raw agricultural commodity, the correction factor is 3. Note that the weight change is already included when calculating the consumption amounts of the modelled foods.

Chronic exposure assessment, daily consumed foods

Model based usual intake

Foods are consumed on a daily basis.

For individual i on day j let Y_{ij} denote the 24 hour recall of a food $(i = 1...n; j = 1...n_i)$. In most cases withinindividual random variation is dependent on the individual mean and has a skewed distribution. It is therefore customary to define a one-way random effects model for Y_{ij} on some transformed scale

$$Y_{ij}^* = g(Y_{ij}) = \mu_i + b_i + w_{ij}$$

with $b_i \sim N(0, \sigma_b^2)$ and $w_{ij} \sim N(0, \sigma_w^2)$

Note that b_i represents variation between individuals and w_{ij} represents variation within individuals between days.

The mean μ_i may depend on a set of covariate $Z_i = (Z_{i1}, ..., Z_{ip})$:

$$\mu_i = \beta_0 + \beta_1^t Z_i$$

where β_0 and β_1 are regression coefficients.

The usual intake T_i for an individual *i* is defined as the mean consumption over many many days. This assumes that the untransformed intakes Y_{ij} are unbiased for true usual intake rather than the transformed intakes Y_{ij}^* . In mathematical terms T_i is the expectation of the intake for this individual where the expectation is taken over the random day effect:

$$T_i = E_w[g^{-1}(\mu_i + b_i + w_{ij})|b_i] = F(b_i)$$

Model based usual intake on the transformed scale

For the model based usual intake first note that the conditional distribution

$$(\mu_i + b_i + w_{ij}|b_i) \sim N(\mu_i + b_i, \sigma_w^2)$$

It follows that the usual intake T_i is given by

$$T_i = E_w[g^{-1}(\mu_i + b_i + w_{ij}|b_i)] = \int_{-\infty}^{\infty} g^{-1}(\mu_i + b_i + w_{ij}) \frac{1}{\sqrt{2\pi\sigma_w^2}} \exp\left(-\frac{w^2}{2\sigma_w^2}\right) \mathrm{d}w$$

Model based using a logarithmic transformation

For the logarithmic transform the usual intake T_i can be written in closed form using the formula for the mean of the lognormal distribution:

$$T_i = \exp(\mu_i + b_i + \sigma_w^2/2)$$

In this case T_i follows a log-normal distribution with mean $\mu_i + \sigma_w^2/2$ and variance σ_b^2 . This fully specifies the usual intake distribution, e.g. the mean and variance of the usual intake are given by

$$\begin{split} \mu_{iT} &= E[T_i] = \exp(\mu_i + \sigma_w^2/2 + \sigma_b^2/2) \\ \sigma_{iT}^2 &= Var[T_i] = [\exp(\sigma_b^2) - 1] \exp(2\mu_i + \sigma_w^2 + \sigma_b^2) \end{split}$$

Model based using a power transformation

For the *power transformation* the integral can be approximated by means of N-point Gauss-Hermite integration. This results in the following usual intake

$$T_i \approx \frac{1}{\sqrt{\pi}} \sum_{j=1}^N w_j (\mu_i + b_i + \sqrt{2} \sigma_w x_j)^p$$

with p the inverse of the power transformation. A similar approximation can be used for the Box-Cox transformation. There can be a small problem with Gauss-Hermite integration. The summation term $(\mu_i + b_i + \sqrt{2}\sigma_w x_j)^p$ can not be calculated when the factor between round brackets is negative and the power p is not an integer. This can happen when $(\mu_i + b_i)$ is small relative to the between day standard error σ_w . In that case the corresponding term is set to zero. This is not a flaw in the numerical method but in the statistical model since the model allows negative intakes on the transformed scale which cannot be transformed back to the natural scale. The mean and variance of T_i can be approximated again by using Gauss-Hermite integration:

$$\begin{split} \mu_{iT} &= E[T_i] = \frac{1}{\sqrt{\pi}} \sum_{k=1}^N w_k \frac{1}{\sqrt{\pi}} \sum_{j=1}^N w_j (\mu_i + \sqrt{2}\sigma_w x_j + \sqrt{2}\sigma_b x_k) \\ \sigma_{iT} &= Var[T_i] = \frac{1}{\sqrt{\pi}} \sum_{k=1}^N w_k \left[\frac{1}{\sqrt{\pi}} \sum_{j=1}^N w_j (\mu_i + \sqrt{2}\sigma_w x_j + \sqrt{2}\sigma_b x_k) \right]^2 - \mu_T^2 \left[\frac{1}{\sqrt{\pi}} \sum_{k=1}^N w_k \left[\frac{1}{\sqrt{\pi}} \sum_{j=1}^N w_j (\mu_i + \sqrt{2}\sigma_w x_j + \sqrt{2}\sigma_b x_k) \right]^2 \right] \right] \\ &= 0$$

An alternative method for obtaining model based usual intakes for the power transformation employs a Taylor series expansion for the power, see e.g. Kipnis (2009) [[Kipnis et al., 2009]]. This is however less accurate than Gauss-Hermite integration. For the power transformation simulation is required to derive the usual intake distribution: simulate a random effect b_i for many individuals and then approximate T_i for these individuals. The T_i values then form a sample form the usual intake distribution.

Model assisted usual intake on the transformed scale

The model assisted approach employs a prediction for the usual intakes of every individual in the study. This requires a prediction of the individual random effect b_i for every individual.

In the one-way random effects model the Best Linear Unbiased Prediction for $(\mu_i + b_i)$ is given by

$$\textit{BLUP}_i = \mu_i + (\bar{Y}^*_i - \mu_i) \left(\frac{\sigma_b^2}{\sigma_b^2 + \sigma_w^2/n_i} \right)$$

in which \bar{Y}_i^* is the mean of the transformed intakes for individual *i*. BLUPs have optimal properties for some purposes, but not for the purpose of representing the variation σ_b^2 between individuals. This can be seen by noting that

$$Var(\bar{Y}_i^*) = \sigma_b^2 + \sigma_w^2/n_i$$

and thus

$$\textit{Var}(\textit{BLUP}_i) = \left(\frac{\sigma_b^4}{\sigma_b^2 + \sigma_w^2/n_i}\right)$$

which is smaller than the between individual variance σ_b^2 . As an alternative a modified BLUP can be defined by means of

$$\textit{modifiedBLUP}_i = \mu_i + (\bar{Y}_i^* - \mu_i) \sqrt{\left(\frac{\sigma_b^2}{\sigma_b^2 + \sigma_w^2/n_i}\right)}$$

which has the correct variance σ_b^2 and also the correct mean μ_i . However these optimal properties disappear when modified BLUPs are directly backtransformed to the original scale.

Model assisted using a logarithmic transformation

For the logarithmic transformation the usual intake T_i follows a log-normal distribution with mean $\mu_i + \sigma_w^2/2$ and variance σ_b^2 . If we can construct a BLUP like stochastic variable with the same mean and variance, then this variable be an unbiased predictor with the correct variance. It is easy to see that the following variable has the same distribution as T_i

$$\textit{modelassistedBLUP}_i = \mu_i + \frac{\sigma_w^2}{2} + (\bar{Y}_i^* - \mu_i) \sqrt{\left(\frac{\sigma_b^2}{\sigma_b^2 + \sigma_w^2/n_i}\right)}$$

So the model assisted individual intake $\exp(modelassistedBLUP_i)$ has the same distribution as the usual intake and is thus the best predictor for usual intake.

Kipnis et al. (2009) [[Kipnis et al., 2009]] employs the conditional distribution of b_i given the observations Y_{i1}, \dots, Y_{in_i} to obtain a prediction. First note that

$$(b_i|Y_{i1},\cdots,Y_{in_i})=(b_i|Y_{i1}^*,\cdots,Y_{in_i}^*)=(b_i|\bar{Y}_i^*)$$

Since all distributions in the one-way random effects model are normal it follows that:

$$(b_i, Y_i^*) \sim BivariateNormal(0, \mu_i, \sigma_b^2, \sigma_b^2 + \sigma_w^2/n_i, \sigma_b^2)$$

where the last parameter represents the covariance between b_i and \bar{Y}_i^* . It follows that the conditional distribution

$$(b_i | \bar{Y}_i^*) \sim N(\mu_c, \sigma_c^2)$$

with

$$\mu_c = \frac{\sigma_b^2}{\sigma_b^2 + \sigma_w^2/n_i} (\bar{Y}_i^* - \mu_i)$$

and

$$\sigma_c^2 = \frac{\sigma_b^2 \sigma_w^2/n_i}{\sigma_b^2 + \sigma_w^2/n_i}$$

A prediction for the usual intake $T_i = F(b_i)$ is then obtained by the expectation

$$E[F(b_i)|\bar{Y}_i^*] = \int F(b)\phi(b;\mu_c,\sigma_c^2)\mathrm{d}b$$

For the logarithmic transform $F(b_i) = \exp(\mu_i + b_i + \sigma_w^2/2)$ and the expectation reduces to

$$E[F(b_i)|\bar{Y}_i^*] = \exp(\mu_i + \mu_c + \sigma_c^2/2 + \sigma_w^2/2)$$

which is a function of \bar{Y}_i^* through μ_c . To obtain the mean and variance of the prediction note that

$$\mu_i + \mu_c + \sigma_c^2/2 + \sigma_w^2/2 \sim N\left(\mu_i + \frac{\sigma_b^2 \sigma_w^2/n_i}{2(\sigma_b^2 + \sigma_w^2/n_i)} + \frac{\sigma_w^2}{2}, \frac{\sigma_b^4}{\sigma_b^2 + \sigma_w^2/n_i}\right)$$

It follows that the expectation of the prediction equals

$$\begin{split} E[E[F(b_i)|\bar{Y}_i^*]] &= \exp\left(\mu_i + \frac{\sigma_b^2 \sigma_w^2/n_i}{2(\sigma_b^2 + \sigma_w^2/n_i)} + \frac{\sigma_w^2}{2} + \frac{\sigma_b^4}{2(\sigma_b^2 + \sigma_w^2/n_i)}\right) \\ &= \exp\left(\mu_i + \frac{\sigma_b^2}{2} + \frac{\sigma_w^2}{2}\right) \end{split}$$

which equals the mean of the usual intake. However the variance of the prediction equals

$$Var[E[F(b_i|\bar{Y}_i^*]] = \left[\exp\left(\frac{\sigma_b^4}{(\sigma_b^2 + \sigma_w^2/n_i)}\right) - 1\right]\exp(2\mu_i + \sigma_b^2 + \sigma_w^2)$$

Which is less than the variance of the usual intake. The approach of Kipnis et al. (2009) [[Kipnis et al., 2009]] will therefor result in too much shrinkage of the model assisted usual intake.

Model assisted using a power transformation

For the *power transformation* a model assisted BLUP with optimal properties, as derived above, cannot be constructed. The approach of Kipnis et al. (2009) [[Kipnis et al., 2009]] can however be used to obtain a prediction in the following way. First approximate $T_i = F(b_i)$ by *Gauss-Hermite integration*:

$$F(b_i) = T_i \approx \frac{1}{\sqrt{\pi}} \sum_{j=1}^N w_i (\mu_i + b_i + \sqrt{2}\sigma_w x_i)^p$$

Secondly again use Gauss-Hermite to approximate the expectation of the conditional distribution giving the prediction P_i .

$$P_i = E[F(b_i)|\bar{Y}_i^*] = \int F(b_i)\phi(b;\mu_c,\sigma_c^2) \mathrm{d}b \approx \frac{1}{\pi} \sum_{k=1}^N w_k \sum_{j=1}^N w_j(\mu_i + \mu_c + \sqrt{2}\sigma_w x_j + \sqrt{2}\sigma_c x_k)^p$$

which is a function of \bar{Y}_i^* through μ_c . It is likely that the thus obtained predictions P_i have a variance that is too small. If we would know the mean μ_{iP} and variance σ_{iP}^2 of the predictions, the predictions could be linearly rescaled to have the correct mean μ_{iT} and variance $\frac{2}{iT}$. The mean and variance of the prediction can be calculated using *Gauss-Hermite integration*.

$$\mu_{iP} = \frac{1}{\sqrt{\pi}} \sum_{l=1}^{N} w_l \frac{1}{\pi} \sum_{k=1}^{N} w_k \sum_{j=1}^{N} w_j (\mu_i + \sqrt{2} \frac{\sigma_b^2}{\sigma_b^2 + \sigma_w^2/n_i} x_l + \sqrt{2} \sigma_w x_j + \sqrt{2} \sigma_c x_k)^p$$

$$\sigma_{iP}^2 = \frac{1}{\sqrt{\pi}} \sum_{l=1}^{N} w_l \left[\frac{1}{\pi} \sum_{k=1}^{N} w_k \sum_{j=1}^{N} w_j (\mu_i + \sqrt{2} \frac{\sigma_b^2}{\sigma_b^2 + \sigma_w^2/n_i} x_l + \sqrt{2} \sigma_w x_j + \sqrt{2} \sigma_c x_k)^p \right]^2 - \mu_{iP}^2$$

The proposed prediction then equals

$$P_i^* = \mu_{iT} + \frac{\sigma_{iT}}{\sigma_{iP}}(P_i - \mu_{iP})$$

Chronic exposure assessment, episodically consumed foods

For episodically consumed foods we need to take the probability of consumption into account. Define p_i as the probability that individual *i* consumes the food on any given day. The usual intake for this individual is then given by the product of p_i and T_i which is now defined as the usual amount on consumption days. Since individuals will vary in their probability pi, besides modelling the amounts as for daily consumed foods, it is also necessary to model the frequency of consumption. A three stage analysis of 24-hour recall data is the necessary:

- 1. A model for the frequency of consumption
- 2. A model for the intakes on consumption days

3. Integration of both models in order to obtain a usual intake distribution.

Step 2 uses the analysis outlined in the previous section for the positive intakes only. For step 1 two popular models which describe between-individual variation for the probability of consumption are the beta-binomial model and the logistic-normal model.

Beta-Binomial model for frequencies (BBN)

Let n_i be the total number of recall days for individual *i* and X_i the number of days with a positive intake. The distribution of X_i , with p_i the probability of consumption for individual *i*, is given by

$$X_i = Binomial(n_i, p_i)$$

In this model the probability p_i varies among individuals according to the Beta distribution:

$$f(p)=B^{-1}(\alpha,\beta)p^{\alpha-1}(1-p)^{\beta-1}$$

with

$$B(\alpha,\beta) = \frac{\Gamma(\alpha)\Gamma(\beta)}{\Gamma(\alpha+\beta)}$$

Combining the binomial and the Beta distribution results in the betabinomial distribution:

$$P(X_i=x) = \left(\begin{array}{c} n_i \\ r \end{array} \right) \frac{B(\alpha+x,n_i+\beta-x)}{B(\alpha,\beta)}$$

The mean and variance of the betabinomial distribution are given by

$$E[X_i] = n_i \frac{\alpha}{\alpha + \beta}$$

and

$$Var[X_i] = n_i \frac{\alpha\beta(\alpha+\beta+n_i)}{(\alpha+\beta)^2(\alpha+\beta+1)}$$

Using the reparameterization $\pi = \alpha/(\alpha + \beta)$ and $\phi = 1/(\alpha + \beta + 1)$, it follows that

$$E[X_i] = n_i \pi$$

and

$$Var[X_i] = n_i \pi (1 - \pi) [1 + (n_i - 1)\phi]$$

This reparameterization enables to model the probability π_i of consumption for individual *i* directly as a logistic regression:

$$logit(\pi_i) = \gamma_0 + \gamma_1^t Z_i$$

Note that the dispersion parameter ϕ : is assumed to be equal for all individuals. The betabinomial logistic regression model can be fitted by means of maximum likelihood.

Model based frequencies for usual intake

For the model based usual intake distribution the estimated parameters π_i and ϕ are backtransformed using $\alpha_i = \pi_i \phi/(1-\phi)$ and $\beta_i = (1-\pi_i)\phi/(1-\phi)$. These can then be used to draw from the Beta distribution.

Model assisted frequencies for usual intake

For the model assisted usual intake distribution a prediction of the consumption probability is required for every individual. Simple predictions are

- 1. the observed frequencies for every individual or
- 2. the fitted probability for every individual. When there are no covariables the fitted probability is the same for every individual.
- 3. Alternatively one can use the approach outlined in Kipnis et al (2009) employing the conditional expectation of the probability given the observed frequency:

$$\begin{split} E(p_i|X_i = x) &= \int_p pf(p|X_i = x)\mathrm{d}p \\ &= \int_p p \frac{f(X_i = x|p)f(p)}{\int f(X_i = x|p)f(p)\mathrm{d}p}\mathrm{d}p \\ &= \frac{1}{P(x_i = x)} \int_p p\left(\begin{array}{c} n_i \\ r \end{array}\right) p^x (1-p)^{n_i-x} B^{-1}(\alpha_i,\beta_i) p^{\alpha_i-1}(1-p)^{\beta_i-1}\mathrm{d}p \\ &= \frac{B^{-1}(\alpha_i,\beta_i)}{P(x_i = x)} \left(\begin{array}{c} n_i \\ r \end{array}\right) \int_p p^{\alpha_i+x}(1-p)^{n_i+\beta_i-x-1}\mathrm{d}p \\ &= \frac{B(\alpha_i + x + 1, n_i + \beta_i - x)}{B(\alpha_i + x, n_i + \beta_i - x)} \\ &= \frac{\alpha_i + x}{\alpha_i + \beta_i - x} \end{split}$$

For individual with zero intakes on all recall days a prediction for the random individual amount effect b_i is not available. There seem to be two option for predicting the usual intake for such individuals:

- Set the individual intake to zero
- Simulate a model based prediction for the amount and combine this with the conditional expected probability given above to obtain an individual usual intake.

Logistic-Normal model for frequencies (LNN0)

In this model the distribution of X_i is again binomial:

$$X_i = Binomial(n_i, p_i)$$

The probability p_i is now given by a logistic regression with a random effect in the linear predictor which represents the between-individual variation in the probability p_i

 $\textit{logit}(p_i) = \lambda_i + v_i \text{ with } v_i \sim N(0, \sigma_v^2) \text{ and the regression equation } \lambda_i = \gamma_0 + \gamma_1^t Z_i$

The marginal probability π_i is obtained by integrating over the random effect v_i , i.e. using Gauss-Hermite integration

$$\pi_i = \int H(\lambda_i + v) f(v) dv \approx \frac{1}{\sqrt{\pi}} \sum_{j=1}^N w_j H(\lambda_i + \sqrt{2}\sigma_v x_j)$$

in which H() is the inverse of the logit transformation. Note that this is different from $logit^{-1}(\lambda_i)$ which is the median probability. The model can be fitted by maximum likelihood using Gauss-Hermite integration. An (approximate) maximum likelihood procedure is implemented in routine glmer of the lme4 package in R. For a new vector of covariates Z_i^* the linear predictor λ_i^* can be calculated along with its standard error $Se(\lambda_i^*)$. The marginal predicted probability π_i^* can be calculated by means of Gauss-Hermite integration and the standard error of the predicted probability can be calculated by means of the usual Taylor series expansion:

$$\mathit{Se}(\pi_i^*) \approx \frac{\mathit{Se}(\lambda_i^*)}{\sqrt{\pi}} \sum_{j=1}^N w_j \frac{d}{d\lambda_i^*} H(\lambda_i^* + \sqrt{2}\sigma_v x_j)$$

$$=\frac{\mathit{Se}(\lambda_i^*)}{\sqrt{\pi}}\sum_{j=1}^N w_j H(\lambda_i^*+\sqrt{2}\sigma_v x_j)[1-H(\lambda_i^*+\sqrt{2}\sigma_v x_j)]$$

Model based frequencies for usual intake

For the model based usual intake distribution the estimated parameters λ_i and σ_v^2 can be used to generate individual probabilities.

Model assisted frequencies for usual intake

=

For the model assisted usual intake distribution simple predictors are (a) the observed frequencies and (b) the marginal probability π_i . The conditional expectation (c) is given by

$$\begin{split} E(p_i|X_i = x) &= \int_v H(\lambda_i + v)f(v|X_i = x)\mathrm{d}v \\ &= \int_v H(\lambda_i + v)\frac{f(X_i = x_i|v)f(v)}{\int f(X_i = x_i|v)f(v)\mathrm{d}v}\mathrm{d}v \\ &\frac{\int_v H(\lambda_i + v)[H(\lambda_i + v)]^{x_i}[1 - H(\lambda_i + v)]^{n_i - x_i}f(v)\mathrm{d}v}{\int_v [H(\lambda_i + v)]^{x_i}[1 - H(\lambda_i + v)]^{n_i - x_i}f(v)\mathrm{d}v} \end{split}$$

and both nominator and denominator can be approximated by means of the *Gauss-Hermite integration*. For individual with zero intakes on all recall days see above for the two options.

Logistic-Normal model for frequencies correlated with amounts (LNN)

This model is extends the LNN0 model with a correlation between the individual random effect b_i for amounts and the individual random effect v_i for frequencies. This model is also known as the NCI model and is introduced by Tooze et al (2006) [[Tooze et al., 2006]] with further mathematical details in Kipnis et al (2009) [[Kipnis et al., 2009]]. The model can be written as

$$\begin{split} logit(P(Y_{ij}>0)) &= \lambda_i + v_i \\ g(Y_{ij}) &= \mu_i + b_i + w_{ij} \\ \text{and} \ (v_i, b_i) \sim \textit{BivariateNormal}(0, 0, \sigma_v^2, \sigma_b^2, \rho) \text{ and } w_{ij} \sim N(0, \sigma_w^2) \end{split}$$

The model can be fitted by maximum likelihood employing two-dimensional Gauss-Hermite integration.

Model based usual intake

Model based usual intake requires generation of the pair (v_i, b_i) for many hypothetical individual. The usual intake U_i for such a hypothetical individual is then given by

$$\begin{split} U_i &= H(\lambda_i + \nu_i)T_i \\ &= H(\lambda_i + \nu_i)E_w[g^{-1}(\mu_i + b_i + w_{ij})|b_i] \\ &= H(\lambda_i + \nu_i)F(b_i) \end{split}$$

The second term can be calculated using the method outlined for daily intakes.

Model assisted usual intake

This requires simultaneous prediction of the random effect for frequency and for amount as outlined in Kipnis et al (2009) [[Kipnis et al., 2009]]. We have for individual *i* in the study $(U_i|Y_{i1}, \dots, Y_{in_i}) = (U_i|Y_{i1}^*, \dots, Y_{in_i}^*) = (U_i|x_i, \bar{Y}_i^*)$ where x_i is the number of positive intakes and \bar{Y}_i^* is the mean of the transformed **positive** intakes. It follows that the required conditional expectation P_i equals

$$\begin{split} P_{i} &= E[U_{i}|x_{i},Y_{i}^{*}] \\ &= E_{v_{i},b_{i}}[H(\lambda_{i}+v_{i})F(b_{i})|x_{i},\bar{Y}_{i}^{*}] \\ &= \frac{\int \int H(\lambda_{i}+v_{i})F(b_{i})f(x_{i},\bar{Y}_{i}^{*}|v_{i},b_{i})\phi(v_{i},b_{i})dv_{i}db_{i}}{\int \int f(x_{i},\bar{Y}_{i}^{*}|v_{i},b_{i})\phi(v_{i},b_{i})dv_{i}db_{i}} \end{split}$$

where

$$f(x_i, \bar{Y}_i^* | v_i, b_i) = [H(\lambda_i + v_i)]^{x_i} [1 - H(\lambda_i + v_i)]^{n_i - x_i} \phi(\bar{Y}_i^* - \mu_i - b_i; 0, \sigma_w^2 / x_i)$$

Both nominator and denominator can be approximated by a *two-dimensional Gauss-Hermite integration*. Note that for the log-transform $F(b_i) = T_i = \exp(\mu_i + b_i + \sigma_w^2)/2)$ can be calculated exactly; for the *power transformation* an approximation must be used. It can be expected that the predicted usual intake will not have the correct variance. This can possibly be remedied by equating the mean and variance of U_i and P_i . These are however rather involved to calculate.

For individual with zero intakes on all recall days the model assisted usual intake can be set to zero, or can be simulated as follows

- 1. Calculate the Model assisted frequency P_0 for usual intake (see LNN0)
- 2. Transform P_0 back to the logistic scale, i.e. $L_0 = logit(P_0)$. Get the conditional distribution of

$$(b|v = L_0 - \lambda_i) \sim N\left(\frac{\sigma_b}{\sigma_v}\rho(L_0 - \lambda_i), (1 - \rho^2)\sigma_b^2\right)$$

3. Simulate a draw b_0 from this conditional distribution and obtain the usual intake as $P_0 \exp(\mu_i + b_0 + \sigma_w^2)$

Note that the backtransformation from P_0 to L_0 is according to the median of the distribution rather than the mean.

Dietary exposures settings

Calculation settings

Name	Description
Dietary exposure calculation	A tier is a pre-specified set of model configurations. By selecting a
tier	model tier, MCRA automatically sets all model settings in this
	module according to this tier. Note that currently tier setting may
	need to be performed separately in sub-modules. Use the Custom
	tier when you want to manually set each model setting.
Risk type	The type of exposure considered in the assessment; acute (short
	term) or chronic (long-term).
Total diet study concentration	Specifies whether exposure is based on sampling data from total
data	diet studies.
Multiple substances analysis	Specifies whether the assessment involves multiple substances.
Express results in terms of	Specifies whether the assessment involves multiple substances and
reference substance equivalents	results should be cumulated over all substances.
(cumulative)	

Table 2.87: Calculation settings for module Dietary exposures.

continues on next page

	2.87 – continued from previous page
Name	Description
Sample based	Include co-occurrence of substances in samples in simulations. If
	checked, substance residue concentrations are sampled using the
	correlations between values on the same sample. If unchecked,
	any correlation between substances is ignored, substance residue
	concentrations are sampled ignoring the correlations between
	values on the same sample.
Consumptions on the same day	if checked, in procedure of EFSA Guidance 2012, section 4.1.1,
come from the same sample	all consumptions of a raw commodity of an individual on the
	same day are assumed to come from the same sample. If
	unchecked, all consumptions of a raw commodity of an individual
	on the same day are assumed to come from different samples.
Maximise co-occurrence of	Within each pattern of substance presence. If checked, substance
high values in simulated	residue concentrations are sorted within co-occurrence patterns of
samples	substances on the same samples. After sorting, high residue values
1	occur more frequently on the same sample. This choice is
	conservative. If unchecked, substance residue concentrations are
	sampled at random, ignoring any co-occurrence patterns of
	substances on the same samples. This choice is less conservative.
Apply processing factors	Specified in table ProcessingFactor. If checked, processing factors
	are applied. Concentrations in the consumed food may be
	different from concentrations in the modelled food in monitoring
	programs (typically raw food) due to processing, such as peeling,
	washing, cooking etc. If unchecked, no processing information is
	used. This is in most (though not all) cases a worst-case
	assumption
Use distribution	Probabilistic specifications of processing factors will be used
Ignore processing factors less	This setting will suppress the use of processing factors lower than
than 1	1 (it is used in the EFSA 2012 Pessimistic tier).
Use unit variability	Controls whether to use unit variability.
Unit variability model	Describes variation between single units when concentration data
5	are from composite samples.
Estimates nature	Simulated unit concentrations can be higher or lower than
	composite value (realistic) or only equal or higher (conservative).
Unit variability parameter	Use Coefficient of variation or Variability factor, specified in
× 1	VariabilityFactor table.
Mean of LogNormal simulated	Unbiased: correct unit simulations for difference between median
values (biasing)	and mean.
Default variability factor for	Default variability factor 1 (unit weight <= 25 g, small crops). Still
unit weight <= 25g	requires specification of unit weight (FoodProperties table) and, in
6 6	case of beta model, also the Number of units in a composite
	sample (UnitVariability table).
Default variability factor for	Default variability factor 5 (unit weight > 25 g, medium/large
unit weight $> 25g$	crops). Still requires specification of unit weight (FoodProperties
6 6	table) and, in case of beta model, also the Number of units in a
	composite sample (UnitVariability table).
Model type	The parametric model for between-and within-individual
· ·) [-	variation, and possibly covariates.
Model-then-add	Specifies whether to create separate exposure models for specific
	groups of foods-as-measured (model-then-add).
Covariate modelling	Specifies whether to model exposures as a function of covariates
covariate modelning	at individual level.
Amount model covariate model	Specifies whether, and how to model exposures amounts as
Amount model covariate model	function of covariates.
	runction of covariates.

Table 2.87 - continued from previous page

continues on next page

Name	Description
Frequency model covariates	Specifies whether, and how to model exposure frequency as
model	function of covariates.
Use occurrence patterns for	When selected, this simulated samples will be based on
generating simulated samples	occurrence patterns.
Details level dietary exposures	Level of detail for summarizing dietary exposure/intakes.
Iterate survey	Instead of (re-)sampling the individual days, loop over the entire
	survey (= 1 iteration). The number of iterations for a survey is
	calculated as round (number of Monte Carlo iterations /(number
	of individuals * surveys days)).
Monte Carlo iterations	The number of iterations for Monte Carlo simulations, e.g.
	100.000 (maximum is 100.000).
Impute exposure distributions	Impute exposure distributions for substances with missing
	concentrations.
Allow conversion using food	Step 3c: try to find read across codes. If unchecked, read across
extrapolations	table is ignored, default is 'Use read across info'. E.g. for
	pineapple no measurements are found but by specifying that
	pineapple is converted to FruitMix (with a default proportion of
	100%), the TDS sample concentration value of FruitMix will be
	used for pineapple (as-eaten or as ingredient). If successful,
	restart at step 1.
Non-detects replacement	How to replace non-detects (when not co-modelled, as in
	censored models).
Default concentration model	The concentration model type that will be used as default for all
	food/substance combinations. If this model type cannot be fitted,
	e.g., due to a lack of data, a simpler model will be chosen
	automatically as a fall-back.

Table 2	2.87 -	continued f	rom	previous page
---------	--------	-------------	-----	---------------

Output settings

Name	Description
Include drill-down on 9	Specifies whether drilldown on 9 individuals is to be included in
individuals around specified	the output.
percentile.	the output.
Summarize simulated data	Specifies whether a summary of the simulated consumptions and
Summarize simulated data	concentrations should be included in the output.
Store simulated individual day	Store the simulated individual day exposures. If unchecked, no
exposures	additional output will be generated. If checked, the output will
	contain an additional section with the simulated individual day
<u>Cl</u> ('1 C	exposures.
Show percentiles for	Give specific percentiles of exposure distribution (%), e.g. 50 90
	95 97.5 99 (space separated).
Percentage for drilldown	Gives detailed output for nine individuals near this percentile of
	the exposure distribution.
Percentage for upper tail	Gives detailed output for this upper percentage of the exposure
	distribution.
Show % of population below	This setting is used for reporting the percentages of individuals
level(s)	(chronic) or individual days (acute) exceeding certain exposure
	levels. These exposure levels can be generated automatically
	based on the observed exposures (Automatic, default) or specified
	explicitly (Manual).
Exposure levels	Specify exposure levels for which to give the percentage of
	exposure below these levels, e.g. 1 10 50 100 200 500.
Number of levels of covariable	Specify the number of levels, e.g. 20. The range of the covariable
to predict exposure	is divided by the number of levels: range = $(max - min)/levels$.
	For these covariable levels exposures are predicted.
Predict exposure at extra	Specify specific prediction levels in addition to the automatically
covariable levels	generated prediction levels (space separated).
Lower percentage for	The default value of 25% may be overruled.
variability (%)	
Upper percentage for	The default value of 75% may be overruled.
variability (%)	
Report consumptions and	Specifies whether body weights should be ignored and
exposures per individual	consumptions and exposures should be expressed per individual.
instead of per kg body weight	Otherwise, the consumptions and exposures are per kg body
1 0 7 0	weight.
Cutoff for ratio total exposure/	For selection of individual(day) exposures specify cutoff for ratio
maximum (MCR plot)	total exposure/ maximum.
Show tail percentiles (MCR	Give specific percentiles of exposure distribution (%), e.g. 97.5
plot) for	99 (space separated).
Set minimum percentage	Set minimum percentage contribution per substance to the tail
contribution per substance to	exposure.
the tail exposure (MCR plot)	- Provence.
the un exposure (merc plot)	

Table 2.88: Output settings for module Dietary exposures.

Uncertainty settings

Table 2.89: Uncertainty settings for module Dietary exposures.

Name	Description
Resample imputation exposure	Specifies whether to resample the imputated exposure
distributions	distributions.

Dietary exposures tiers

Overview

	2.90: Tier over		•	-	
Name	EFSA	EFSA	EFSA	EC 2018	EC 2018
	2012 Op-	2012	2012	Tier 1	Tier 2
	timistic	Pes-	Pes-		
		simistic -	simistic -		
		Acute	Chronic		
Total diet study	false		false	false	false
concentration data					
Sample based	true	true	true	true	true
Consumptions on the	false	true	true	false	false
same day come from					
the same sample					
Apply processing	true	true	true	true	true
factors					
Use distribution	false	false	false	false	false
Ignore processing	false	true	true	false	false
factors less than 1					
Use unit variability	false	true		true	true
Unit variability model	NoUnit-	BetaDis-		BetaDis-	BetaDis-
	Variability	tribution		tribution	tribution
Model type	OIM		OIM	OIM	OIM
Model-then-add	false		false	false	false
Covariate modelling	false	false	false	false	false
Iterate survey	false	false	false	false	false
Report consumptions	false	false	false	false	false
and exposures per					
individual instead of					
per kg body weight					
Default concentration	Empirical	NonDe-	NonDe-	Empirical	Empirical
model		tect-	tect-		
		SpikeLog-	SpikeLog-		
		Normal	Normal		
Include MRL fallback	false	true	true	false	false
model					
Non-detects	Replace-	Replace-	Replace-	Replace-	Replace-
replacement	ByZero	ByLOR	ByLOR	ByLOR	ByLOR
Impute missing values	false	true	true	true	true
from available values					
(if unchecked, missing					
values are imputed					
with 0)					

Table 2.90: Tier overview for module Dietary exposures.

continues on next page

Ta	able 2.90 – c	ontinued fro	m previous p	age	
Name	EFSA	EFSA	EFSA	EC 2018	EC 2018
	2012 Op-	2012	2012	Tier 1	Tier 2
	timistic	Pes-	Pes-		
		simistic -	simistic -		
		Acute	Chronic		
Correlate imputed	false	true	true	true	false
values with sample	Tuise	liue	liue	true	laise
potency					
Use occurrence	false			true	true
frequencies for	Taise			uue	liue
-					
imputation	false		false	false	false
Parametric uncertainty	Taise	true	Chronic	Taise	Taise
Risk type		Acute	Chrome	Destruct	Destruction
Estimates nature		Realistic		Realistic	Realistic
Unit variability		Variabili-		Variabili-	Variabili-
parameter		tyFactor		tyFactor	tyFactor
Restrict LOR		false	false	false	false
imputation to					
authorised uses					
Factor f (f x LOR)		1	1	0.5	0.5
MRL Factor (f x		1	1		
MRL)					
Apply occurrence				false	true
pattern percentages					
Substance conversion				UseMost-	DrawRan-
method				Toxic	dom
Retain all allocated				true	true
substances after active					
substance allocation					
Account for substance				false	true
authorisations in				iuise	liue
substance conversions					
Use extrapolation rules				tmio	tmio
Threshold for				true	true
				10	10
extrapolation					
Restrict extrapolations				true	true
to equal MRLs					
Restrict extrapolations				true	true
to authorised uses					
Impute water				true	true
concentrations					
Water concentration				0.1	0.05
value (µg/kg)					
Restrict water				true	true
imputation to the five					
most toxic substances					
Restrict water				false	false
imputation to					
authorised uses					
Scale up use frequency					true
to 100%					
Restrict use percentage					true
up-scaling to					liue
authorised uses					
autionseu uses					<u> </u>

Table	2.90 -	continued	from	previous	page

EFSA 2012 Optimistic

Use the optimistic model settings according to the EFSA Guidance 2012. Concentration values are sampled using a sample-based empirical distribution. Available processing factors are applied. No unit variability model should be applied.

Name	Setting	From	In
		input tier	module
Total diet study concentration data	false		
Sample based	true		
Consumptions on the same day come	false		
from the same sample			
Apply processing factors	true		
Use distribution	false		
Ignore processing factors less than 1	false		
Use unit variability	false		
Unit variability model	NoUnit-		
2	Variability		
Model type	OIM		
Model-then-add	false		
Covariate modelling	false		
Iterate survey	false		
Report consumptions and exposures	false		
per individual instead of per kg body			
weight			
Default concentration model	Empirical	EFSA	Concen-
		2012	tration
		Optimistic	models
Include MRL fallback model	false	EFSA	Concen-
		2012	tration
		Optimistic	models
Non-detects replacement	Replace-	EFSA	Concen-
	ByZero	2012	tration
		Optimistic	models
Sample based	true	EFSA	Concen-
		2012	tration
		Optimistic	models
Impute missing values from available	false	EFSA	Concen-
values (if unchecked, missing values		2012	tration
are imputed with 0)		Optimistic	models
Correlate imputed values with sample	false	EFSA	Concen-
potency		2012	tration
		Optimistic	models
Use occurrence frequencies for	false	EFSA	Concen-
imputation		2012	tration
		Optimistic	models
Parametric uncertainty	false	EFSA	Concen-
		2012	tration
		Optimistic	models

Table 2.91:	Tier definition	for EFSA	2012 Optimistic.
-------------	-----------------	----------	------------------

EFSA 2012 Pessimistic - Acute

Acute probabilistic exposure assessment using the pessimistic model settings according to the EFSA Guidance 2012. Only processing factors > 1 are applied. For unit variability, the Beta distribution is applied.

Name	Setting	From input tier	In module
Risk type	Acute		
Sample based	true		
Consumptions on the same day come	true		
from the same sample			
Apply processing factors	true		
Use distribution	false		
Ignore processing factors less than 1	true		
Use unit variability	true		
Unit variability model	BetaDis-		
	tribution		
Estimates nature	Realistic		
Unit variability parameter	Variabili-		
	tyFactor		
Covariate modelling	false		
Iterate survey	false		
Report consumptions and exposures per individual instead of per kg body weight	false		
Default concentration model	NonDe-	EFSA	Concen-
	tect-	2012	tration
	SpikeLog-	Pessimistic	models
	Normal	- Acute	
Include MRL fallback model	true	EFSA	Concen-
		2012	tration
		Pessimistic	models
		- Acute	
Restrict LOR imputation to	false	EFSA	Concen-
authorised uses		2012	tration
		Pessimistic	models
		- Acute	
Non-detects replacement	Replace-	EFSA	Concen-
	ByLOR	2012	tration
		Pessimistic	models
		- Acute	
Factor f (f x LOR)	1	EFSA	Concen-
	1	2012	tration
		Pessimistic	models
		- Acute	moucis
MRL Factor (f x MRL)	1	EFSA	Concen-
	1	2012	tration
		2012 Pessimistic	models
		- Acute	mouco
Sample based	true	- Acute EFSA	Concen-
Sample Dased	true	2012	tration
		2012 Pessimistic	tration models
			models
Impute missing line farmer 111	tmac	- Acute	Concern
Impute missing values from available	true	EFSA 2012	Concen-
values (if unchecked, missing values		2012 Deceminatio	tration
are imputed with 0)		Pessimistic	models
		- Acute	
Correlate imputed values with sample	true	EFSA	Concen-
potency		2012	tration
		Pessimistic	models
		- Acute	
Parametric uncertainty	true	EFSA	Concen-
		2012	tratichapter 2. Modul
		Pessimistic	models
		- Acute	

Table 2.92:	Tier definition	for EFSA	2012 Pessimistic	- Acute.
-------------	-----------------	----------	------------------	----------

EFSA 2012 Pessimistic - Chronic

Chronic probabilistic exposure assessment using the pessimistic model settings according to the EFSA Guidance 2012. Only processing factors > 1 are applied.

Name	Setting	From input tier	In module
Risk type	Chronic		
Total diet study concentration data	false		
Sample based	true		
Consumptions on the same day come	true		
from the same sample			
Apply processing factors	true		
Use distribution	false		
Ignore processing factors less than 1	true		
Model type	OIM		
Model-then-add	false		
Covariate modelling	false		
Iterate survey	false		
Report consumptions and exposures per individual instead of per kg body weight	false		
Default concentration model	NonDe-	EFSA	Concen-
	tect-	2012	tration
	SpikeLog-	Pessimistic	models
	Normal	- Chronic	
Include MRL fallback model	true	EFSA	Concen-
		2012	tration
		Pessimistic	models
		- Chronic	
Restrict LOR imputation to	false	EFSA	Concen-
authorised uses		2012	tration
		Pessimistic	models
		- Chronic	
Non-detects replacement	Replace-	EFSA	Concen-
	ByLOR	2012	tration
		Pessimistic	models
		- Chronic	~
Factor f (f x LOR)	1	EFSA	Concen-
		2012	tration
		Pessimistic	models
	1	- Chronic	
MRL Factor (f x MRL)	1	EFSA	Concen-
		2012 Description	tration
		Pessimistic	models
Sample based	trac	- Chronic	Concer
Sample based	true	EFSA 2012	Concen- tration
		2012 Pessimistic	tration models
		- Chronic	models
Impute missing values from available	true	- Chronic EFSA	Concen-
mipute missing values nom available	true	2012	tration
values (if unchecked missing values		2012	
values (if unchecked, missing values are imputed with 0)		Passimistic	modele
values (if unchecked, missing values are imputed with 0)		Pessimistic	models
are imputed with 0)	true	- Chronic	
are imputed with 0) Correlate imputed values with sample	true	- Chronic EFSA	Concen-
are imputed with 0)	true	- Chronic EFSA 2012	Concen- tration
are imputed with 0) Correlate imputed values with sample	true	- Chronic EFSA 2012 Pessimistic	Concen-
are imputed with 0) Correlate imputed values with sample potency		- Chronic EFSA 2012 Pessimistic - Chronic	Concen- tration models
are imputed with 0) Correlate imputed values with sample	false	- Chronic EFSA 2012 Pessimistic - Chronic EFSA	Concen- tration models Concen-
are imputed with 0) Correlate imputed values with sample potency		- Chronic EFSA 2012 Pessimistic - Chronic	Concen- tration models

Table 2.93: Tier definition for EFSA 2012 Pessimistic - Chronic.

EC 2018 Tier 1

Name	Setting	From input tier	In module
Total diet study concentration data	false		
Sample based	true		
Consumptions on the same day come	false		
from the same sample	luise		
Apply processing factors	true		
Use distribution	false		
Ignore processing factors less than 1	false		
Use unit variability	true		
Unit variability model	BetaDis-		
	tribution		
Estimates nature	Realistic		
Unit variability parameter	Variabili-		
Sint variability parameter	tyFactor		
Model type	OIM		
Model-then-add	false		
Covariate modelling	false		
Iterate survey	false		
Report consumptions and exposures	false		
per individual instead of per kg body weight	Taise		
Default concentration model	Empirical	EC 2018	Concen-
		Tier 1	tration
			models
Include MRL fallback model	false	EC 2018	Concen-
		Tier 1	tration
			models
Restrict LOR imputation to	false	EC 2018	Concen-
authorised uses		Tier 1	tration
			models
Non-detects replacement	Replace-	EC 2018	Concen-
	ByLOR	Tier 1	tration
			models
Factor f (f x LOR)	0.5	EC 2018	Concen-
		Tier 1	tration
			models
Sample based	true	EC 2018	Concen-
		Tier 1	tration
			models
Impute missing values from available	true	EC 2018	Concen-
values (if unchecked, missing values		Tier 1	tration
are imputed with 0)			models
Correlate imputed values with sample	true	EC 2018	Concen-
potency		Tier 1	tration
			models
Use occurrence frequencies for	true	EC 2018	Concen-
imputation		Tier 1	tration
			models
Parametric uncertainty	false	EC 2018	Concen-
		Tier 1	tration
			models
		continues of	n next page

Table 2.94: Tier definition for EC 2018 Tier 1.

Name	Setting	From	In
		input tier	module
Apply occurrence pattern percentages	false	EC 2018	Occur-
		Tier 1	rence
			patterns
Substance conversion method	UseMost-	EC 2018	Concen-
	Toxic	Tier 1	trations
Retain all allocated substances after	true	EC 2018	Concen-
active substance allocation		Tier 1	trations
Account for substance authorisations	false	EC 2018	Concen-
in substance conversions		Tier 1	trations
Use extrapolation rules	true	EC 2018	Concen-
		Tier 1	trations
Threshold for extrapolation	10	EC 2018	Concen-
		Tier 1	trations
Restrict extrapolations to equal MRLs	true	EC 2018	Concen-
		Tier 1	trations
Restrict extrapolations to authorised	true	EC 2018	Concen-
uses		Tier 1	trations
Impute water concentrations	true	EC 2018	Concen-
		Tier 1	trations
Water concentration value (µg/kg)	0.1	EC 2018	Concen-
		Tier 1	trations
Restrict water imputation to the five	true	EC 2018	Concen-
most toxic substances		Tier 1	trations
Restrict water imputation to	false	EC 2018	Concen-
authorised uses		Tier 1	trations

Table	2.94 -	continued	from	previous page
-------	--------	-----------	------	---------------

EC 2018 Tier 2

Name	Setting	From	In
	_	input tier	module
Total diet study concentration data	false		
Sample based	true		
Consumptions on the same day come	false		
from the same sample			
Apply processing factors	true		
Use distribution	false		
Ignore processing factors less than 1	false		
Use unit variability	true		
Unit variability model	BetaDis-		
	tribution		
Estimates nature	Realistic		
Unit variability parameter	Variabili-		
	tyFactor		
Model type	OIM		
Model-then-add	false		
Covariate modelling	false		
Iterate survey	false		
Report consumptions and exposures	false		
per individual instead of per kg body			
weight			
		continues o	n next page

Table 2.95: Tier definition for EC 2018 Tier 2.

Table 2.95 – continued from previous page					
Name	Setting	From input tier	In module		
Default concentration model	Empirical	EC 2018	Concen-		
		Tier 2	tration		
			models		
Include MRL fallback model	false	EC 2018	Concen-		
		Tier 2	tration		
			models		
Restrict LOR imputation to	false	EC 2018	Concen-		
authorised uses		Tier 2	tration		
			models		
Non-detects replacement	Replace-	EC 2018	Concen-		
	ByLOR	Tier 2	tration		
			models		
Factor f (f x LOR)	0.5	EC 2018	Concen-		
		Tier 2	tration		
			models		
Sample based	true	EC 2018	Concen-		
		Tier 2	tration		
			models		
Impute missing values from available	true	EC 2018	Concen-		
values (if unchecked, missing values		Tier 2	tration		
are imputed with 0)			models		
Correlate imputed values with sample	false	EC 2018	Concen-		
potency		Tier 2	tration		
		77.0010	models		
Use occurrence frequencies for	true	EC 2018	Concen-		
imputation		Tier 2	tration		
Description	C. 1	EC 2019	models		
Parametric uncertainty	false	<i>EC 2018</i>	Concen-		
		Tier 2	tration models		
Apply accurrance pattern percentages	truo	EC 2018	Occur-		
Apply occurrence pattern percentages	true	<i>EC</i> 2018 <i>Tier</i> 2	rence		
		Tier 2			
Scale up use frequency to 100%	true	EC 2018	patterns Occur-		
Scale up use frequency to 100 %	liuc	<i>LC 2010</i> <i>Tier 2</i>	rence		
		11012	patterns		
Restrict use percentage up-scaling to	true	EC 2018	Occur-		
authorised uses	liuc	Tier 2	rence		
		1107 2	patterns		
Substance conversion method	DrawRan-	EC 2018	Concen-		
	dom	Tier 2	trations		
Retain all allocated substances after	true	EC 2018	Concen-		
active substance allocation		Tier 2	trations		
Account for substance authorisations	true	EC 2018	Concen-		
in substance conversions		Tier 2	trations		
Use extrapolation rules	true	EC 2018	Concen-		
		Tier 2	trations		
Threshold for extrapolation	10	EC 2018	Concen-		
1		Tier 2	trations		
Restrict extrapolations to equal MRLs	true	EC 2018	Concen-		
		Tier 2	trations		
Restrict extrapolations to authorised	true	EC 2018	Concen-		
uses		Tier 2	trations		
<u> </u>	1		n next page		

Table 2.95 – continued from previous page		Table	2.95 -	continued	from	previous	page
---	--	-------	--------	-----------	------	----------	------

continues on next page

Name	Setting	From	In
		input tier	module
Impute water concentrations	true	EC 2018	Concen-
		Tier 2	trations
Water concentration value (µg/kg)	0.05	EC 2018	Concen-
		Tier 2	trations
Restrict water imputation to the five	true	EC 2018	Concen-
most toxic substances		Tier 2	trations
Restrict water imputation to	false	EC 2018	Concen-
authorised uses		Tier 2	trations

Table 2.95 - continued from previous page

EFSA 2012 Pessimistic

Note: This tier is deprecated and has been replaced by separate acute/chronic tiers.

Probabilistic exposure assessment using the pessimistic model settings according to the EFSA Guidance 2012. Only processing factors > 1 are applied. For unit variability, the Beta distribution is applied.

Setting	From	In
	input tier	module
true		
true		
true		
false		
true		
true		
BetaDis-		
Realistic		
Variabili-		
tyFactor		
OIM		
false		
NonDe-	EFSA	Concen-
tect-	2012	tration
SpikeLog-	Pessimistic	models
Normal		
true	EFSA	Concen-
	2012	tration
	Pessimistic	models
false	EFSA	Concen-
	2012	tration
	Pessimistic	models
Replace-	EFSA	Concen-
-	2012	tration
	Pessimistic	models
1	EFSA	Concen-
	2012	tration
	Pessimistic	models
1	EFSA	Concen-
	2012	tration
	Pessimistic	models
true	EFSA	Concen-
	2012	tration
	Pessimistic	models
true	EFSA	Concen-
	2012	tration
1	Pessimistic	models
	1 costinustic	
true	EFSA	Concen-
true		Concen- tration
true	EFSA	
true	EFSA 2012	tration
	EFSA 2012 Pessimistic	tration models
	SettingfalsetruetruefalsetruefalsetrueBetaDis- tributionRealisticVariabili- tyFactorOIMfalsefalsefalsefalsefalsefalsefalsefalsefalsefalsefalsefalsefalsefalsefalsefalsefalseltruefalse <t< td=""><td>input tierfalsetruetruetruefalsetruefalsetruefalsetruefalsetruefalsetruefalse</td></t<>	input tierfalsetruetruetruefalsetruefalsetruefalsetruefalsetruefalsetruefalse

Table 2.96: Tier definition	for EFSA 2012 Pessimistic.
-----------------------------	----------------------------

Calculation of dietary exposures

Dietary exposures are calculated from consumptions per modelled food and concentration models. Optionally, also processing factors and unit variability models are applied.

• Dietary exposures calculation

Inputs used: Consumptions by modelled food Concentration models Processing factors Unit variability factors High exposure food-substance combinations Active substances Occurrence patterns Relative potency factors Food conversions Concentration distributions

Settings used

• Calculation Settings

2.4.3 High exposure food substance combinations

Identification of food-as-eaten/modelled food/substance combinations that have the highest expected contribution to exposure based on a simple screening model.

This module has as primary entities: Foods Substances Effects

Output of this module is used by: Dietary exposures

High exposure food substance combinations calculation

A full Monte Carlo analysis can be unwieldy for large cumulative assessment groups (CAGs) and/or large number of foods or concentration data. An algorithmic approach was developed to handle large CAGs. Two unique features of MCRA are:

- contributions to the exposure results can be seen both in terms of foods as eaten (e.g. white bread) and foods as measured (e.g. wheat), and
- a drill-down can be made into the exact foods and substances contributing for simulated individuals or individual-days in the upper tail.

The number of combinations of simulation, substance, modelled food and food as eaten can be very large. To avoid memory problems with very large datasets, an additional optional modelling step, named *screening*, was added to MCRA. *Screening* should be used if the data dimensions are too large for a direct analysis. Screening identifies risk drivers. A full analysis based on screened risk drivers will still retain all food/substance combinations in the exposure calculation, and will therefore produce exactly the same cumulative exposure distribution, and allow to see contributions of all substances and all modelled foods. Details with respect to foods as eaten are however restricted to the risk drivers selected in the screening step. For more details see *screening calculation for large Cumulative Assessment Groups*.

The two-step approach consists of:

• Step 1: Data screening and selection of risk drivers Run a simple analysis for each potential source/substance combination (SCC). Here source means the combination of food as eaten and modelled food, for example apple in apple pie. The screening is based on this combination, and not just foods as measured, to avoid problems with potentially multi-modal consumption distributions as much as possible (see van der Voet et al. 2014). SCCs are also referred to as risk driver components. The screening step in MCRA implements a simple model that is applied to each SCC. The model calculates a percentile of interest in a distribution, consisting of a spike of zeroes (non-consumptions), and a mixture of two lognormal distributions for the exposure related to non-detects and positive concentrations, respectively. SCCs (risk driver components) can be combined to measured source/substance combinations (MSCCs, risk drivers). For example APPLE/apple juice/captan and APPLE/apple pie/captan combine to APPLE/captan. MCRA has an interface which identifies the Top-*N* SCCs (based on a chosen exposure percentile, e.g. p95) with an option to select *N* based on cumulative importance according to some criterion. Remark: Screening is performed before concentration modelling. Therefore there is no correction for processing at the screening stage. Note, originally SCC stands for Source Compound Combination, MSCC for Measured Source Compound Combination.

• Step 2: Full MC analysis Perform the standard MC to all combinations of substances and foods, but restrict the stored information regarding foods as eaten to the SCCs selected in step 1.

The screening method requires to specify:

- Which percentile to consider for each single source/substance combination (SCC, potential risk driver component) (default p95)
- Which percentage of all exposures (according to the screening model) should be covered by the selected set of SCCs (default 95%)
- How to impute non-detect concentrations (c < LOR) in the screening step. The choice of a factor 0 (default) represents optimistic imputation, the choice of a factor 1 represents a pessimistic imputation. It may be noted that a factor 1 (pessimistic imputation) may select many SCCs (risk driver components) with relatively high LORs and high RPFs, but with only nondetect measurements. Choosing a lower fraction, e.g. 0.25 can be useful if a more realistic method is sought.

Based on limited experience with the EFSA test data, useful settings of these three screening parameters were found to be (95, 95, 0) in preparation for an EFSA optimistic run, and (50, 95, 0.25) in preparation for an EFSA pessimistic run. See also screening calculation *acute exposure* and *chronic exposure*.

Screening calculation for large Cumulative Assessment Groups

Statistical model for the screening step (acute exposure)

The screening step implements a simple model that is applied to each SCC. Assume independent *NonDetectSpike-LogNormal* (NDS-LN) models for both the consumptions of modelled food in source S and the concentrations of substance C in source S. A non-detect consumption is assumed to be a zero consumption. A non-detect concentration will be imputed by a user-specified fraction f of the Limit of Reporting. Then the model for consumption has 3 parameters and the model for concentration has four parameters, as specified in Table 2.97. Note that the parameters of the consumption distribution are estimated from the consumption data using sampling weights if these have been provided in the consumption data set.

	U I	/
parameter	consumptions	concentrations
probability of a positive	π_x	π_c
mean positives (ln scale)	μ_x	μ_c
standard deviation positives (ln scale)	σ_x	σ_c
value to use for NonDetects (ln scale)		$f \cdot L_c$

 Table 2.97: Parameters for screening models (per source/substance)

Exposure is consumption times concentration, so on logarithmic scale they can be added:

$$e = x + c.$$

The assessment will focus on a chosen percentile of exposure, e.g. p95. The relevant fraction will be denoted by p, for example p = 0.95 for the 95th percentile. The two NDS-LN models combine to three possibilities, depending on whether there is consumption and if so, whether the concentration is non-detect or positive. In the screening model the two possibilities that lead to potential exposure are modelled with a mixture of two lognormal distribution. For the non-detect case the positive exposure distribution equals the positive consumption distribution modified by the multiplication of a user-chosen factor times an estimate of the average worst-case limit value for concentration (LOR):

$$\pi_1 = \pi_x (1 - \pi_c); \mu_1 = \mu_x + f \cdot L_c; \sigma_1 = \sigma_x$$

where L_c is the logarithm of the LOR, or, if there are multiple analytical methods with different LOR, a weighted average of these different LORs.

For the detect case the positive exposure distribution is easily combined from the positive consumption distribution and the positive concentration distribution:

$$\pi_{2} = \pi_{x}\pi_{c}; \mu_{2} = \mu_{x} + \mu_{c}; \sigma_{12} = \sqrt{\sigma_{x}^{2} + \sigma_{c}^{2}}$$

p can be corrected for the non-consumptions to the appropriate fraction needed in the mixture of the two positive distributions:

$$p' = \frac{p-(1-\pi_x)}{\pi_x}$$

If $p' \leq 0$ then all positive exposures are beyond the requested fraction, and the estimated exposure is just 0.

If p' > 0 then the relevant log exposure e_p satisfies

$$(1 - \pi_c) \cdot \Phi\left(\frac{e_p - \mu_1}{\sigma_1}\right) + \pi_c \cdot \Phi\left(\frac{e_p - \mu_{12}}{\sigma_2}\right) = p'$$

where $\Phi(\cdot)$ represents the cumulative standard normal distribution function. The value of e_p can easily be found in a bisection search within the interval

$$[\mu_{min} - 4\sigma_{max}, \mu_{max} + max(0, z_{p'}\sigma_{max})].$$

The final exposure percentile estimate then is $\exp(e_n)$.

Denote by $e_{(p,max)}$ the highest estimate (for the SCC denoted by $SSC_{highest}$). Then evaluate for each SCC the probability to exceed $e_{(p,max)}$.

$$P_i = Pr(e > e_{p,max}) = \pi_x \cdot \left[(1 - \pi_c) * \Phi\left(\frac{e_{p,max} - \mu_1}{\sigma_1}\right) + \pi_c \cdot \Phi\left(\frac{e_{p,max} - \mu_2}{\sigma_1}\right) \right]$$

 P_i is a tentative measure for the 'probability of a high exposure'. For $SSC_{highest}$ $P_i = 1 - p$, for all other SCCs it will be lower. The sum of all these probabilities is not a meaningful probability in itself. However, this sum is used to scale the individual P_i values to measures of relative importance for the SCCs

$$Imp_i = P_i / \sum P_i$$

Rank all SCCs according to Imp_i and calculate cumulative importance. The relative importance of the two mixture components at e_p can be estimated as

$$w_{1,2} = \frac{\pi_{1,2} \cdot \phi\left(\frac{e_p - \mu_{1,2}}{\sigma_{1,2}}\right) / \sigma_{1,2}}{\pi_1 \cdot \phi\left(\frac{e_p - \mu_1}{\sigma_1}\right) / \sigma_1 + \pi_2 \cdot \phi\left(\frac{e_p - \mu_2}{\sigma_2}\right) / \sigma_2}$$

where $\phi(.)$ represent the standard normal probability density function. The user interface should allow to select the top-*N* SCCs from the list, based on a chosen percentage (e.g. 95%) of cumulative importance included. The full analysis will calculate exactly the same exposure distribution as a full analysis without screening. However, less information is retained in the output. This concerns tables with information on foods-as-eaten, which is only shown for the selected risk driver components (SCCs). Risk drivers are groupings of SCCs (risk driver components) at the level of measured-source-substance combinations (MSCCs). Note that output for an MSSC (e.g. APPLE/captan) only covers the selected SCCs (e.g. APPLE from apple juice/captan and APPLE from apple pie/captan), but not unselected SCCs (e.g. APPLE from fruit yoghurt/captan).

Statistical model for the screening step (chronic exposure)

In chronic exposure assessments, the mean concentration of chemicals is calculated first, and combined with the consumption distribution. For this reason a chronic calculation uses less memory, and therefore larger datasets can be handled.

The model described under *acute exposure* can be simplified for a chronic screening. The concentration distribution is only used to estimate a mean exposure, incorporating any effect from the imputation of non-detects. The exposure distribution is therefore only a scaled version of the consumption distribution.

$$\pi_2 = \pi_x \pi_c; \mu_2 = \mu_x + \mu_c; \sigma_2 = \sigma_x$$

The parameters of the consumption distribution (π_x, μ_x, σ_x) are calculated from the observed individual means *(OIM)*, i.e. the mean daily consumptions over the survey days of each person in the data (allowing for sampling weights). The percentiles are calculated as $e_p = \mu_2 + z_p$ where z is a percentile of the standard normal distribution. The exceedances of the maximum percentile are calculated as

$$P_i = Pr(e > e_{p,max}) = \pi_x \cdot \Phi\left(\frac{e_{p,max} - \mu_2}{\sigma_2}\right)$$

High exposure food substance combinations settings

Calculation settings

Table 2.98: Calculation settings for module High exposure food-substance	
combinations.	

Name	Description
Percentage defining the	Percentage defining the exposure percentile of interest per
exposure percentile of interest	food-as-eaten/food-as-measured/substance combination.
per food-as-eaten/food-as-	
measured/substance	
combination	
Include risk drivers to include	The selection criterion for the risk drivers. The cumulative
minimally a percentage	contribution percentage of the selected risk drivers will be this
	percentage.
Non-detect replacement: factor	A constant between 0 and 1. A value 0 can be used for an
x LOR	optimistic screening (LOR not used). Note that a factor 0.5
	(pessimistic) may result in many and often high contributions
	from food-substance combinations with only non-detects.

Calculation of high exposure food-substance combinations

Screening results are computed for each combination of source (being a specific combination of food-aseaten/modelled food) and substance by combining simple approximations of the consumption and the concentration distribution.

• High exposure food-substance combinations calculation

Inputs used: Consumptions by modelled food Concentration models Active substances Relative potency factors

Settings used

Calculation Settings

2.4.4 Exposures

Exposures are amounts of substances, typically expressed per mass unit and per day, to which individuals in a population are exposed at a chosen target level. This target level may be external exposure (dietary exposure, expressed per unit body weight, or per person) or internal exposure (expressed per unit organ weight). Internal exposures may be aggregated from dietary and non-dietary exposures using either absorption factors or kinetic models to translate the external exposures to internal exposures. Exposures can be short-term/acute exposures and then contain exposures for individual-days, or they can be long-term/chronic exposures, in which case they represent the average exposure per day over an unspecified longer time period.

This module has as primary entities: Populations Foods Substances

Output of this module is used by: Exposure mixtures Human monitoring analysis Risks

Exposures calculation

Calculation of exposures comprises two main steps:

- 1. Linking dietary and non-dietary individual/individual-day exposures.
- 2. Computing the (aggregated) internal exposures at the specified target compartment.

Both steps are optional in this module. If none is selected, exposures are external *dietary exposures*, i.e the target level is external/dietary. However, when multiple routes of exposure are considered, then the target level should be an internal compartment (organ). In the latter case, *absorption factors* or *kinetic model* are needed to aggregate the exposures from multiple routes into exposure at the target compartment. It is also possible to only provide dietary exposures and compute internal exposures at some target compartment.

In cumulative exposure calculations two simple approaches are used to identify and select mixtures contributing to the exposure of a target population:

- 1. qualitative approach: counting of co-exposure. To which combinations of substances are individuals exposed?
- 2. quantitative approach: *maximum cumulative ratio (MCR)*. To what degree are mixtures more important than single substances?

A quantitative approach is available in the *exposures mixtures module*.

Combining dietary and non-dietary exposures

If *dietary* and *non-dietary exposures* are available for the same individuals or individual-days, the non-dietary exposures can be matched to specific individuals of the food survey from which the dietary exposures are generated. More commonly, dietary and non-dietary exposures are available from separate surveys, in which case they can be randomly combined. If both dietary and non-dietary information is available for a known population of individuals, the user may select the *matching option* such that specific dietary and non-dietary exposures that do not correspond to individuals from the food survey population. If matching is enabled, any non-dietary exposures that do not correspond to individuals from the food survey will be ignored (see *Example 2*), unless an individual is specified with id = General. In that case, the dietary individual should meet the criteria of the non-dietary survey, specified by the survey properties, to be assigned. If the non-dietary data relates instead to a population in which individuals have no corresponding records in the food survey (unmatched case), the user may choose to randomly assign the non-dietary exposures to the individuals from the food survey.

When multiple non-dietary surveys are available, the options with or without correlation are important (not relevant when matching is switched on). When correlation is chosen, the exposure contributions of non-dietary individuals with identical ids in different surveys are combined and allocated to a randomly selected dietary individual. When the correlation is not chosen, the non-dietary exposures of randomly selected individuals from different surveys are combined and allocated to a dietary individual.

The user may also define demographic criteria for the assignment (for each source of non-dietary exposure) to indicate that those exposures are relevant only to a defined sub-population. Only those individuals in the food survey who meet the criteria of the non-dietary survey will be assigned non-dietary exposures from that source e.g. only males aged 18 to 65 (see *Example 1*). The simplest assessment consists of a single (deterministic) non-dietary exposure estimate which is assigned to all individuals in the food survey (*idIndividual = General*). This case, and more complex possibilities are illustrated below using hypothetical examples.

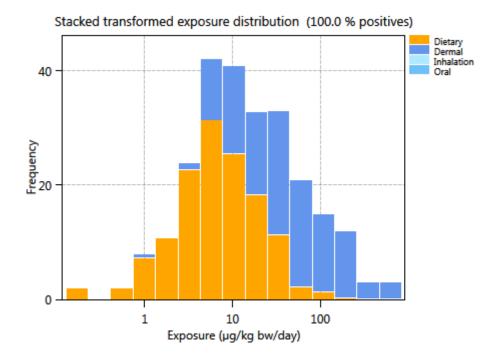


Figure 2.23: Aggregate exposure distributions.

Example 1

Deterministic cumulative (multi-substance) non-dietary exposure input, adult male sub-population. Unmatched case.

idIndividual	idNonDietarySurvey	idSubstance	Dermal	Oral	Inhalation
General	1	011003001	10	5	17
General	1	011003002	34	20	18
General	1	011003002	56	43	19

Table 2.99: NonDietaryExposures

idNonDietary- Survey	Description	Location	Date	NonDietary- IntakeUnit
1	BROWSE, acute,	York	09/10/2012	$\mu g/day$
	cumulative,			
	operators			

Table 2.101: NonDietarySurveyProperties

idNonDietary- Survey	Property- IndividualName	Individual- PropertyText- Value	Individual- Property- DoubleMin- Value	Individual- Property- DoubleMax- Value
1	Age		18	65
1	Gender	Male		

In this example, there are exposure values for multiple substances in Table 2.99 and the user has provided Table 2.101 which specifies that the non-dietary exposures given in survey number 1 relate to males aged 18 to 65.

When this assessment is performed, only those individuals whose property values fit the criteria in Table 2.101 will receive the non-dietary exposures in survey 1. The use of *idIndividual* = *General* indicates that this is the

default exposure. All individuals in the dietary survey meeting the criteria receive all exposures given in the 3 rows, corresponding to 3 substances. The following should be noted:

- There should only ever be one General entry in the dietary exposures table per substance, survey combination.
- The property names and the values of any text properties must precisely match those given in the **Individual**-**Properties** and **IndividualPropertyValues** tables for this to work.
- The minimum and maximum values for numeric properties are both inclusive boundaries.

So in this example, all males aged 18 to 65 will receive the given exposures of all three substances and the other members of the population will receive no non-dietary exposure. Note that example 1 describes the unmatched case.

Example 2

Variability (but no uncertainty) in cumulative non-dietary exposure input (matched to dietary survey individuals).

idIndividual	idNonDietarySurvey	idSubstance	Dermal	Oral
5432	1	011003001	10	5
5432	1	011003002	33	22
5433	1	011003001	12	7
5433	1	011003002	34	23
5434	1	011003001	18	9
5434	1	011003002	35	25
5435	1	011003001	10	5
5435	1	011003002	33	21

Table 2.102: NonDietaryExposures

Table 2.103: NonDietarySurveys

idNonDietary-	Description	Loca-	Date	NonDietaryIntakeU-
Survey		tion		nit
1	BROWSE, acute, cumulative, opera-	York	09/10/2012	$\mu g/day$
	tors			

In this example, the non-dietary exposures are being specified explicitly for individuals in the dietary population. Switch 'matching' on to allow exposure variability to be specified at the individual level. For the purposes of illustration, the population is extremely small, consisting of only four individuals. The values in the *idIndividual* column of Table 2.102 match those in the **Individuals** table (dietary population).

It is not mandatory to specify exposures for every individual in the population. Those not included will simply receive a zero non-dietary exposure, unless there is also a default exposure value (*General* row(s) in Table 2.102) and the individual matches the specified demographic criteria for the survey, as specified in Table 2.101. (In this example, a default exposure would apply to all individuals not listed in Table 2.102 because Table 2.101 has not been used).

There is variability between individuals in this example, but no uncertainty in exposure. Note that these data could also be used with matching switched off. This would be the same as treating the *idIndividual* values as generic individuals, so that each pair of exposure lines would be assigned at random to individuals meeting the criteria.

Example 3

Variability (no uncertainty) in cumulative non-dietary exposure input (unmatched individuals).

idIndividual	idNonDietarySurvey	idSubstance	Dermal	Oral	Inhalation
ND1	1	011003001	10	5	17
ND1	1	011003002	33	22	45
ND2	1	011003001	12	7	18
ND2	1	011003002	34	23	47
ND3	1	011003001	18	9	19
ND3	1	011003002	35	25	49
ND4	1	011003001	10	5	17
ND4	1	011003002	33	21	45

Table 2.104: NonDietaryExposures

Table 2.105: NonDietarySurveys

idNonDietary-	Description	Loca-	Date	NonDietaryIntakeU-
Survey		tion		nit
1	BROWSE, acute, cumulative, opera-	York	09/10/2012	$\mu g/day$
	tors			

Table 2.106: NonDietarySurveyProperties

idNonDi- etarySurvey	PropertyIndi- vidualName	IndividualProper- tyTextValue	IndividualProperty- DoubleMinValue	IndividualProperty- DoubleMaxValue
1	Age		50	65
1	Gender	Female		

This example is similar to example 2, except that the values in the *idIndividual* column of Table 2.104 do not match those in the **Individuals** table. In this instance, 'matching' would not be an option, and the non-dietary exposures would be randomly assigned to individuals who meet the criteria in Table 2.106. (In fact for the same result rather than changing the values in the *idIndividual* column in Table 2.102 from the previous example may be used with matching switched off). Exposures in Table 2.104 will be recycled if the number of exposure rows is less than the number of dietary records with which to aggregate exposures.

Again, there is variability between individuals in this example, but no uncertainty in exposure.

By allowing generic *idIndividual* values in this way, correlations between different sources (within individual) can be accounted for even in the unmatched case. If the same *idIndividual* value is used in different surveys, then the corresponding exposure values will be kept together and assigned to an eligible individual as a combined exposure.

So for option matching switched of, it is relevant whether individuals are correlated or not. In the following example, two non-dietary surveys are available, per survey three individuals are specified.

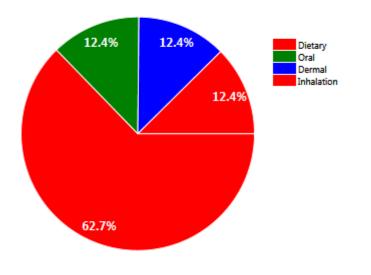
idIndividual	idNonDietarySurvey	idSubstance	Dermal	Oral	Inhalation
ND0	1	011003001	10	5	17
ND1	1	011003001	23	22	45
ND2	1	011003001	12	7	18
ND0	1	011003001	34	23	47
ND3	1	011003001	18	9	19
ND4	1	011003001	33	16	35

Table 2.107: matching switched of, with correlation or without.

• When a correlation is applied, the non-dietary exposure for individual ND0 from survey 1 and 2 are combined and allocated to a dietary individual. For individual ND1, ND2, ND3 and ND4 just a single non-dietary exposure is found and allocated to a dietary individual.

• When no correlation is applied, the exposure for individual ND0 from survey 1 is combined with one of the exposures of ND0, ND3 or ND4 from survey 2; exposure of ND1 from survey 1 is combined with one of the exposures of ND0, ND3 or ND4 from survey 2, etc.

When the intention is to sample just one exposure for a dietary individual, specify per survey different codes, e.g. ND1, ND2, ND3 for survey 1, ND4, ND5, ND6 for survey 2 and apply correlation, or specify 6 different individual codes and just one *idNonDietarySurvey*. Then, options with or without correlation are irrelevant and sampling results are identical no matter which option is chosen.



Contribution to the total exposure distribution by route

Figure 2.24: Contributions by route to aggregate exposure distributions.

See non-dietary exposure settings.

Internal exposures calculation

Computation of internal exposures (internal substance amounts and concentrations) requires a *kinetic model* to translate external doses, possibly from multiple routes, to internal doses at the target compartment/organ of interest.

Calculation of internal concentrations using absorption factors

In the simplest form, internal concentrations are derived from external exposure concentrations using multiplication factors (or, absorption factors) that can be specified by substance and by route. That is, for a given substance, the internal exposure exp_{int} is computed as

$$exp_{\text{int}} = \sum_{r \in \textit{Routes}} f_{\text{abs},r} \cdot exp_{\text{ext},r}$$

Here, *Routes* denotes the set external exposure routes, $exp_{ext,r}$ denotes the external exposure for route r and $f_{abs,r}$ denotes the absorption factor of route r. Note that this model assumes that both external and internal exposures refer to amounts or concentrations depending on the *dietary exposures* setting (External exposure: substance amount per individual, or substance amount divided by body weight; internal exposure: substance amount per organ, or substance amount divided by organ weight.) Also, both external and internal exposures are expressed per day.

Calculation of internal concentrations using kinetic models

A more detailed alternative to using absorption factors is to use one of the *advanced kinetic models* available in MCRA. In this approach, for each substance independently, the external exposures of an individual (chronic) or individual-day (acute) are presented for a number of simulated day to a PBK model of the individual. This yields a time course of the internal substance amount at the specified target compartment/organ from which a long term average substance amount (chronic) or peak substance amount (acute) can be obtained. An example of such a time course is given in Figure 2.25 for acute exposure assessments, and in Figure 2.26 for chronic exposure assessments. By dividing this substance amount by the weight of the compartment, an internal concentration is obtained. Notice that this procedure also changes the unit of the exposures from exposure per day to long term exposure.

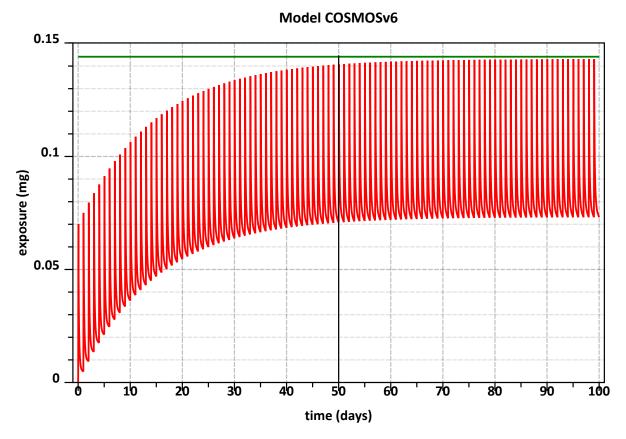


Figure 2.25: Time course of the internal substance amount when applying the same single dose on each day. The acute internal concentration is derived as the peak substance amount (the green line in the figure) divided by the compartment weight. The vertical line at 50 indicates the selected end of an assumed non-stationary period, defining a burn-in period that is to be ignored for computing the peak substance amount.

Mathematically, the calculation of the peak substance amount (d_{peak}) for deriving acute internal exposures is as follows:

$$d_{\mathrm{peak}} = \max_{i=0,\ldots,n_{\mathrm{stop}}} \left\{ d(t_{\mathrm{start}} + i\Delta t) \right\}.$$

Here, d(t) denotes the substance amount at time t, t_{start} denotes the starting time of the evaluation window (defined by the *non-stationary period*), Δt denotes the time resolution of the kinetic model (e.g., hours or minutes), and n_{stop} denotes the total number of time-points, marking the end of the evaluation window (defined by the specified number of simulation days), which is computed as

$$n_{\rm stop} = \left\lfloor \frac{t_{\rm stop} - t_{\rm start}}{\Delta t} \right\rfloor.$$

Likewise, chronic long term average substance amounts (d_{avq}) are computed as:

$$d_{\rm avg} = \frac{\sum_{i=0}^{n_{\rm stop}} d(t_{\rm start} + i\Delta t)}{n_{\rm stop}}. \label{eq:davg}$$

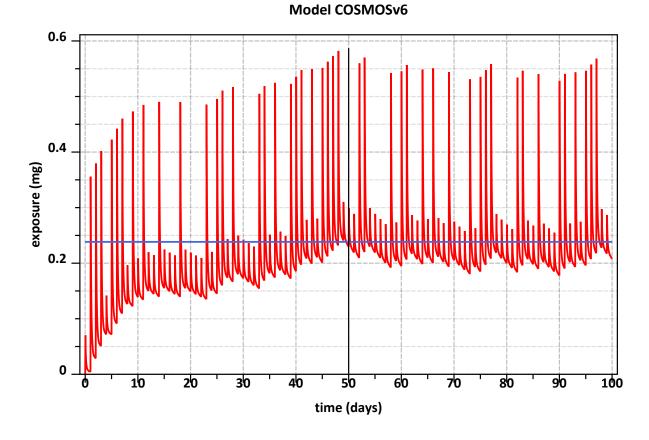


Figure 2.26: Time course of the internal substance amount when randomly applying one of the individual-day doses for a number days. The chronic internal concentration is derived as the average substance amount (the blue line in the figure), divided by the compartment weight. The vertical line at 50 indicates the selected end of an assumed non-stationary period, defining a burn-in period that is to be ignored for computing the average substance amount.

Chapter 2. Modules

Dosing patterns

In MCRA, the dietary and non-dietary exposures are computed at the level of exposures per day. However, when applying advanced PBK models, dosing patterns may be specified at a much finer resolution (e.g., hours or minutes). For this, a method is needed to translate external exposures provided per day to dosing patterns of substance amounts during the day. The simplest, yet not very realistic model is to apply, per route, the full exposure amount in one single dose at the beginning of the day. Alternatively, MCRA offers the possibility to specify, per route, the *number of exposure events per day*. If it is specified to use multiple doses per day, then the total substance amount of each day is divided into equal portions which are applied at regular time-intervals during the day.

Non-stationary period

Especially in the case of chronic exposure assessments, where a long term average exposure is computed based on the simulated time-course, it is important to realise that at time zero, the substance is commonly considered to be completely absent in the simulated system. However, this is not a realistic assumption. It is much more likely that the substance was already present in the system, and that the level is equal to the level obtained from applying the same chronic exposures to the system. For this, a specification of the *number of days skipped* (or burn-in period) is required in order to come to these initial concentration levels. This period is not used for computing the long term average or peak exposures, but just to determine initial (background) concentration levels.

Counting of co-exposure

In this qualitative approach, the number of combinations of substances to which an individual is exposed are recorded, see Table 2.108. There is no cut-off level, the only criterion is the presence of a substance in the simulated daily diet or not. For an *acute* or short term exposure assessment, a simulated individual day is the smallest entity to determine co-exposure. For a *chronic* or long term exposure assessment, co-exposures are summarized at the individual level, e.g. co-exposure is determined combining all consumption days of an individual.

Substance	day 1	day 2	day 3	 day n
Tebuconazole	Х	X		
Bitertanol	Х		Х	 Х
Triadimefon	Х			 Х

Table 2.108: Counting combinations of substances in the exposure matrix: for example, on day 1 there is coexposure to substances Tebuconazole, Bitertanol and Triadimefon

In Table 2.109, the frequency and percentage for the number of substances occurring together are shown.

Tuble 2.109. Co exposure of substances					
Number of substances	Frequency	Percentage			
0	337	3.4			
1	959	9.6			
2	1207	12.1			
3	1275	12.8			

Table 2.109: Co	-exposure of substances
-----------------	-------------------------

In Table 2.110, the mixtures containing the substance(s) including all other combinations with the specified combination of substance(s), (a maximum number of 15 records is shown).

Number of substances	Percentage	Substances
1	5.88	Tebuconazole
2	3.88	Imazalil (aka enilconazole), Tebuconazole
0	3.37	
3	2.20	Difenoconazole, Imazalil (aka enilconazole), Tebuconazole
1	1.78	Imazalil (aka enilconazole)
3	1.76	Imazalil (aka enilconazole), Tebuconazole, Triadimenol

Table 2.110: Mixtures containing substances

Maximum Cumulative Ratio

Price and Han [[Price et al., 2011]] propose the Maximum Cumulative Ratio (MCR) which is defined as the ratio of the cumulative exposure received by an individual on an intake day to the largest exposure received from a single substance:

$MCR = \frac{Cumulative exposure}{Maximum exposure}$

This MCR statistic is also picked up as a practical device in a recent JRC report [[Bopp et al., 2015]] to investigate cumulative exposure. If MCR is large, it is important to consider cumulative effects. If MCR is close to 1, the individual exposure will not be much different from a single-substance assessment. The MCR can therefore be interpreted as the degree to which the risk of being exposed is underestimated by not performing a cumulative risk assessment.

The MCR statistic is implemented in MCRA for both the *acute* risk and the *chronic* risk cases. In the acute risk case the short-term (single-day) exposures are used. For the chronic case long-term individual exposures (estimated by aggregating over the available survey days of each individual) are used.

Table 2.111 shows an artificial example how the MCR is calculated in the acute risk case. First the cumulative exposure per day is calculated by cumulating the exposure of each substance multiplied by the *relative potency factors* (RPF). Then, for each day, the cumulative exposure (in equivalents of the reference substance) is divided by the maximum exposure of a single substance on that day. The last column shows the MCR values, with the substance with the highest exposure in parenthesis. The MCR has a value of 1 or close to 1 for mixtures where the exposure (e.g. day 1, substance B). When all substances have approximately equal exposure (e.g. day 3) the MCR value is equal or close to the number of substances, here 4. Day 2 represents an intermediate case. The MCR suggest that for exposure days (or persons) with MCR values close to 1, the need for a cumulative risk assessment is low.

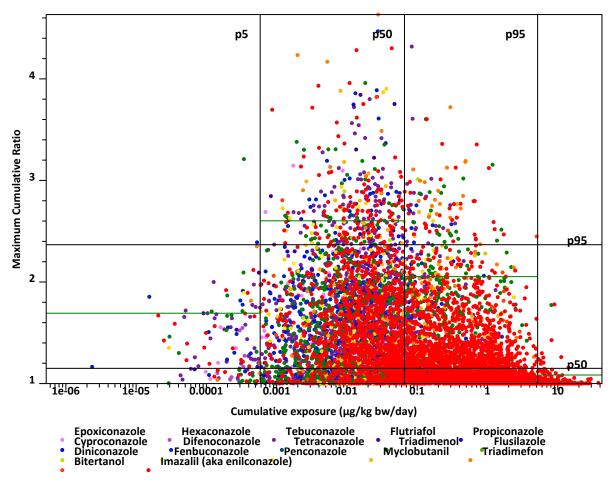
	Substance A	Substance B	Substance C	Substance D	total exposure	ratio
day 1	0.01	0.99	0	0	1	1.01 (B)
day 2	0.1	0.2	0.3	0.4	1	2.50 (D)
day 3	0.25	0.25	0.24	0.26	1	3.99 (D)

Table 2.111: Maximum Cumulative Ratios

In the example, all days have equal values for total exposure. For real data, total exposure will vary. It is obviously of interest to know if the MCR is high or low at those days (or individuals) where the total exposure is highest.

In Figure 2.27, French steatosis data (39 substances, 4079 persons) are used to calculate the chronic exposure matrix. For each individual the MCR is calculated and plotted against the total exposure. The different colours are used to identify the single substances with maximum exposure. From the original 39 substances, 10 different substances have the largest exposures. For the total exposure and MCR, the p_5 , p_{50} and p_{95} percentiles are indicated with the black line segments. The red line indicates the ratio with value 5. The dashed green lines indicate the p_{95} percentiles for the MCR value for different ranges of the total exposure.

The plot shows that MCR values with Imazalil as risk driving substance (purple) are predominantly found in the lower part of the plot for relatively high values of the total exposure. A second finding is that MCR values decline when total exposure increases. This implies that cumulative exposure for most individuals is driven by multiple substances. At

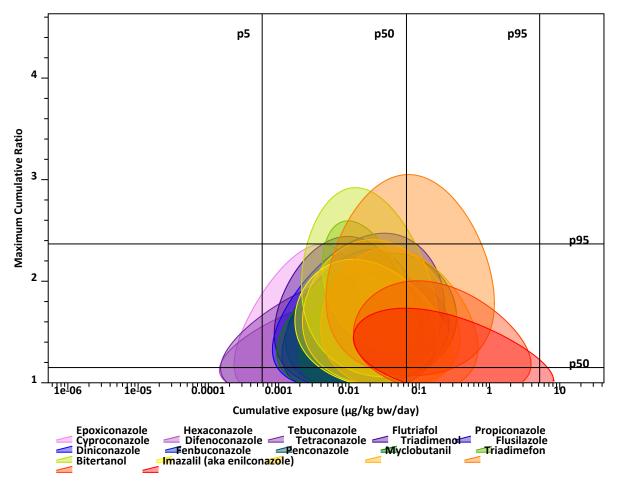


Using MCR to identify substances that drive cumulative exposures

Figure 2.27: Maximum Cumulative Ratios vs total exposure

the right site of the plot, individuals are found with high exposure. Because MCR values tend to be lower here, higher exposures are received from one predominant substance and not because many substances are above the average level. For those individuals a cumulative risk assessment has less value.

Because Figure 2.27 can be very dense, in Figure 2.28, 95% confidence regions representing bivariate lognormal distributions of MCR and total exposure are plotted. The latter figure facilitates interpretation of the first figure. Note that substances with just one or two observations cannot be plotted in this display (substances with 2 observations are represented by a line).



Bivariate distributions

Figure 2.28: Bivariate distributions MCR vs total exposure

In Figure 2.29 and Figure 2.30 scattered MCR distributions for the total and upper tail (here 37%) that drive the cumulative exposure are shown. The red line indicates the MCR threshold, 1.5. The black lines represent the regression lines MCR vs ln(Cumulative exposure) for each tail. Substances with an exposure contribution less than 15% are not displayed.

In Table 2.112 contributions to tail exposures at various percentile are shown. Column MCR = 1 shows the percentage of tail exposure due to individual(day)s with a single substance. Column $1 < MCR \le 2$ shows the percentage of tail exposure due to individual(day)s with multiple substances, but the MCR ≤ 2 . Column MCR > 2 shows the percentage of tail exposure due to individual(day)s with multiple substances with MCR > 2.

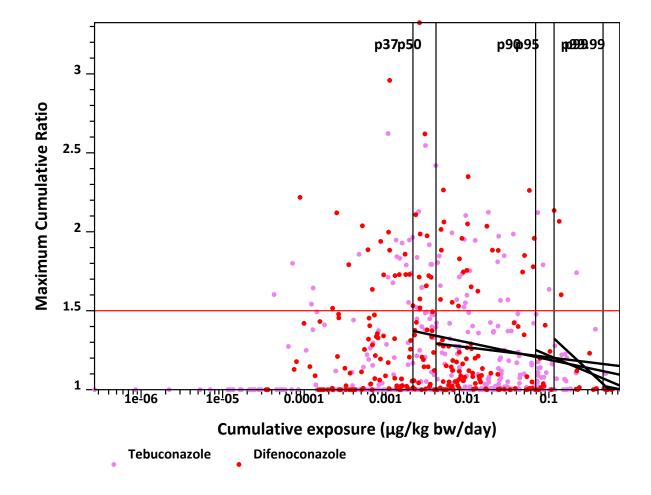


Figure 2.29: Using MCR to identify substances that drive cumulative exposures, scatter distributions (total).

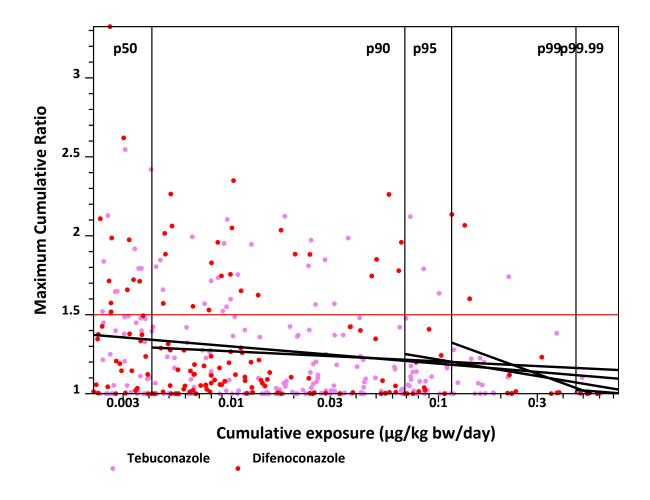


Figure 2.30: Using MCR to identify substances that drive cumulative exposures, scatter distributions (upper tail 37%).

Tail	% with MCR	Sub-	% with 1 <	Substances	% with MCR	Sub-
%	= 1	stances	MCR<=2		> 2	stances
37	20.6	Difeno,	73.7	Difeno, Tebu	5.7	Difeno,
		Tebu				Tebu
50	19.2	Difeno,	75.6	Difeno, Tebu	5.2	Difeno,
		Tebu				Tebu
90	16.3	Difeno,	78.8	Difeno, Tebu	5.0	Difeno,
		Tebu				Tebu
95	15.0	Difeno,	82.5	Difeno, Tebu	2.5	Difeno,
		Tebu				Tebu
99	25.0	Difeno	75.0	Difeno, Tebu	0.0	
				Propi		

nary
nary

For MCR settings, see *exposure mixture settings*.

Exposures settings

Calculation settings

1000 2.11	5. Calculation settings for module Exposures.		
Name	Description		
Risk type	The type of exposure considered in the assessment; acute (short		
	term) or chronic (long-term).		
Multiple substances analysis	Specifies whether the assessment involves multiple substances.		
Express results in terms of	Specifies whether the assessment involves multiple substances and		
reference substance equivalents	results should be cumulated over all substances.		
(cumulative)			
Include dietary and non-dietary	Specifies whether the assessment involves both dietary and		
routes of exposure	non-dietary (oral, inhalatory or dermal) routes of exposure.		
Target level	Select to express hazard characterisations at external or internal		
	exposure level.		
Match non-dietary to dietary	Specifies whether the individuals of one or more non-dietary		
survey individuals	surveys should be matched to individuals in the dietary survey		
	according to the individual codes (idIndividual). If unchecked,		
	nondietary exposures are randomly allocated to dietary survey		
	individuals.		
Match individuals of multiple	If checked, exposures from identical individuals in non-dietary		
non-dietary surveys	surveys are aggregated to obtain the overall nondietary exposures.		
	If unchecked, exposures from random individuals in all		
	non-dietary surveys are aggregated.		
Model-then-add	Specifies whether to create separate exposure models for specific		
	groups of foods-as-measured (model-then-add).		

Table 2.113: Calculation settings for module Exposures.

Output settings

Name	Description		
Include drill-down on 9	Specifies whether drilldown on 9 individuals is to be included in		
individuals around specified	the output.		
percentile.			
Summarize simulated data	Specifies whether a summary of the simulated consumptions and		
	concentrations should be included in the output.		
Store simulated individual day	Store the simulated individual day exposures. If unchecked, no		
exposures	additional output will be generated. If checked, the output will		
	contain an additional section with the simulated individual day		
	exposures.		
Show percentiles for	Give specific percentiles of exposure distribution (%), e.g. 50 90		
	95 97.5 99 (space separated).		
Percentage for drilldown	Gives detailed output for nine individuals near this percentile of		
	the exposure distribution.		
Percentage for upper tail	Gives detailed output for this upper percentage of the exposure		
	distribution.		
Show % of population below	This setting is used for reporting the percentages of individuals		
level(s)	(chronic) or individual days (acute) exceeding certain exposure		
	levels. These exposure levels can be generated automatically		
	based on the observed exposures (Automatic, default) or specified		
	explicitly (Manual).		
Exposure levels	Specify exposure levels for which to give the percentage of		
	exposure below these levels, e.g. 1 10 50 100 200 500.		
Number of levels of covariable	Specify the number of levels, e.g. 20. The range of the covariable		
to predict exposure	is divided by the number of levels: range = $(max - min)/levels$.		
	For these covariable levels exposures are predicted.		
Predict exposure at extra	Specify specific prediction levels in addition to the automatically		
covariable levels	generated prediction levels (space separated).		
Lower percentage for	The default value of 25% may be overruled.		
variability (%)			
Upper percentage for	The default value of 75% may be overruled.		
variability (%)			
Report consumptions and	Specifies whether body weights should be ignored and		
exposures per individual	consumptions and exposures should be expressed per individual.		
instead of per kg body weight	Otherwise, the consumptions and exposures are per kg body		
	weight.		

Table 2.114: Output settings for module Exposures.

Uncertainty settings

Table 2.115:	Uncertainty	settings for	module H	Exposures.

Name	Description
Resample kinetic model	Specifies whether kinetic model parameter values are resampled.
parameter values	

Calculation of exposures

Exposures are computed by linking dietary and (if available) non-dietary individual/individual-day exposures and computing the (aggregated) internal exposures at the specified target compartment.

• Exposures calculation

Inputs used: Dietary exposures Non-dietary exposures Active substances Relative potency factors Kinetic models

Settings used

Calculation Settings

2.4.5 Exposure mixtures

Exposure mixtures are mixtures of substances that contribute relatively much to the overall cumulative exposure (potential risk drivers).

This module has as primary entities: Foods Substances Effects

Exposure mixtures calculation

The most common model of cumulative risk assessment is to focus on substances that belong to the same common assessment groups (CAG). *Substances* in such a group belong to the same chemical family and may or may not have a similar mode of action. In assessing the risk, possible interactions between substances are often ignored and, moreover, little information is available about synergistic effects at low doses. More information is needed about the combined effects of such substances, but it is impractical to investigate all possible mixtures, and therefore instruments are needed to select the most relevant substances for further investigation. This selection should not only be based on the hazard (highest relative potencies) but also on the exposure of the population to these substances. The potential risk of being exposed to chemicals in a mixture depends on the food *consumption* patterns of *individuals* in a population. A regular diet can contain hundreds of substances, so the number of combinations of substances to which an individual in a population is exposed can be numerous. The exposures mixtures module can be used to identify the most relevant mixtures to which a population is exposed.

Exposure mixtures are identified using a quantitative approach: *sparse non-negative matrix underapproximation* (*SNMU*). What mixtures predominantly determine the underlying patterns in the exposure matrix (substance x person (day))?

Sparse nonnegative matrix underapproximation

Starting point to identify major mixtures of substances using exposure data was to use Non-negative Matrix Factorization (NMF). Non-negative Matrix Factorization was proposed by Lee & Seung [[Lee et al., 1999]] and Saul & Lee [[Saul et al., 2002]] and deals specifically with non-negative data that have excess zeros such as exposure data. Zetlaoui et al. [[Zetlaoui et al., 2011]], introduced this method in food risk assessment to define diet clusters.

The NMF method was then implemented by Béchaux et al. [[Béchaux et al., 2013]] in order to identify food consumption patterns and main mixtures of pesticides to which the French population was exposed using *TDS* exposure to 26 priority pesticides.

At the start of the Euromix project ideas had evolved: to obtain less components per mixture experiments were made using Sparse Nonnegative Matrix Factorization (SNMF) [[Hoyer, 2004]]. This method was found not to give stable solutions. Better results were obtained with Sparse Nonnegative Matrix Underapproximation (SNMU) [[Gillis et al., 2013]]. This model also fits better to the problem situation because also the residual matrix after extracting some mixtures should be nonnegative. The SNMU method has been implemented in MCRA.

Indeed, NMF may be used to identify patterns or associations between substances in exposure data. NMF can be described as a method that finds a description of the data in a lower dimension. There are standard techniques available such as principal components analysis or factor analysis that do the same, but their lower rank representation is less suited because they contain negative values which makes interpretation hard and because of the modelling with a Gaussian random vectors which do not correctly deal with the excess of 0 and non-negative data. The NMF solution

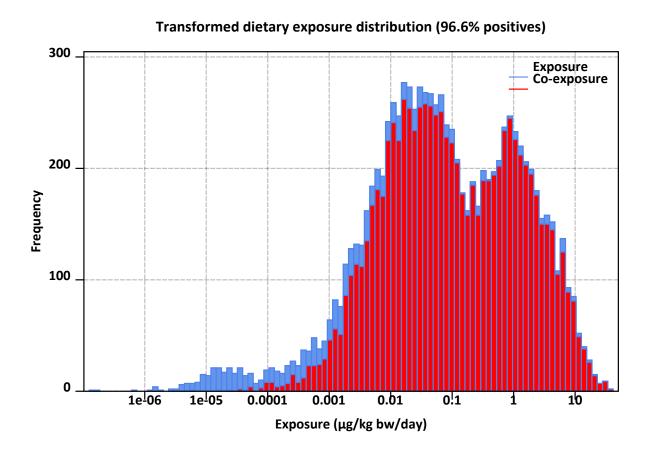


Figure 2.31: Example of co-exposure distribution (from >1 substance per individual-day, red) super-imposed on the total exposure distribution (blue).

approximates a non-negative input matrix (i.c. exposure data) by two constrained non-negative matrices in a lower dimension such that the product of the two is as close as possible to the original input matrix. In Figure 2.32, the exposure matrix M with dimensions m (number of substances) and n (number of intake days or persons) is approximated by matrix U and V with dimensions $(m \times k)$ and $(k \times n)$ respectively, where k represents the number of mixtures. Matrix U contains weights of the substances per mixture, matrix V contains the coefficients of presence of mixtures in exposure per intake day or person. M is non-negative (zero or positive) and U and V are constraint to be non-negative. The minimization criterion is: $||M-UV||^2$ such that $U \ge 0$ and $V \ge 0$.

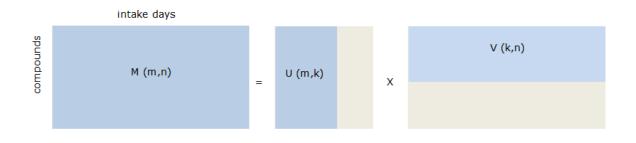


Figure 2.32: NMF approximation of exposure data

The solution found by NMF contains zeros, but mixtures still contain many components which complicates interpretability. Therefore, the Sparse Nonnegative Matrix Underapproximation (SNMU) [[Gillis et al., 2013]] which also provide sparse results was investigated. The SNMU has also some nice features well adapted to the objective of the Euromix project: the solution is independent of the initialization and the algorithm is recursive. Consequently, the SNMU method which is based on the same decomposition process as the NMF was the one implemented in MCRA.

SNMU is initialized using an optimal nonnegative rank-one approximation using the power method (https://en. wikipedia.org/wiki/Power_iteration). This initialization is based on a singular value decomposition and results in general in a unique solution that is sparse. The SNMU algorithm is called recursive because after identifying the first optimal rank-one underapproximation u_1v_1 , the next rank-one factor is recovered by subtracting u_1v_1 from M and applying the same factorization algorithm to the remainder $M - u_1v_1$. The solution u_1v_1 is called a rank-one underapproximation because an upper bound constraint is added to ensure that the remainder $M - u_1v_1$ is non-negative. For Matlab code see: https://sites.google.com/site/nicolasgillis/code.

For each mixture, a percentage of explained variance is calculated. M is the exposure matrix with m rows (substances) and n columns (individuals or individual days) S_t is sum of squared elements of M:

$$S_t = ||M||^2 = \sum_{i,j}^{m,n} e_{i,j}^2$$

Apply SNMU on *M*, then for the first mixture:

- u is $m \times 1$ vector, contains weights of the mixture.
- v is $1 \times n$ vector, contains presence of mixture in exposure per individual.

Calculate residual matrix R:

$$R = M - uv$$

Calculate S_r , residual sum of squared elements of R:

$$S_r = ||R||^2 = \sum_{i,j}^{m,n} e_{i,j}^2$$

Percentage explained variance first mixture (k = 1) is:

$$V_k = (1 - S_r)/S_t \cdot 100$$

For the second mixture:

- 1. continue with residual matrix R (replace M by R),
- 2. estimate u and v,
- 3. calculate new residual matrix R
- 4. calculate S_r , residual sum of squared elements of R

Percentage explained variance second mixture (k = 2) is:

$$V_k = (1-S_r)/S_t) \cdot 100 - \sum_{l=1}^{k-1} V_l$$

The last term is de percentage explained variance of the first mixture. Continue with the third mixture etc....

Exposure matrix

Basically, exposure is calculated as consumption x concentration. By summing the exposures from the different foods for each substance per person day separately, the exposure matrix for mixture selection is estimated:

$$y_{ijc} = \frac{\sum_{k=1}^{p} x_{ijk} c_{ijkc}}{bw_i}$$

where y_{ijc} is the exposure to substance c by individual i on day j (in microgram substance per kg body weight), x_{ijk} is the consumption by individual i on day j of food k (in g), c_{ijkc} is the concentration of substance c in food k eaten by individual i on day j (in mg/kg), and bw_i is the body weight of individual i (in kg). Finally, p is the number of foods accounted for in the model. More precisely, for an *acute* or short term risk assessment, the exposure matrix summarises the y_{ijc} with columns denoting the individual-days (ij) and rows denoting the substances (c). Each cell represents the sum of the exposures for a substance on that particular day. For a *chronic* or long term risk assessment, the exposure matrix is constructed as the sum of all exposures for a particular substance averaged over the total number of intake days of an individual (substances x persons). So each row represents a substance and a column an individual. Each cell represents the observed individual mean exposure for a substance for that individual. Note that in the exposure calculation, the concentration is the average of all residue values of a substance.

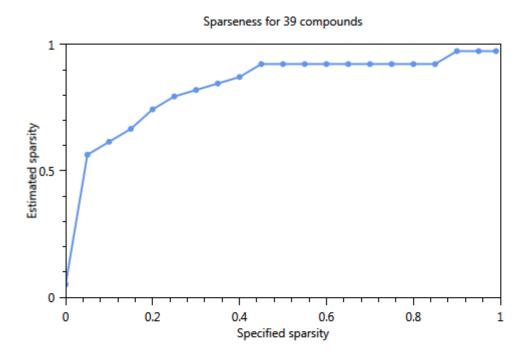
When *relative potency factors* (RPF) are available then exposures are multiplied by the RPF and thus exposures to the different substances are on the same and comparable scale (toxicological scale). In this case, the selection of the mixture is risk-based. In some cases, RPFs may not be available. In this case exposure to different substances, even in the same unit, may lead to very different effect. To give all substances an equal weight a priori and avoid scaling effect, a normalization of the data can be applied as done in Béchaux et al. [[Béchaux et al., 2013]]. In this case, all substances are assigned equal mean and variance, and the selection of the mixtures will work on patterns of correlation only. Then mixture selection is not risk-based anymore but, what could be called, co-exposure-based.

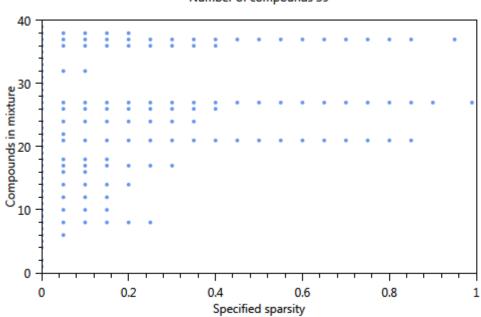
Finally, in the mixture selection module of MCRA, the SNMU expects RPF data for a risk-based selection. If not available, the user should provide alternative RPF data, e.g. values 1 for a purely exposure-based selection, or lower-tier estimates (e.g. a maximum value on RPF thought possible).

Mechanisms to influence sparsity

Two mechanisms to *influence sparsity* are available. The SNMU algorithm incorporates a sparsity parameter and by increasing the value, the final mixtures will be more sparse (sparsity close to 0: not sparse; sparsity close to 1: sparse). The other approach is by using a subset of the exposure matrix based on a cut-off value for the *MCR*. High ratios correspond to high co-exposure, so it is reasonable to focus on these (person) days and not include days where exposure is received from a single substance (ratio close to 1). To illustrate the combined use of MCR and the sparsity parameter, the French steatosis data (39 substances, 4079 persons) are used and person days with a ratio > 5 (see Figure 2.27) are selected yielding a subset of 139 records.

In Figure 2.33, the effect of using a cut-off level is immediately clear. The number of substances of the first mixture is 17 whereas in the unselected case only 4 substances were found The three plots show the influence of increasing the sparsity parameter from 0 to 1 on the number of substances in the mixture. For values close to 0, the mixture contains 17 substances. For values > 0.4 the number of substances in the mixture drops to 3.





Number of compounds 39

Number of compounds 39

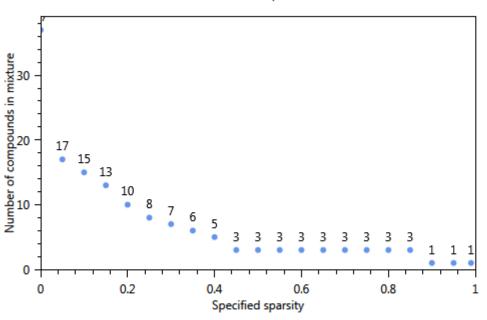


Figure 2.33: Influence of the specified sparsity parameter on the realized sparsity, n = 139

In Figure 2.34 and Figure 2.35 the sparsity parameter is set to 0.0001 (not sparse) and 0.4 (sparse), respectively. This leads to mixtures containing different number of substances.

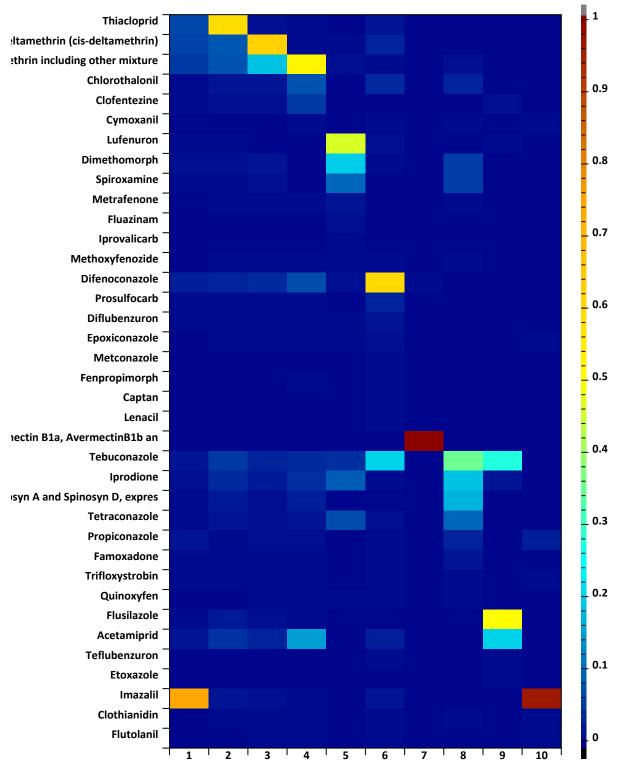
As mentioned before, one of the nice features of the SNMU algorithm is its recursive character which results in identical mixtures. In Figure 2.36 and Figure 2.37, two U matrices are visualized. In the upper plot 4 mixtures are estimated, in the lower plot the solution for 5 mixtures is shown. Because of the ordering the plots look different, but a closer inspection of the first 4 mixtures of each solution shows that they are the same. In both figures, mixture 1 contains Imazalil, Thiacloprid, Deltamethrin (cis-deltamethrin) and Deltamethrin including other mixture.

Exposure mixtures settings

Calculation settings

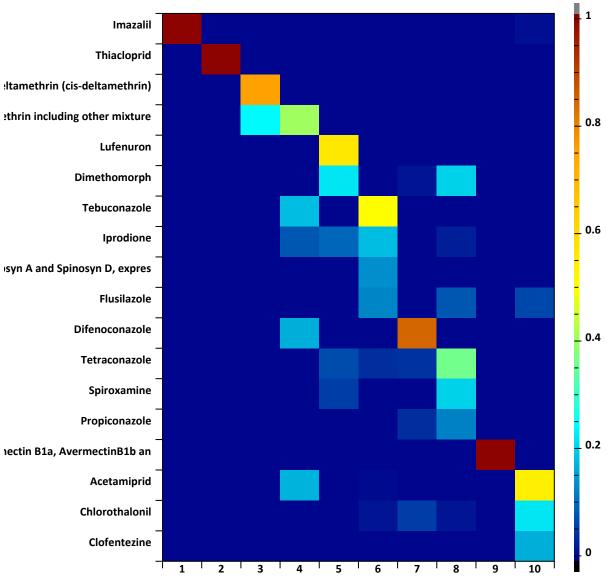
Name	Description		
Sparseness constraints	Sparseness parameter. Should be a value between 0 (not sparse,		
	many substances) and 1 (sparse, few substances).		
Number of mixtures	The number of mixtures.		
Number of iterations	Number of iterations, e.g. 1000.		
Seed for pseudo-random	Random seed for initializing matrix W and H.		
number generator.			
Exposures are	Exposures are risk based (expressed in equivalents of the		
	reference substance) or standardized.		
Convergence criterion	Convergence criterion for factorization algorithm.		
Cutoff for ratio total exposure/	For selection of individual(day) exposures specify cutoff for ratio		
maximum	total exposure/ maximum.		
Cutoff percentage (%) for total	For selection of individual(day) exposures specify cutoff		
exposure	percentage (%) for total exposure.		

Table 2.116: Calculation settings for module Exposure mixtures.



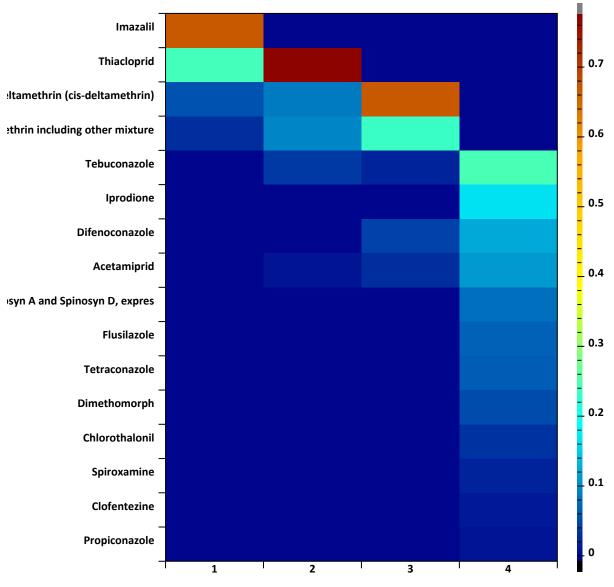
Co-exposure of substances

Figure 2.34: Heatmap for a solutions with 10 mixtures. The sparsity is set to 0 (not sparse). Each mixture contains many substances (see also Figure 2.35).



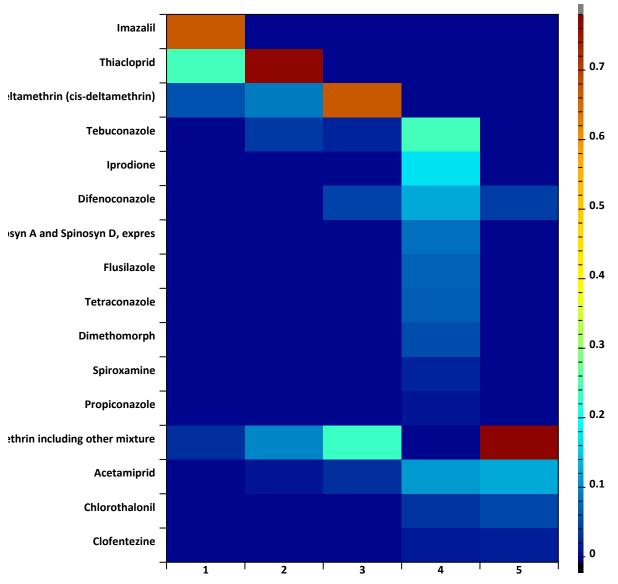
Co-exposure of substances

Figure 2.35: Heatmap for a solutions with 10 mixtures. The sparsity is set to 0.4 (sparse). Mixtures contain less substances compared to Figure 2.34.



Co-exposure of substances

Figure 2.36: Heatmap for solution with 4 mixtures. The first 4 mixtures in Figure 2.36 and Figure 2.37 are identical.



Co-exposure of substances

Figure 2.37: Heatmap for solution with 5 mixtures. The first 4 mixtures in Figure 2.36 and Figure 2.37 are identical.

Calculation of exposure mixtures

A multivariate decomposition method, sparse non-negative matrix underestimation (SNMU), is applied to the matrix of exposures per substance and per individual (chronic) or individual-day (acute) to find substance combinations that contribute most to the cumulative exposure.

• Exposure mixtures calculation

Inputs used: Exposures

Settings used

• Calculation Settings

2.4.6 Food conversions

Food conversions relate foods-as-eaten, as found in the consumption data, to modelled foods (foods-as-measured), which are the foods for which concentration data are available. A food-as-eaten can be linked to one, or multiple modelled foods using various conversion steps (e.g., using food recipes to translate a composite food into its ingredients). There are several types of conversion steps, and a conversion path may comprise multiple conversion steps between a food-as-eaten and a modelled food.

This module has as primary entities: Foods Substances

Output of this module is used by: Consumptions by modelled food Dietary exposures

Food conversions calculation

Food conversions are computed using a recursive search algorithm to link foods-as-eaten to modelled foods, possibly through intermediate conversion steps. For instance, if (unpeeled) apple and grapes are the modelled foods, the foodas-eaten apple pie contains peeled apple and raisins, peeled apple is linked to unpeeled apple, and raisins are dried grapes. Hence, for apple pie, there are two conversions, one to apple (with processing type 'peeled') and one to grapes (with processing type 'dried'), each with its own conversion path of intermediate conversion steps.

Substance independent conversion

The current implementation of the food conversion algorithm can be run substance independent. The *Find processing link (deprecated)* is skipped from the algorithm (default = false) and is only retained for backwards compatibility reasons only (see Advanced, set to true). Processed foods are easily recognized in the food translation step and retrieving the processing factor that belong to a certain the processing type is done outside the algorithm. In fact, finding processing types with corresponding processing factors is not a task of the conversion algorithm: conversion is about converting food codes to other food codes.

When the processing step is skipped, there is no need to run the conversion algorithm on a substance basis. The only information that is needed is whether a food code is a modelled food or not (i.c. is there a concentration available or not). This information can be computed before hand: for each substance all modelled foods are collected and supplied to the conversion algorithm in a common dictionary containing all modelled food codes. As soon as a food code is found in the dictionary, the conversion ends and the next food code is converted.

For each food-as-eaten, the food conversion algorithm recursively builds up the conversion paths using the following procedure:

- 1. *Substance independent conversion:* the conversion algorithm is substance independent. Check whether the current food is a modelled food. If successful, the food has been found, and the current search stops.
- 2. Check whether the current food translates to one or more foods through composition or read-across. Identify any processing types.
 - a. *Food recipe link:* try to find food translations for the current food (i.e., the ingredients of a composite food). This may result in one or more food codes for ingredients, and the iterative algorithm

will proceed with each of the ingredient food codes in turn. Simultaneously check, whether the current food is a processed food or not. If so, determine the processing type or facets.

- b. *TDS food sample composition link:* try to find the code in the TDSFoodSampleCompositions table (column idFood), a default translation proportion of 100% is assumed. The iterative algorithm will proceed with a TDS food (column idTDSFood) sample.
- c. *Read-across link:* try to find a food extrapolation rule for the current food, a default translation proportion of 100% for 'idToFood' is assumed.

Note that in the *food recipe link* processed foods are recognized and that the translation proportion to correct for a weight reduction or increase is stored.

If successful, restart at the first step with each of the new codes of the ingredient foods, TDS foods or Read Across foods.

- 3. *Marketshares link:* try to find subtype codes, e.g. 'xxx\$*' in the MarketShares table. In general, marketshares should sum to 100%. Foods with marketshares not summing to 100% are ignored in the analysis unless the checkbox *Allow marketshares not summing to 100%* is checked. This step is optional, see advanced settings. If successful, restart at step 1 with each of the new codes of the subtype foods.
- 4. *Supertype link:* try to find supertypes, e.g. 'xxx\$yyy' is converted to 'xxx'. This step is optional, see advanced settings if you want to use this. If successful, restart at step 1 with the new code of the supertype food.
- 5. *Default processing factor:* remove processing part (-xxx) of the code. If successful, restart at step 1 with the new code without processing part.

Substance dependent conversion

The original conversion algorithm contains two steps which are substance dependent. For each substance all food codes are supplied to the conversion algorithm and for each food code it is checked whether there is:

- a concentration,
- a processing type.

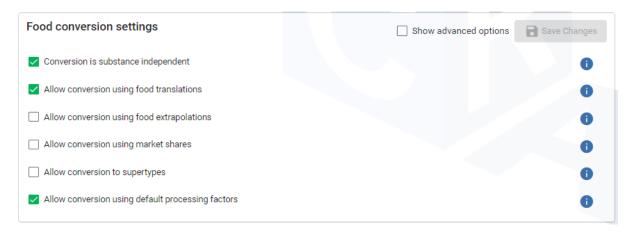
When a concentration is available for the food, this food is a modelled food (formerly known as food as measured). The food may be a food as eaten as such, like apple, or an ingredient of a food as eaten like tomato sauce on pizza which is converted to tomato. If concentrations are available, the food code is found and the conversion algorithm starts with converting the next food code. Otherwise, the conversion proceeds to the *processing link (deprecated)*:. Here, basically, processed foods are converted to an unprocessed food and processing type with corresponding processing factor. This processing step may be substance specific and, occasionally, this results in different conversion paths for different substances. This is undesirable behaviour and normally not the case (dependent on the supplied data in the food processing factor table). However, on rare occasions this might happen.

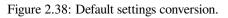
Find processing link (deprecated): Check whether the current food can be considered to be a processed variant (e.g., cooked or peeled) of another food.

Match processing factor: try to find the code in the processing factors table.

If successful, try to find the corresponding food translation proportion in the food recipes data to correct for a weight reduction or increase. Then, restart at the first step with the new code of the unprocessed food.

Warning: the *find processing link (deprecated):* step is not recommended and is currently maintained for backwards compatibility reasons only. Finding different conversions paths depending on the substance is undesirable behaviour.





Food conversion settings

Calculation settings

Name	Description
Allow conversion using processing info	Warning, the processing step is deprecated and is currently only maintained for backwards compatibility reasons. See documentation for more details how processed foods are converted in the upgraded conversion algorithm. Step 2a: try to find the code in the processing table. Try to find the code in the FoodTranslation table (step 3a) to account for weight reduction/increase (translation proportion). If unchecked (default), processing table is ignored. If successful, restart at step 1.
Allow conversion using food translations	Step 3a: try to find food translations for the current food (i.e., the ingredients of a composite food). This may result in one or more food codes for ingredients, and the iterative algorithm will proceed with each of the ingredient food codes in turn.
Allow conversion using TDS food sample compositions	Step 3b: try to find the code in the TDS food sample compositions table (idFood), a default translation proportion of 100% is assumed. The iterative algorithm will proceed with a TDS food (column idTDSFood) sample.
Allow conversion using food extrapolations	Step 3c: try to find read across codes. If unchecked, read across table is ignored, default is 'Use read across info'. E.g. for pineapple no measurements are found but by specifying that pineapple is converted to FruitMix (with a default proportion of 100%), the TDS sample concentration value of FruitMix will be used for pineapple (as-eaten or as ingredient). If successful, restart at step 1.
Allow conversion using market shares	Step 4: try to find subtype codes, e.g. 'xxx\$*' in the market shares table.
Allow marketshares not summing to 100%	Specify whether to rescale market share percentages that do not sum to 100%. If checked, then foods with marketshares not summing to 100% are allowed. If not, then these foods are ignored in the analysis.
Allow conversion to supertypes	Step 5: try to find supertypes, e.g. 'xxx\$yyy' is converted to 'xxx' (optional, check box if you want to use this). If checked, allows for linkage of consumed foods coded at a lower hierarchical level to foods with measured concentrations at a higher hierarchical level e.g. consumed is Apple (code PF\$Apple) -> measured is Pome Fruit (code PF). Note: food codes are split on '\$'. Measurements of substances on food are available at a less detailed food coding level than consumption data. MCRA allows to use the concentration data of a supertype for all underlying food codes. If successful, restart at step 1.
Allow conversion using default processing factors	Step 6: remove processing part. If unchecked, no default processing factors are assumed, default is 'Use default processing factors'. If successful, restart at step 1.
Conversion is substance independent	Conversion of foods is independent of the substance.

Table 2.117: Calculation settings for module Food conversions.

Calculation of food conversions

Food conversions are computed recursively, starting with a food-as-eaten and following a path to ingredients (food recipes), super/sup-type foods, etc. until either arriving at a modelled food (commonly the raw primary commodity) or concluding that the path does not lead to a modelled food.

• Food conversions calculation

Inputs used: Consumptions Modelled foods Processing factors Food recipes Market shares Food extrapolations Total diet study sample compositions Active substances

Settings used

• Calculation Settings

2.4.7 Human monitoring analysis

Human monitoring analysis compares observed human monitoring data with predictions made for the same population of individuals from dietary survey data, concentration data and (optionally) non-dietary exposure data.

This module has as primary entities: Populations Substances

Human monitoring analysis calculation

Human monitoring analysis computes internal substance concentration estimates based on provided human monitoring data. These estimates are specified at the level of long term average concentrations for individuals in case of *chronic assessments*, or the average concentrations for individual-days in case of *acute assessments*. The internal concentrations are computed independently for each substance, compartment, and sampling type.

The main steps for computing the human monitoring concentration estimates are:

- 1. Imputation of non-detects.
- 2. Imputation of missing values.
- 3. Calculation of individual concentrations (chronic) or individual day concentrations (acute).
- 4. Comparison of monitoring versus modelled exposures by substance and compartment (optional).

Imputation of non-detects

Similar to *concentrations measurements in food*, human monitoring measurements can also contain measurements below the limit of reporting and similar to *concentrations modelling in foods*, human monitoring analysis needs to address these non-detects and replace them with imputed concentration values. For this, two approaches are available:

- 1. Replace non-detects by zero.
- 2. Replace non-detects by a factor times LOR, in which the factor is set between zero and one.

Imputation of missing values

Concentration measurements may be missing. The following imputation methods are available for imputation of missing values:

- 1. Replace missing values by zero.
- 2. For each substance, sampling type, and compartment, replace missing values by a random other sample of this substance, sampling type, and compartment.

Note: For the second imputation method, more refined methods could be useful as well. E.g., when for a given day multiple samples are available, of which one is missing, then it may alternatively be sensible to leave this sample out when computing an average exposure. Also, when samples have been taken at different times during the day, it may be better to impute missing records using samples approximately from the same time-slot.

Calculation of acute human monitoring concentrations

For acute assessments, the monitoring concentrations are computed for each substance, compartment, and sampling type as average individual-day concentrations. That is, for a given substance, compartment, and sampling type, the acute individual-day concentration c_{ij} for individual *i* on day *j* is:

$$c_{ij} = \frac{\sum_{k=1}^{n_{\text{samples}}} c_{ijk} \cdot \textit{sg}_{ijk}}{n_{\text{samples}}},$$

where n_{samples} is the number of samples available for individual *i* on day *j*, and c_{ijk} and sg_{ijk} denote the concentration and specific gravity, respectively, of the *k*-th sample of the individual day.

Note: Note that currently, the acute concentrations are computed as mean concentrations when multiple samples are available for one day. In acute scenarios, one may be more interested in peak concentrations. I.e., the highest concentration of a day.

Calculation of chronic human monitoring concentrations

Note: The implementation for chronic is not yet available. Below is a description of the foreseen implementation.

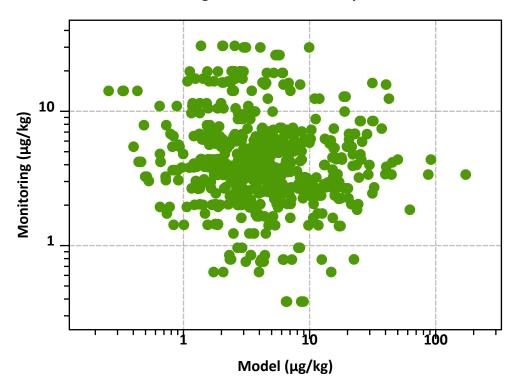
For chronic assessments, the monitoring concentrations are computed as the average monitoring concentrations of multiple individual-days for each substance, compartment, and sampling type. That is, for a given substance, compartment, and sampling type, the chronic concentration c_i for individual *i* is:

$$c_i = \frac{\sum_{j=1}^{n_{\rm days}} c_{ij}}{n_{\rm days}}, \label{eq:ci}$$

where n_{days} is the number of days that individual *i* was monitored, and c_{ij} denotes the average monitoring concentration of individual *i* on day *j*.

Compare measured and modelled exposures

An optional step of the human monitoring analysis is to compare the monitoring concentrations with *modelled exposures* that were obtained from *dietary* (and optionally *non-dietary*) exposure assessments. This comparison may provide insight in the coherence between modelled exposures and the measured reality. A requirement is that both monitoring data and dietary/non-dietary use data is available for the same individuals or individual-days. An example of a graphical output of these comparison is given in Figure 2.39.



Monitoring versus modelled exposures BPA

Figure 2.39: Measured exposures from monitoring versus modelled exposures

Human monitoring analysis settings

Calculation settings

Table 2.118: Calculation settings for module Human monitoring ana	lysis.
---	--------

Name	Description	
Non-detects handling method	Method for dealing with non-detects samples in human	
	monitoring data.	
Fraction of LOR	Factor for replacing non-detects with factor times LOR.	
Missing value imputation	Imputation method for missing values.	
method		
Correlate monitoring with	Correlate monitoring with modelled exposures.	
modelled exposures		

Calculation of human monitoring analysis

Human monitoring analysis calculations comprise two parts. The first part is to compute estimates of the human monitoring concentrations based on the human monitoring data. The second part, which is optional, is to relate the human monitoring concentrations to modelled concentrations from exposure assessments.

• Human monitoring analysis calculation

Inputs used: Human monitoring data Exposures

Settings used

• Calculation Settings

2.4.8 Human monitoring data

Human monitoring data quantify substance concentrations found in humans collected in human monitoring surveys.

This module has as primary entities: *Substances*

Output of this module is used by: Human monitoring analysis

Human monitoring data data formats

Data are provided on the survey, the individuals in the survey, the samples taken, the analyses performed, the analytical methods used, the properties for substances analysed, and the concentrations found.

Data are provided in the following relational tables:

- Human monitoring surveys
- Human monitoring individuals
- Human monitoring samples
- Human monitoring sample analyses
- Sample concentrations
- Analytical methods
- Analytical method properties for substances

Human monitoring samples

Suggested table definitions for human monitoring data.

Human monitoring surveys

Contains the survey definitions.

Name	Туре	Description	Aliases	Required
idSurvey	AlphaNumeric(50)	Unique identification code of	idSurvey	Yes
Name	AlphaNumeric(100)	the survey.	Name	No
	1 , ,	Name of the survey.		
Description	AlphaNumeric(200)	Description of the survey.	Description	No
Location	AlphaNumeric(50)	The location or country where	Location,	No
		survey is held. It is	Country	
		recommended to use ISO		
		Alpha-2 country codes.		
BodyWeight-	AlphaNumeric(50)	The unit of bodyweight of the	BodyWeight-	No
Unit		individuals of the survey: kg	Unit,	
		(default) or g.	UnitBody-	
			Weight,	
			WeightIn	
AgeUnit	AlphaNumeric(50)	The unit of age, i.e., year or	UnitAge, agein,	No
		month.	AgeUnit	
StartDate	DateTime	The starting date of the	StartDate	No
		survey.		
EndDate	DateTime	The end date of the survey.	EndDate	No
NumberOf-	Integer	The number of days each	NumberOf-	Yes
SurveyDays		individual participated in the	SurveyDays,	
		survey.	NDaysInSurvey	
idPopulation	AlphaNumeric(50)	Unique identification code of	IdPopulation,	No
		the population.	PopulationId	

 $Table\ aliases:\ HumanMonitoringSurveys,\ HumanMonitoringSurvey.$

Human monitoring individuals

The individuals of a survey are recorded in the individuals table.

Name	Туре	Description	Aliases	Required
idIndividual	AlphaNumeric(50)	Unique identification code of the individual.	idIndividual, IndividualId, Individual, Id	Yes
idSurvey	AlphaNumeric(50)	The identification code / short name of survey.	idSurvey	Yes
BodyWeight	Numeric	The body weight of the individual.	BodyWeight, Weight	Yes
Sampling- Weight	Numeric	The sampling weight for an individual (default = 1).	SamplingWeight	No
NumberOf- DaysInSurvey	Integer	The number of days the individual participated in the survey.	NumberOf- SurveyDays, NumberOfDays- InSurvey, DaysInSurvey, NDaysInSurvey	No
Age	Numeric	The age of the individual.	Age	No
Gender	AlphaNumeric(50)	The gender of the individual. Recommendation: use the codes Male/Female for coding the gender.	Gender	No
Other individual properties		Other individual properties can be added just like the fields age and gender. These properties are automatically parsed as co-factors or co-variables.		No

Table aliases: HumanMonitoringIndividuals.

Human monitoring samples

Contains the samples taken during the study.

Name	Туре	Description	Aliases	Required
idSample	AlphaNumeric(50)	Unique identification code of the monitoring sample.	idSample, Sample	Yes
idIndividual	AlphaNumeric(50)	Unique identification code of the individual.	idIndividual, IndividualId, Individual, Id	Yes
DateSampling	DateTime(50)	Date of sampling.	DateSampling, DateOf- Sampling, SamplingDate	No
DayOfSurvey	AlphaNumeric(50)	Identification code of the day of measurement.	Day, idDay, DayId, DayOfSurvey	Yes
TimeOf- Sampling	AlphaNumeric(50)	Identification code of the time of sampling.	TimeOf- Sampling, SamplingTime, TimeSampling	No
SampleType	AlphaNumeric(50)	Type of sample (e.g., pooled, 24h urine, spot urine, serum from blood, etc.).	SampleType, SamplingType	No
Compartment	AlphaNumeric(50)	If applicable, the measured compartment of the human body (e.g., blood, urine). When specified, the measurements are considered at the level of internal doses.	Compartment	No
ExposureRoute	AlphaNumeric(50)	If applicable, the measured exposure route, e.g., dermal (in case of skin wipes). When specified, the measurements are considered at the level of external doses.	ExposureRoute	No
SpecificGravity	Numeric	Specific gravity of the measured person for this particular sample.	SpecificGrafity, SpecificGravity	No
SpecificGravity- Correction- Factor	Numeric	Specific gravity of the measured person for this particular sample.	SpecificGravity- Correction- Factor	No

Table aliases: HumanMonitoringSamples, HumanMonitoringSample.

Human monitoring sample analyses

Contains the measurements of the samples of human monitoring studies.

Name	Туре	Description	Aliases	Required
idSample-	AlphaNumeric(50)	Unique identification code of	idSample-	Yes
Analysis		the sample analysis.	Analysis,	
			SampleAnalysis	
idSample	AlphaNumeric(50)	Code of the measured	idSample,	Yes
		monitoring sample.	Sample	
idAnalytical-	AlphaNumeric(50)	The code of method of	idAnalytical-	Yes
Method		analysis.	Method,	
			Analytical-	
			MethodName,	
			Analytical-	
			MethodId	
AnalysisDate	AlphaNumeric(50)	Date of analysis.	AnalysisDate,	No
			DateAnalysis	
Substance	AlphaNumeric(100)	One or more columns with		Yes
concentration(s)		the measured concentrations		
		of the substances in the unit		
		as specified by the analytical		
		method. The column name(s)		
		should match the substance		
		codes of the substances		
		measured by the analytical		
		methods. Empty fields for		
		substances that should have		
		been measured by the		
		analytical method are		
		considered to be non-detects		
		with measurement values		

Table 2.122: Table definition for HumanMonitoringSampleAnalyses.

 $Table\ aliases:\ Human Monitoring Sample Analyses,\ Human Monitoring Sample Analysis.$

Sample concentrations

The positive concentration values for substances from analysis in the unit specified in table human monitoring sample analyses. Non-detects (i.e. results 'less than LOR') are not included, their existence can be inferred from the tables AnalysisSamples and AnalyticalMethodSubstances, and the LOR itself from the analytical method.

Table 2.123: Table definition for HumanMonitoringSampleConcentrations.

Name	Туре	Description	Aliases	Required
idAnalysis-	AlphaNumeric(50)	The identification number of	idAnalysis-	Yes
Sample		the analysed sample.	Sample,	
			AnalysisSample-	
			Id	
idSubstance	AlphaNumeric(50)	The substance code.	idSubstance,	Yes
			SubstanceId,	
			Substance	
Concentration	Numeric	The measured concentration.	Concentration	Yes

Table aliases: HumanMonitoringSampleConcentrations, HumanMonitoringSampleConcentration.

Analytical methods

The analytical methods used for analyzing the samples are recorded in the analytical methods table. Each analytical method should have a unique identification code (idAnalyticalMethod). The description field may be used for a more detailed description of the analytical method. The records of this table should be linked to one or more analytical method substance properties table, which record the substances that are measured by this method (and their limits of reporting).

Name	Туре	Description	Aliases	Required
idAnalytical-	AlphaNumeric(50)	The code for the method of	idAnalytical-	Yes
Method		analysis.	Method,	
			Analytical-	
			MethodId,	
			Analytical-	
			MethodName,	
			Id	
Description	AlphaNumeric(200)	Additional description of	Description	No
		method of analysis.		

Table aliases: AnalyticalMethod, AnalyticalMethods.

Analytical method properties for substances

Name	Туре	Description	Aliases	Required
idAnalytical- Method	AlphaNumeric(50)	The code of method of analysis.	idAnalytical- Method,	Yes
			Analytical- MethodName,	
			Analytical-	
			MethodId	
idSubstance	AlphaNumeric(50)	The substance code.	idSubstance,	Yes
			SubstanceId,	
			Substance	
LOR	Numeric	The limit of reporting (LOR).	LOR	Yes
		In MCRA, LOR just means		
		the limit below which no		
		quantitative result has been		
		reported. Depending on a		
		laboratory's format of		
		reporting, LOR may be a		
		limit of detection (LOD), a		
		limit of quantification (LOQ)		
		or another limit.		
Concentration-	AlphaNumeric(50)	The code of the unit as used	Concentration-	No
Unit		for substance concentration	Unit, Units, Unit	
		data. Allowed code: kg/kg or		
		kilogram/kilogram; g/kg or		
		gram/kilogram; mg/kg or		
		milligram/kilogram (default);		
		µg/kg or		
		microgram/kilogram; ng/kg		
		or nanogram/kilogram; pg/kg		
		or picogram/kilogram.		

 Table 2.125: Table definition for AnalyticalMethodCompounds.

 $Table \ a liases: \ Analytical Method Substances, \ Analytical Method Substance, \ Analytical Method Compounds, \ Analytical Method Compound.$

Human monitoring data settings

Selection settings

Name	Description		
Surveys	The surveys that should be included in the action.		
Sampling methods	The sampling methods that should be included in the action.		

Table 2.126: Selection settings for module Human monitoring data.

Human monitoring data as data

Data are provided in the form of surveys consisting of individuals from which the human monitoring samples taken. Substance concentration measurements are linked to analyses performed on the human monitoring samples. The data should also include information about the analytical methods that were used.

• Human monitoring data data formats

2.4.9 Non-dietary exposures

Non-dietary exposures are the amounts of substances to which individuals in a population are exposed via any of three non-dietary routes: dermal, inhalation or oral, per day. Non-dietary exposures can be used for *computing aggregate exposure distributions* from both dietary and non-dietary routes of exposure. Depending on the exposure type, non-dietary exposures can be short-term/acute exposures and then contain exposures for individual-days, or they can be long-term/chronic exposures, in which case they represent the average exposure per day over an unspecified longer time period. Examples are presented as case studies in Kennedy et al. ([[Kennedy et al., 2012]], [[Kennedy et al., 2015b]], [[Kennedy et al., 2015b]], and R code to generate these examples is available for general use.

Datasets are typically generated by external programs, e.g. Browse, Bream2 or PACEM. The Browse and Bream2 models both simulate distributions of potential exposure of residents and bystanders to pesticides sprayed on crops. Probability distributions are included to quantify variations in input parameters representing conditions during a spray event. PACEM is a probabilistic exposure model for substances present in consumer products. Browse was an EU FP7 project (https://secure.fera.defra.gov.uk/browse/software), that in addition to bystanders and residents from boom-sprayers includes various arable and orchard scenarios. It includes dermal, oral and inhalation routes of exposure and can generate exposure files in the correct format for MCRA non-dietary exposure. The underlying simulation of dermal spray deposits on bystanders and residents was taken from Bream, although Browse includes post-processing to model indirect exposures, multiple routes and long-term exposure [[Kennedy et al., 2017]]. Volatilisation is also included through the PEARL-OPS model [[van den Berg et al., 2016]] to account for inhalation of vapours. Bream2 is an updated version of the original Bream model [[Kennedy et al., 2012]] and software is available online (http: //www.ssau.co.uk/bream2-calculator). Results from Bream had been used as part of EFSA guidance on bystander and resident exposure [[Butler et al., 2018]].

This module has as primary entities: Populations Substances

Output of this module is used by: *Exposures*

Non-dietary exposures data formats

Non-dietary exposures may be specified for multiple routes of exposure (dermal, oral and inhalation), for multiple substances, and for multiple exposure sources. Also, they can be provided as single deterministic exposure levels or as probabilistic exposure estimates and it is possible, but not mandatory, to specify uncertainty. The non-dietary exposures may be short term (acute) or longer term averages (chronic), and the user must ensure to supply appropriate non-dietary data for the type of exposure assessment of interest. For chronic assessments this means the non-dietary exposure is averaged over an appropriate time interval.

Non-dietary exposures are defined by non-dietary surveys to which dietary exposures are linked. For these surveys, individual properties can be specified to define non-dietary exposures for particular sub-groups of the population (e.g., specific age groups, or a specific gender). For each non-dietary survey a percentage of the target population that is not exposed from this source can be specified by means of a percentage. Uncertainty about non-dietary exposures can be specified by specifying multiple records for each individual in an additional table.

The use of multiple surveys can be used when multiple sources are relevant. For example, when modelling individuals taking part in various activities involving pesticide use or incidental exposures as a resident. Each non-dietary source is characterised in a particular user-selected or user-supplied non-dietary survey. By default, exposures from separate non-dietary surveys (sources) are considered to be independent events, but as explained below correlations between substances and/or activity types per individual can be represented if generated prior to uploading to MCRA. When

including multiple non-dietary surveys it is possible to supply some with uncertainty/variability and others without variability/uncertainty according to the requirements and data availability.

When the user supplies probabilistic non-dietary exposure estimates (i.e., there is a distribution for the non-dietary exposure rather than a single nominal value), then this information will be propagated as part of the *exposure assessment*. Distributions may be included to represent variability, uncertainty or both, and in these cases the aggregate exposure estimates are reported with variability and/or uncertainty as appropriate. Multiple (uncertain) values from the non-dietary exposure distribution may be supplied per individual and per substance.

Exposures within a non-dietary survey may be expressed as correlated or independent for the different substances. For example, if the exposures are a mixture of substances in a known ratio (e.g. from a specific tank mix of pesticides), or if exposure to one substance strongly implies that exposure to another is likely, these relationships may be included in the non-dietary data supplied by the user. Inference for the matched-case scenario with uncertainty analysis can use exposure sets. These are specific sets of exposures defined for each individual and in any uncertainty iteration an individual will receive exactly one of the exposure sets for that individual. Alternatively, independence may be represented by generating sets drawn from independent distributions when generating these tables.

Non-dietary exposures

Non-dietary exposure data is provided per non-dietary surveys. Each non-survey has some general information about the exposed population and the origin of the non-dietary exposure data. Also, a number of properties, such as specific age groups, can be specified for a survey. To each non-dietary survey, non-dietary exposures can be linked. These exposures may originate from dermal, oral and/or inhalatory exposure routes.

Non-dietary surveys

This table provides detail about non-dietary surveys (source of non-dietary exposure): description, location, date and unit of exposure).

Name	Туре	Description	Aliases	Required
idNonDietary-	AlphaNumeric(50)	The survey identification	idNonDietary-	Yes
Survey		number.	Survey	
Description	AlphaNumeric(200)	Description of non-dietary	Description	No
		survey.		
Location	AlphaNumeric(50)	The location of survey.	Location	No
Date	DateTime	The date of survey.	Date	No
NonDietary-	AlphaNumeric(50)	The unit of the non-dietary	Unit,	Yes
IntakeUnit		exposure.	NonDietary-	
			IntakeUnit,	
			NonDietary-	
			ExposureUnit	
Percentage-	Numeric	The proportion zeros,	PercentageZeros	No
Zeros		specified as a percentage (%).		
idPopulation	AlphaNumeric(50)	Unique identification code of	IdPopulation,	No
		the population.	PopulationId	

Table aliases: NonDietarySurveys, NonDietarySurvey.

Non-dietary survey properties

This table specifies demographic properties that apply to the individuals in the surveys. These properties could be used to link the individuals of a non-dietary survey with individuals from dietary surveys. That is, if demographic criteria are defined, only those individuals in the dietary survey that meet these criteria will be assigned non-dietary exposures. This table is not relevant when matching is switched on (i.e., when individuals are matched based on individual id).

Name	Туре	Description	Aliases	Required
Individual-	AlphaNumeric(50)	Name of demographic criteria	Individual-	Yes
PropertyName		for non-dietary exposures in a	PropertyName	
		particular survey e.g. age,		
		gender, height (must		
		correspond to a column name		
		in Individuals table).		
idNonDietary-	AlphaNumeric(50)	The code of survey (must	idNonDietary-	Yes
Survey		correspond to values in id	Survey	
		column of the non-dietary		
		surveys table).		
Individual-	AlphaNumeric(50)	Text value of the property e.g.	Individual-	No
PropertyText-		male or female, smoker or	PropertyText-	
Value		non-smoker.	Value	
Individual-	Numeric	Inclusive lower bound value	Individual-	No
Property-		of the property. E.g., a value	PropertyDouble-	
DoubleValue-		of "18" for an individual	ValueMin	
Min		property name called Age		
		would mean that only		
		individuals aged 18 and above		
		receive the non-dietary		
		exposures.		
Individual-	Numeric	Inclusive upper bound value	Individual-	No
Property-		of property e.g. a value of	PropertyDouble-	
DoubleValue-		"65" for an	ValueMax	
Max		IndividualPropertyName		
		called Age would mean that		
		only individuals aged 65 and		
		below receive the non-dietary		
		exposures.		

 $Table\ aliases:\ NonDietary Survey Properties,\ NonDietary Survey Property.$

Non-dietary exposures

This table defines nominal non-dietary exposure values (such as means) for individuals within the non-dietary surveys. It can also be used to specify non-dietary exposures for individuals within the food surveys. Each exposure comprises a non-dietary survey (source of exposure); a string identifying an individual, which may or may not correspond to the ID of an individual in a food survey; a substance; and dermal, oral and inhalation exposure values. Exposures are assumed to be external doses.

Name	Туре	Description	Aliases	Required
idIndividual	AlphaNumeric(50)	Non-dietary individual identification number. This id may 1) match with the individual ids of the dietary survey (dietary exposures matched to food survey individuals), 2) not match with the individual ids of the dietary survey (unmatched individuals), or contain a default exposure (indicated by idIndividual = 'General') linking the dietary exposures to individuals based on the demographic criteria defined in the non-dietary survey properties table.	idIndividual	Yes
idNonDietary- Survey	AlphaNumeric(50)	The code of the survey (must correspond to values in id column of non-dietary surveys table).	idNonDietary- Survey	Yes
idSubstance	AlphaNumeric(50)	The substance code.	idSubstance, SubstanceId, SubstanceCode, Substance	Yes
Dermal	Numeric	The dermal (non-dietary) exposure value.	Dermal	No
Oral	Numeric	The oral (non-dietary) exposure value.	Oral	No
Inhalation	Numeric	The inhalation (non-dietary) exposure value.	Inhalation	No

Table aliases: NonDietaryExposures, NonDietaryExposure.

Non-dietary exposure uncertainty records

This table may be used to supply uncertainty sets of multiple (uncertain) non-dietary exposure values for individuals within the non-dietary surveys. Multiple non-dietary values are generated by probabilistic exposure calculations i.e. when there is a distribution for the non-dietary exposure rather than a single nominal value. If this table is supplied, aggregate exposure estimates will be reported with uncertainty using the uncertainty set approach. Each exposure set comprises a non-dietary survey (source of exposure); an individual ID; a substance; and dermal, oral and inhalation exposure values. In addition, the id column is used to define the uncertainty set. Summarizing, an uncertainty set is identified by column id and contains all exposure sets defined for each individual. In each uncertainty run (outer loop) an uncertainty set is sampled and in each iteration (inner loop) nondietary individuals are sampled from this set.

Name	Туре	Description	Aliases	Required
idIndividual	AlphaNumeric(50)	Non-dietary individual	idIndividual	Yes
		identification number. The		
		idIndividual value may		
		correspond to an id in the		
		Individuals table (dietary		
		exposures matched to food		
		survey individuals), may not		
		correspond to an id in the		
		Individuals table (unmatched		
		individuals), or may contain a		
		default exposure (indicated by		
		idIndividual = 'General' -		
		demographic criteria for the		
		assignment of exposures are		
		defined in the		
		NonDietarySurveyProperties		
		table). For matching to occur,		
		the user will need to tick the		
		option to 'match specific		
		dietary survey individuals' in		
		the user-interface. The		
		software will then assign		
		non-dietary exposures to the		
		dietary individuals according		
		to the values in this column.		
		Any idIndividual values that		
		do not correspond to		
		individuals in the food survey		
		will be ignored, unless a value		
		'General' is specified. Then		
		the individual should meet the		
		demographic criteria as		
		defined in the		
		NonDietarySurveyProperties		
		table. If this box is left		
		unticked, the non-dietary		
		exposures will be randomly		
		allocated to the dietary		
		population provided they meet		
		the demographic criteria.		
idNonDietary-	AlphaNumeric(50)	code of survey (must	idNonDietary-	Yes
Survey		correspond to values in id	Survey	
		column of	2	
		NonDietarySurveys table)		
idCompound	AlphaNumeric(50)	Substance code (must	idSubstance,	Yes
r	r	correspond to values in id	SubstanceId,	
		column of Substances table).	SubstanceCode,	
			Substance	
id	AlphaNumeric(50)	Uncertainty set identification	id	Yes
		number.		
Dermal	Numeric	Dermal non-dietary exposure	Dermal	No
Donnai		value.	Domai	
Oral	Numeric	Oral non-dietary exposure	Oral	No
Jiai		value.	Jiai	110
Inhalation	Numeric		Inhalation	No
malation		Inhalation (non-dietary) exposure value.	minaration	

Table 2.130: Table definition for NonDietaryExposuresUncertain.

Table aliases: NonDietaryExposuresUncertain, NonDietaryExposureUncertain.

Non-dietary exposures settings

Uncertainty settings

Table 2.151. Oncertainty settings for module twon-detaily exposures.				
Name	Description			
Resample non-dietary Specifies whether non-dietary exposures are resampled. Note				
exposures non-dietary uncertainty is only ignored when individual				
uncertainty is set to false (uncheck box: do NOT resample				
	individuals).			

Table 2.131: Uncertainty settings for module Non-dietary exposures.

Non-dietary exposures uncertainty

In an aggregate exposure assessment, dietary and nondietary data are combined into an aggregate exposure distribution. The nondietary data are supplied in table NonDietaryExposures. In an uncertainty analysis, MCRA provides two ways to assess the uncertainty:

- 1. the uncertainty set approach
- 2. the bootstrap algorithm.

When table **NonDietaryExposuresUncertain** is not supplied, the nondietary data in table **NonDietaryExposures** is resampled and the bootstrapped sets are used in the uncertainty run. More precisely, in each outer loop of the 2D Monte Carlo, within each nondietary survey (multiple surveys may be supplied), the nondietary individuals are resampled. Each individual represents a nondietary exposure set containing dermal and/or oral and/or inhalation exposure values for multiple substances. Bootstrapping is the default behaviour when the **NonDietaryExposure-sUncertain** table is missing. When uncertainty distributions supplied in this table represent sampling uncertainty (individual exposure sets are repeatedly sampled using the same nondietary exposure generator without changing the input parameters), then bootstrapping the data performs equally well and is more efficient.

Non-dietary exposures as data

Non-dietary exposures are collected in non-dietary surveys. Data may be specified on population level or individual level, and may or may not include variability and uncertainty.

• Non-dietary exposures data formats

Inputs used: Active substances

See also Combining dietary and non dietary exposures.

2.4.10 Single value dietary exposures

Single value dietary exposures are based on the single value concentrations of substances, expressed per standard (kg) bodyweight and/or single value amounts of consumed modelled food. Depending on the exposure type, dietary exposures can be short-term/acute exposures.

This module has as primary entities: Populations Foods Substances

Output of this module is used by: Single value risks

Single value dietary exposures data formats

Single value dietary exposures are IESTI etc.

Dietary exposures

Dietary exposure data is specified through dietary exposure models. To each dietary exposure model, exposure distributions are linked.

Dietary exposure models

High level description of the dietary exposure models, specifying the id, name, description and the (reference) substance and exposure unit used for reporting the exposures. To this models, exposure percentiles and bootstrap values of the percentile may be linked.

Name	Туре	Description	Aliases	Required
idDietary- ExposureModel	AlphaNumeric(50)	Identifier of the dietary exposure model.	id, idDietary- Exposure, idExposure- Model	Yes
Name	AlphaNumeric(100)	The name of the dietary exposure model.	Name	No
Description	AlphaNumeric(200)	Description of dietary exposure model.	Description	No
idSubstance	AlphaNumeric(50)	The code of the substance.	idSubstance, SubstanceId, SubstanceCode, Substance, idCompound, CompoundId, Compound- Code, Compound	Yes
ExposureUnit	AlphaNumeric(50)	The intake/exposure unit of the dietary exposures reported by this model. If not specified, then a default exposure unit of mg/kg BW/day is assumed.	Unit, ExposureUnit, IntakeUnit	Yes

Table aliases: DietaryExposureModels, DietaryExposures.

Dietary exposure percentiles

Exposure percentiles linked to a dietary exposure model. The percentiles are reported in the unit specified by the exposure model to which they belong.

Name	Туре	Description	Aliases	Required
idDietary-	AlphaNumeric(50)	The code of the dietary	idDietary-	Yes
ExposureModel		exposure model to which this	ExposureModel	
		record belongs.		
Percentage	Numeric	The percentage to which the	Individual-	Yes
		percentile value belongs.	PropertyDouble-	
			ValueMin	
Exposure	Numeric	The percentile value. I.e., the	Exposure	Yes
		exposure value belonging to		
		the specified percentage.		

Table 2.133:	Table	definition	for	Dietary	Exposurel	Percentiles.
14010 211001				210000192		

Table aliases: DietaryExposurePercentiles.

Dietary exposure percentile bootstrap values

Uncertainty values, obtained from bootstrap runs, of the dietary exposure percentiles.

Name	Туре	Description	Aliases	Required
idDietary-	AlphaNumeric(50)	The code of the dietary	idDietary-	Yes
ExposureModel		exposure model to which this record belongs.	ExposureModel	
idUncertainty- Set	AlphaNumeric(50)	The uncertainty set identifier.	idUncertainty- Set, UncertaintyId	Yes
Percentage	Numeric	The percentage to which the percentile value belongs.	Individual- PropertyDouble- ValueMin	Yes
Exposure	Numeric	The percentile value. I.e., the exposure value belonging to the specified percentage.	Exposure	Yes

Table 2.134: Table definition for DietaryExposurePercentilesUncertain.

 $Table\ aliases:\ Dietary Exposure Percentiles Uncertain,\ Dietary Exposure Percentile Uncertains.$

Single value dietary exposures calculation

Either the short-(acute) or long-term (chronic) dietary exposure to a substances via food can be estimated as a single value calculated from single value inputs. This is often referred to as deterministic estimation. MCRA implements the IESTI, TMDI, IEDI and NEDI (Rees-Day) calculation methods that are also available in the EFSA PRIMo (Pesticide Residue Intake Model) tool revision 3, [[EFSA, 2018]].

The implementation in MCRA allows more choices than EFSA PRIMo by choosing other inputs or input combinations for the calculation formula. Moreover, the calculations can in all cases be adapted for processing factors or occurrence frequencies. For the chronic estimates, also the contributions per food or processed food are reported.

Acute single value dietary exposure assessment

The short term (acute) exposure assessment is usually the exposure related to a consumption of food over a single day. MCRA applies in principle the IESTI equations as shown in EFSA PRIMo revision 3 [[EFSA, 2018]], but the equations are extended with a factor OF to allow adaptation for an occurrence frequency lower than 1. So the inputs to the equations are not necessarily the same as used in PRIMo. For example, the large portion (LP) and body weight (BW) can be computed instead of just being standard values.

IESTI (International Estimated Short-Term Intake)

The IESTI (International Estimated Short-Term Intake) is calculated according to different equations depending on the unit weight of the raw agricultural commodity (RAC) and the unit weight of the edible portion (EP). The following cases are distinguished.

- **Case 1** refers to commodities with unit weight of the raw agricultural commodity $U_{RAC} \le 25$ g (e.g. walnuts, strawberries and peas. It is also used for meat, liver, kidney, edible offal, eggs and for post-harvest uses in cereal grains, oilseeds and pulses).
- **Case 2a** for food product with a $U_{RAC} > 25$ g, where the meal portion is $> U_{ep}$ (unit weight edible portion).
- Case 2b for food products with a ${\rm U_{RAC}}>25$ g, where the meal portion is $\leq {\rm U_{ep}}.$

Case 3 for food products that are usually bulked or blended before they are consumed (e.g. cereals, pulses, oilseeds and milk).

The calculations are as follows.

Case 1

Case 2a

$$\frac{U_{ep} \cdot HR \cdot PF \cdot CF \cdot VF \cdot OF + (LP - U_{ep}) \cdot HR \cdot PF \cdot CF \cdot OF}{BW}$$

Case 2b

Case 3

$$\frac{\text{LP} \cdot \text{STMR} \cdot \text{PF} \cdot \text{CF} \cdot \text{OF}}{\text{BW}}$$

New Case 1 and 3:

$$\frac{\text{LP} \cdot \text{MRL} \cdot \text{PF} \cdot \text{CF} \cdot \text{OF}}{\text{BW}}$$

New Case 2a and 2b

 $\frac{\text{LP} \cdot \text{MRL} \cdot \text{PF} \cdot \text{CF} \cdot \text{VF} \cdot \text{OF}}{\text{BW}}$

Parameters used in the equations

MRL: Maximum residue level for the RAC concerned (default in mg/kg);

STMR: Supervised Trials Median Residue for raw agricultural commodity (RAC) concerned (default in mg/kg);

CF: Conversion factor residue definition enforcement to residue definition risk assessment (calculated as the ratio of residues according to the residue definition for risk assessment divided by the residue concentration according to the residue definition for enforcement);

OF: Use Frequency of the raw agricultural commodity (RAC),

BW: body weight of the population related to the LP (default in kg);

LP: Large portion reported (in kg/day) (97.5th percentile of eaters (or alternative percentile, depending on the number of reported eating occasions);

HR: Highest residue according to residue definition for enforcement in composite sample (default in mg/kg);

 U_{ep} : Unit weight of edible portion (in kg), provided by the country from which the LP was reported (or mean unit weight calculated from all available unit weight data, if no unit weight is available from the country matching the highest LP;

PF: Processing factor or peeling factor (calculated as the ratio of residues in processed/peeled product, divided by residue concentration in unprocessed/unpeeled product);

VF: variability factor, depending on the unit weight of the whole product (U_{RAC}) , different default VF are used in the calculations.

 $(\rm U_{RAC}) < 25$ g, the calculations are performed according to case 1 (VF = 1).

 (U_{RAC}) between 25 and 250 g: VF = 7.

 (U_{RAC}) greater than 250: VF = 5.

In $IESTI_{new}$, a default VF of 3 is used.

In case the empirically derived variability factors are available, the default VF is to be replaced.

Alternative IESTI-styled assessments

If consumption survey data for a specific population are available, the LP values in the IESTI equations may be replaced by statistics calculated from these data (at the consumed modelled food level).

If concentration monitoring data (retrospective) or concentration field trial data (prospective) are available, the MRL, HR, STMR values in the IESTI equations may be replaced by statistics calculated from these data (at the consumed modelled food level).

In the current use of IESTI, the occurrence frequency (use frequency) OF is assumed to be 1. In alternative assessments, a more realistic estimate may be used. Such an estimate could be derived for example as the highest occurrence frequency observed in a retrospective assessment for either the same substance or the same food.

IESTI special cases

For some foods, substances are applied after harvest, i.c. post-harvest use. For those combinations of food and substance, Case 1 should be used in the calculation. However, commodities with post-harvest use like cereal grains, oilseeds and pulses are typically bulked or blended (Case 3). To overrule Case 3, specify in table *IESTISpecialCases* the food and substance combination with 'PostHarvest' as application type. For those food and substance combinations with a unit weight of the raw agricultural commodity $U_{RAC} \leq 25$ g, Case 1 is applied. When substances are applied before harvest, i.c. pre-harvest use, Case 1 should be overruled by Case 3. Specify in table *IESTISpecialCases* the food and substance combination with 'PreHarvest' as application type. See also *IESTISpecialCases table format*.

Chronic single value dietary exposure assessment

The long term (chronic) exposure assessment is usually the exposure related to a consumption over a longer period of time. MCRA applies in principle the TMDI, IEDI or NEDI (Rees-Day) equations as shown in EFSA PRIMo revision 3 [[EFSA, 2018]]. However, the equations are extended with factors PF_i and OF_i to allow adaptation for processing factors and occurrence frequencies lower than 1. Also, the inputs to the equations are not necessarily the same as used in PRIMo. For example, the consumption statistics (MC, $p_{97.5}$) and body weight (BW) can be computed instead of just being standard values. Note that TMDI, IEDI and NEDI (Rees-Day) estimates are summations over foods (raw agricultural products). In addition to the summations, MCRA will also report the individual terms (single value dietary exposures per food).

TMDI (Theoretical Maximum Dietary Intake)

$$\sum_{X=i}^{n} \frac{MRL_i \cdot CF_i \cdot PF_i \cdot OF_i \cdot MC_i}{BW}$$

i, j, k, ...n: individual raw agricultural products

IEDI (International Estimated Dietary Intake)

$$\sum_{X=i}^{n} \frac{\textit{STMR}_i \cdot \textit{CF}_i \cdot \textit{PF}_i \cdot \textit{OF}_i \cdot \textit{MC}_i}{BW}$$

i, j, k, ...n: individual raw agricultural products

NEDI (National Estimated Dietary Intake): Rees-Day model (I)

$$\sum_{X=i}^{j} \frac{\textit{MRL}_i \cdot \textit{CF}_i \cdot \textit{PF}_i \cdot \textit{OF}_i \cdot \textit{p}_{97.5}\textit{consumption}_i}{\textit{BW}} + \sum_{X=k}^{n} \frac{\textit{MRL}_k \cdot \textit{CF}_k \cdot \textit{PF}_i \cdot \textit{OF}_i \cdot \textit{MC}_k}{\textit{BW}}$$

i, *j*: two raw agricultural products leading to the highest intake;

k, l, m, ...n: remaining raw agricultural commodities consumed

NEDI (National Estimated Dietary Intake): Rees-Day model (II)

$$\sum_{X=i}^{j} \frac{STMR_i \cdot CF_i \cdot PF_i \cdot OF_i \cdot p_{97.5} consumption_i}{BW} + \sum_{X=k}^{n} \frac{STMR_k \cdot CF_i \cdot PF_i \cdot OF_i \cdot MC_k}{BW}$$

i, *j*: two raw agricultural products leading to the highest intake;

k, l, m, ...n: remaining raw agricultural commodities consumed

Parameters used in the equations

MRL_i: Maximum residue level for the RAC concerned (default in mg/kg);

 $STMR_i$: Supervised Trials Median Residue for raw agricultural commodity (RAC) concerned (default in mg/kg);

 CF_i : Conversion factor residue definition enforcement to residue definition risk assessment (calculated as the ratio of residues according to the residue definition for risk assessment divided by the residue concentration according to the residue definition for enforcement);

 MC_i : mean consumption for a given raw agricultural product (RAC) calculated for the whole survey/subgroup of the survey, including processed products (recalculated to the unprocessed RAC) (default in kg/day);

 $p_{97.5}$ consumption_i for a given raw agricultural product (RAC), calculated from the individual consumption reported by the participants of the whole survey/subgroup of the survey, including processed products (recalculated to the unprocessed RAC) (default in kg/day);

BW: mean body weight of the population related to the LP or mean consumption (default in kg).

 OF_i : Occurrence Frequency of the substance on the food (typically, a raw agricultural commodity, RAC),

 PF_i : Processing factor or peeling factor (calculated as the ratio of residues in processed/peeled product, divided by residue concentration in unprocessed/unpeeled product);

Alternative TMDI-, IEDI- or NEDI-styled assessments

If consumption survey data for a specific population are available, the MC, $p_{97.5}$ consumption values in the IESTI equations may be replaced by statistics calculated from these data (at the consumed modelled food level).

If concentration monitoring data (retrospective) or concentration field trial data (prospective) are available, the MRL, STMR values in the IESTI equations may be replaced by statistics calculated from these data (at the consumed modelled food level).

In the current use of IESTI, the occurrence frequency (use frequency) OF is assumed to be 1. In alternative assessments, a more realistic estimate may be used. Such an estimate could be derived for example as the highest occurrence frequency observed in a retrospective assessment for either the same substance or the same food.

Single value dietary exposures settings

Selection settings

Name	Description
Risk type	The type of exposure considered in the assessment; acute (short
	term) or chronic (long-term).
Dietary exposure calculation	A tier is a pre-specified set of model configurations. By selecting a
tier	model tier, MCRA automatically sets all model settings in this
	module according to this tier. Note that currently tier setting may
	need to be performed separately in sub-modules. Use the Custom
	tier when you want to manually set each model setting.

Table 2 135.	Selection setting	for module Single	value dietary exposures.
14010 2.155.	Sciection settings	s for moune single	value uletary exposures.

Calculation settings

54165.	
Name	Description
Single value dietary exposure calculation method	Method for computing single value dietary exposures.
Apply processing factors	Specified in table ProcessingFactor. If checked, processing factors are applied. Concentrations in the consumed food may be different from concentrations in the modelled food in monitoring programs (typically raw food) due to processing, such as peeling, washing, cooking etc. If unchecked, no processing information is used. This is in most (though not all) cases a worst-case assumption
Use occurrence frequencies	Account for occurrence frequencies for combinations of food and substance in the exposure calculations.

Table 2.136: Calculation settings for module Single value dietary exposures.

Calculation of single value dietary exposures

Single value dietary exposures are calculated from single value consumptions per modelled food and single value concentrations. Optionally, also processing factors, unit variability models and use frequencies are applied.

• Single value dietary exposures calculation

Inputs used: Single value consumptions Single value concentrations Processing factors Unit variability factors Occurrence frequencies

Settings used

• Calculation Settings

2.5 Hazard modules

Hazard data exist at two levels: at a lower level *dose response data* give *responses* measured in *test systems* from doses of *active substances*. Such data can be modelled with *dose response models*.

At a higher level *responses* can be linked to *effects*, optionally via *AOP networks*, using *effect representations*. If benchmark responses (BMRs) have been specified, *dose response models* can calculate Benchmark Doses (BMDs), which are the preferred Points of departure in hazard assessments. In addition, or alternatively, external *points of departure* can be specified for *active substances* and *effects*.

BMDs from *dose response models* and/or other *points of departure* can be converted to *hazard characterisations* at the intended level (external or internal dose, without or with safety factors), using *kinetic models*, *inter-species conversions* and/or *intra-species factors*. Finally, *hazard characterisations* can be translated to *relative potency factors*.

2.5.1 Active substances

Active substances are the substances that may lead with non-zero probability (P (AG)>0) to a specific *health effect* (adverse outcome). In the simplest case, all substances in the scope of the action will form one assessment group (AG) of active substances. In more advanced cases, the list of active substances is derived from possibly multiple assessment group memberships, which are scores for substances that determine whether a substance is included (score > 0) or excluded (score = 0) in the set of active substances. Substances with membership 0 are excluded from the list of active substances. Memberships scores between 0 and 1 are treated as probabilities of being in the set of active substances. Assessment group memberships can be either specified directly as data or derived from *QSAR membership models*, *molecular docking models*, or from availability of *points of departure*.

This module has as primary entities: Effects Substances

Output of this module is used by: Concentrations Single value concentrations Occurrence patterns Occurrence frequencies Substance conversions Non-dietary exposures Kinetic models Relative potency factors Hazard characterisations Inter-species conversions Intra species factors Food conversions High exposure food-substance combinations Dietary exposures Exposures

Active substances data formats

Active substances as data have to be specified via assessment group (AG) memberships in an AG membership model. For each effect one or more AG membership models can be available, one of which should be chosen in assessments. The AG memberships can be crisp, i.e. a positive list of active substances (with default memberships 1, although it is also allowed to include the negative memberships with membership 0 explicitly) or probabilistic ($0 \le P \le 1$).

Assessment group membership models

Assessment group membership models contain substance membership definitions for a given (health) effect. This data is described using two tables: the assessment group membership models table and the assessment group memberships table. The groups for a specified health effect are defined in the assessment group membership models table. The assessment group memberships table describes the substance memberships (or membership probabilities) in each group.

Assessment group membership models

This table contains the definitions of the assessment group membership models. Each model contains a id, name, an optional description, and refers to its related health effect.

Туре	Description	Aliases	Required
AlphaNumeric(50)	The unique identification code of the assessment group membership model	id, idModel, Model, idAssessment-	Yes
	membership model.		
		-	
		± .	
		Model,	
		Group-	
		Membership-	
		Model	
AlphaNumeric(100)	The name of the assessment	Name	No
	group membership model.		
AlphaNumeric(200)	Description of the assessment	Description	No
AlphaNumeric(50)	The effect code.	,	Yes
Numeric		Accuracy	No
	1		
	-	0	N
Numeric		Sensitivity	No
Numeria		Spacificity	No
Inumeric		specificity	NO
AlphaNumeric(200)	-	References	No
		References	
	miormunon about the		1
	AlphaNumeric(50) AlphaNumeric(100)	AlphaNumeric(50)The unique identification code of the assessment group membership model.AlphaNumeric(100)The name of the assessment group membership model.AlphaNumeric(100)The name of the assessment group membership model.AlphaNumeric(200)Description of the assessment group membership model.AlphaNumeric(50)The effect code.NumericIf applicable, the accuracy of the assessment group membership modelNumericIf applicable, the sensitivity of the assessment group membership model.NumericIf applicable, the sensitivity of 	AlphaNumeric(50)The unique identification code of the assessment group membership model.id, idModel, Model, idAssessment- GroupModel, Assessment- GroupModel, idGroup- Membership- ModelAlphaNumeric(100)The name of the assessment group membership model.NameAlphaNumeric(200)Description of the assessment group membership model.DescriptionAlphaNumeric(50)The effect code.idEffect, EffectId, EffectNumericIf applicable, the accuracy of the assessment group memberships.AccuracyNumericIf applicable, the sensitivity of the assessment group membership model.SensitivityNumericIf applicable, the specificity of the assessment group membership model.SensitivityNumericIf applicable, the specificity of the assessment group membership model.SpecificityAlphaNumeric(200)External reference(s) to sources containing moreReferences

Table 2.137: Table definition for AssessmentGroupMembershipModel	s
ruble 2:157. ruble definition for 7 issessment Group Membership Model	10.

 $Table\ aliases:\ Assessment Group Membership Models,\ Assessment Group Membership Model.$

Assessment group memberships

Substances belong to an assessment group with certainty (probability 1), or the membership are uncertain. This table allows to specify membership probabilities for assessment group membership models. The probability should be a value between zero and one. For example, set to 1 or 0, or prior probabilities, or probabilities or 0/1 values estimated from QSAR, from Molecular Docking or from expert elicitation. The table can contain prior or posterior memberships. Default membership are specified with an empty idSubstance field.

Name	Туре	Description	Aliases	Required
Name idGroup- Membership- Model	AlphaNumeric(50)	Description The id of the assessment group memberships model or source.	Allases Model, idModel, idAssessment- Group- Membership- Model, Assessment- Group- Membership- Model, idGroup- Membership- Model, Group- Membership-	Yes
idSubstance	AlphaNumeric(50)	The code of the substance.	Model, idGroup idSubstance, SubstanceId, SubstanceCode, Substance	Yes
Group- Membership	Numeric	Probability of the substance for belonging to the assessment group for the effect. If omitted, the default is 1, i.e. certain membership.	Group- Membership, Membership, Probability, Probability, Assessment- Group- Membership	Yes

Table 2.138: Table definition for AssessmentGroupMemberships.

Table aliases: AssessmentGroupMemberships, AssessmentGroupMembership.

Active substances calculation

Depending on the *model settings*, the set of active substances for a specified effect can be computed in several ways:

- 1. From the list of substances with available *points of departure (POD) data* for the specified effect. If there is a POD, then the substance is considered an active substance, with membership 1. If not, the membership is 0, and the substance is excluded from the list of active substances.
- 2. From one or more in-silico (QSAR and/or molecular docking) models. The results of the in-silico models should be provided as *QSAR membership models data* and/or *molecular docking models data*. Binding energies from molecular docking models are first translated to crips memberships using a threshold value. The results from multiple in-silico models can be combined in any of four membership calculation methods:
 - 1. (crisp, any) the substance is considered an active substance if any in-silico model indicates activity;
 - 2. (crisp, majority) the substance is considered an active substance if the majority of in-silico models indicates activity;
 - 3. (probabilistic, ratio) the membership probability is the fraction of in-silico models that indicate activity;
 - 4. (probabilistic, Bayesian) the membership probability is calculated using a Bayesian model according to Kennedy et al. [[Kennedy et al., 2020]] and a specified prior probability (which is by default 0.5).

For substances within the scope of the assessment but without in-silico data, the default is to omit them in the AG. There is an option however to include such substances with a default membership probability.

3. From a combination of 1 and 2, using either the union (OR) method or the intersection (AND method) of results.

Active substances settings

Calculation settings

Name	Description
Filter by certain assessment	Filter substances by certain assessment group membership.
group membership	
Filter by possible assessment	Filter substances by possible assessment group membership.
group membership	
Restrict active substances to	Restrict assessment group membership based on presence/absence
substances with available PODs	of points of departure.
Restrict active substances to	Restrict assessment group membership based on presence/absence
substances with available	of hazard characterisations.
hazard characterisations	
Derive memberships from	Specifies whether QSAR membership data is used for computing
QSAR membership data	the assessment group memberships.
Derive memberships from	Specifies whether molecular docking data is used for computing
molecular docking data	the assessment group memberhips.
Include substances without	For non-probabilistic methods: specifies whether substances for
membership information	which no membership information is available in the specified
	inputs should be included in the assessment group.
Combination method	Specifies whether to take the intersection or the union of the set of
memberships from available	substances with available PoDs and the set of substances with
PODs and in-silico data	positive/probable (in-silico) membership score.
Membership calculation	Calculation method for computing assessment group
method	memberships: majority/any (crisp methods), ratio/Bayesian
	(probabilistic methods)
Default/prior membership	Default substance membership probability for which no
probability	membership information is available in the specified inputs. Prior
	probability for Bayesian method.
Use probabilistic assessment	Specifies whether substance memberships should be expressed in
group memberships	terms of probabilities (probabilistic). Otherwise, substance
	memberships are expressed as in or out (crisp).

Table 2.139: Calculation settings for module Active substances.

Uncertainty settings

Table 2.140: Uncertainty settings for module Active substances.

Name	Description
Resample assessment group	Specifies whether assessment group memberships of substances
memberships	should be resampled using the assessment group membership
	probabilities.

Active substances as data

When provided as data, in the form of assessment group memberships, the active substances are derived from the specified memberships.

• Active substances data formats

Inputs used: AOP networks Points of departure Hazard characterisations

Calculation of active substances

Active substances and assessment group memberships may be computed from PoD presence of in-silico data.

• Active substances calculation

Inputs used: Molecular docking models QSAR membership models

Settings used

• Calculation Settings

2.5.2 AOP networks

Effects are related to each other using the toxicological concept of adverse outcome pathways (AOPs) and adverse outcome pathway networks (see https://aopwiki.org). Adverse Outcome Pathway (AOP) Networks specify how biological events (effects) can lead to an adverse outcome (AO) in a qualitative way through relations of upstream and downstream key events (KEs), starting from molecular initiating events (MIEs). Using AOPs, the adverse outcome (AO), e.g., liver steatosis, is linked to key events (KEs), e.g., triglyceride accumulation in the liver, and to molecular initiating events (MIEs), e.g., PPAR-alpha receptor antagonism. In general, multiple AOPs may lead to the same AO, and therefore AOP networks can be identified.

This module has as primary entities: *Effects*

Output of this module is used by: *QSAR membership models Molecular docking models Active substances Relative potency factors Hazard characterisations Points of departure Effect representations*

AOP networks data formats

AOP networks

AOP networks are described using two tables: the AOP networks table, and the effect relations table. The AOP networks table records the ids, names, descriptions, and other metadata of the AOP networks. The effect relations table describes the effects and effect relations (i.e., upstream and downstream key event relations) that are part of the AOP network.

AOP networks

Data format for specification of adverse outcome pathway (AOP) networks.

Name	Туре	Description	Aliases	Required
idAdverse-	AlphaNumeric(50)	Unique identification code of	idAOPN,	Yes
Outcome-		the AOP network.	idAOPNetwork,	
Pathway-			AOPN,	
Network			AOPNetwork,	
			Id	
Name	AlphaNumeric(100)	Name of the AOP network.	Name	No
Description	AlphaNumeric(200)	Additional description or label	Description	No
		of the AOP network.		
Reference	AlphaNumeric(200)	External reference(s) to	Reference,	No
		sources containing more	References	
		information about the AOP		
		network. E.g., the AOP wiki,		
		and the associated AOP wiki		
		Ids.		
idAdverse-	AlphaNumeric(50)	The identification code of the	idAdverse-	Yes
Outcome		effect representing the adverse	Outcome, idAO,	
		outcome of this AOP	idEffect,	
		network.	Adverse-	
			Outcome	
RiskType	AlphaNumeric(100)	The risk type of the adverse	RiskType	No
		outcome.		

Table 2.141: Table definition for AdverseOutcomePathwayNetworks.

Table aliases: AOPNetworks, AOPNetwork.

Effect relations

Dataformat for specification of the effect (key event) relationships of adverse outcome pathway (AOP) networks.

Name	Туре	Description	Aliases	Required
idAdverse-	AlphaNumeric(50)	Identification code of the	idAdverse-	Yes
Outcome-		AOP network for which this	Outcome-	
Pathway-		link is defined.	Pathway-	
Network			Network,	
			idAOPN,	
			idAOPNetwork,	
			AOPN,	
			AOPNetwork	
idDownstream-	AlphaNumeric(50)	Identification code of the	idDownstream-	Yes
KeyEvent		(triggered) effect of this	KeyEvent,	
		relationship.	idEffect,	
			idKeyEvent,	
			Effect, KeyEvent	
idUpstream-	AlphaNumeric(50)	Identification code of the	idTrigger,	Yes
KeyEvent		triggering effect of this	idUpstreamKey-	
		relationship.	Event, Trigger	
Reference	AlphaNumeric(200)	External reference(s) to	Reference,	No
		sources containing more	References	
		information about the effect		
		(key event) relationships.		

Table aliases: EffectRelations, EffectRelation, EffectRelationships, EffectRelationship, KeyEventRelationships, KeyEventRelationship.

AOP networks settings

Selection settings

Table 2.145. Selection settings for module AOF networks.		
Name	Description	
AOP Network	The AOP networks of interest.	
Restrict AOP network by focal	Restrict the AOP network to a specific sub-network, containing	
upstream event	only the AOPs that include both the focal key event (KE) defined	
	here (which must be upstream of the AO) and the focal effect	
	(adverse outcome, AO)	
Focal upstream event	The focal key event used for restricting the AOP network to a	
	specific sub-network of interest.	

Table 2.143: Selection settings for module AOP networks.

AOP networks as data

AOP networks can only be provided as data in the form of network definitions containing effect relations (key-event relationships) collections.

• AOP networks data formats

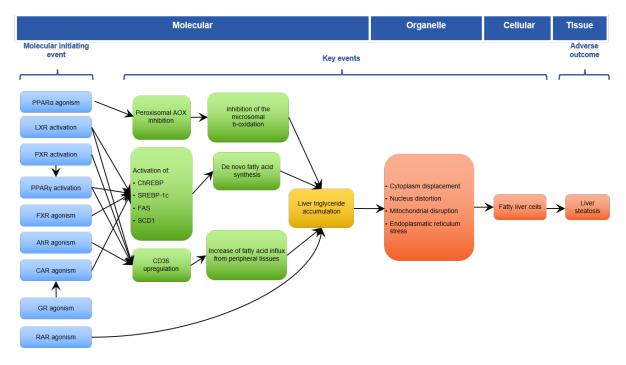


Figure 2.40: AOP network

2.5.3 Dose response data

Dose response data are data on response values of test systems at specified doses of substances (or mixtures of substances) from dose response experiments.

This module has as primary entities: *Substances Test systems Responses*

Output of this module is used by: Dose response models

Dose response data data formats

The meta-data of dose response experiments (such as name, description, etc.) are specified in the DoseResponseExperiments table.

For presenting the data of these experiments to the system, there are two formats: a single table format (DoseResponseData) and a relational data format (three tables DoseResponseExperimentDoses, ExperimentalUnitProperties, DoseResponseExperimentMeasurements). Usually, the single table format will be the easier one. For internal use in MCRA, this single table data is converted to the relational data format.

Dose response data

Dose response data are used to extract assessment group membership or hazard doses. The meta-data of dose response experiments (such as name, description, etc.) are specified in the DoseResponseExperiments table. For presenting the data of these experiments to the system, there are two formats: a single table format (DoseResponse-Data) and a relational data format (three tables). Usually, the single table format will be the easier one. For internal use in MCRA, this single table data is converted to the relational data format.

Dose response experiments

General information about the dose response experiments, such as the (unique) identifier, name, description, the used test-system, and the dose unit is stored in the table DoseResponseExperiments. If the data of an experiment is provided in a single table format, then the fields Time, Covariates, Substances, and Responses are used to map the column header names of the columns of the single data table to these their respective types.

Neme		e definition for DoseResponseExp		Deminut
Name	Type	Description	Aliases	Required
idExperiment	AlphaNumeric(50)	Unique identification code of	idExperiment,	Yes
<u></u>		the dose effect experiment.	Id, Code	NT.
Name	AlphaNumeric(100)	Name of the dose effect experiment.	Name	No
Description	AlphaNumeric(200)	Description of the dose effect experiment.	Description	No
Date	DateTime	The starting date of the experiment.	Date	No
Reference	AlphaNumeric(200)	External reference, for instance, to the experiment protocol and/or supporting material.	Reference	No
Experimental- Unit	AlphaNumeric(100)	The name of the experimental unit of the experiment, e.g., rat, cage, litter, vial, cup, petridish.	Experimental- Unit	No
DoseRoute	AlphaNumeric(100)	For in-vivo test systems, the route in which the dose was administered	DoseRoute	No
Substances	AlphaNumeric	Code or comma separated list of the codes of the substances measured in the experiment. E.g., 'Cyproconazole, Thiram'. Required when presenting the dose-response data in a single table. Make sure that in table DoseResponseData the column headers exactly match these names.	idSubstance, SubstanceId, SubstanceCode, Substance, idSubstances, SubstanceIds, SubstanceCodes, Substances	Yes
DoseUnit	DoseUnits	Unit of the doses administered in this experiment.	DoseUnit	Yes
Responses	AlphaNumeric	Code or comma separated list of codes of the responses measured in the experiment. E.g., 'AngleM_PQ, Mortality'. Required when presenting the dose-response data in a single table. Make sure that in table DoseResponseData the column headers exactly match these names.	Responses, Response, idResponses, idResponse	Yes
Time	AlphaNumeric(100)	Identifier of the time field of the experiment. Required when presenting the dose-response data in a single table and responses are measured at multiple times. Make sure that in the table DoseResponseData the column header of the time-column exactly matches this name.	Time, Times	No
TimeUnit	TimeUnit	Unit of the time scale used in	TimeUnit	No
Covariates	AlphaNumeric(200)	the experiments. Comma separated list of the names/codes of the covariates of the experiment. E.g. 'Gender, Inhibitor,	Covariates, Chapte Covariate	er _{Ro} Modu

Table 2.144: Table definition for DoseResponseExperiments.

Table aliases: DoseResponseExperiments, DoseResponseExperiment.

Dose response data

Single (two-way) table data format for specifying data of dose response experiments (as alternative for the relational format). The column headers are dynamic and should be defined in the table DoseResponseExperiments through fields Substances and Responses (and, optionally, Covariates and Time). For responses given as aggregated statistics, also SD, CV, N and Uncertainty are specified as [Datatype:Response]. E.g., 'SD:Y', 'CV:Y', 'N:Y'. Uncertainty upper 95%limits are specified as 'UncertaintyUpper:Y'. For each quantal response an additional column 'N:[responsename]'is required with binomial totals (e.g. Mortality = 3, N:Mortality = 10).

Name	Туре	Description	Aliases	Required
idExperiment	AlphaNumeric(50)	Unique identification code of the dose effect experiment.	idExperiment, Experiment, Code	No
Experimental unit	AlphaNumeric(50)	Experimental unit numbers or identifiers. The column name of the experimental unit should be as specified in the dose response experiment record.	Experimental- Unit, Experimental- Units, Experimental unit	No
Substance(s)	AlphaNumeric(100)	One or more columns with doses for each substance, in the unit as specified in the dose response experiment table. The column name(s) should match the substance codes listed in the comma-separated list of the substances field of the dose response experiment record.		Yes
Response(s)	AlphaNumeric(100)	One or more columns with results for each response, in the unit(s) as specified in the dose response experiment table. The column name(s) should match the response codes listed in the comma-separated list of the responses field of the dose response experiment record.		Yes
Time	Numeric	The column containing the observed response times. The column name (header) should match that of the Time column in the dose response experiment record.		No
Covariate(s)	AlphaNumeric(100)	The column(s) containing additional properties of the experimental unit. The column name (header) should match the codes of the comma-separated covariates list in the dose response experiment record.		No

Table 2.145:	Table definition	for DoseRes	sponseData.
14010 2.1 10.	raole acimition		poince aua.

Table aliases: TwoWayDoseResponseData, DoseResponseDataTwoWay, DoseResponseData.

Relational dose response data

In the relational data format, dose response experiment data are specified using the triplet of tables: DoseResponseExperimentDoses, DoseResponseExperimentMeasurements, and ExperimentalUnitProperties. These tables describe, respectively, the experiment designs (including the administered substance doses), the response measurements, and additional properties of the experimental units of the experiment.

Dose response experiment doses

The table DoseResponseExperimentDoses describes the experiment design, being a complete specification of which doses of which substances were applied to which experimental unit and if relevant at what time.

Name	Туре	Description	Aliases	Required
idExperiment	AlphaNumeric(50)	Identification code of the	idExperiment,	Yes
		experiment to which this	Experiment	
		design record belongs.		
idExperimental-	AlphaNumeric(50)	Identification code of the	idExperimental-	Yes
Unit		experimental unit to which	Unit,	
		the dose is applied.	Experimental-	
			Unit	
Time	Numeric	The time of administration of	Time	No
		the dose.		
idSubstance	AlphaNumeric(50)	Code of the substance that	idSubstance,	Yes
		was administered.	SubstanceId,	
			SubstanceCode,	
			Substance	
Dose	Numeric	The dose that was	Dose	Yes
		administered.		

Table 2.146: Table definition for DoseResponseExperimentDoses.

 $Table\ aliases:\ Dose Response Experiment Doses,\ Dose Response Experiment Dose.$

Experimental unit properties

The table ExperimentalUnitProperties are used to specify additional properties of the experimental units of the experiment. For instance, the gender of the rat, in case rats are the experimental units.

Name	Туре	Description	Aliases	Required
idExperiment	AlphaNumeric(50)	Identification code of the	idExperiment,	Yes
		experiment.	Experiment	
idExperimental-	AlphaNumeric(50)	Identification code of the	idExperimental-	Yes
Unit		experimental unit.	Unit,	
			Experimental-	
			Unit	
PropertyName	AlphaNumeric(50)	Name of the experimental	Property, Name	Yes
		unit property.		
Value	AlphaNumeric(100)	Value of the experimental	PropertyValue	No
		unit property.		
OtherProperty		Other properties of		No
		experimental units are		
		automatically parsed, using		
		the column name (header) as		
		property name.		

Table 2.147: Table definition for ExperimentalUnitProperties.

Table aliases: ExperimentalUnitProperties, ExperimentalUnitProperty.

Dose response experiment measurements

The table DoseResponseMeasurements describes the measurements that were done in the experiments. That is, for each response and experimental unit, at each observation time, one measurement should be recorded. If the response is an aggregated statistic, then this record may also include a standard deviation and number of units over which was aggregated.

Name	Туре	Description	Aliases	Required
idExperiment	AlphaNumeric(50)	Identification code of the experiment to which this measurement belongs.	idExperiment, Experiment	Yes
idExperimental- Unit	AlphaNumeric(50)	Identification code of the experimental unit from which the measurement is taken.	idExperimental- Unit, Experimental- Unit	Yes
idResponse	AlphaNumeric(50)	Identifier of the response that is measured.	idResponse, Response	Yes
Time	Numeric	Time of observation.	Time	No
ResponseValue	Numeric	The measured response.	ResponseValue, Value	Yes
SD:Response	Numeric	For aggregated responses, the standard deviation of the measurement.	SD:Response, ResponseSD	No
CV:Response	Numeric	For aggregated responses, the coefficient of variation (cv) of the measurement.	CV:Response, ResponseCV	No
N:Response	Numeric	For aggregated responses, the number of units over which was aggregated.	N:Response, ResponseN	No
Uncertainty- Upper:Response	Numeric	Optionally, measurement uncertainty quantification in terms of the upper value (i.e., an estimate of 95th percentile).	Uncertainty- Upper:Response, Response- Uncertainty- Upper, Uncertainty- Upper, Upper	No

Table 2.148:	Table definition for DoseResponseExperimentMe	easure-
ments.		

 $Table\ aliases:\ DoseResponseExperimentMeasurements,\ DoseResponseExperimentMeasurement,\ DoseResponseMeasurements,\ DoseResponseMeasurement.$

Dose response data settings

Selection settings

Name	Description
Experiments	The dose response experiments of interest.
Merge dose response data of	Specifies whether the dose response data of multiple experiments
multiple experiments	should be merged into one large dose response data set.

Dose response data as data

Dose response data are provided per experiment or study in which several responses (on in-vitro or in-vivo test systems) are measured from several administered substance doses.

• Dose response data data formats

2.5.4 Dose response models

Dose response models are models fitted to dose response data and can be provided as data or calculated using a local or remote version of PROAST. The main results for hazard and risk assessment are benchmark doses (BMDs), related to a specified substance, response, optionally covariate value, and the benchmark response (BMR). Dose response models can be uploaded as data, retrieved from PROASTweb through *linked remote repositories*, or *calculated using an internal version of PROAST*.

This module has as primary entities: Test systems Responses Substances

Output of this module is used by: Hazard characterisations

Dose response models data formats

Dose response models

Dose response models are specified using three tables: the dose response models table holds the dose response model definitions (id, name, description) and other information about the dose response models. The dose response model benchmark doses table records the benchmark doses and (optionally) the model parameters for specific substances and covariates. The dose response model benchmark doses uncertainty table records results from bootstrap runs for the benchmark doses per substance/covariate combination.

Dose response models

Each dose response model has a unique id, a name (optional), and description (optional). Also, each dose response model is associated with a specific dose response experiment (idExperiment) from which the data used to create the model is obtained, a response (idResponse), one or more substances, and, optionally, specific covariates considered by the dose response model. The combination of the benchmark response type and the associated value define the benchmark response of the model. The dose unit specifies the unit used for the doses, and if applicable, the model equation can be specified.

Name	Туре	Description	Aliases	Required
idDose- ResponseModel	AlphaNumeric(50)	The unique identification code of the fitted dose response model.	idDose- ResponseModel, idModel	Yes
idExperiment	AlphaNumeric(50)	The identification code of the experiment from the dose response model.	experiment- Code, experimentId	Yes
Name	AlphaNumeric(100)	The name of the dose response model.	Name	No
Description	AlphaNumeric(200)	Description of the dose response model.	Description	No
Substances	AlphaNumeric	Code or comma separated list of the codes of the substances in the Dose Response Model. E.g., 'Cyproconazole, Thiram'.	Substances	Yes
idResponse	AlphaNumeric	The response of the dose response model.	idResponse, Response	Yes
Covariates	AlphaNumeric	The covariates considered by the dose response model.	Covariates, Covariate	No
Benchmark- Response	Numeric	The value of the benchmark response or critical effect size.	Benchmark- Response, CriticalEffect- Size, CES	Yes
Benchmark- ResponseType	Benchmark- ResponseTypes	Specifies how the benchmark response is expressed. E.g., using a percent change in mean response or, for quantal response types, in terms of extra risk, additional risk, or ED50.	Benchmark- ResponseType, HazardEffect- SizeType, CriticalEffect- SizeType	No
LogLikelihood	Numeric	Loglikelihood of the model fit.	LogLikelihood	No
DoseUnit	AlphaNumeric(50)	The dose unit (if not specified, then mg/kg is assumed).	DoseUnit, UnitDose	No
ModelEquation	AlphaNumeric(500)	If available, the model equation of the dose response model (R model equation) or the identifier of the dose response model type.	ModelEquation, DoseResponse- ModelEquation, Equation	No

Table 2.150: Table definition for DoseResponseModels.

Table aliases: DoseResponseModels, DoseResponseModel.

Dose response model benchmark doses

The benchmark responses and benchmark doses belonging to the dose response models are recorded per substance/covariate in the dose response model benchmark doses table. Optionally, if the model equation of the dose response model has been specified in the dose response models table, the model parameter values for this specific substance/covariate can be specified here.

Name	Туре	Description	Aliases	Required
idDose-	AlphaNumeric(50)	The identification code of the	idDose-	Yes
ResponseModel		dose response model to which	ResponseModel	
		this record belongs.		
idSubstance	AlphaNumeric(50)	The code of the substance.	idSubstance,	Yes
			SubstanceId,	
			SubstanceCode,	
			Substance	
Covariates	AlphaNumeric(500)	Comma separated list of the	Covariates,	No
		covariate values for which this	Covariate	
		benchmark dose applies.		
Benchmark-	Numeric	The (nominal) benchmark	Benchmark-	Yes
Dose		dose (BMD).	Dose, BMD,	
			CED	
Benchmark-	Numeric	Benchmark dose lower	Benchmark-	No
DoseLower		uncertainty bound (BMDL).	DoseLower,	
			BMDL, CEDL	
Benchmark-	Numeric	Benchmark dose upper	Benchmark-	No
DoseUpper		uncertainty bound (BMDU).	DoseUpper,	
			BMDU, CEDU	
Model-	AlphaNumeric(500)	Parameter values for dose	ParameterValues	No
Parameter-		response models.		
Values				

Table aliases: DoseResponseModelBenchmarkDoses.

Dose response model benchmark dose bootstraps

Empirical uncertainty values of the benchmark benchmark doses of dose response models can be recorded in the dose response model benchmark doses bootstraps table. The uncertainty set identifier (idUncertaintySet) can be specified to retain correlations between uncertainty records that originate from the same bootstrap run.

Name	Туре	Description	Aliases	Required
idDose-	AlphaNumeric(50)	The identification code of the	idDose-	Yes
ResponseModel		dose response model to which	ResponseModel	
		this record belongs.		
idUncertainty-	AlphaNumeric(50)	The uncertainty set identifier.	idUncertainty-	Yes
Set			Set,	
			UncertaintyId	
idSubstance	AlphaNumeric(50)	The code of the substance.	idSubstance,	Yes
			SubstanceId,	
			SubstanceCode,	
			Substance	
Covariates	AlphaNumeric(500)	Comma separated list of the	Covariates	No
		covariate values for which this		
		benchmark dose applies.		
Benchmark-	Numeric	Benchmark dose (BMD).	Benchmark-	Yes
Dose			Dose, BMD,	
			CED	

Table 2.152: Table definition for DoseResponseModelBenchmarkDosesUncertain.

 $Table\ a liases:\ Dose Response Model Benchmark Doses Bootstraps,\ Dose Response Model Benchmark Doses Uncertain.$

Dose response models calculation

Besides uploading dose response models as data or retrieving them from PROASTweb through *linked remote repositories*, there is also a possibility to compute dose response models using an integrated version of PROAST. When computing dose response models using the integrated version, MCRA will attempt to fit a dose response model for each response of each dose response experiment. Depending on the type of data (e.g., response type, covariates y/n, single or multiple substances) a PROAST run is configured and executed. If *effect representations* are provided, then benchmark responses specified by the effect representations data are used, otherwise only the model fits will be computed without benchmark doses.

Dose response models as data

Dose response models as data contain the details of fitted dose response models. The main elements for hazard and risk assessment are the benchmark doses (BMDs) related to specified substances, responses, and optionally covariate values for specified benchmark responses (BMR). These specifications can be provided in data files or can be retrieved/imported from PROAST output files on the PROAST website https://proastweb.rivm.nl/user/login using a PROASTweb user account and an application access key.

• Dose response models data formats

Inputs used: Dose response data

Calculation of dose response models

Used as a calculator, dose response models are fitted to dose response data using an MCRA-internal version of PROAST. Currently, all available models appropriate for the response type will be fitted, and for the Hill and Exponential model families, the best fitting model based on maximum likelihood will be selected. The set of results for the calculation will include BMDs etc. for all fitted models.

• Dose response models calculation

Inputs used: Effect representations

2.5.5 Effect representations

Effect representations specify the responses that can be used to measure specified effects and which response levels, the benchmark response (BMR), define the hazard limits for the effects.

This module has as primary entities: Effects Responses

Output of this module is used by: Hazard characterisations Dose response models

Effect representations data formats

Effect representations

Effect representations specify responses that may represent the effect.

Effect representations

One response can be set as the canonical response (golden standard). For a quantitative or stochastically qualitative canonical response a benchmark response should be defined.

Name	Туре	Description	Aliases	Required
idEffect	AlphaNumeric(50)	Identifier of the effect	idEffect	Yes
idResponse	AlphaNumeric(50)	Identifier of the response	idResponse	Yes
Benchmark-	Numeric	The threshold response value	BenchMark-	No
Response		that defines a hazard. For	Response,	
		numeric responses	HazardEffect-	
		(Continuous, Quantal, Count)	Size, BMR,	
		the value that defines a	CriticalEffect-	
		hazard. For Binary responses	Size, CES	
		1 defines a hazard by default,		
		unless redefined here.		
Benchmark-	Benchmark-	Specifies how the	Benchmark-	No
ResponseType	ResponseTypes	BenchMarkResponse is	ResponseType,	
		expressed, relative to the	HazardEffect-	
		response at zero dose, or	SizeType,	
		absolute. Required for	CriticalEffect-	
		numeric response types	SizeType	
		(Continuous, Quantal,		
		Count). For qualitative		
		responses (Ordinal,		
		Categorical) Absolute is used.		

Table aliases: EffectRepresentations, EffectRepresentation.

Effect representations as data

Effect representations are provided as data in the form of specified combinations of effect and response, optionally with a benchmark response that defines a hazard limit for the effect.

• Effect representations data formats

Inputs used: AOP networks

2.5.6 Hazard characterisations

Hazard characterisations are reference exposure values for active substances at the chosen biological target level (external or internal). Hazard characterisations may be specified for specific effects or for the critical effect as defined in hazard characterisation. Hazard characterisations are specified as external values (e.g. human based guidance values, such as ADI or ARfD) or are based on points of departure, such as BMDs from dose-response models or externally specified points of departure (NOAEL, LOAEL, MDS). The computation may involve assessment factors, e.g. for inter-species conversion, intra-species variation or additional sources of uncertainty. The calculation may also use kinetic models or absorption factors to convert external doses to internal doses or vice versa.

This module has as primary entities: Substances Effects

Output of this module is used by: Active substances Relative potency factors Risks Single value risks

Hazard characterisations data formats

Hazard characterisations

Hazard characterisations provide reference threshold values associated with the hazard of interest. Examples are health-based guidance values such as ADI or ARfD, and points of departure such as BMD or NOAEL.

Hazard characterisations

Hazard characterisations are specified for combinations of hazard characterisation type, effect, substance, population type, target level, and exposure route (for external) or target organ (for internal). Effects can be specific, but can also be labelled as being the critical effect and used as such if this has been specified in the hazard characterisation settings.

Name	Туре	Description	Aliases	Required
idHazard-	AlphaNumeric(50)	Id of the hazard	id, idHazard-	Yes
Characterisation		characterisation.	Characterisation	
idEffect	AlphaNumeric(50)	Code of the (critical) effect	idEffect,	No
		linked to this hazard	EffectId, Effect	
		characterisation.		
idSubstance	AlphaNumeric(50)	The code of the substance.	idSubstance	Yes
idPopulation-	AlphaNumeric(50)	The code of the population	idPopulation-	No
Туре		type for which this reference	Туре	
		value is defined. If not		
		specified, PS06A, Consumers		
		is assumed.		
TargetLevel	TargetLevelType	The target level. I.e., internal	TargetLevel	No
		or external. If omitted,		
		external is assumed		
ExposureRoute	ExposureRouteTypes	The exposure route (only	ExposureRoute	No
		applicable if target level is		
		external). If not specified,		
		Dietary is assumed.		
TargetOrgan	AlphaNumeric(50)	The target organ (should be		No
		specified when target level is		
		internal).		
IsCriticalEffect	Boolean	Specifies whether this value is	IsCriticalEffect	No
		the value associated with the		
		critical effect. If omitted, No		
		is assumed		
ExposureType	ExposureTypes	The exposure type associated	ExposureType	Yes
Exposurerype	Exposition ypes	with the hazard	Exposurerype	105
		characterisation (i.e., chronic		
		or acute).		
Hazard-	Hazard-	The type of the hazard	Hazard-	Yes
Characterisation-	Characterisation-	characterisation (e.g., ARfD,	Characterisation-	103
Туре	Types	ADI, NOAEL, BMD).	Туре	
Qualifier	QualifierType	Qualifier of the hazard	QualifierType	No
Quanner	Quanner Type	characterisation value, e.g.	QuannerType	INO
		equal-to (=) or smaller-than		
Value	Numeric	(<). If omitted, = is assumed.Reference value that	Value, Hazard-	Yes
value	INUITIENC	characterises the hazard.	Characterisation-	105
		characterises the hazard.		
DoseUnit	DoseUnits	Unit of the hazard	Value DoseUnit, Unit	Yes
DoseOnit	DoseOnus		DoseOnit, Onit	ies
	AlahaNI and (70)	characterisation value.		N.
idPointOf-	AlphaNumeric(50)	The code of the point of	idHazardDose,	No
Departure		departure from which this	idPod	
		hazard characterisation was		
0 1: 1		derived.		N
Combined-	Numeric	Combined assessment factor	Combined-	No
Assessment-		(includes, e.g., safety factor,	Assessment-	
Factor		but also other extrapolation	Factor	
		factors that may be used to		
		derive the hazard		
		characterisation from the		
		underlying PoD).		
PublicationTitle	AlphaNumeric	Title of the publication of the	PublicationTitle,	No
		study in which this hazard	Title	
		characterisation was		
		established.		
Publication- Hazard modu Authors	AlphaNumeric	Author(s) of the publication	Publication-	No
Authors	lies	of the study in which this	Authors,	
		hazard characterisation was	Publication-	
		established.	Author, Author,	

Table 2.154: Table definition for HazardCharacterisations.

Table aliases: HazardCharacterisations.

Hazard characterisations calculation

Hazard characterisations are defined as deterministic threshold values (e.g. ADI, ARfD) or as distributions (using probabilistic models). They are linked to an effect of interest or alternatively are defined for the critical effect. Hazard characterisations depend on the *risk type* (acute or chronic) and the biological *target level* of the human body (external via some route of exposure or internal for a specific defined organ or compartment). Hazard characterisations are derived from *points of departure* provided as data and/or from *dose-response models*. The procedure for computing hazard characterisations has two main phases: 1) collection of all available hazard characterisations and imputation of missing hazard characterisations.

Collection of available hazard characterisation candidates involves collecting the appropriate points of departure data and/or dose-response models that are used for deriving the hazard characterisations. In MCRA, a distinction is made between three *methods for computing hazard characterisations*:

- 1. Calculation of hazard characterisations from externally specified in-vivo points of departure (PoD, e.g. BMD, NOAEL, LOAEL).
- 2. Calculation of hazard characterisations from PoDs (in this case BMD) calculated from dose response data.
- 3. Calculation of hazard characterisations based on an *in-vivo PoD for the index substance and in-vitro RPFs from dose-response models for the other substances (IVIVE model)* (cumulative assessments only).

For all three methods, the collected points of departure and benchmark doses should be aligned with the target system. This alignment may involve various conversion steps for each point of departure and specific substance, and can be formally specified as:

$$HC = f_{\text{expression-type}} \cdot f_{\text{kinetic}} \cdot \frac{1}{f_{\text{inter-species}}} \cdot \frac{1}{f_{\text{intra-species}}} \cdot \frac{1}{f_{\text{additional}}} \cdot PoD$$

where:

- *HC* denotes the hazard characterisation.
- $f_{\text{expression-type}}$ denotes the *expression type correction factor*, e.g., for extrapolation from LOAEL or NOAEL, or from NOAEL to BMD.
- *f*_{kinetic} denotes the kinetic conversion factor for *conversion from internal to external or external to internal hazard characterisations*.
- *f*_{inter-species} denotes the inter-species factor for *extrapolation from animal to human (inter-species)*.
- $f_{intra-species}$ denotes the intra-species factor for *extrapolation from the average to the sensitive human or* probabilistic calculation of the distribution of human individuals (intra-species).
- $f_{additional}$ denotes the additional assessment factor for *extrapolation from the POD to the hazard characterisation in humans for sources where appropriate data or information is scarce or missing (additional).*
- PoD denotes the point of departure.

Note that inter- and intra-species extrapolation and the use of an additional assessment factor are optional. However, expression type correction and the kinetic conversion are always applied (when relevant) whatever option is chosen.

Occasionally, for some substances multiple hazard characterisations are available (e.g., obtained from multiple experiments) and for others substance hazard characterisations are still missing. Hence, two final steps remain to come to the final set of hazard characterisation:

- Aggregation over multiple available hazard characterisations. Set the selection method in case of multiple candidate hazard characterisations from MostToxic to Aggregate.
- Imputation of missing hazard characterisations. Check the option Imput missing hazard characterisations, to select the Imputation method.

Hazard characterisation type extrapolation

Hazard doses, or points of departure can be of *various types*. E.g., BMDs, NOAELs, or LOAELs. When computing hazard characterisations, the type in which the hazard characterisations are expressed (i.e., the *hazard characterisation expression type*) should be specified explicitly. When points of departure from types different from the expression type are provided, these should be translated to the specified expression level. In the current implementation, the simple conversion factors shown in Table 2.155 are used, roughly based on the WHO guidance document on evaluating and expressing uncertainty in hazard characterization [[WHO, 2018]].

From	То	Conversion factor
BMD	NOAEL	1/3
BMD	LOAEL	1
NOAEL	BMD	3
NOAEL	LOAEL	1/3
LOAEL	BMD	1
LOAEL	NOAEL	1/3

Table 2.155: Conversion factors for hazard characterisation types.

Inter-species extrapolation

Hazard doses, or points of departure, are commonly only determined for animals, not for humans. In order to derive hazard characterisations for humans, the animal hazard doses need to be converted to toxicologically equivalent doses for humans. This extrapolation is usually expressed as a multiplication factor, and traditionally a factor of 10 is used (which is roughly obtained from the product of a factor of 3.2 for toxicokinetic variability and a factor 3.2 for toxicodynamic variability).

The following methods are available within the toolbox:

- 1. No inter-species extrapolation: Assume that for all available points of departure, the animal hazard dose is equal to the human hazard dose. Effectively, this is equivalent to using a conversion factor of 1.
- 2. **Default distribution:** Use a conversion factor drawn from a default, substance and species independent lognormal uncertainty distribution specified (as *model settings*) by a geometric mean (GM) and geometric standard deviation (GSD). In the *nominal run*, the nominal value of this distribution (i.e., the geometric mean) is used as a conversion factor. In the *uncertainty analysis loop*, provided that inter-species extrapolation uncertainty is *included in the uncertainty analysis*, a single factor is drawn from the lognormal distribution.
- 3. Substance/species specific distributions: Use conversion factors drawn from substance/species specific lognormal uncertainty distributions specified (as *data*) by a geometric mean (GM) and geometric standard deviation (GSD). In the *nominal run*, a factor equal to the geometric mean is used for all combinations of substance and species. In the *uncertainty analysis loop*, provided that inter-species extrapolation uncertainty is *included in the uncertainty analysis*, a uncertainty factor is drawn from the lognormal distribution with $\mu = 0$ and $\sigma^2 = 1$, which is used to obtain correlated draws for all available inter-species conversion factor distributions. If the distribution parameters are missing for a specific substance/species, then the default distribution is used as a fallback.

Intra-species extrapolation of hazard characterisations

There is variation between individuals concerning their individual sensitivities to experience health effects. In some scenarios the aim is to perform assessments for the sensitive individuals instead of the average individuals for which the points of departure are derived. If this is the case, then extrapolation is required to translate hazard characterisations derived for the average individual to hazard characterisations for a sensitive individual. In traditional exposure assessments, a safety of 100 is commonly used as a margin of safety, that is assumed to be composed of a interspecies extrapolation factor (factor 10), and inter-individual extrapolation factor (factor 10). I.e., the hazard characterisation defined for the sensitive individual is defined as

$$HC_{\rm sens} = \frac{1}{f_{\rm intra-species}} \cdot HC_{\rm avg}$$

Here $f_{inter-species}$ denotes the intra-species factor. An alternative to using a fixed safety factor, is to define intraspecies variability may be explicitly *a lognormal distribution*, characterised by a geometric mean (GM) equal to 1 and a geometric standard deviation (GSD). For *risks calculations*, this distribution could be used to sample individual hazard characterisations. This effectively converts the description of hazard characterisations to include variability, with an unbiased central value.

Additional assessment factors

In cases where where appropriate data or information is scarce or missing, an additional assessment factor is used for extrapolation from the POD to the hazard characterisation in humans. This factor can be used as a worst case value (preferably AF = 100) for inter- and intra-species extrapolation, but it may equally well serve as an additional extrapolation factor next to inter- and/or intra-species extrapolation. In the latter case the factor merely serves to account for differences in for example route to route, metabolic rates in interspecies, completeness and consistency of available data, reliability of alternative data (e.g. read-across) or quality of data in general.

In-vitro in-vivo extrapolation (IVIVE)

The in-vitro in-vivo extrapolation method implemented in MCRA is based on the following prerequisites:

- 1. For one substance, the index substance, a reliable point of departure is available for the adverse outcome of interest obtained from an in-vivo assay (i.e., external dose).
- 2. There are other substances for which there is a dose-response model available from an in-vitro assay on a response representing an early key event of the adverse outcome for these substances and the index substance.

In IVIVE, these RPFs, in combination with the known hazard characterisation of the index substance, can be used to derive hazard characterisations for the other substances as well. Figure 2.41 shows the conceptual model that forms the basis of the IVIVE methodology of MCRA.

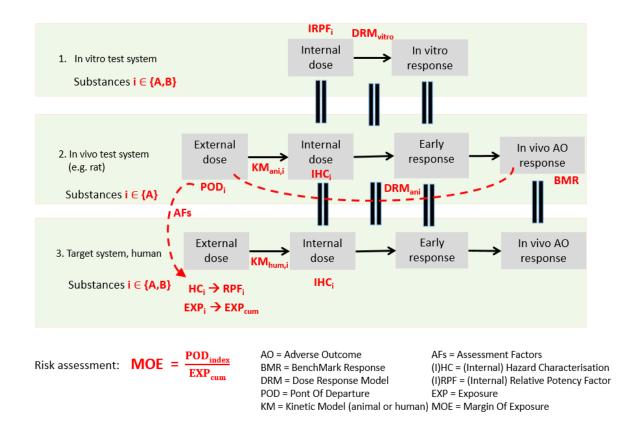


Figure 2.41: Conceptual model IVIVE.

IVIVE for calculating internal hazard characterisations

- 1. Translate the (external) PoD of the index substance substance to an internal hazard characterisation for the human target system/compartment.
- 2. If the RPFs are obtained are obtained using mol-based specification of the doses, then convert the mol-based RPFs to mass-based RPFs. I.e.,

$$RPF_{mass-based,i} = RPF_{mol-based,i} \cdot \frac{MW_{ref}}{MW_i}$$

3. Derive internal hazard characterisations for the human target system for the other substances by multiplying the RPF obtained from dose-response modelling with the hazard characterisation of the index substance. I.e.,

$$HC_i = HC_{ref} \cdot RPF_{mass-based,i}$$

IVIVE for calculating external hazard characterisations

- 1. Translate the PoD of the index substance to an external human hazard characterisation (dietary/oral exposure route).
- 2. Derive an internal hazard characterisation for the index substance, with an target organ/compartment representative for the response of the dose-response model.
- 3. If the RPFs are obtained are obtained using mol-based specification of the doses, then convert the mol-based RPFs to mass-based RPFs.
- 4. Derive internal hazard characterisations for the human target system for the other substances by multiplying the RPF obtained from dose-response modelling with the hazard characterisation of the index substance.
- 5. Convert the internal hazard characterisations of the other substance to external hazard characterisations for the dietary/oral exposure route using.

Kinetic conversion of hazard characterisations

When the *hazard characerisation level* is internal and points of departure are available for external exposures (e.g., NOAELs from in-vivo animal studies) or when the hazard characterisation level is external and benchmark doses are available at the internal level, then *kinetic conversion models* are needed to *translate the external doses to equivalent internal doses at the target compartment/organ* of interest or *vice-versa*.

In the toolbox, this alignment from internal to external or from external to internal is generally termed *kinetic conversion*, associated with a *kinetic conversion factor*. The kinetic conversion factor is a multiplication factor needed to obtain a hazard characterisation on the target level from a hazard characterisation of the point of departure or benchmark dose. Depending on the chosen kinetic modelling tier, this kinetic conversion factor may be 1) assumed to be one, 2) derived from absorption factors, or 3) derived using PBPK models.

An important detail in the use of kinetic conversion factors for computing hazard characterisations is the order between kinetic conversion and inter-species extrapolation. Notice that when points of departure are determined for animals, a choice should be made regarding the order of inter-species extrapolation and kinetic modelling. That is, one may first choose to convert animal external point of departure to an internal hazard characterisation for that animal, using the available animal kinetic model. Alternatively, one may first extrapolate the animal external point of departure to a human external hazard characterisation, and thereafter apply the human kinetic model to obtain internal hazard characterisations. In the toolbox, only the latter approach is currently implemented.

Extrapolation from external to internal hazard characterisations

The calculation of internal hazard characterisations based on external hazard characterisations is similar to the procedure for *computing internal exposures*. In the simplest tier, equivalence can be assumed between internal and external hazard characterisations, and in higher tiers absorption factors, respectively PBPK models can be used.

Calculation of internal doses using absorption factors

In the simplest form, internal doses are obtained from external exposure concentrations using multiplication factors (or, absorption factors) that can be specified by substance and by route. That is, for a given substance, the internal hazard characterisation HC_{int} can be derived from an external hazard characterisation HC_{ext} as

$$HC_{\text{int}} = f_{\text{abs},r} \cdot HC_{\text{ext},r}$$

Here, r denotes the route of the external exposure HC_{ext} , and $f_{abs,r}$ denotes the absorption factor of route r for the specified compartment. Note that this model assumes that the external hazard characterisations are specified as concentrations (i.e., substance amount divided by the body weight).

Calculation of internal doses using human PBPK models

A more detailed alternative to using absorption factors is to use one of the *advanced PBPK models* available in MCRA. In this approach, for each substance independently, an external exposure equivalent to the dose of the external hazard characterisation is presented to a representative simulated individual for a number of simulated days to the PBPK model of the individual. This representative individual should represent the "average" individual of the population, with nominal physiological properties (e.g., an average bodyweight of 70kg). This yields a time course of the internal substance amount at the specified target compartment/organ from which a long term average substance amount (chronic) or peak substance amount (acute) can be obtained. By dividing this substance amount by the weight of the compartment, an internal concentration is obtained, which then represents the internal hazard characterisation.

More details on computing internal doses from external doses can be found in the description of the *calculation of internal exposures from external exposures*. For both tasks, the procedure for computing internal exposures/doses is exactly and the same *kinetic model settings*, such as *dosing patterns* and *non-stationary period* period apply for calculation of internal hazard characterisations as well.

Calculation of internal doses using animal PBPK models

In the above methods, the assumption is that the external points of departure (often obtained from experiments on animals) are first converted to external hazard characterisations for humans, and a human kinetic model is used for obtaining the internal hazard characterisations. As mentioned, an alternative approach is to use first the animal PBPK models to derive an internal hazard characterisation specific for the tested animal species and thereafter extrapolate to humans. When there are more precise kinetic models available for the animal used in the experiments for obtaining the point of departure, this could be a preferred path.

Note: Notice that this procedure is not yet implemented.

Extrapolation from internal to external hazard characterisations

In some cases, hazard characterisations are available at the internal level whereas the specified *hazard characerisation level* is external. This situation may occur, for instance, in *in-vitro in-vivo extrapolation (IVIVE)*. In this case, conversion is needed from the internal level to the external level, where the external level is implicitly defined as coming from the dietary/oral route of exposure.

When using absorption factors, the external (dietary) hazard characterisation of a substance is simply computed by dividing the internal hazard characterisation by the dietary absorption factor. I.e.,

$$HC_{\text{ext,diet}} = \frac{HC_{\text{int}}}{f_{\text{abs.diet}}}$$

When using PBPK models, reverse dosimetry is needed to find for the available internal hazard characterisation, the corresponding external (dietary) doses that yield the internal concentrations specified by the internal hazard characterisation. In MCRA, this is done using a bisection method, in which external doses are systematically fed to the PBPK model in order to converge to an external dose that yields the specified internal hazard characterisation with some level of precision.

Hazard characterisation imputation

In some cases it may be that there are substances that are known to cause (or may possibly cause) the effect of interest, but for which there are no data available for obtaining hazard characterisations. I.e., for these substances, there are no points of departure or dose response models. Instead of excluding these substances in quantitative analyses, it is also possible to impute hazard characterisations for these substances based on hazard characterisations of other (similar) substances, and use these for calculating, e.g., relative potency factors or for risk assessment.

Munro P5 (TTC approach)

The Threshold of Toxicological Concern (TTC) is an example of a tier for extrapolation of hazard characterisations from other substances that is already in common use (see [[Munro et al., 1996]]). The *Munro collection of NOELs/LOAELs* is a collection of NOELs/LOAELs for chemicals for the critical (i.e., first occurring) effect. In the TTC approach, the toxicity of an unknown substance is, depending on its Cramer class (see [[Cramer et al., 1976]]), imputed by the 5th percentile NOAEL of the sub-collection of the corresponding Cramer class.

Two variations of this approach are to use the empirical NOAEL distribution itself (just sample from the NOAEL data), or to fit a distribution (e.g. lognormal) to the empirical data and sample from the parametric distribution. MCRA provides an implementation of the TTC approach that uses the empirical distribution. In the nominal run, this implementation imputes the hazard characterisations with a value equivalent to the TTC. In the uncertainty runs, NOAELs are sampled from the empirical distribution.

The TTC is a conservative estimate of the NOAEL for at least two reasons:

- 1. TTCs are calculated from a collection of NOELs for the critical (i.e., first occurring) effect within each study and often the effect of interest will not be the critical effect, and therefore higher NOAELs are expected.
- 2. The TTC is a low percentile and therefore a conservative estimate for a random class member with unknown NOAEL.

Munro central value

To avoid the conservatism of taking the 5th percentile in the Munro P5 approach, alternatively, a nominal (or central) value could be taken from the Munro collection for each Cramer class. For a nominal run without uncertainty, the expected contribution of a substance with missing hazard characterisation to the risk as quantified in the hazard index is obtained from

$$HI = SF \cdot \sum_{i}^{n} \frac{\exp_{i}}{HC_{i}}$$

Here SF are all combined safety factors. It follows from this equation that an unbiased estimate for the contribution from a substance with missing hazard characterisations is obtained by taking the harmonic mean from the available NOAELs:

$$NOAEL = \left(\sum_{i=1}^{n} \frac{1}{NOAEL_i}\right)^{-1}$$

This is the value to use in a nominal run without uncertainty for the Munro central value approach. For the uncertainty runs, this approach also uses random sampling from the empirical distribution of the corresponding Cramer class.

Available hazard characterisations distribution P5

Another conservative aspect of the TTC approach is the fact that the Munro set lists NOELs/LOAELs for critical effects, not for the specific effect under study. Therefore an alternative is to use the effect-specific hazard characterisations of the substances for which these are available. This collection will have on average higher NOAELs than those of the Munro NOEL collection, because for many substances, the effect of interest will not be the critical effect.

Available hazard characterisations distribution central value

Similar to the Munro central value approach, a central value could also be obtained from the set of effect-specific hazard characterisations distribution for imputation of hazard characterisations. This approach may yield the most realistic, or unbiased imputation value for missing hazard characterisations.

Aggregation over multiple available hazard characterisations

In some scenarios, it may be that for a given substance and effect there are multiple available hazard characterisations. This can happen, for instance, if there are two different NOAELs originating from different studies. In such cases, a single hazard characterisation should be derived from the available candidates.

A conservative approach is to choose the lowest hazard characterisation (HC) of the available hazard characterisations. I.e.,

$$\mathrm{HC} = \min_{i=1,\dots,n} \mathrm{HC}_i$$

Alternatively, it is possible to aggregate the candidates into a new "average" hazard characterisation. For this, the harmonic mean, also used for obtaining central value estimates in the *imputation of missing hazard characterisations*, is a suitable approach.

$$\mathrm{HC} = \left(\sum_{i}^{n} \frac{1}{\mathrm{HC}_{i}}\right)^{-1}$$

Hazard characterisations settings

Selection settings

Table 2.156: Selection settings for module Hazard characterisations.

Name	Description
Risk type	The type of exposure considered in the assessment; acute (short
	term) or chronic (long-term).
Target level	Select to express hazard characterisations at external or internal
	exposure level.
Consider critical effect	Specifies whether the analysis should look at critical effects such
	as specified in the Hazard characterisation data source.

Calculation settings

Name	Description
Method	Choose method for computing the hazard characterisations: from
	in-vivo or in-vitro points of departure or both.
Expression type	Specifies how hazard characterisations are expressed: as BMD, as
	NOAEL, or the expression type is ignored.
Selection method in case of	Choose either the most toxic (default) or an aggregated hazard
multiple candidate hazard	characterisation when in nominal runs there are multiple available
characterisations	candidates. In uncertainty runs, multiple candidates are
	resampled.
Impute missing hazard	If checked, missing hazard characterisations are imputed based on
characterisations	Munro NOELs or on other available points of departure.
Imputation method	Imputation of Hazard characterisations: use low percentile (P5) or
	unbiased central estimate from either the Munro set or the
	available POD collection.
Use BMDs from dose response	If checked, preferably BMDs from dose response models will be
models	used. If these data are not available, other POD data are used.
Use inter-species conversions	Use inter-species conversion factors (default value, e.g. 10, or
	data).
Use intra-species factors	Use intra-species conversion factors (default value, e.g. 10, or
	data).
Use additional assessment	Use additional assessment factor for extrapolation of PODs to
factor	(human) hazard characterisations.
Include dietary and non-dietary	Specifies whether the assessment involves both dietary and
routes of exposure	non-dietary (oral, inhalatory or dermal) routes of exposure.

Table 2.157: Calculation settings for module Hazard characterisations.

Uncertainty settings

Table 2.158: Uncertainty settings for module Hazard characterisations.
--

Name	Description
Resample intra-species factor	Specifies whether intra-species factors are resampled from a
	parametric uncertainty distribution.
Resample hazard	Specifies whether to resample the hazard characterisations or
characterisations or RPFs	relative potency factors. Requires hazard characterisation or RPF
	uncertainty to be quantified in DoseResponseModelsUncertain or
	RelativePotencyFactorsUncertain tables.

Hazard characterisations as data

Hazard characterisations can be provided as data e.g., in the form of ADI or ARfD.

• Hazard characterisations data formats

Inputs used: AOP networks Active substances Points of departure

Calculation of hazard characterisations

Hazard characterisations can be computed from points of departure. The computation may involve assessment factors, e.g. for inter-species conversion, intra-species variation or additional sources of uncertainty. The additional assessment factor can be used to bypass inter- and intra species conversion, or as an additional assessment factor to account for extrapolation for sources where appropriate data or information is scarce or missing (e.g. to implement a mixture assessment factor). The hazard characterisation calculation may also use kinetic models or absorption factors to convert external doses to internal doses or vice versa.

• Hazard characterisations calculation

Inputs used: Dose response models Effect representations Inter-species conversions Intra species factors Kinetic models

Settings used

• Calculation Settings

2.5.7 Inter-species conversions

Inter-species conversions specify how to convert a hazard characterisation for a given species to a hazard characterisation for humans. In the simplest approach, this specifies a fixed inter-species factor. In a higher tier, this specifies a geometric mean (GM) and geometric standard deviation (GSD) for a lognormal uncertainty distribution of the interspecies factor. Inter-species conversion are specified per effect and can be general or substance-specific.

This module has as primary entities: Substances Effects

Output of this module is used by: Hazard characterisations

Inter-species conversions data formats

Inter-species conversions

Inter-species conversion models specify how to convert a hazard dose for a given species to a hazard dose for humans.

Inter-species model parameters

Inter-species extrapolation factors are described using a lognormal distribution specified by a geometric mean (GM) and geometric standard deviation (GSD). Inter-species factors are defined for an effect and a species and may optionally be specified specifically for a substance.

Name	Туре	Description	Aliases	Required
idEffect	AlphaNumeric(50)	The code of the effect for which this inter-species model is defined.	idEffect, EffectId, Effect	Yes
idSubstance	AlphaNumeric(50)	The code of the substance for which this inter-species model is defined.	idSubstance, SubstanceId, SubstanceCode, Substance	No
Species	AlphaNumeric(50)	Species	Species	Yes
InterSpecies- GeometricMean	Numeric	Interspecies geometric mean.	InterSpecies- GeometricMean, InterSpeciesGM	Yes
InterSpecies- Geometric- Standard- Deviation	Numeric	Interspecies geometric standard deviation.	InterSpecies- Geometric- Standard- Deviation, InterSpeciesGS- D	Yes
Standard- HumanBody- Weight	Numeric	The standard human body weight.	Standard- HumanBody- Weight	Yes
HumanBody- WeightUnit	AlphaNumeric(50)	The unit of the human body weight specification (kg is assumed if not defined).	HumanBody- WeightUnit	No
Standard- AnimalBody- Weight	Numeric	The standard animal body weight.	Standard- AnimalBody- Weight	Yes
AnimalBody- WeightUnit	AlphaNumeric(50)	The unit of the animal body weight specification (kg is assumed if not defined).	AnimalBody- WeightUnit	No

Table aliases: InterSpeciesModelParameters, InterSpeciesModelParameter, InterSpeciesFactors, InterSpeciesFactor.

Inter-species conversions settings

Selection settings

Name	Description
Interspecies factor geometric	Interspecies factor geometric mean.
mean	
Interspecies factor geometric	Interspecies factor geometric standard deviation.
standard deviation	

Table 2.160: Selection settings for module Inter-species conversions.

Uncertainty settings

Table 2.161: Uncertainty settings for module Inter-species conversions.

Name	Description	
Resample inter-species factor	Specifies whether inter-species factors are resampled from a	
	parametric uncertainty distribution.	

Inter-species conversions as data

Data are provided in the form of a geometric mean (GM) and geometric standard deviation (GSD)

• Inter-species conversions data formats

Inputs used: Active substances

2.5.8 Intra species factors

Intra-species factors specify how to convert a hazard characterisation from the average to a sensitive human individual.

This module has as primary entities: Substances Effects

Output of this module is used by: Hazard characterisations

Intra-species factors data formats

Intra-species factors

Intra-species factors.

Intra-species model parameters

Intra species factors.

Name	Туре	Description	Aliases	Required
idEffect	AlphaNumeric(50)	The effect code.	idEffect,	Yes
			EffectId, Effect	
idSubstance	AlphaNumeric(50)	The code of the substance.	idSubstance,	No
			SubstanceId,	
			SubstanceCode,	
			Substance	
IntraSpecies-	Numeric	The lower variability factor.	IntraSpecies-	No
Lower-		The lower and upper factor	LowerVariation-	
VariationFactor		are used to derive a geometric	Factor	
		standard deviation (gsd) and		
		degrees of freedom (df).		
IntraSpecies-	Numeric	The upper variability factor.	IntraSpecies-	Yes
UpperVariation-		The lower and upper factor	UpperVariation-	
Factor		are used to derive a geometric	Factor	
		standard deviation (gsd) and		
		degrees of freedom (df).		
idPopulation	AlphaNumeric(50)	Unique identification code of	IdPopulation,	No
		the population.	PopulationId	

Table 2.162: Table definition for IntraSpeciesModelParameters.

Table aliases: IntraSpeciesModelParameters, IntraSpeciesModelParameter, IntraSpeciesFactors, IntraSpeciesFactor.

Intra species factors settings

Selection settings

Table 2.163:	Selection	settings f	or module	Intra specie	s factors.
14010 211001		o o com Bo i	01 1110 4410	mina opeere	0 14000101

Name	Description
Intra-species factor	Intra-species factor.

Intra species factors as data

In the simplest approach, intra-species factors are fixed factors. In a higher tier, lower and upper values for the intraspecies factor are used to derive a variability distribution (log-normal around 1) and an uncertainty distribution for the geometric standard deviation related to human variability in sensitivity.

• Intra species factors data formats

Inputs used: Active substances

2.5.9 Points of departure

Externally specified points of departure can be used as an alternative to calculated BMDs from dose response models. Points of departure can be of various types, such as NOAEL, LOAEL or BMD. They can be used to construct the list of active substances, to derive relative potency factors, and to perform health impact assessments.

This module has as primary entities: Substances Effects

Output of this module is used by: Active substances Hazard characterisations

Points of departure data formats

Points of departure

Points of departure, such as NOAELS and BMDs, describe the critical/reference levels of substance dose in relation to the presence or absence of an effect. If available, the uncertainty of externally specified points of departure can be specified with uncertainty sets (empirical distributions representing possible values) in the points of departure uncertainty table.

Points of departure

Nominal points of departure should be presented in this table. Each point of departure should be linked to an effect using the effect code (idEffect) and to substances using the substance code (idSubstance).

Name	Туре	Description	Aliases	Required
idModel	AlphaNumeric(50)	The dose response model code.	idDose- ResponseModel, idModel, idPod, idPointOf- Departure, Pod, PointOf-	No
idEffect	AlphaNumeric(50)	The effect code.	Departure idEffect, EffectId, Effect	Yes
idSubstance	AlphaNumeric(50)	The code of the substance.	idSubstance, SubstanceId, SubstanceCode, Substance	Yes
Species	AlphaNumeric(50)	The species used to obtain this point of departure.	Species, System	No
Point of departure	Numeric	Point of departure, can be of various types, e.g. NOAEL, LOAEL, BMD, CED	PointOf- Departure, LimitDose, HazardDose, Value, CED	Yes
Point of departure type	HazardDoseTypes	The type of the point of departure: e.g. NOAEL, LOAEL, BMD (default).	PODType, HazardDose- Type, LimitDoseType	No
DoseUnit	AlphaNumeric(50)	The dose unit (if not specified, then mg/kg is assumed).	DoseUnit, UnitDose	No
Benchmark response (BMR)	AlphaNumeric(100)	The effect size.	Benchmark- Response, CriticalEffect- Size, HazardEffect- Size	No
ExposureRoute	AlphaNumeric(100)	The route of dose administration used in the study to obtain this point of departure. If not specified exposure route = Dietary is assumed.	ExposureRoute, RouteExposure	No
IsCriticalEffect	Boolean	Specifies whether this value is the value associated with the critical effect. If omitted, No is assumed	IsCriticalEffect	No

Table 2.164:	Table def	inition for	HazardDoses.
--------------	-----------	-------------	--------------

Table aliases: PointsOfDeparture, PointOfDeparture, HazardDoses, HazardDose.

Points of departure uncertainty

Often, the PODs found for a substance/effect combination are uncertain. This table facilitates in specifying the POD uncertainty in the form of a set of uncertainty values that may additionally be specified for a substance/effect combination.

Name	Туре	Description	Aliases	Required
idDose-	AlphaNumeric(50)	The dose response model	idDose-	Yes
ResponseModel		code (must correspond to	ResponseModel	
		values in id column of		
		DoseResponseModels table).		
idUncertainty-	AlphaNumeric(50)	The identification code of the	idUncertainty-	Yes
Set		uncertainty set. During an	Set,	
		uncertainty iteration one set	UncertaintyId	
		will be picked to be the POD		
		value.		
idEffect	AlphaNumeric(50)	The effect code.	idEffect,	Yes
			EffectId, Effect	
idSubstance	AlphaNumeric(50)	The code of the substance.	idSubstance,	Yes
			SubstanceId,	
			SubstanceCode,	
			Substance	
Point of	Numeric	Point of departure, can be of	PointOf-	Yes
departure		various types, e.g. NOAEL,	Departure,	
		LOAEL, BMD, CED	HazardDose,	
			LimitDose,	
			CED	
DoseResponse-	AlphaNumeric(200)	A comma separated list of the	DoseResponse-	No
Model-		values of the parameters of	Model-	
Parameter-		the model, format:	Parameter-	
Values		a=1.2,b=3.4,c=5.6	Values,	
			ParameterValues	

Table aliases: PointsOfDepartureUncertain, PointOfDepartureUncertain, HazardDosesUncertain, HazardDoseUncertain.

Points of departure settings

Uncertainty settings

Name	Description
Resample hazard	Specifies whether to resample the hazard characterisations or
characterisations or RPFs	relative potency factors. Requires hazard characterisation or RPF
	uncertainty to be quantified in DoseResponseModelsUncertain or
	RelativePotencyFactorsUncertain tables.

Table 2.166: Uncertainty settings for module Points of departure.

Points of departure as data

Points of departure are provided as data for combinations of substance and effect and each is minimally described by a reference value and a type (e.g., NOAEL or LOAEL). In addition, the exposure route, specifies, and references may be specified.

• Points of departure data formats

Inputs used: *AOP networks*

2.5.10 Relative potency factors

Relative potency factors (RPFs) quantify potencies of substances with respect to a defined effect, relative to the potency of a chosen index substance. RPFs can be used to express combined exposures of multiple substances in terms of a the exposure value of the chosen index substance (i.e., in index substance equivalents). In MCRA, hazard characterisations, and therefore also RPFs are based on mass units (e.g., μ g), and not on mol units. RPFs can be different for different levels of the human organism (external, internal, specific compartment). RPFs can be given as data or computed from hazard characterisations. RPFs can be specified with uncertainty. Computation from uncertain hazard characterisations allows to include correlations between uncertain RPFs which originate from using the same index substance.

This module has as primary entities: Substances Effects

Output of this module is used by: Concentrations Concentration models High exposure food-substance combinations Dietary exposures Exposures

Relative potency factors data formats

Relative potency factors

Relative potency factors quantify relative potencies of substances with respect to an effect and can be used to express combined exposures of multiple substances in terms of the exposure value of the chosen index substance (i.e., as index substance equivalents). Relative potency factors can be provided in case hazard characterisations are missing. If available, the uncertainty of externally specified RPFs can be specified with uncertainty sets (empirical distributions representing possible values) in an additional table.

Relative potency factors

Relative potency factors are linked to an effect using the effect code (idEffect) and to substances using the substance code (idSubstance).

Name	Туре	Description	Aliases	Required
idSubstance	AlphaNumeric(50)	The code of the substance.	idSubstance,	Yes
			SubstanceId,	
			SubstanceCode,	
			Substance	
idEffect	AlphaNumeric(50)	The effect code.	idEffect,	Yes
			EffectId, Effect	
RPF	Numeric	The relative potency factor.	RPF, Relative-	Yes
			PotencyFactor	

Table 2.167: Table definition for RelativePotencyFactors.

Table aliases: RelativePotencyFactors, RelativePotencyFactor.

Relative potency factor uncertainty

This table contains sets of values representing the uncertainty for relative potency factors.

Name	Туре	Description	Aliases	Required
idUncertainty-	AlphaNumeric(50)	The uncertainty set	idUncertainty-	Yes
Set		identification number. During	Set,	
		each uncertainty iteration one	UncertaintyId	
		set is used.		
idEffect	AlphaNumeric(50)	The effect code (must	idEffect,	Yes
		correspond to values in id	EffectId, Effect	
		column of Effects table).		
idSubstance	AlphaNumeric(50)	The substance code (must	idSubstance,	Yes
		correspond to values in id	SubstanceId,	
		column of Substances table).	SubstanceCode,	
			Substance	
RPF	Numeric	The relative potency factor.	RPF, Relative-	Yes
			PotencyFactor	

 Table 2.168: Table definition for RelativePotencyFactorsUncertain.

Table aliases: RelativePotencyFactorsUncertain, RelativePotencyFactorUncertain.

Relative potency factors calculation

Relative potency factors (RPFs) describe the potency of substances with respect to a defined effect, relative to the potency of a chosen index substance. RPFs can be given as data or computed from *hazard characterisations*. The RPF for substance i is defined by the ratio of hazard characterisation value for the index substance (ref) and the hazard characterisation value for substance i. That is,

$$RPF_i = POD_{ref} / POD_i$$
.

When the hazard characterisations are resampled in the uncertainty runs, RPFs are also recomputed based on the bootstrapped hazard characterisations. In this way, RPF uncertainty can also included in the uncertainty analysis.

Relative potency factors settings

Calculation settings

Table 2.169: Calculation settings for module Relative potency factors.

Name	Description
Index substance	The substance of interest or index substance.

Uncertainty settings

Name	Description
Resample hazard	Specifies whether to resample the hazard characterisations or
characterisations or RPFs	relative potency factors. Requires hazard characterisation or RPF
	uncertainty to be quantified in DoseResponseModelsUncertain or
	RelativePotencyFactorsUncertain tables.

Relative potency factors as data

Data are provided in the form of a RPF for a specific substance and effect.

• *Relative potency factors data formats*

Inputs used: Active substances AOP networks

Calculation of relative potency factors

RPFs are computed from hazard characterisations.

• Relative potency factors calculation

Inputs used: Hazard characterisations

Settings used

• Calculation Settings

2.6 In-silico modules

Two types of in-silico models are available: QSAR models specify assessment group memberships for active substances, as numbers in the interval [0,1]. This allows both crisp (0 or 1) and probabilistic memberships. Molecular docking models specify binding energies and thresholds which can be used to convert binding energies to assessment group memberships for active substances.

2.6.1 Molecular docking models

Molecular docking models specify binding energies for substances in specific molecular docking models related to a specific health effect (adverse outcome).

This module has as primary entities: Substances Effects

Output of this module is used by: Active substances

Molecular docking models data formats

Required data tables:

- Molecular docking models, to identify models for a specified effect (receptor)
- Molecular docking binding energies, to specify the binding energies per substance for the receptor

Molecular docking models

Contains definitions of molecular docking models for a given effect (molecular initiating event), for example parameters needed in the conversion of binding energies to group memberships or to relative potency factors. Substance specific binding energies are specified in the binding energies table.

Molecular docking models

Each docking model has a unique identifier, and optionally a name and a description. Each model is linked to an effect using the idEffect field and optionally a binding threshold and the number of receptors can be added. A reference to the source of the data can be stored in the reference field.

Name	Туре	Description	Aliases	Required
id	AlphaNumeric(50)	The unique identification code of the molecular docking model.	idMolecular- DockingModel, idBinding-	Yes
Name	AlphaNumeric(100)	The name of the molecular docking model.	EnergyModel Name	No
Description	AlphaNumeric(200)	Description of the molecular docking model.	Description	No
idEffect	AlphaNumeric(50)	The effect code, typically for the Molecular Initiating Event that is modelled	idEffect, EffectId, Effect	Yes
Threshold	Numeric	Threshold Molecular Docking binding energy (group membership = 1 when higher).		No
NumberOf- Receptors	Integer	Example parameter needed for translating Molecular Docking binding energies to RPFs.		No
Reference	AlphaNumeric(200)	External reference(s) to sources containing more information about the molecular docking model.	References	No

Table 2.171:	Table definition for	MolecularDockingModels.
--------------	----------------------	-------------------------

Table aliases: MolecularDockingModels, MolecularDockingModel, BindingEnergyModels, BindingEnergyModel.

Molecular docking binding energies

Molecular docking model binding energies per substance

Name	Туре	Description	Aliases	Required
idMolecular-	AlphaNumeric(50)	The id of the molecular	idMolecular-	No
DockingModel		docking model or source.	Docking,	
			Molecular-	
			DockingModel	
idSubstance	AlphaNumeric(50)	The code of the substance.	idSubstance,	Yes
			SubstanceId,	
			SubstanceCode,	
			Substance	
BindingEnergy	Numeric	Molecular Docking binding	Molecular-	Yes
		energy.	Docking-	
			BindingEnergy	

Table 2.172: Table definition for MolecularBindingEnergies.

Table aliases: MolecularBindingEnergies, MolecularBindingEnergy, BindingEnergies, BindingEnergy, MolecularDockingBindingEnergies, MolecularDockingBindingEnergy.

Molecular docking models as data

Binding energies for substances in specific molecular docking models related to a specific health effect (adverse outcome) are provided as data.

• Molecular docking models data formats

Inputs used: AOP networks

2.6.2 QSAR membership models

QSAR membership models specify assessment group memberships for active substances related to a specific health effect (adverse outcome). Memberships should be derived externally from Quantitative Structure-Activity Relationship (QSAR) models.

This module has as primary entities: Substances Effects

Output of this module is used by: Active substances

QSAR membership models data formats

Required data tables:

- QSAR membership models, to identify QSAR models for a specified health effect
- QSAR membership scores, to specify the memberships per substance per QSAR model

Note that only memberships 1 (include) and 0 (exclude) are allowed.

QSAR membership models

Substance membership models obtained from QSAR for a given (health) effect. The models are defined in the membership models table, and substance specific memberships are specified in the QSAR memberships table.

QSAR membership models

This table contains the definitions of the QSAR membership models. Each model contains a id, name, an optional description, and refers to its related health effect.

Name	Туре	Description	Aliases	Required
id	AlphaNumeric(50)	The unique identification code of the QSAR membership	id, Model, ModelCode,	Yes
		model.	idModel,	
			QSARModel,	
			idQSARModel,	
			QSAR-	
			Membership-	
			Model,	
			idQSAR-	
			Membership-	
			Model,	
			Membership-	
			Model,	
			idMembership-	
			Model	
Name	AlphaNumeric(100)	The name of the QSAR	Name	No
Description		membership model.	Decemintian	No
Description	AlphaNumeric(200)	Description of the QSAR membership model.	Description	INO
idEffect	AlphaNumeric(50)	The effect code.	idEffect, EffectId, Effect	Yes
Accuracy	Numeric	Accuracy of the QSAR	Accuracy	No
Accuracy	Numeric	membership model.	Accuracy	
Sensitivity	Numeric	Sensitivity of the QSAR membership model.	Sensitivity	No
Specificity	Numeric	Specificity of the QSAR membership model.	Specificity	No
Reference	AlphaNumeric(200)	External reference(s) to sources containing more information about the QSAR model.	References	No

 $Table \ a liases: \ QSAR, \ QSARMembership Models, \ QSARMembership Model, \ QSARModels, \ QSARModel.$

QSAR membership scores

Substance membership score according to the QSAR model.

Name	Туре	Description	Aliases	Required
idQSAR-	AlphaNumeric(50)	The id of the QSAR model.	Model,	Yes
Membership-			ModelCode,	
Model			idModel,	
			QSARModel,	
			idQSARModel,	
			QSAR-	
			Membership-	
			Model,	
			idQSAR-	
			Membership-	
			Model,	
			Membership-	
			Model,	
			idMembership-	
			Model	
idSubstance	AlphaNumeric(50)	The code of the substance.	idSubstance,	Yes
			SubstanceId,	
			SubstanceCode,	
			Substance	
Membership-	Numeric	QSAR membership score.	Membership-	Yes
Score		Value should be 1 for positive	Score,	
		membership, or 0 for negative	Membership,	
		membership.	QSARScore,	
			Score	

Table aliases: QSARMembershipScores, QSARMembershipScore, QSARMemberships, QSARMembership.

QSAR membership models as data

QSAR memberships models are provided as data, per QSAR model assessment group memberships for active substances related to a specific health effect are specified.

• QSAR membership models data formats

Inputs used: AOP networks

2.7 Kinetic modules

Kinetic models convert exposures or hazard characterisations from one or more external routes or compartments to an internal (target) compartment. The reverse conversion from internal to external can also be made (reverse dosimetry).

In a simple tier, kinetic models are specified as absorption factors. In a higher tier, physiologically based toxicokinetic (PBTK) models of a specified type (currently available is the EuroMix generic PBTK model) are linked to MCRA. Kinetic model instances for specific substances and test systems (e.g. cypermethrin in the rat) are specified with parameter sets for the chosen kinetic model.

2.7.1 Kinetic models

External exposure can be from on more more exposure routes: oral (dietary or non-dietary), dermal or inhalation. Internal exposure can be systemic or related to a specific compartment in a kinetic model. There are four tiers for relating external to internal exposures (doses):

- 1. Assume 100% absorption: internal exposures are equal to external exposures.
- 2. Assume conservative absorption factors as suggested by EFSA ([[EFSA, 2014]], [[EFSA, 2017a]]): oral and inhalation 100%, dermal 50%.
- 3. Use externally provided absorption factors (absorption factors data tables).
- 4. Use one of the *implemented kinetic models*, with instances for specific substances defined in data table *kinetic* model instances and model parameters specified in data table kinetic model instance parameters.

Given a chosen tier, the calculation will fall back to the next lower tier in case of missing data.

This module has as primary entities: Substances

Output of this module is used by: Exposures Hazard characterisations

Kinetic models data formats

Data tables:

- · Absorption factors
- · Kinetic model instances
- · Kinetic model instance parameters

Kinetic models

Kinetic models may be specified as kinetic model instances that contain parameter specifications of built in kinetic models or as simple absorption factors.

Kinetic model instances

Kinetic model instances.

	Table 2.175: Ta	ble definition for KineticModelIn	stances.	
Name	Туре	Description	Aliases	Required
idModel-	AlphaNumeric(50)	Unique identification code of	idModel-	Yes
Instance		the kinetic model instance.	Instance, Id,	
			Code	
idModel-	KineticModelType	Identifier of the kinetic model	idModel-	Yes
Definition		definition for which this is an	Definition,	
		instance.	ModelDefinition	
idTestSystem	AlphaNumeric(200)	The species on which the	System,	Yes
		experiment was performed.	TestSystem	
idSubstance	AlphaNumeric(50)	Unique identification code of	idSubstance,	No
		substance, Default: valid for	SubstanceId,	
		all substances. Should be	SubstanceCode,	
		omitted for parameters in the	Substance	
		class Physiological		
Reference	AlphaNumeric(100)	Reference or author.	References	No

Table aliases: KineticModelInstances, KineticModelInstance.

Kinetic model instance parameters

Kinetic model parameters

Name	Туре	Description	Aliases	Required
idModel-	AlphaNumeric(50)	Unique identification code of	Id, Code	Yes
Instance		the kinetic model instance to		
		which this parameter belongs		
Parameter	AlphaNumeric(100)	Name of the parameter in the		Yes
		kinetic model.		
Description	AlphaNumeric	Description of or reference		No
		for the parameter values in		
		the kinetic model.		
Value	Numeric	Mean.	MEAN, mean	Yes
Distribution-	AlphaNumeric(20)	Distribution.	Distribution-	No
Туре			Туре,	
			Distribution	
CvVariability	Numeric	Variability.		No
CvUncertainty	Numeric	Uncertainty.		No

Table 2.176: Table definition for KineticModelInstanceParameters.

Table aliases: KineticModelInstanceParameters, KineticModelInstanceParameter.

Kinetic model absorption factors

Kinetic absorption factors

Name	Туре	Description	Aliases	Required
idCompound	AlphaNumeric(50)	code of substance (must	idSubstance,	No
		correspond to values in id	SubstanceId,	
		column of Substances table)	SubstanceCode,	
			Substance	
Route	AlphaNumeric(50)	Non-dietary route or pathway, use 'Oral', 'Dermal', or 'Inhalation' to specify the	Route, Pathway	No
		route.		
Absorption-	Numeric	absorption factor value	Absorption-	No
Factor			Factor, Factor	

Table 2.177: Table definition for KineticAbsorptionFactors.

Table aliases: KineticAbsorptionFactors, KineticAbsorptionFactor, AbsorptionFactors, AbsorptionFactor.

Kinetic models settings

Calculation settings

Name	Description
Default oral absorption factor	When there is no kinetic model and absorption factors are not
for non-dietary exposure	specified in file, non-dietary oral exposures (external doses) are
	multiplied by this factor to determine the absorbed (internal) dose.
Default oral absorption factor	When there is no kinetic model and absorption factors are not
for dietary exposure	specified in file, dietary exposures (external doses) are multiplied
	by this factor to determine the absorbed (internal) dose .
Default dermal absorption	When there is no kinetic model and absorption factors are not
factor for non-dietary exposure	specified in file, dermal exposures (external doses) are multiplied
	by this factor to determine the absorbed (internal) dose.
Default inhalation absorption	When there is no kinetic model and absorption factors are not
factor for non-dietary exposure	specified in file, inhalation exposures (external doses) are
	multiplied by this factor to determine the absorbed (internal) dose.
Number of days	The number of days.
Number of events per day for	The daily dose is administered in equal portions (dose / number of
the ORAL dietary dose	events) per event.
Number of initial days skipped	This period is skipped in the calculation of the mean internal
	exposure.
Kinetic model	Code Kinetic Model.
Use parameter variability	When specified, use parameter variability.

Table 2.178: Calculation settings for module Kinetic models.

Uncertainty settings

Table 2.179:	Uncertainty	settings	for module	Kinetic models.
--------------	-------------	----------	------------	-----------------

Name	Description
Resample kinetic model	Specifies whether kinetic model parameter values are resampled.
parameter values	

Kinetic models as data

Specify nondietary absorption factors as data.

• Kinetic models data formats

Inputs used: Active substances

Available kinetic models

Physiologically based toxicokinetic (PBTK) models, or kinetic models for short, are mathematical representations of the animal or human body aimed at describing and predicting the time course distribution of chemicals in tissues and organs. Those internal dose metrics can usefully replace external exposure dose in the derivation of the quantitative dose-response relationships and following risk assessments. PBTK models can simulate both internal doses from exposure scenarios (forward dosimetry) and external dose from biomonitoring data (reverse dosimetry).

The following generic PBTK models are currently implemented in MCRA:

- EuroMix generic PBTK model [[Tebby et al., 2019]].
- *bisphenol model ETHZ* [[Karrer et al., 2019]].

EuroMix Generic PBTK model v6

Cosmos version 6 (received 3/27/2019)

Table 2.180: Exposure routes (forcings)

ld	Description	Unit	Order
Dietary	Dietary exposure	mmoles	0
Dermal	Dermal exposure	mmoles	1
Inhalation	Inhalatory exposure	mmoles	2

ld	Description	ScalingFactor	Multiplication- Factor	Unit	Order
CTotal	Total concentration			mM	0
CVen	Venous blood concentration	scVBlood	0.66667	mM	1
CArt	Arterial blood concentration	scVBlood	0.33333	mM	2
CFat	Fat (adipose) tissue concentration	scVFat		mM	3
CPoor	Poorly perfused tissue (muscle) concentration			mM	4
CRich	Richly perfused tissue (viscera) concentration	scVRich		mM	5
CLiver	Liver concentration	scVLiver		mM	6
CSkin_u	Viable unexposed skin concentration			mM	7
CSkin_e	Viable exposed skin concentration	BSA, Height_vs, fsA_exposed		mM	8
CSkin_sc_u	Skin unexposed stratum corneum concentration			mM	9
CSkin_sc_e	Skin exposed stratum corneum concentration	BSA, Height_vs, fsA_exposed		mM	10

Table 2.181: Output

ld	Description	Unit	Туре	Order
BM	Body mass	kg	Physiological	0
BSA	Body surface area (internally scaled by an allometric scaling factor s = 70/BM^0.3)	dm2	Physiological	1
scVFat	Fat as fraction of total body volume		Physiological	2
scVRich	Richly perfused tissues (viscera) as fraction of total body volume		Physiological	3
scVLiver	Liver as fraction of total body volume		Physiological	4
scVBlood	Blood as fraction of total body volume		Physiological	5
Height_sc	Skin thickness	decimeter	Physiological	6
Height_vs	Viable skin		Physiological	7
scFBlood	Total blood flow per unit mass	L/h/kg	Physiological	8
scFFat	Fat fraction of total blood flow going to compartments		Physiological	9
scFPoor	Poorly perfused tissues (muscles) fraction of total blood flow going to compartments		Physiological	10
scFLiver	Liver fraction of total blood flow going to compartments		Physiological	11
scFSkin	Skin fraction of total blood flow going to compartments		Physiological	12
Falv	Alveolar ventilation rate	L/h	Physiological	13
mic	Microsomal proteins content	mg/gr liver	Physiological	14
Kp_sc_vs	Diffusion rate from stratum corneum to viable skin	decimeter/h	Metabolic	22
Ke	Renal excretion rate	L/h	Metabolic	23
Michaelis	Flag for Michaelis-Menten vs linear metabolism (0 = linear)		Metabolic	24
Vmax	Maximum rate of metabolism	mmoles/h/L liver	Metabolic	25
Km	Michaelis-Menten constant for metabolism	mM	Metabolic	26
CLH	Hepatic metabolic clearance		Metabolic	27
fub	Unbound fraction in blood		Metabolic	28
(inetic modules	Fraction absorbed by the gut		Metabolic	29
kGut	Oral 1st order absorption rate constant	1/h	Metabolic	30

Table 2.182: Input

Model aliases: Cosmos6, CosmosV6.

EuroMix Generic PBTK model v5

Cosmos version 5 (adapted 9/11/2018)

Table 2.183:	Exposure routes	(forcings)
--------------	-----------------	------------

ld	Description	Unit	Order
Dietary	Dietary exposure	mmoles	0
Dermal	Dermal exposure	mmoles	1
Inhalation	Inhalatory exposure	mmoles	2

		Table 2.184: Output			
ld	Description	ScalingFactor	Multiplication- Factor	Unit	Order
CVen	Venous blood	scVBlood	0.66667	mM	0
CArt	Arterial blood	scVBlood	0.33333	mM	1
CFat	Fat tissues	scVFat		mM	2
CPoor	Muscle tissues			mM	3
CRich	Viscera	scVRich		mM	4
CLiver	Liver	scVLiver		mM	5
CSkin_u	Viable skin, unexposed			mM	6
CSkin_e	Viable skin, exposed	BSA, Height_vs, fsA_exposed		mM	7
CSkin_sc_u	Skin stratum corneum, unexposed			mM	8
CSkin_sc_e	Skin stratum corneum, exposed	BSA, Height_vs, fsA_exposed		mM	9

259

ld	Description	Unit	Туре	Order
BM	Body mass	kg	Physiological	0
BSA	Body skin surface area	dm2	Physiological	1
scVFat	Fat as fraction of total body volume		Physiological	2
scVRich	Richly perfused tissues (viscera) as fraction of total body volume		Physiological	3
scVLiver	Liver as fraction of total body volume		Physiological	4
scVBlood	Blood as fraction of total body volume		Physiological	5
Height_sc	Skin thickness	decimeter	Physiological	6
Height_vs	Viable skin		Physiological	7
scFBlood	Total blood flow per unit mass	L/h/kg	Physiological	8
scFFat	Fat fraction of total blood flow going to compartments		Physiological	9
scFPoor	Poorly perfused tissues (muscles) fraction of total blood flow going to compartments		Physiological	10
scFLiver	Liver fraction of total blood flow going to compartments		Physiological	11
scFSkin	Skin fraction of total blood flow going to compartments		Physiological	12
Falv	Alveolar ventilation rate	L/h	Physiological	13
mic	Microsomal proteins content	mg/gr liver	Physiological	14
PCAir	Partition coefficient: blood over air		Partition coefficient	15
Kp_sc_vs	Diffusion rate from stratum corneum to viable skin	decimeter/h	Metabolic	22
Ke	Renal excretion rate	L/h	Metabolic	23
Michaelis	Flag for Michaelis-Menten vs linear metabolism (0 = linear)		Metabolic	24
Vmax	Maximum rate of metabolism	mmoles/h/L liver	Metabolic	25
Km	Michaelis-Menten constant	mM	Metabolic	26
CLH	Hepatic clearance		Metabolic	27
fup	Unbound fraction in blood		Metabolic	28
Frac	Fraction absorbed by the gut		Metabolic	29
kGut	Oral 1st order	1/h	Metabolic	30
inetic modules	absorption rate constant			
fSA_exposed	Fraction of skin surface area actually exposed		Metabolic	35

Table 2.185: Input

Model aliases: Cosmos4, CosmosV4, Cosmos5, CosmosV5.

Generic Model BPA

Generic model Cecile Karrer 23 juli 2018

ld	Description	Unit	Order
Dietary	Dietary exposure	nmoles	0
Oral	Oral exposure	nmoles	1
Dermal	Dermal exposure	nmoles	2
Inhalation	Inhalation exposure	nmoles	3

Table 2.186: Exposure routes (forcings)

Table	2.187:	Output
rable	2.107.	Output

ld	Description	ScalingFactor	Multiplication- Factor	Unit	Order
CPlasmaOut	Concentration in plasma			nmol/L	0
CGonadOut	Concentration in gonads			nmol/L	1
AurinebpaOut	Cumulative excretion of BPA in urine			nmol/L	2
AurinegOut	Cumulative excretion of BPA-g in urine			nmol/L	3
AurineTotalOut	Cumulative excretion of BPA and metabolites in urine			nmol/L	4

Table 2.188: Input

ld	Description	Unit	Туре	Order
BW	Bodyweight	kg	Physiological	0
QCC	Cardiac output	L/min	L/min Physiological 1	
QgonadC	Fractional blood		Physiological	2
	flow to gonads			
QliverC	Fractional blood		Physiological	3
	flow to liver			
QfatC	Fractional blood		Physiological	4
	flow to fat tissue			
QbrainC	Fractional blood		Physiological	5
	flow to brain			
QskinC	Fractional blood		Physiological	6
	flow to skin			
QmuscleC	Fractional blood		Physiological	7
	flow to gonads			
VplasmaC	Fractional volume of		Physiological	8
	plasma			
VfatC	Fractional volume of		Physiological	9
	fat tissue			
VliverC	Fractional volume of		Physiological	10
	liver tissue			

		continued from prev	1 8	
ld	Description	Unit	Туре	Order
VbrainC	Fractional volume of		Physiological	11
	brain tissue			
VskinC	Fractional volume of		Physiological	12
	skin tissue			
VgonadC	Fractional volume of		Physiological	13
	gonads			
VmuscleC	Fractional volume of		Physiological	14
	muscle tissue			
VrichC	Fractional volume of		Physiological	15
	richly perfused			
	tissue			
VbodygC	Distribution volume		Physiological	16
	of BPA-g			
MW	Molecular weight	g/mol	Chemical property	18
pliver	Partition coefficient		Partition coefficient	19
	liver to blood			
pfat	Partition coefficient		Partition coefficient	20
1	fat to blood			
pslow	Partition coefficient		Partition coefficient	21
1	slowly perfused			
	tissue to blood			
prich	Partition coefficient		Partition coefficient	22
F	richly perfused			
	tissue to blood			
pgonad	Partition coefficient		Partition coefficient	23
r8	gonads to blood			
pbrain	Partition coefficient		Partition coefficient	24
porum	brain to blood		Turthion coefficient	21
pskin	Partition coefficient		Partition coefficient	25
pskii	skin to blood		I ditution coefficient	25
geC	Gastric emptying	1/h/kg bw^-0.25	Metabolic	26
k0C	Oral uptake from the	1/h/kg bw^-0.25	Metabolic	20
RUC	stomach into the	1/11/Kg UW -0.25	Wietabolie	21
	liver			
k1C	Oral uptake from the	1/h/kg bw^-0.25	Metabolic	28
ĸic	small intestine into	1/11/Kg Uw -0.25	Wietabolic	20
	the liver			
k4C	Fecal elimination	1/h/kg bw^-0.25	Metabolic	29
K4C	from small intestine	1/11/Kg 0w^-0.25	Metabolic	29
	after oral administration			
		1/1./1. 1. 0.025	M. (. 1 1' .	20
kGIingC	Transport of	1/h/kg bw^-0.25	Metabolic	30
	glucuronide from			
	enterocytes into			
1.01. 0	serum	100 1 4005		21
kGIinsC	Transport of sulfate	1/h/kg bw^-0.25	Metabolic	31
	from enterocytes			
-	into serum			-
kmgutg	Km of	nM	Metabolic	32
	Glucuronidation in			
	the gut			
vmaxgutgC	Vmax of	nmol/h/kg bw	Metabolic	33
	Glucuronidation in			
	the gut			

Table 2.188 - continued from previous page

continues on next page

		continued from previo		Order
ld	Description	Unit	Туре	Order
fgutg	Correction factor of		Metabolic	34
	glucuronidation in			
	the gut			
kmguts	Km of Sulfation in	nM	Metabolic	35
	the gut			
vmaxgutsC	Vmax of Sulfation in	nmol/h/kg bw	Metabolic	36
	the gut			
fguts	Correction factor of		Metabolic	37
	sulfation in the gut			
met1g	Fraction of		Metabolic	38
	glucuronide in the			
	liver taken up			
	directly into serum			
	(the rest undergoes			
	EHR)			
met1s	Fraction of sulfate in		Metabolic	39
	the liver taken up			
	directly into serum			
enterocytes	Sum of enterocytes	L	Metabolic	40
2	weights in			
	duodenum, jejunum			
	and ileum			
kmliver	Km of	nM	Metabolic	41
	Glucuronidation in			
	the liver			
vmaxliverC	Vmax of	nmol/h/g liver	Metabolic	42
	Glucuronidation in		1.1000000	
	the liver			
fliverg	Correction factor of		Metabolic	43
niverg	glucuronidation in		Metabolie	
	the liver			
kmlivers	Km of Sulfation in	nM	Metabolic	44
KIIIIVels	the liver	11111	Wieddoone	
vmaxliversC	Vmax of Sulfation in	nmol/h/g liver	Metabolic	45
villaxiiverse	the liver		Wietabolie	45
flivers	Correction factor of		Metabolic	46
liiveis	sulfation in the liver		Wietabolie	40
EHRtime	Time until EHR	h	Metabolic	47
ERRuitie		11	Metabolic	47
EHRrateC	OCCURS	1 /h /h = h = A 0 25	Metabolic	4.0
	EHR of glucuronide	1/h/kg bw^-0.25		48
k4C_IV	Fecal elimination of	1/h/kg bw^-0.25	Metabolic	49
	glucuronide from the			
1 1 0	EHR compartment	1 1 1 1 10 75		
kurinebpaC	Clearance, urine	L/h/kg bw^0.75	Metabolic	50
	excretion of parent			
~	compound	.		
kurinebpagC	Clearance, urine	L/h/kg bw^0.75	Metabolic	51
	excretion of			
	glucuronide			
kurinebpasC	Clearance, urine	L/h/kg bw^0.75	Metabolic	52
	excretion of sulfate			
vreabsorptiong-	Vmax for renal	nmol/h/kg bw^0.75	Metabolic	53
vieuosorptiong				
C	reabsorption of glucuronide			

Table 2.188 - continued from previous page

continues on next page

ld	Description	Unit	Туре	Order
vreabsorptionsC	Vmax for renal reabsorption of sulfate	nmol/h/kg bw^0.75	Metabolic	54
kreabsorptiong	kreabsorptiong Km for renal reabsorption of glucuronide		Metabolic	55
kreabsorptions	Km for renal reabsorption of sulfate	nM	Metabolic	56
kenterobpagC	EHR of parent compound due to biliary excretion of glucuronide	1/h/kg bw^-0.25	Metabolic	57
kenterobpasC	EHR of parent compound due to biliary excretion of sulfate	1/h/kg bw^-0.25	Metabolic	58
EoA_O	Extent of oral absorption		Physiological	61
period_O	uptake period	h	External	63
t0_0	time point at which dosing starts	h	External	65
EoA_D	Extent of dermal absorption from TP		Physiological	68
aHL_D	Half-life for dermal penetration	h	External	70
period_D	Uptake period dermal exposure from TP	h	External	72
t0_D	Time points at which dermal dosing from TP starts	h	External	74
EoA_D2	Extent of dermal absorption from PCPs		Physiological	77
aHL_D2	Half-life for dermal penetration from PCPs	h	External	79
period_D2	Uptake period dermal exposure from PCPs	h	External	81
t0_D2	Time points at which dermal dosing from PCPs starts	h	External	83
BW075	BW^0.75	kg^0.75	External	103
BW025	BW^0.25	kg^0.25	External	104

Table	2.188 -	continued from	m previous j	bage
-------	---------	----------------	--------------	------

Model aliases: PBPKModel_BPA, PBPKModelBPA, ModelBPA, BPA.

Note: Additional kinetic models can be implemented, please contact the MCRA administrator.

EuroMix generic PBTK model

Reference: Tebby et al, 2019: [[Tebby et al., 2019]]

In MCRA updated versions (version 4b, 6) of the PBTK model developed at INERIS in the framework of the COS-MOS project is used. The model describes the distribution of chemicals in venous blood, arterial blood, adipose tissues, poorly perfused tissues (muscles), gut lumen, liver, richly perfused tissues (other viscera), and skin. Each of those is described as a compartment (homogeneous virtual volume) in which distribution is instantaneous and limited only by the incoming blood flow or rate of entry in the compartment. Exposure can occur through the dermal route, ingestion or inhalation. The absorbed molecules can be excreted to urine, exhaled through the lung, or metabolized in liver.

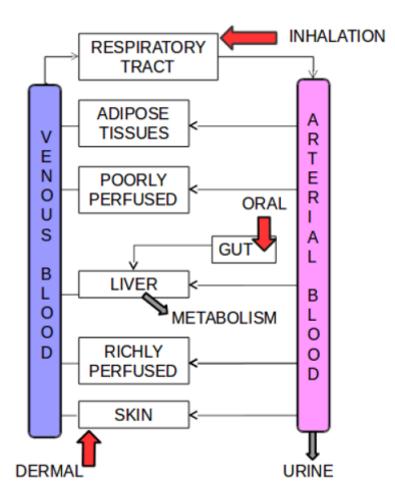
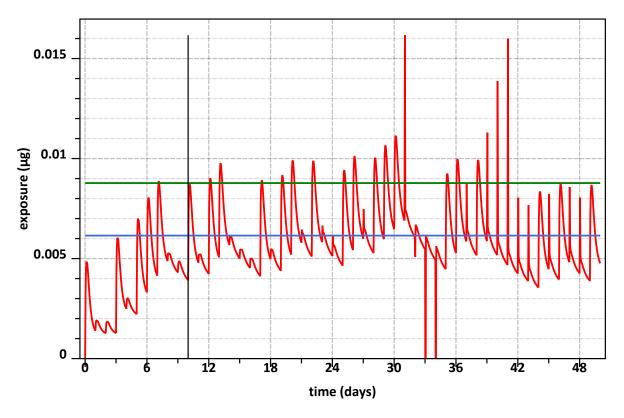


Figure 2.42: Schematic representation of the EuroMix Generic PBTK model.

The EuroMix generic PBTK model is coded as a set of ordinary differential equations. There is one such equation per time-dependent chemical quantity of the model (so-called state variables). There are 13 state variables in the model: the quantity of chemical in venous blood (Q_{ven}) , in arterial blood (Q_{art}) , in adipose tissues (Q_{fat}) , in poorly perfused tissues (Q_p) , in well perfused tissues (Q_r) , in liver (Q_{liv}) , in unexposed skin $(Q_{s,u})$, in exposed skin $(Q_{s,e})$, in the stratum corneum of unexposed skin $(Q_{sc,u})$, in exposed stratum corneum $(Q_{sc,e})$, in gut lumen (Q_{gut}) , the quantity excreted to urine (Q_{ex}) , and the quantity metabolized (Q_{met}) . The model can predict, as a function of time, for given oral, dermal and/or inhalation exposures, all the above quantities and the corresponding concentrations as a function of time. Concentrations are obtained by dividing quantities by compartment volumes (cited: Bois, Tebby & Brochot).

In Figure 2.43 a time course of the internal substance amount (μg) for Clothianidin in the liver is shown. For 50 consecutive days a bolus per day is submitted. The red line shows the substance amount varying over time. The green line displays the average of the peaks representing acute exposure, the blue line displays the steady state representing chronic exposure, all after skipping a nonstationary period of 10 days (the vertical black line).

From the substance amount, a concentration is computed by dividing it by the total compartment weight (i.e., the mass/volume of the compartment/organ).



Model CosmosV6

Figure 2.43: Time course of exposure (μg) for Clothianidin in the liver (EuroMix generic PBTK model version 6).

In Figure 2.44, for a large number of individuals the internal exposure (acute, green dots) in the liver is plotted versus the external exposure ($\mu g/kgbw$). The diagonal represents the 1:1 ratio of internal vs external exposure.

Bisphenol model

Reference: Karrer et al. 2019: [[Karrer et al., 2019]]

'Structural analogs such as the bisphenols S, F, and AF (BPS, BPF, BPAF) are used to replace the endocrine disrupting chemical bisphenol A (BPA), but they exert estrogenic effects in the same order of magnitude. In order to investigate the consequences of BPA restrictions, we assessed the cumulative risk from BPA, BPS, BPF, and BPAF in Europe before and after the first BPA restrictions in 2011. We modelled external exposures from food, personal care products (PCPs), thermal paper, and dust, using the models MCRA and PACEM for food and PCPs, respectively. We calculated internal concentrations of unconjugated BPs with substance-specific PBPK models and cumulated concentrations by taking into account relative estrogenic potencies. Average cumulative exposure to unconjugated BPs was 3.8 and 2.1 ng/kg bw/day before and after restrictions, respectively. The decline was mostly caused by the replacement of BPA with BPS in thermal paper. Therefore, the margins of exposure (MOEs) for estrogenic effects were mostly higher after the restrictions. However, in high uncertainty percentiles the MOEs were partly lower than before (e.g. the MOEs for the uncertainty P97.5 of the variability P99 were 2.6 and 1.9 before and after restrictions, respectively), which shows the higher uncertainty around exposures for substitutes compared to BPA.'

Abstract: Linking probabilistic exposure and pharmacokinetic modelling to assess the cumulative risk from the bisphenols BPA, BPS, BPF, and BPAF for Europeans. Authors: Cecile Karrer, Waldo de Boer, Christiaan Delmaar, Yaping Cai, Amélie Crépet, Konrad Hungerbühler, Natalie von Goetz

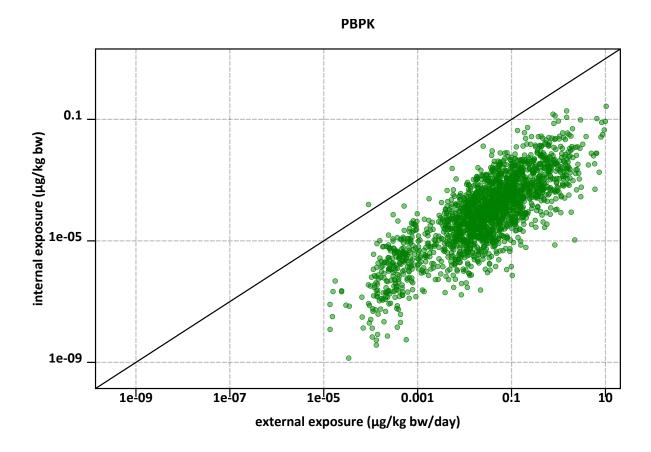


Figure 2.44: Internal versus external exposure for Clothianidin in the liver (EuroMix Generic PBTK model version 6).

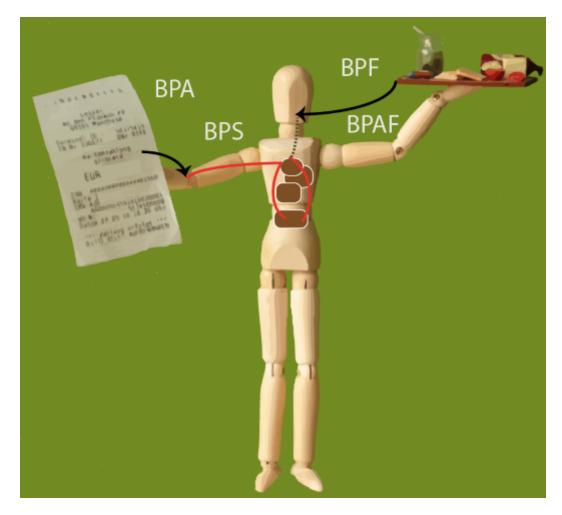


Figure 2.45: Graphical abstract 'Linking probabilistic exposure and pharmacokinetic modelling to assess the cumulative risk from the bisphenols BPA, BPS, BPF, and BPAF for Europeans.'

2.8 Risk modules

Exposures and *hazard characterisations* are compared in risk metrics. If both exposure and hazard characterisation are characterised by a single value, the risk metric (e.g. a traditional margin of exposure, hazard quotient or hazard index) can be calculated using module Single value risks. Module Risks allows for probabilistic risk calculations. In both cases a threshold can be specified to assist in interpretation. The threshold value should be chosen in relation to the assessment factors used in the hazard characterisation, e.g. a threshold MOE=100 is often used if no assessment factors have been used, but a threshold 1 would be appropriate if assessment factors have already been used to address relevant uncertainties.

2.8.1 Risks

Risks (health impacts) are defined as a function of exposure and hazard characterisation at a chosen biological level (external or internal). Risk metrics are margins of exposure (MOE) or hazard indices (HI) or more generalised MOE or HI distributions.

This module has as primary entities: Substances Effects Populations

Output of this module is used by: Single value risks

Risks calculation

Probabilistic risk is calculated as a distribution of either margin of exposure (MOE) or hazard index (HI), if at least one of the inputs exposure and hazard characterisation is a distribution. The risk distribution is characterised by percentiles. To accommodate for matching results of MOE and HI=1/MOE in the case of percentiles, there is an option to calculate percentiles via the complementary percentile of the inverse distribution in order to handle numerical differences when calculating percentiles for a left or right tail. E.g. p1 of the MOE distribution can optionally be calculated as 1 divided by p99 of the corresponding 1/MOE distribution. For more details about the graphical displays and the calculations see Individual risks.

Individual risks

A (cumulative) risk assessment aims to characterise the health impact due to one or multiple substances present in food causing one or more health effects. The health impact is characterized by a distribution of individual risks: exposures and hazard characterizations are compared at the chosen level (external or internal) via margins of exposure (MOE) or hazard indices (HI = 1/MOE). Hazard characterisations are included as single values or in a probabilistic way.

The aim is to specify the probability that a random individual from a defined (sub)population will have an exposure high enough to cause a particular health effect of a predefined magnitude, the critical effect size. The exposure level that results in exactly that critical effect in a particular person is that person's individual critical hazard dose. Individuals in a population typically show variation, both in their individual exposure and in their hazard characterization. Both the variation in exposure and the variation in hazard characterization are quantified in the form of probability distributions. Assuming independence between both distributions, they are combined by Monte Carlo methods.

The proportion of the MOE distribution below unity is the probability of critical exposure (*PoCE*) in the particular (sub)population. Uncertainties involved in the overall risk assessment (i.e., both regarding exposure and effect assessment) are quantified using Monte Carlo and bootstrap methods. This results in an uncertainty distribution for any statistic of interest, such as the probability of critical exposure (*PoCE*).

In Figure 2.46, margin of exposures for a number of substances are shown. As shown, the distinction between variability (grey bars, 90% probability) and uncertainty (whiskers) is retained. This is discussed in [[van der Voet et al., 2007]] and [[van der Voet et al., 2009]].

In Figure 2.47, hazards versus exposures are plotted for the same substances.

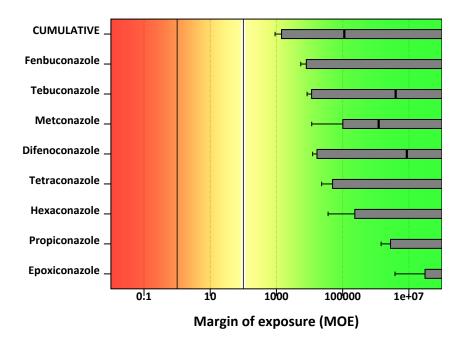


Figure 2.46: Individual margin of exposure (MOE) plot for multiple substances.

Risks settings

Calculation settings

Name	Description		
Multiple substances analysis	Specifies whether the assessment involves multiple substances.		
Express results in terms of	Specifies whether the assessment involves multiple substances and		
reference substance equivalents	results should be cumulated over all substances.		
(cumulative)			
Health effect type	Specifies whether the health effect is a risk (negative) or benefit		
	(positive).		
Risk metric type	Report risks in terms of hazard index (HI = 1/MOE) or margin of exposure.		
Show equivalent animal dose	Specifies whether equivalent animal doses should be reported in		
output	the output.		
Threshold safety plot	Threshold for interpretation in the margin of exposure or hazard		
	index plot, e.g. $MOE = HI = 1$ or $MOE = 100$.		
Use inverse distribution to	ibution to Calculate percentile via the complementary percentage of the		
calculate percentile	inverse distribution (default: no). Description: E.g., P0.1 of MOE		
	distribution is calculated via P99.9 of 1/MOE distribution. Note:		
	This option is provided because percentile calculation in small		
	data sets is asymmetric in both tails.		
Target level	Select to express hazard characterisations at external or internal		
	exposure level.		
Risk type	The type of exposure considered in the assessment; acute (short		
	term) or chronic (long-term).		

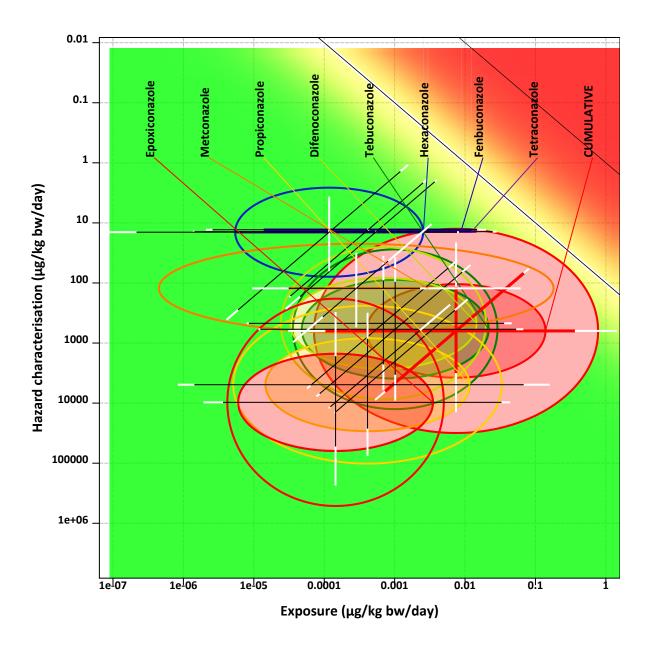


Figure 2.47: Hazard vs. exposure plot for multiple substances. 95% bivariate confidence areas for target hazard dose distribution and exposure distribution. Inner ellipses express variability, outer ellipses uncertainty.

Output settings

Name	Description		
Number of plot labels	Maximum number of labels to plot in hazard vs exposure plot.		
Number of substances in	Maximum number of substances to plot in hazard vs exposure		
hazard vs. exposure plot	plot.		
Left margin safety plot	Left margin of the plot for margins of exposure or hazard indices.		
Right margin safety plot	Right margin of the plot for margins of exposure of hazard		
	indices.		
Inclusion percentage variability	The central percentage of the variability distribution to include in		
interval	intervals for exposure, hazard and MOE (e.g. 90 means p5-p95).		
Include drill-down on 9	Specifies whether drilldown on 9 individuals is to be included in		
individuals around specified	the output.		
percentile.			
Summarize simulated data	Specifies whether a summary of the simulated consumptions and		
	concentrations should be included in the output.		
Store simulated individual day	Store the simulated individual day exposures. If unchecked, no		
exposures	additional output will be generated. If checked, the output will		
	contain an additional section with the simulated individual day		
	exposures.		
Show percentiles for	Give specific percentiles of exposure distribution (%), e.g. 50 90		
	95 97.5 99 (space separated).		
Percentage for drilldown	Gives detailed output for nine individuals near this percentile of		
	the exposure distribution.		
Percentage for upper tail	Gives detailed output for this upper percentage of the exposure		
	distribution.		
Number of levels of covariable	Specify the number of levels, e.g. 20. The range of the covariable		
to predict exposure	is divided by the number of levels: range = $(max - min)/levels$.		
Predict exposure at extra	For these covariable levels exposures are predicted. Specify specific prediction levels in addition to the automatically		
covariable levels	generated prediction levels (space separated).		
Lower percentage for	generated prediction levels (space separated). The default value of 25% may be overruled.		
variability (%)	The default value of 25 /0 may be overfuled.		
Upper percentage for	The default value of 75% may be overruled.		
variability (%)	The default value of 7570 may be overfuled.		
Report consumptions and	Specifies whether body weights should be ignored and		
exposures per individual			
instead of per kg body weight	Otherwise, the consumptions and exposures are per kg body		
r	weight.		

Table 2.190: Output settings for module Risks.

Calculation of risks

Risk (health impact) is quantified as exposure relative to hazard characterisation, which in MCRA is called a hazard index (HI) for any type of inputs, or as hazard characterisation relative to exposure, which in MCRA is called a margin of exposure (MOE) for any type of inputs. Exposures or hazards can be single values or distributions, the risk metric is a distribution if at least one of the inputs is a distribution (if both are single values, see the module Single value risks). Risk metrics are valid for a specific biological level (external or internal at a specific organ).

• Risks calculation

Inputs used: Exposures Hazard characterisations

Settings used

• Calculation Settings

Risks are expressed as distribution of margin of exposure or hazard index. The distribution is summarised by percentiles, and by the probability to exceed the specified threshold value (e.g. 1 or 100). The hazard vs. exposure plot compares the exposures and the hazard characterisation for individuals or individual-days in a population. Exposures, hazard characterisations and risks can be acute or chronic. The default unit for exposures and hazard characterisations is $\mu g/kgBW/day$, but this can be changed by choosing non-default units for consumptions, concentrations and/or body weight.

By using probabilistic tiers for both exposure and hazard characterisation, the calculated MOE distribution is equal to the Integrated Margin Of Exposure (IMOE) distribution, as described for the Integrated Probabilistic Risk Assessment (IPRA) approach in [[van der Voet et al., 2007]] and [[van der Voet et al., 2009]].

2.8.2 Single value risks

Single value risks are risk estimates obtained from combining single value exposures with single value hazard characterisations or as a percentile from a risk distribution.

This module has as primary entities: Substances Effects Populations

Single value risks calculation

Single value risks can be calculated in two ways.

- If the option 'Single value risk calculation method' is set to 'From single value risks' then *single value exposures* are combined with (single value) *hazard characterisations*.
- If the option 'Single value risk calculation method' is set to 'As percentile from risks distribution' then a percentile can be selected from a *risks* distribution.

See below for a more detailed explanation.

Combining single value exposures and hazard characterisations

Single value risks are computed by combining *single value exposures* by route/source and substance with (single value) *hazard characterisations* by substance. They are computed as margin of exposure (hazard characterisation / exposure), hazard quotient or hazard index (exposure / hazard characterisation), or as a percentage of the reference dose (100 * exposure / hazard characterisation).

Single value risks from individual risks

In this option, a percentage point can be specified for the chosen risk metric (margin of exposure (MOE) or hazard index (HI)). The corresponding percentile is calculated from the distribution of individual *risks*. The default percentiles are a margin of exposure at 0.1% or a hazard index at 99.9%, but another value can be chosen. It can also be indicated whether the risk metric is calculated using the inverse distribution or not. This option is provided because percentile calculation in small data sets is asymmetric in both tails. When this option is set, the percentile is calculated as the inverse of the complementary percentage of the inverse distribution. E.g., the $p_{0.1}$ of the MOE distribution is calculated as $1/(p_{99.9}$ of 1/MOE distribution); the $p_{99.9}$ of the HI distribution is calculated as $1/(p_{0.1}$ of 1/HI distribution).

Adjustment factors and uncertainty specification

Many sources of uncertainty that may affect input data, model assumptions and assessment methodology do not enter the assessment. In [[EFSA, 2020a]] and [[EFSA, 2020b]], thirty-four sources of uncertainty were identified and the impact of each source on the MOE was quantified. Some uncertainties tend to overestimate the MOE, others tend to underestimate it. Following the guidance of the EFSA Scientific Committee, specific MOE and/or HI percentiles are adjusted using adjustment factors for exposure and hazard, e.g. from expert elicitation. They may be available as fixed values or as parametric uncertainty distributions. In the nominal run, the percentile is adjusted with the median of the uncertainty distribution. In each uncertainty run, adjustment factors are sampled from the uncertainty distribution is specified. The available parametric uncertainty distributions are the same as available in the SHELF package that was used by EFSA. The SHeffield ELicitation Framework (SHELF) is a package of documents, templates and software to carry out elicitation of probability distributions for uncertain quantities from a group of experts (http://www.tonyohagan.co.uk/shelf/).

Options for specifying uncertainty distributions are:

- Lognormal(μ , s) with offset c. Parameters μ and s specify the mean and standard deviation of the underlying normal.
- Log Student t(μ, s, ν) with offset d. Parameters μ and s specify the mean and standard deviation of the underlying normal, ν the degrees of freedom, ν > 0
- Beta(a, b) scaled to the interval [c, d], with shape parameters a and b > 0.
- Gamma(a, b) with offset c, with shape and rate parameters a and b > 0.

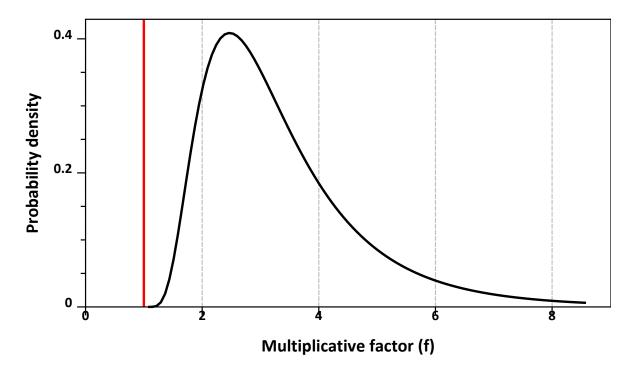


Figure 2.48: Scaled lognormal ($\mu = 0.705$, s = 0.566, offset=1), table 8, EFSA 2020 [[EFSA, 2020b]].

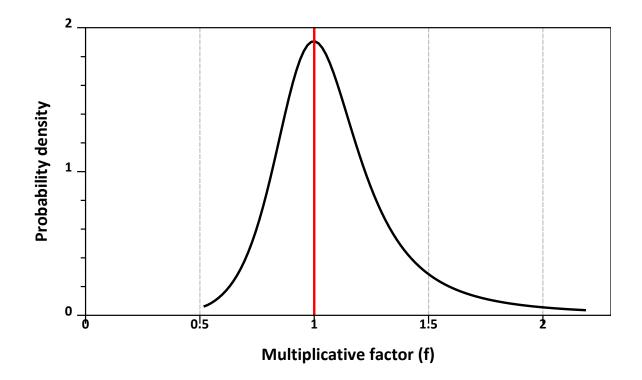


Figure 2.49: Scaled logstudents t ($\mu = -0.593$, s = 0.367, $\nu = 3$, offset=0.5), table 9, EFSA 2020 [[EFSA, 2020b]].

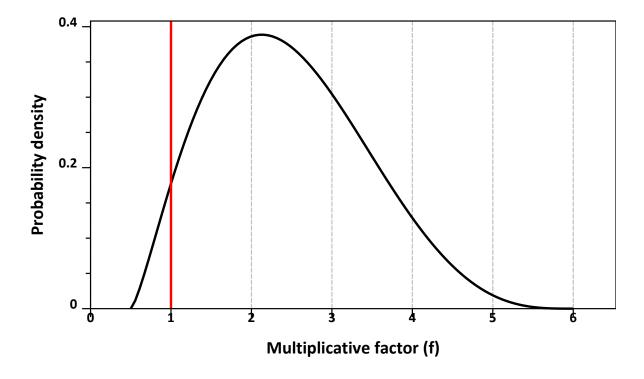


Figure 2.50: Scaled beta (a=2.37, b=4.26, lowerbound=0.5, upperbound=6), table 7, EFSA 2020 [[EFSA, 2020a]].

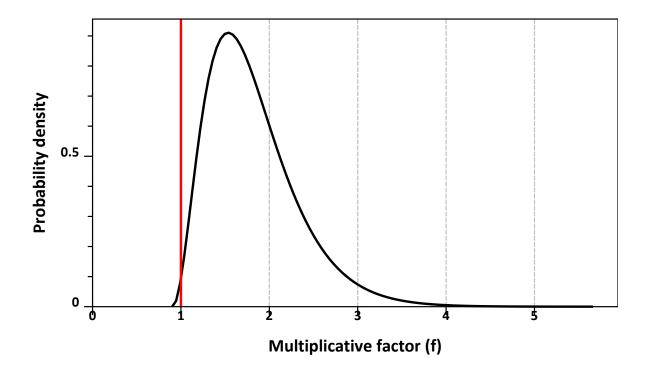


Figure 2.51: Scaled gamma (a=3.26, b=3.56, offset=0.9), table 6, EFSA 2020 [[EFSA, 2020a]].

Background-only adjustment factor

When exposures are calculated by *combining focal food/substance concentrations with background concentrations*, it may be appropriate to have a separate adjustment for the foreground and background. A pragmatic solution agreed with EFSA is to estimate the contribution of the foreground in the tail above the selected percentile. Suppose this contribution is c. Note that c will vary in uncertainty runs. Then, the adjustment factor should be multiplied by (1-c), i.e. no adjustment for the focal part.

The calculation proceeds as follows:

$p_{\mathrm{MOE,adjusted}}$	$= p_{\text{MOE}} \cdot (c + (1 - c) \cdot \text{AdjustmentFactor}_{\text{exposure}} \cdot \text{AdjustmentFactor}_{\text{hazard}})$
$p_{\mathrm{HI,adjusted}}$	$=$ $p_{\rm HI}$
r mi,adjusted	$-\frac{1}{c+(1-c)}$ · AdjustmentFactor _{exposure} · AdjustmentFactor _{hazard}

Note that when the focal substance measurements are converted to active substances using *substance conversions* or *deterministic substance conversions*, then *c* is the sum of the contributions of the focal food in and all active substances to which the substance translates.

In Figure 2.52, an example is shown where the margin of exposure is adjusted for the exposure and hazard distribution based on expert elicitation. The median adjustment factors for exposure and hazard are respectively, 1.77 and 3.01. The overall adjustment factor is 5.33.

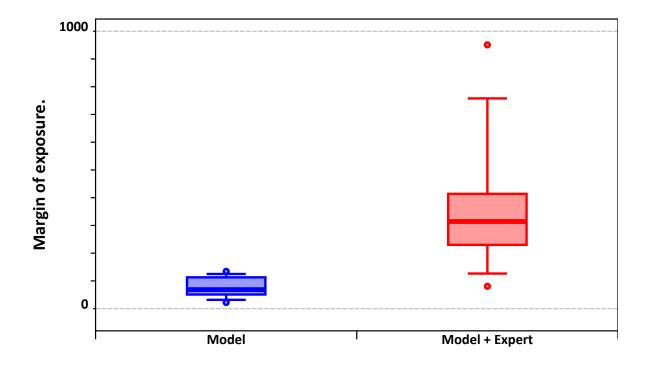


Figure 2.52: Margin of exposure (model) and adjusted margin of exposure (model + expert) with uncertainty bounds.

Single value risks settings

Calculation settings

Name Multiple substances analysis	Description Specifies whether the assessment involves multiple substances.			
Express results in terms of reference substance equivalents	Specifies whether the assessment involves multiple substances. Specifies whether the assessment involves multiple substances and results should be cumulated over all substances.			
(cumulative) Risk type	The type of exposure considered in the assessment; acute (short			
	term) or chronic (long-term).			
Risk metric type	Report risks in terms of hazard index (HI = 1/MOE) or margin of exposure.			
Single value risk calculation method	Calculate single value from exposures and hazard or from an individual risks distribution.			
Percentage for percentile	Percentage for percentile (default 0.1 for MOE or 99.9 for HI).			
Use inverse distribution to calculate percentile	Calculate percentile via the complementary percentage of the inverse distribution (default: no). Description: E.g., P0.1 of MOE distribution is calculated via P99.9 of 1/MOE distribution. Note: This option is provided because percentile calculation in small data sets is asymmetric in both tails.			
Apply adjustment factors to the specified risk percentile	 Specify adjustment factors, e.g. based on expert knowledge elicitation, to a specified MOE percentile (default 0.1%). If the selected risk metric is HI, the adjustment factors should still b specified for the complementary percentile of MOE (e.g. P0. MOE if P99.9 of HI is selected). 			
Adjustment type related to exposure	Specify the factor and/or distribution of the adjustment factor for the MOE percentile. Default is no adjustment. Alternatives are a fixed factor or an uncertainty distribution. If distributions are selected, default values are set based on EFSA cumulative risk reports 2020.			
Parameter A (Fixed factor, mean Lognormal or LogStudent-t, or shape parameter Beta or Gamma)	This parameter can be: 1) the fixed adjustment factor; 2) for Lognormal or LogStudent-t, the mean of the underlying normal distribution; 3) For Beta or Gamma. the shape parameter.			
Parameter B (standard	This parameter can be: 1) for Lognormal or LogStudent-t, the			
deviation Lognormal or	standard deviation of the underlying normal distribution; 2) For			
LogStudent-t or second shape parameter Beta or rate	Beta, the second shape parameter; 3) for Gamma, the rate parameter.			
parameter Gamma) Parameter C (Lower bound	This parameter can be: 1) for Beta, the lower bound value; 2) for			
Beta, offset Gamma or Lognormal or degrees of freedom Logstudent-t)	Gamma or Lognormal, the offset; 3) for LogStudent-t, the degrees of freedom.			
Parameter D (Upper bound Beta or offset LogStudent-t)	This parameter can be: 1) for Beta, the upper bound value; 2) for LogStudent-t, the offset.			
Adjustment type related to hazard	Specify the factor and/or distribution of the adjustment factor for the MOE percentile. Default is no adjustment. Alternatives are a fixed factor or an uncertainty distribution. If distributions are selected, default values are set based on EFSA cumulative risk			
	reports 2020.			
Parameter A (Fixed factor,	This parameter can be: 1) the fixed adjustment factor; 2) for			
mean Lognormal or LogStudent-t, or shape parameter Beta or Gamma)	Lognormal or LogStudent-t, the mean of the underlying normal distribution; 3) For Beta or Gamma. the shape parameter.			
Parameter B (standard	This parameter can be: 1) for Lognormal or LogStudent-t, the			
deviation Lognormal or	standard deviation of the underlying normal distribution; 2) For			
LogStudent-t or second shape parameter Beta or rate	Beta, the second shape parameter; 3) for Gamma, the rate parameter.			
Riskameter (Fashma)				
Parameter C (Lower bound Beta, offset Gamma or	This parameter can be: 1) for Beta, the lower bound value; 2) for Gamma or Lognormal, the offset; 3) for LogStudent-t, the degrees			
Lognormal or degrees of freedom Logstudent t)	of freedom.			

Table 2.191: Calculation settings for module Single value risks.

Calculation of single value risks

Single value risk can be computed by route and substance in the form of hazard quotients or margins of exposure. Single value risks are risk estimates obtained from combining single value exposures with single value hazard characterisations or as a percentile from a risk distribution. Optionally, the exposure and hazard behind the percentile calculation can be adjusted with fixed adjustment factors or factors drawn from parametric uncertainty distributions.

• Single value risks calculation

Inputs used: Single value dietary exposures Hazard characterisations Risks

Settings used

• Calculation Settings

continues on next page

280

			nued from previ	
Category	Module	Inputs	Used by	Description
			view of MCRA m	
Category	Module	Inputs	Used by	Description
Primary	Foods		Consump-	Foods are uniquely defined
entities			tions, Single	sources of dietary exposure to
			value con-	chemical substances. Foods
			sumptions, Market	may refer to 1) foods as eaten, foods as coded in food
			shares, Food	consumption data (e.g. pizza);
			recipes, Con-	2) modelled foods, foods as
			centrations,	coded in concentration data
			Concentra-	(e.g. wheat, tomato); 3) any
			tion	other type of food (e.g.
			distributions,	ingredients like flour, tomato
			Single value	sauce).
			concentra-	
			tions,	
			Processing	
			factors, Unit	
			variability	
			factors,	
			Occurrence	
			patterns, Occurrence	
			frequencies,	
			Substance	
			authorisa-	
			tions,	
			Deterministic	
			substance	
			conversion	
			factors, Con-	
			centration	
			limits, Con-	
			centration	
			models,	
			Modelled foods, Focal	
			food concen-	
			trations,	
			Total diet	
			study sample	
			composi-	
			tions, Food	
			extrapola-	
			tions, Food	
			conversions,	
			Consump-	
			tions by	
			modelled	
			food, High	
			exposure food-	
			substance	
			combina-	
			tions,	
			Dietary	
			exposures,	
			Single value	
			dietary	Chapter 2. Module
			exposures,	
			Exposures,	
			Exposure	

Table 2.192 - continued from previous page

Category	Module	Inputs	tinued from previ	Description
	Substances		Concentra-	Substances are chemical
			tions,	entities that can refer to: 1)
			Concentra-	active substances such as
			tion	investigated in toxicology; 2)
			distributions,	measured substances such as
			Single value	defined in specific analytical
			concentra-	methods. MCRA assessments
			tions,	can have one or more
			Processing	substances as the scope. When
			factors, Unit	more than one substance is
			variability	specified, there is an option to
			factors,	perform a cumulative
			Occurrence	assessment. In that case one of
			patterns,	the substances has to be
			Occurrence	indicated as the index/reference
			frequencies,	substance, and results will be
			Substance	expressed in equivalents of the index substance.
			authorisa- tions,	muex substance.
			substance	
			conversions,	
			Deterministic	
			substance	
			conversion	
			factors, Con-	
			centration	
			limits, Con-	
			centration	
			models,	
			Modelled	
			foods, Focal	
			food concen-	
			trations,	
			Food	
			conversions,	
			Consump-	
			tions by	
			modelled	
			food, High	
			exposure	
			food- substance	
			combina-	
			tions,	
			Dietary	
			exposures,	
			Single value	
			dietary	
			exposures,	
			Non-dietary	
			exposures,	
			Exposures,	
			Exposure	
			mixtures,	
			Human	
			monitoring	
			data,	
			Human	
sk modules			monitoring	
			analysis,	
			QSAR membership	
	1	i .		

Category	Module	Inputs	Used by	Description
	Effects	•	Concentra-	Effects are biological or
			tion models,	toxicological consequences for
			High	human health, that may result
			exposure	from chemical exposure and
			food-	are the focus of hazard or risk
			substance	assessment.
			combina-	
			tions,	
			Dietary	
			exposures,	
			Exposure	
			mixtures,	
			QSAR	
			membership	
			models,	
			Molecular	
			docking	
			models,	
			Active	
			substances,	
			Relative	
			potency	
			factors,	
			Hazard	
			characteri-	
			sations,	
			Points of	
			departure,	
			Effect repre-	
			sentations,	
			Inter-species	
			conversions,	
			Intra species	
			factors, AOP	
			networks,	
			Risks, Single	
			value risks.	continues on next nade

Table 2.192 - continued from previous page

			ued from previ	
Category	Module	Inputs	Used by	Description
	Populations		Consump-	Populations are groups of
			tions, Single	human individuals that are the
			value con-	scope of exposure or risk
			sumptions,	assessments. Optional
			Consump-	descriptors of populations are
			tions by	location (e.g. a country), time
			modelled	period (start date, end date),
			food,	age range and gender.
			Dietary	Example: the French
			exposures,	population in 2005-2007 of
			Single value	women of child-bearing age
			dietary	(18-45 yr).
			exposures,	
			Non-dietary	
			exposures,	
			Exposures,	
			Human	
			monitoring	
			analysis,	
			Risks, Single	
			value risks.	
	Test systems		Responses,	Test systems are biological or
	i est systemis		Dose	artificial systems used for
			response	assessing hazard in relation to
			models,	chemical exposure from
			Dose	substances in varying doses.
				Test systems may refer to 1)
			response data.	
			aaa.	in-vivo test systems (e.g. a rat
				90-day study, a human
				biomonitoring study); 2)
				in-vitro test systems (e.g.
	D		D	HepaRG cells).
	Responses	Test systems.	Dose	Responses are measurable
			response	entities in test systems.
			models,	Responses are used to
			Dose	represent effects (see effect
			response	representations) and their
			data, Effect	measured values are collected
			representa-	in dose response data.
		N N N	tions.	
Consumption	Consump-	Populations,	Food	Consumptions data are the
	tions	Foods.	conversions,	amounts of foods consumed on
			Consump-	specific days by individuals in a
			tions by	food consumption survey. For
			modelled	acute exposure assessments,
			food.	the interest is in a population of
				person-days, so one day per
				individual may be sufficient.
				For chronic exposure
				assessments, the interest is in a
				population of persons, so
				preferably two or more days
				per individual are needed.
			I	continues on next page

Table 2.192 - continued from previous page

		2.192 - contin		
Category	Module	Inputs	Used by	Description
	Single value	Consump-	Single value	Single value consumption data
	consump-	tions by	dietary	are the single value amounts
	tions	modelled	exposures.	(Large Portion, Mean
		food.		Consumption,
				p97.5Consumption) of
				modelled foods
				(foods-as-measured) consumed
				in a population.
	Market	Foods.	Food	Market shares data specify for
	shares		conversions.	a given food, percentages of
				more specific foods (subfoods,
				e.g. brands) representing their
				share in a market. Market
				shares are used when
				consumption data are available
				at a more generalised level than
				concentration data.
	Food recipes	Foods.	Food	Food recipes data specify the
			conversions.	composition of specific foods
				(typically: foods-as-eaten) in
				terms of other foods
				(intermediate foods or
				foods-as-measured) by
				specifying proportions in the
				form of a percentage.
Occurrence	Concentra-	Foods,	Single value	Concentrations data are
	tions	Substances,	concentra-	analytical measurements of
		Focal food	tions,	chemical substances occurring
		concentra-	Occurrence	in food samples. In their
		tions, Food	patterns,	simplest form, concentration
		extrapola-	Concentra-	data can just be used as
		tions,	tion models.	provided by datasets.
		Substance	Modelled	Optionally, concentrations data
			£] .	
		conversions.	TOOAS.	can be manipulated for active
		conversions, Deterministic	foods.	can be manipulated for active substances, extrapolated to
		Deterministic	<i>J00as</i> .	substances, extrapolated to
			<i>Jooas</i> .	_
		Deterministic substance conversion	Jooas.	substances, extrapolated to other foods, and/or default
		Deterministic substance	<i>Jooas</i> .	substances, extrapolated to other foods, and/or default
		Deterministic substance conversion factors, Relative	<i>Jooas</i> .	substances, extrapolated to other foods, and/or default
		Deterministic substance conversion factors, Relative potency	jooas.	substances, extrapolated to other foods, and/or default
		Deterministic substance conversion factors, Relative	<i>Jooas</i> .	substances, extrapolated to other foods, and/or default
		Deterministic substance conversion factors, Relative potency factors,	<i>Jooas</i> .	substances, extrapolated to other foods, and/or default
		Deterministic substance conversion factors, Relative potency factors, Substance authorisa-	<i>Jooas</i> .	substances, extrapolated to other foods, and/or default
		Deterministic substance conversion factors, Relative potency factors, Substance authorisa- tions, Active	<i>Jooas</i> .	substances, extrapolated to other foods, and/or default
		Deterministic substance conversion factors, Relative potency factors, Substance authorisa- tions, Active substances,	jooas.	substances, extrapolated to other foods, and/or default
		Deterministic substance conversion factors, Relative potency factors, Substance authorisa- tions, Active substances, Concentra-	jooas.	substances, extrapolated to other foods, and/or default
	Concentra-	Deterministic substance conversion factors, Relative potency factors, Substance authorisa- tions, Active substances, Concentra- tion limits.		substances, extrapolated to other foods, and/or default values can be added for water.
	Concentra- tion	Deterministic substance conversion factors, Relative potency factors, Substance authorisa- tions, Active substances, Concentra- tion limits. Foods,	Concentra-	substances, extrapolated to other foods, and/or default values can be added for water.
	tion	Deterministic substance conversion factors, Relative potency factors, Substance authorisa- tions, Active substances, Concentra- tion limits.	Concentra- tion models,	substances, extrapolated to other foods, and/or default values can be added for water.
		Deterministic substance conversion factors, Relative potency factors, Substance authorisa- tions, Active substances, Concentra- tion limits. Foods,	Concentra-	substances, extrapolated to other foods, and/or default values can be added for water.

Table 2.192 – continued from previous page

Category	Module	Inputs	Used by	Description
<u> </u>	Single value	Active	Modelled	Single value concentrations
	concentra-	substances,	foods, Single	data are the single value
	tions	Concentra-	value dietary	estimates (High Residue,
		tions,	exposures.	Maximum Residue Limit,
		Concentra-		Supervised Trials Median
		tion limits,		Residue) of residue
		Deterministic		concentrations on modelled
		substance		foods.
		conversion		
		factors.		
	Processing	Foods,	Food	Processing factors are
	factors	Substances.	conversions,	multiplication factors to derive
			Dietary	the concentration in a
			exposures,	processed food from the
			Single value	concentration in an
			dietary	unprocessed food and can be
			exposures.	specified for identified
				processing types (e.g., cooking,
				washing, drying). Processing
				factors are primarily used in
				dietary exposure assessments
				to correct for the effect of
				processing on substance
				concentrations in dietary
				exposure calculations.
	Unit	Foods,	Dietary	Unit variability factors specify
	variability	Substances.	exposures,	the variation in concentrations
	factors		Single value	between single units of the
			dietary	same food, which have been
			exposures.	put together in a mixture
				sample on which the
				concentration measurements
				have been made. Unit
				variability factors are used to
				account for the fact that
				concentration data often relate
				to composite samples, whereas
				an acute risk may result from
				single food units.

Table 2.192 - cont	inued from previous page	
--------------------	--------------------------	--

Onterio		2.192 – contin		
Category	Module	Inputs	Used by	Description
	Occurrence	Substance	Occurrence	Occurrence patterns (OPs) are
	patterns	authorisa-	frequencies,	the combinations (or mixtures)
		tions, Active	Dietary	of substances that occur
		substances,	exposures.	together on foods and the
		Concentra-		frequencies of these mixtures
		tions.		occurring per food, expressed
				in percentages. In the context
				of pesticides, occurrence
				patterns are associated with
				agricultural use percentages.
				Occurrence patterns are
				relevant to account for
				co-occurrence of active
				substances in exposed
				individuals. Occurrence
				patterns may be specified as
				data or modelled based on
				observed patterns of positive
				concentrations.
	Occurrence	Active	Concentra-	Occurrence frequencies specify
	frequencies	substances,	tion models,	how often substances occur on
		Occurrence	Single value	foods. Frequencies are
		patterns.	dietary	expressed as percentages.
		^	exposures.	
	Substance	Foods,	Concentra-	Substance authorisations
	authorisa-	Substances.	tions,	specify which food/substance
	tions		Occurrence	combinations are authorised
			patterns,	for (agricultural) use. If
			Concentra-	substance authorisations are
			tion models.	used, then only the
				food/substance combinations
				that are specified in the data
				are assumed to be authorised
				and all other combinations are
				assumed to be not authorised.
				This information may, for
				instance, be used to determine
				whether concentration
				measurements below the LOR
				could be assumed true zeros.
				I.e., if a food/substance
				combinations is assumed to be
				unauthorised, then the LOR
				may be assumed to be a zero.
	Substance	Substances,	Concentra-	Substance conversions specify
	conversions	Active	tions.	how measured substances are
	Conversions	substances.		converted to active substances,
		substances.		which are the substances
				assumed to cause health effects.
				In the pesticide legislation such
				measured substances and the
				substance conversion rules are
				known as residue definitions.

Table 2.192 – continued from previous page

Table 2.192 – continued from previous page				
Category	Module	Inputs	Used by	Description
	Deterministic substance conversion factors	Substances, Foods.	Concentra- tions, Single value con- centrations.	Deterministic substance conversion factors.
	Concentra- tion limits	Foods, Substances.	Concentra- tions, Single value con- centrations, Concentra- tion models, Modelled foods.	Concentration limits specify (legal) limit values for substance concentrations on foods and are sometimes used as conservative values for concentration data. In the framework of pesticides the legal Maximum Residue Limit (MRL) is the best known example.
	Concentra- tion models	Concentra- tions, Concentra- tion limits, Modelled foods, Substance authorisa- tions, Occurrence frequencies, Relative potency factors, Con- centration distributions, Total diet study sample composi- tions.	High exposure food- substance combina- tions, Dietary exposures.	Concentration models are distributional models of substance concentrations on foods. They describe both the substance presence (yes/no, with no representing an absolute zero concentration) and the substance concentrations. Concentration models are specified per food/substance combination.
	Modelled foods	Concentra- tions, Single value con- centrations, Concentra- tion limits.	Concentra- tion models, Food conversions.	Modelled foods are foods within the foods scope for which concentration data or MRLs of substances are available (or expected).
	Focal food concentra- tions	Foods, Substances.	Concentra- tions.	In some cases the attention in an assessment is on a specific food (focal food), against the background of other foods. Focal food concentrations are separate concentration data for one or more focal food commodities, that will take the place of any other concentration data for the focal food in the ordinary concentration data.

Table	2.192 -	continued from	previous page
-------	---------	----------------	---------------

Table 2.192 – continued from previous page					
Category	Module	Inputs	Used by	Description	
	Total diet	Foods.	Concentra-	Total diet study sample	
	study sample		tion models,	compositions specify the	
	compositions		Food	composition of mixed food	
			conversions.	samples, such as used in a total	
				diet study (TDS), in terms of	
				their constituting foods.	
	Food extrap-	Foods.	Concentra-	Food extrapolations data	
	olations		tions, Food	specify which foods (data rich	
			conversions.	foods) can be used to impute	
				concentration data for other	
				foods with insufficient data	
				(data poor foods).	
Exposure	Food	Consump-	Consump-	Food conversions relate	
	conversions	tions,	tions by	foods-as-eaten, as found in the	
		Modelled	modelled	consumption data, to modelled	
		foods,	food,	foods (foods-as-measured),	
		Processing	Dietary	which are the foods for which	
		factors,	exposures.	concentration data are	
		Food recipes,		available. A food-as-eaten can	
		Market		be linked to one, or multiple	
		shares, Food		modelled foods using various	
		extrapola-		conversion steps (e.g., using	
		tions, Total		food recipes to translate a	
		diet study		composite food into its	
		sample com-		ingredients). There are several	
		positions,		types of conversion steps, and a	
		Active		conversion path may comprise	
		substances.		multiple conversion steps	
				between a food-as-eaten and a	
	C	C	C'a a la combra	modelled food.	
	Consump-	Consump-	Single value	Consumptions by modelled	
	tions by modelled	tions, Food conversions.	consump-	food are consumptions of individuals expressed on the	
		conversions.	tions, High	level of the foods for which	
	food		exposure		
			food- substance	concentration data are available (i.e., the modelled-foods).	
			combina-	(i.e., the modelled-loods). These are calculated from	
			tions, Dietary	consumptions of foods-as-eaten and food	
			exposures.	conversions that link the	
			corposures.	foods-as-eaten amounts to	
				modelled-foods amounts.	
	High	Consump-	Dietary	Identification of	
	exposure	tions by	exposures.	food-as-eaten/modelled	
	food-	modelled	corposition.	food/substance combinations	
	substance	food, Con-		that have the highest expected	
	combina-	centration		contribution to exposure based	
	tions	models,		on a simple screening model.	
	10103	Active		on a simple screening model.	
		substances,			
		Relative			
		potency			
		factors.			
		juciois.		continues on next page	

Table 2.192 – continued from previous page

Category	Module	2.192 – contini Inputs	Used by	Description
Jaleyory	Dietary	Consump-	Exposures.	Dietary exposures are the
	Dietary exposures	tions by	Exposures.	amounts of substances,
	exposures	modelled		,
				expressed per kg bodyweight or per individual, to which
		food, Con-		1 ·
		centration		individuals in a population are
		models,		exposed from their diet per
		Processing		day. Depending on the
		factors, Unit		exposure type, dietary
		variability		exposures can be
		factors, High		short-term/acute exposures and
		exposure		then contain exposures for
		food-		individual-days, or they can be
		substance		long-term/chronic exposures,
		combina-		in which case they represent
		tions, Active		the average exposure per day
		substances,		over an unspecified longer time
		Occurrence		period.
		patterns,		
		Relative		
		potency		
		factors,		
		Food .		
		conversions,		
		Concentra-		
		tion		
		distributions.		
	Single value	Single value	Single value	Single value dietary exposures
	dietary	consump-	risks.	are based on the single value
	exposures	tions, Single		concentrations of substances,
		value con-		expressed per standard (kg)
		centrations,		bodyweight and/or single value
		Processing		amounts of consumed
		factors, Unit		modelled food. Depending on
		variability		the exposure type, dietary
		factors,		exposures can be
		Occurrence		short-term/acute exposures.
		frequencies.		
	Non-dietary	Populations,	Exposures.	Non-dietary exposures are the
	exposures	Substances,		amounts of substances to which
		Active		individuals in a population are
		substances.		exposed via any of three
				non-dietary routes: dermal,
				inhalation or oral, per day.

Catogory	Module	2.192 – contin	Used by	
Category		Inputs		Description
	Exposures	Dietary	Exposure	Exposures are amounts of
		exposures,	mixtures,	substances, typically expressed
		Non-dietary	Human	per mass unit and per day, to
		exposures,	monitoring	which individuals in a
		Active	analysis,	population are exposed at a
		substances,	Risks.	chosen target level. This target
		Relative		level may be external exposure
		potency		(dietary exposure, expressed
		factors,		per unit body weight, or per
		Kinetic		person) or internal exposure
		models.		(expressed per unit organ
				weight). Internal exposures
				may be aggregated from
				dietary and non-dietary
				exposures using either
				absorption factors or kinetic
				models to translate the external
				exposures to internal
				exposures. Exposures can be
				short-term/acute exposures and
				then contain exposures for
				individual-days, or they can be
				long-term/chronic exposures,
				in which case they represent
				the average exposure per day
				over an unspecified longer time
				period.
	Exposure	Exposures.		Exposure mixtures are
	mixtures			mixtures of substances that
				contribute relatively much to
				the overall cumulative exposure
				(potential risk drivers).
	Human	Substances.	Human	Human monitoring data
	monitoring		monitoring	quantify substance
	data		analysis.	concentrations found in
				humans collected in human
				monitoring surveys.
	Human	Human		Human monitoring analysis
	monitoring	monitoring		compares observed human
	analysis	data,		monitoring data with
		Exposures.		predictions made for the same
				population of individuals from
				dietary survey data,
				concentration data and
				(optionally) non-dietary
				exposure data.
	1	1	1	

Table 2.192 – continued from previous page

×	Module	Inputs	Used by	Description
In-silico	QSAR	Substances,	Active	QSAR membership models
	membership	Effects, AOP	substances.	specify assessment group
	models	networks.		memberships for active
				substances related to a specific
				health effect (adverse
				outcome). Memberships
				should be derived externally
				from Quantitative
				Structure-Activity Relationship
				(QSAR) models.
	Molecular	Substances,	Active	Molecular docking models
	docking	Effects, AOP	substances.	specify binding energies for
	models	networks.		substances in specific
				molecular docking models
				related to a specific health
				effect (adverse outcome).
Kinetic	Kinetic	Substances,	Exposures,	Kinetic models convert
	models	Active	Hazard	exposures or hazard
		substances.	characteri-	characterisations from one or
			sations.	more external routes or
				compartments to an internal
				(target) compartment. The
				reverse conversion from
				internal to external can also be
				made (reverse dosimetry).

Table 2.192 – continued from previous page

Category	Module	Inputs	Used by	Description
Hazard	Active	AOP	Concentra-	Active substances are
	substances	networks,	tions, Single	substances that may lead (P>0)
		Points of	value con-	to a specific health effect
		departure,	centrations,	(adverse outcome). Active
		Hazard	Occurrence	substances are specified
		characteri-	patterns,	directly as data or calculated
		sations,	Occurrence	from POD presence, QSAR
		Molecular	frequencies,	models or Molecular docking
		docking	Substance	models. Active substances can
		models,	conversions,	have an assessment group
		QSAR	Non-dietary	membership 1 (crisp), or
		membership	exposures,	values in the range (0,1]
		models.	Kinetic	(probabilistic).
			models,	
			Relative	
			potency	
			factors,	
			Hazard	
			characteri-	
			sations,	
			Inter-species	
			conversions,	
			Intra species	
			factors,	
			Food	
			conversions,	
			High	
			exposure	
			food-	
			substance	
			combina-	
			tions,	
			Dietary	
			exposures,	
			Exposures.	

Table 2.192 – continued from previous page

Category	Module	Inputs	Used by	Description
Calegory	Relative potency factors	Active substances, AOP networks,	Concentra- tions, Concentra- tion models,	Relative potency factors (RPFs) quantify potencies of substances with respect to a defined effect, relative to the
		Hazard characteri- sations.	High exposure food- substance combina- tions, Dietary exposures, Exposures.	potency of a chosen index substance. RPFs can be used to express combined exposures of multiple substances in terms of a the exposure value of the chosen index substance (i.e., in index substance equivalents). In MCRA, hazard characterisations, and therefore also RPFs are based on mass units (e.g., μ g), and not on mol units. RPFs can be different for different levels of the human organism (external, internal, specific compartment). RPFs can be given as data or computed from hazard characterisations. RPFs can be specified with uncertainty. Computation from uncertain hazard characterisations allows to include correlations between
				uncertain RPFs which originate from using the same index substance.

Table 2.192 – continued from previous page				
Category	Module	Inputs	Used by	Description
	Hazard	AOP	Active	Hazard characterisations are
	characteri-	networks,	substances,	reference exposure values for
	sations	Active	Relative	active substances at the chosen
		substances,	potency	biological target level (external
		Points of	factors,	or internal). Hazard
		departure,	Risks, Single	characterisations may be
		Dose	value risks.	specified for specific effects or
		response		for the critical effect as defined
		models,		in hazard characterisation.
		Effect repre-		Hazard characterisations are
		sentations,		specified as external values
		Inter-species		(e.g. human based guidance
		conversions,		values, such as ADI or ARfD)
		Intra species		or are based on points of
		factors,		departure, such as BMDs from
		Kinetic		dose-response models or
		models.		externally specified points of
				departure (NOAEL, LOAEL,
				MDS). The computation may
				involve assessment factors, e.g.
				for inter-species conversion,
				intra-species variation or
				additional sources of
				uncertainty. The calculation
				may also use kinetic models or
				absorption factors to convert
				external doses to internal doses
				or vice versa.
	Points of	Substances,	Active	Externally specified points of
	departure	Effects, AOP	substances,	departure can be used as an
	acpanta c	networks.	Hazard	alternative to calculated BMDs
		nerworks.	characteri-	from dose response models.
			sations.	Points of departure can be of
			suitons.	various types, such as NOAEL,
				LOAEL or BMD. They can be
				used to construct the list of
				active substances, to derive
				relative potency factors, and to
				perform health impact
	Dosa	Dose	Hazard	assessments.
	Dose		characteri-	Dose response models are
	response	response		models fitted to dose response
	models	data, Effect	sations.	data and can be provided as
		representa-		data or calculated using a local
		tions.		or remote version of PROAST.
				The main results for hazard
				and risk assessment are
				benchmark doses (BMDs),
				related to a specified substance,
				response, optionally covariate
				value, and the benchmark
				response (BMR).

Table 2.192 - continued from previous page

Category	Module	Inputs	Used by	Description
	Dose	Substances,	Dose	Dose response data are data on
	response	Test systems,	response	response values of test systems
	data	Responses.	models.	at specified doses of substances
				(or mixtures of substances)
				from dose response
				experiments.
	Effect repre-	Effects,	Hazard	Effect representations specify
	sentations	Responses,	characteri-	the responses that can be used
		AOP	sations, Dose	to measure specified effects
		networks.	response	and which response levels, the
			models.	benchmark response (BMR),
				define the hazard limits for the
				effects.
	Inter-species	Substances,	Hazard	Inter-species conversions
	conversions	Effects,	characteri-	specify how to convert a hazard
		Active	sations.	characterisation for a given
		substances.		species to a hazard
				characterisation for humans. In
				the simplest approach, this
				specifies a fixed inter-species
				factor. In a higher tier, this
				specifies a geometric mean
				(GM) and geometric standard
				deviation (GSD) for a
				lognormal uncertainty
				distribution of the interspecies
				factor. Inter-species conversion
				are specified per effect and can
				be general or
				substance-specific.
	Intra species	Substances,	Hazard	Intra-species factors specify
	factors	Effects,	characteri-	how to convert a hazard
		Active	sations.	characterisation from the
		substances.		average to a sensitive human
				individual.

Table 2.192 – continued from previous page

Category	Module	2.192 – contin Inputs	Used by	Description
	AOP	Effects.	QSAR	Effects are related to each
	networks		~ membership	other using the toxicological
			models,	concept of adverse outcome
			Molecular	pathways (AOPs) and adverse
			docking	outcome pathway networks
			models,	(see https://aopwiki.org).
			Active	Adverse Outcome Pathway
			substances,	(AOP) Networks specify how
			Relative	biological events (effects) can
			potency	lead to an adverse outcome
			factors,	(AO) in a qualitative way
			Hazard	through relations of upstream
			characteri-	and downstream key events
			sations,	(KEs), starting from molecular
			Points of	initiating events (MIEs). Using
			departure,	AOPs, the adverse outcome
			Effect repre-	(AO), e.g., liver steatosis, is
			sentations.	linked to key events (KEs),
				e.g., triglyceride accumulation
				in the liver, and to molecular
				initiating events (MIEs), e.g.,
				PPAR-alpha receptor
				antagonism. In general,
				multiple AOPs may lead to the
				same AO, and therefore AOP
				networks can be identified.
Risks	Risks	Exposures,	Single value	Risks (health impacts) are
		Hazard	risks.	defined as a function of
		characteri-		exposure and hazard
		sations.		characterisation at a chosen
				biological level (external or
				internal). Risk metrics are
				margins of exposure (MOE) or
				hazard indices (HI) or more
				generalised MOE or HI
				distributions.
	Single value	Single value		Single value risks are risk
	risks	dietary		estimates obtained from
		exposures,		combining single value
		Hazard		exposures with single value
		characteri-		hazard characterisations or as a
		sations,		percentile from a risk
		Risks.		distribution.

Table 2.192 – continued from previous page

STANDARD ACTIONS

A standard action is a user friendly way to perform a complex probabilistic calculation. By using a standard action predefined settings are used and the user can set only a limited number of selections. All settings (pre-defined and set by the user) are visible in the output. As a result a short output is presented. More detailed output is still available.

3.1 Chronic cumulative exposure assessment PFAS

This standard action is of type: Risks

This standard action can be used to calculate a chronic cumulative exposure for four PFAS. Assuming equipotency, and using proposed RPFs.

Table Group	Name	Repository	Туре
Survey	ConsumptionData-1-2yr.mdb	Standard Actions/Chronic cumulative expo- sure assessment PFAS	Fixed
Survey	ConsumptionData-10- 17yr.mdb	Standard Actions/Chronic cumulative expo- sure assessment PFAS	Fixed
Survey	ConsumptionData-18- 64yr.mdb	Standard Actions/Chronic cumulative expo- sure assessment PFAS	Fixed
Survey	ConsumptionData-3-9yr.mdb	Standard Actions/Chronic cumulative expo- sure assessment PFAS	Fixed
Survey	ConsumptionData-65- 74yr.mdb	Standard Actions/Chronic cumulative expo- sure assessment PFAS	Fixed
Survey	ConsumptionData- 75plusyr.mdb	Standard Actions/Chronic cumulative expo- sure assessment PFAS	Fixed
Concentrations	PFAS-Occurrencedata-EFSA- LB.mdb	Standard Actions/Chronic cumulative expo- sure assessment PFAS	Vari- able
Concentrations	PFAS-Occurrencedata-EFSA- UB.mdb	Standard Actions/Chronic cumulative expo- sure assessment PFAS	Vari- able
AssessmentGroup- Memberships	PFAS-OtherData.mdb	Standard Actions/Chronic cumulative expo- sure assessment PFAS	Fixed
Compounds	PFAS-OtherData.mdb	Standard Actions/Chronic cumulative expo- sure assessment PFAS	Fixed
Effects	PFAS-OtherData.mdb	Standard Actions/Chronic cumulative expo- sure assessment PFAS	Fixed
Foods	PFAS-OtherData.mdb	Standard Actions/Chronic cumulative expo- sure assessment PFAS	Fixed
HazardCharacterisa- tions	PFAS-OtherData.mdb	Standard Actions/Chronic cumulative expo- sure assessment PFAS	Fixed
RelativePotencyFac- tors	PFAS- RelativePotencyFactorsEquipote	Standard Actions/Chronic cumulative expo- ncysumelassessment PFAS	Vari- able
RelativePotencyFac- tors	PFAS- RelativePotencyFactorsProposal	Standard Actions/Chronic cumulative expo-	Vari- able

Table 3.1: Datasources for Chronic cumulative exposure assessment PFAS.

3.2 Chronic cumulative exposure assessment PA

This standard action is of type: Risks

In 2017 EFSA published a statement *Risks for human health related to the presence of pyrrolizidine alkaloids in honey, tea, herbal infusions and food supplements*, see [[EFSA, 2017b]]. Occurrence data used in this opinion was published in 2015 in the external scientific report *Occurrence of Pyrrolizidine Alkaloids in food*, see [[Mulder et al., 2015]]. The occurrence data in tea and herbal infusions is linked to the Consumption of 6 different population groups of the DNFCS 2012-2016. An lower bound (LB) and upper bound (UB) chronic cumulative exposure assessment can be calculated for different PAs. One scenario is assuming equipotency, another scenario is using provisional RPFs [[Merz et al., 2016]].

3.3 Demo acute cumulative risk assessment

This standard action is of type: Risks

In this demo with fictitious data, acute cumulative risk assessments can be performed following various calculation methods (EFSA 2012 Optimistic and Pessimistic, EC 2018 Tier 1 and Tier 2). Here, also the effect of applying processing factors can be assessed.

Table Group	Name	Repository	Туре
MaximumResidue-	DemoConcentrationLim-	Standard Actions/Demo Acute Cumulative Risk	Fixed
Limits	its.mdb	Assessment	
Concentrations	DemoConcentrations.mdb	Standard Actions/Demo Acute Cumulative Risk	Fixed
		Assessment	
Survey	DemoConsumptions.mdb	Standard Actions/Demo Acute Cumulative Risk	Fixed
		Assessment	
Effects	DemoEffects.mdb	Standard Actions/Demo Acute Cumulative Risk	Fixed
		Assessment	
FoodExtrapolations	DemoFoodExtrapola-	Standard Actions/Demo Acute Cumulative Risk	Fixed
	tions.mdb	Assessment	
FoodTranslations	DemoFoodRecipes.mdb	Standard Actions/Demo Acute Cumulative Risk	Fixed
		Assessment	
Foods	DemoFoods.mdb	Standard Actions/Demo Acute Cumulative Risk	Fixed
		Assessment	
HazardDoses	DemoPointsOfDepar-	Standard Actions/Demo Acute Cumulative Risk	Fixed
	ture.mdb	Assessment	
Processing	DemoProcessingFac-	Standard Actions/Demo Acute Cumulative Risk	Vari-
	tors.mdb	Assessment	able
Processing	DemoProcessingfac-	Standard Actions/Demo Acute Cumulative Risk	Vari-
	torOri.mdb	Assessment	able
AuthorisedUses	DemoSubstanceAuthorisa-	Standard Actions/Demo Acute Cumulative Risk	Fixed
	tions.mdb	Assessment	
ResidueDefinitions	DemoSubstanceConver-	Standard Actions/Demo Acute Cumulative Risk	Fixed
	sions.mdb	Assessment	
Compounds	DemoSubstances.mdb	Standard Actions/Demo Acute Cumulative Risk	Fixed
		Assessment	
UnitVariability	DemoUnitVar36.mdb	Standard Actions/Demo Acute Cumulative Risk	Vari-
		Assessment	able
UnitVariability	DemoUnitVarPRIMo.mdb	Standard Actions/Demo Acute Cumulative Risk	Vari-
		Assessment	able

Table 3.2: Datasources for Demo acute cumulative risk assessment.

3.4 EU acute cumulative exposure assessment (2018) Tier I and Tier II

This standard action is of type: Single value risks

This standard action is based on work done in 2018 [[van Klaveren et al., 2019a]]. In the context of the second framework partnership agreement between the National Institute for Public Health and the Environment of the Netherlands (RIVM) and the European Food Safety Authority (EFSA) acute cumulative dietary exposure assessments were performed for two cumulative assessment groups (CAGs) of pesticides that affect the nervous system: pesticides causing brain and/or erythrocyte AChE inhibition (CAG-NAN, 47 pesticides) and pesticides causing functional alterations of the motor division (CAG-NAM, 100 pesticides). The exposure assessments used monitoring data collected by the Netherlands under their official monitoring programmes in 2014, 2015 and 2016 and individual Dutch food consumption data. Exposure estimates were obtained for each group of pesticides using the MCRA software. The Standing Committee on Plants, Animals, Food and Feed (SC PAFF) discussed the scope of the assessment in 2018 and agreed on the parameters to be used for the cumulative exposure assessment. Based on that discussion, a very conservative tier I modelling approach and a refined, but still conservative tier II modelling approach were used. In these assessments, common risk assessment practice was followed and the cumulative exposure was expressed as the total margin of exposure (MOET) at the 50th, 90th, 95th, 99th and 99.9th percentile of the exposure distribution.

Table Group	Name	Repository	Туре
AssessmentGroup-	LeanCAGsAssessmentGroup-	Standard Actions/EU 2018 Acute Cumula-	Fixed
Memberships	Membership10%p999.xlsx	tive Exposure Assessment	
AuthorisedUses	SecondaryInputData.mdb	Standard Actions/EU 2018 Acute Cumula-	Fixed
		tive Exposure Assessment	
Compounds	SecondaryInputData.mdb	Standard Actions/EU 2018 Acute Cumula-	Fixed
-		tive Exposure Assessment	
Effects	SecondaryInputData.mdb	Standard Actions/EU 2018 Acute Cumula-	Fixed
		tive Exposure Assessment	
FoodExtrapolations	SecondaryInputData.mdb	Standard Actions/EU 2018 Acute Cumula-	Fixed
-		tive Exposure Assessment	
FoodTranslations	SecondaryInputData.mdb	Standard Actions/EU 2018 Acute Cumula-	Fixed
		tive Exposure Assessment	
Foods	SecondaryInputData.mdb	Standard Actions/EU 2018 Acute Cumula-	Fixed
		tive Exposure Assessment	
HazardDoses	SecondaryInputData.mdb	Standard Actions/EU 2018 Acute Cumula-	Fixed
		tive Exposure Assessment	
MaximumResidue-	SecondaryInputData.mdb	Standard Actions/EU 2018 Acute Cumula-	Fixed
Limits		tive Exposure Assessment	
Processing	SecondaryInputData.mdb	Standard Actions/EU 2018 Acute Cumula-	Fixed
-		tive Exposure Assessment	
ResidueDefinitions	SecondaryInputData.mdb	Standard Actions/EU 2018 Acute Cumula-	Fixed
		tive Exposure Assessment	
UnitVariability	UnitVar36.mdb	Standard Actions/EU 2018 Acute Cumula-	Vari-
		tive Exposure Assessment	able
UnitVariability	UnitVarPrimo.mdb	Standard Actions/EU 2018 Acute Cumula-	Vari-
		tive Exposure Assessment	able
Concentrations	a_ConcentrationsSSD_NAM.mdb	Standard Actions/EU 2018 Acute Cumula-	Vari-
		tive Exposure Assessment	able
Concentrations	a_ConcentrationsSSD_NAN.mdb	Standard Actions/EU 2018 Acute Cumula-	Vari-
		tive Exposure Assessment	able
Survey	a_ConsumptionsNL2.mdb	Standard Actions/EU 2018 Acute Cumula-	Fixed
•		tive Exposure Assessment	
Survey	a_ConsumptionsNL3_6.mdb	Standard Actions/EU 2018 Acute Cumula-	Fixed
-	· · -	tive Exposure Assessment	

Table 3.3: Datasources for EU acute cumulative exposure assessment (2018) Tier I and Tier II.

3.5 EU chronic cumulative exposure assessment (2018) Tier I and Tier II

This standard action is of type: Risks

This standard action is based on research done in 2018 [[van Klaveren et al., 2019b]].

This standard action will enable you to reproduce the exposure assessment of chronic cumulative effects of pesticide residues in food affecting the thyroid. These are retrospective exposure assessments of the cumulative exposure for the thyroid using monitoring data from 2014, 2015 and 2016. In this standard action Dutch monitoring and consumption data are used. The results, data used and methodology are reported in a scientific report following published on the EFSA website in September 2019. The methodology fulfils the requirements set by the European Commission.

Table Group	Name	Repository	Туре
AssessmentGroup-	LeanCAGsAssessmentGroup-	Standard Actions/EU 2018 Chronic Cumu-	Fixed
Memberships	Membership10%p999.xlsx	lative Exposure Assessment	
AuthorisedUses	SecondaryInputData.mdb	Standard Actions/EU 2018 Chronic Cumu-	Fixed
		lative Exposure Assessment	
Compounds	SecondaryInputData.mdb	Standard Actions/EU 2018 Chronic Cumu-	Fixed
		lative Exposure Assessment	
Effects	SecondaryInputData.mdb	Standard Actions/EU 2018 Chronic Cumu-	Fixed
		lative Exposure Assessment	
FoodExtrapola-	SecondaryInputData.mdb	Standard Actions/EU 2018 Chronic Cumu-	Fixed
tions		lative Exposure Assessment	
FoodTranslations	SecondaryInputData.mdb	Standard Actions/EU 2018 Chronic Cumu-	Fixed
		lative Exposure Assessment	
Foods	SecondaryInputData.mdb	Standard Actions/EU 2018 Chronic Cumu-	Fixed
		lative Exposure Assessment	
HazardDoses	SecondaryInputData.mdb	Standard Actions/EU 2018 Chronic Cumu-	Fixed
		lative Exposure Assessment	
MaximumResidue-	SecondaryInputData.mdb	Standard Actions/EU 2018 Chronic Cumu-	Fixed
Limits		lative Exposure Assessment	
Processing	SecondaryInputData.mdb	Standard Actions/EU 2018 Chronic Cumu-	Fixed
		lative Exposure Assessment	
ResidueDefinitions	SecondaryInputData.mdb	Standard Actions/EU 2018 Chronic Cumu-	Fixed
		lative Exposure Assessment	
Concentrations	c_ConcentrationsSSD_TCF.mdb	Standard Actions/EU 2018 Chronic Cumu-	Vari-
		lative Exposure Assessment	able
Concentrations	c_ConcentrationsSSD_TCP.mdb	Standard Actions/EU 2018 Chronic Cumu-	Vari-
		lative Exposure Assessment	able
Survey	c_ConsumptionsNL2.mdb	Standard Actions/EU 2018 Chronic Cumu-	Fixed
		lative Exposure Assessment	
Survey	c_ConsumptionsNL3_6.mdb	Standard Actions/EU 2018 Chronic Cumu-	Fixed
		lative Exposure Assessment	

Table 3.4: Datasources for EU chronic cumulative exposure assessment (2018) Tier I and Tier II.

3.6 IPGF Acute Cumulative Risk Assessment No Background

This standard action is of type: Risks

IPGF base action for single sample analyses without a background concentration dataset.

Bac	ckground.		
Table Group	Name	Repository	Туре
AssessmentGroup-	ActiveSubstances-	Standard Actions/IPGF Acute Cumulative Risk As-	Fixed
Memberships	Recoded.zip	sessment No Background	
Effects	Effects.xlsx	Standard Actions/IPGF Acute Cumulative Risk As- sessment No Background	Fixed
HazardCharacterisa- tions	HazardCharacterisations- Recoded.zip	Standard Actions/IPGF Acute Cumulative Risk As- sessment No Background	Fixed
Compounds	IpgfSubstances.accdb	Standard Actions/IPGF Acute Cumulative Risk As- sessment No Background	Fixed
FoodTranslations	NL-Otherdata.mdb	Standard Actions/IPGF Acute Cumulative Risk As- sessment No Background	Fixed
Foods	NL-Otherdata.mdb	Standard Actions/IPGF Acute Cumulative Risk As- sessment No Background	Fixed
Survey	NL-VCP2007-2010.mdb	Standard Actions/IPGF Acute Cumulative Risk As- sessment No Background	Fixed
Processing	ProcessingFac- torsRIVM.xlsx	Standard Actions/IPGF Acute Cumulative Risk As- sessment No Background	Fixed
AuthorisedUses	SubstanceAuthorisations- Recoded.zip	Standard Actions/IPGF Acute Cumulative Risk As- sessment No Background	Fixed
ResidueDefinitions	SubstanceConversions- Recoded.zip	Standard Actions/IPGF Acute Cumulative Risk As- sessment No Background	Fixed
UnitVariability	UnitVar36.mdb	Standard Actions/IPGF Acute Cumulative Risk As- sessment No Background	Fixed

Table 3.5: Datasources for IPGF Acute Cumulative Risk Assessment No
Background.

3.7 Risk steatosis from imazalil

This standard action is of type: Risks

In-vitro tests may help to reduce the use of test animals. What is the impact of this for risk assessment? Traditional risk assessment often uses animal data to evaluate toxicological limit values, such as the acceptable daily intake (ADI). In this standard action such a traditional risk calculation for the example of steatosis as health effect and imazalil as chemical substance can be compared to a similar calculation, but with the toxicological hazard characterisation based on in-vitro data from AdipoRed measurements in human liver cells. In-vitro concentrations are assumed to represent internal liver concentrations and a human physiologically based kinetic model for imazalil is used to extrapolate from in-vitro to in-vivo doses.

In this standard action, risk for steatosis from exposure to imazalil is estimated from in-vivo or in-vitro based hazard characterisations.

Two parameter sets are available:

- 1. estimates based on QSAR models only, and
- 2. estimates based on QSAR models and in-vitro experiments.

The latter parameter set shows that in the long term concentration levels of imazalil in the liver are stationary.

Table Group	Name	Repository	Туре
AdverseOut-	AOPN-Effects-EffectRelations-181017.xlsx	Standard Actions/Risk	Fixed
comePathwayNet-		steatosis from imazalil	
works			
Effects	AOPN-Effects-EffectRelations-181017.xlsx	Standard Actions/Risk	Fixed
		steatosis from imazalil	
DoseResponseData	BfR-HepaRG-AdipoRed-Single.xlsx	Standard Actions/Risk	Fixed
		steatosis from imazalil	
Concentrations	ConcentrationsSSD_20190129.zip	Standard Actions/Risk	Fixed
		steatosis from imazalil	
Compounds	EuroMix Substances Inventory (v8) (PPPs).zip	Standard Actions/Risk	Fixed
		steatosis from imazalil	
KineticModels	EuroMix-KineticModels-CosmosV6 9 sub-	Standard Actions/Risk	Vari-
	stances ParA QSAR.xlsx	steatosis from imazalil	able
KineticModels	EuroMix-KineticModels-CosmosV6 9 sub-	Standard Actions/Risk	Vari-
	stances ParB QSAR-Invitro.xlsx	steatosis from imazalil	able
UnitVariability	Foods coded in FoodEx1.mdb	Standard Actions/Risk	Fixed
-		steatosis from imazalil	
HazardCharacterisa-	HazardCharacterisation Imazalil.xlsx	Standard Actions/Risk	Fixed
tions		steatosis from imazalil	
Survey	NL-VCP-RPC 2005-2006 2-6yr.mdb	Standard Actions/Risk	Fixed
-		steatosis from imazalil	
Survey	NL-VCP-RPC 2007-2010 7-69yr.mdb	Standard Actions/Risk	Fixed
·		steatosis from imazalil	
Survey	NL-VCP-RPC 2010-2012 70+yr.mdb	Standard Actions/Risk	Fixed
·		steatosis from imazalil	
FoodTranslations	NL-VCP-RPC Foods.mdb	Standard Actions/Risk	Fixed
		steatosis from imazalil	
Foods	NL-VCP-RPC Foods.mdb	Standard Actions/Risk	Fixed
		steatosis from imazalil	
EffectRepresentations	TestSystems-Responses-	Standard Actions/Risk	Fixed
1	EffectRepresentations-181026.xlsx	steatosis from imazalil	
Responses	TestSystems-Responses-	Standard Actions/Risk	Fixed
ĩ	EffectRepresentations-181026.xlsx	steatosis from imazalil	
TestSystems	TestSystems-Responses-	Standard Actions/Risk	Fixed
	EffectRepresentations-181026.xlsx	steatosis from imazalil	
	1		

Table 3.6: Datasources for Risk steatosis from imazali	Table 3.6	Datasources	for Risk	steatosis f	from	imazalil.
--	-----------	-------------	----------	-------------	------	-----------

3.8 Training prospective risk assessment acute Tier II

This standard action is of type: Single value risks

This standard action allows you to run and compare the background and different Tier II approaches to probabilistic prospective risk assessment for a newly proposed use (focal substance/food) as defined by the EC working group. This is an acute risk assessment of a neurological effect (functional alternations of the motor division) for the Dutch toddlers, the Dutch other children, the Bulgarian other children and Italian adults populations. The focal food and substance combination that can be selected are emamectin with peach, acrinathrin with wheat (fictitious data) or pirimicarb with lettuces (fictitious data). The focal exposure can be based on a MRL, GAP or actual exposure scenario as defined by the working group, or it can be omitted (for comparison) so that only the background exposure is calculated. The 99.9th percentile of the exposure has been adjusted using uncertain adjustment factors from the EFSA cumulative risk assessment report 2020. The exposure is calculated probabilistically and the risk is expressed as margin of exposure (MOE) and is compared to a threshold value of 100.

Table Group	Name	Repository	Туре
FocalFoods	FocalConcentrations-	Standard Actions/Training Acute Prospec-	Vari-
i obuli obus	ActualScenario.mdb	tive Risk Assessment Tier II	able
FocalFoods	FocalConcentrations-	Standard Actions/Training Acute Prospec-	Vari-
	GapScenario.mdb	tive Risk Assessment Tier II	able
FocalFoods	FocalConcentrations-	Standard Actions/Training Acute Prospec-	Vari-
	MrlScenario.mdb	tive Risk Assessment Tier II	able
AssessmentGroupMem-	SecondaryInput-	Standard Actions/Training Acute Prospec-	Fixed
berships	DataMRL2016.mdb	tive Risk Assessment Tier II	
AuthorisedUses	SecondaryInput-	Standard Actions/Training Acute Prospec-	Fixed
	DataMRL2016.mdb	tive Risk Assessment Tier II	
Compounds	SecondaryInput-	Standard Actions/Training Acute Prospec-	Fixed
	DataMRL2016.mdb	tive Risk Assessment Tier II	
Effects	SecondaryInput-	Standard Actions/Training Acute Prospec-	Fixed
	DataMRL2016.mdb	tive Risk Assessment Tier II	
FoodExtrapolations	SecondaryInput-	Standard Actions/Training Acute Prospec-	Fixed
	DataMRL2016.mdb	tive Risk Assessment Tier II	
FoodTranslations	SecondaryInput-	Standard Actions/Training Acute Prospec-	Fixed
	DataMRL2016.mdb	tive Risk Assessment Tier II	
Foods	SecondaryInput-	Standard Actions/Training Acute Prospec-	Fixed
	DataMRL2016.mdb	tive Risk Assessment Tier II	
HazardDoses	SecondaryInput-	Standard Actions/Training Acute Prospec-	Fixed
	DataMRL2016.mdb	tive Risk Assessment Tier II	
MaximumResidueLimits	SecondaryInput-	Standard Actions/Training Acute Prospec-	Vari-
	DataMRL2016.mdb	tive Risk Assessment Tier II	able
Processing	SecondaryInput-	Standard Actions/Training Acute Prospec-	Fixed
	DataMRL2016.mdb	tive Risk Assessment Tier II	
ResidueDefinitions	SecondaryInput-	Standard Actions/Training Acute Prospec-	Fixed
	DataMRL2016.mdb	tive Risk Assessment Tier II	
MaximumResidueLimits	SecondaryInput-	Standard Actions/Training Acute Prospec-	Vari-
	DataMRL2019.mdb	tive Risk Assessment Tier II	able
DeterministicSubstance-	SingleValueCalcula-	Standard Actions/Training Acute Prospec-	Fixed
ConversionFactors	tions.mdb	tive Risk Assessment Tier II	
UnitVariability	UnitVar36.mdb	Standard Actions/Training Acute Prospec-	Fixed
		tive Risk Assessment Tier II	
Concentrations	a_ConcentrationsSSD_NAM	I.Stabbdard Actions/Training Acute Prospec-	Fixed
2		tive Risk Assessment Tier II	D : 1
Survey	a_ConsumptionsBU.mdb	Standard Actions/Training Acute Prospec-	Fixed
0		tive Risk Assessment Tier II	D ' 1
Survey	a_ConsumptionsIT.mdb	Standard Actions/Training Acute Prospec-	Fixed
0		tive Risk Assessment Tier II	D
Survey	a_ConsumptionsNL2.mdb	Standard Actions/Training Acute Prospec-	Fixed
0		tive Risk Assessment Tier II	D ' 1
Survey	a_ConsumptionsNL3_6.md	b Standard Actions/Training Acute Prospec-	Fixed
		tive Risk Assessment Tier II	

Table 3.7: Datasources for Training prospective risk assessment acute Tier II.

3.9 Training prospective risk assessment chronic Tier II

This standard action is of type: Single value risks

This standard action allows you to run and compare the background and different Tier II approaches to probabilistic chronic prospective risk assessment for a newly proposed use (focal substance/food) as defined by the EC working group. This is a chronic risk assessment of an effect (hypothyroidism) for the Dutch toddlers, Dutch other children and German adults populations. The focal food and substance combination that can be selected are cyprodinil with apples (fictitious data), valifenalate with lettuce or valifenalate with aubergines. The focal exposure can be based on a GAP or actual exposure scenario as defined by the working group, or it can be omitted (for comparison) so that only the background exposure is calculated. The 99.9th percentile of the exposure has been adjusted using uncertain adjustment factors from the EFSA cumulative risk assessment report 2020. The exposure is calculated probabilistically and. the risk is expressed as margin of exposure (MOE) and is compared to a threshold value of 100.

Table Group	Name	Repository	Туре
FocalFoods	FocalConcentrations-	Standard Actions/Training Chronic Prospec-	Vari-
	ActualScenario.mdb	tive Risk Assessment Tier II	able
FocalFoods	FocalConcentrations-	Standard Actions/Training Chronic Prospec-	Fixed
	GAPScenario.mdb	tive Risk Assessment Tier II	
FocalFoods	FocalConcentrations-	Standard Actions/Training Chronic Prospec-	Vari-
	GapScenario.mdb	tive Risk Assessment Tier II	able
AssessmentGroupMem-	SecondaryInput-	Standard Actions/Training Chronic Prospec-	Fixed
berships	DataMRL2016.mdb	tive Risk Assessment Tier II	
AuthorisedUses	SecondaryInput-	Standard Actions/Training Chronic Prospec-	Fixed
	DataMRL2016.mdb	tive Risk Assessment Tier II	
Compounds	SecondaryInput-	Standard Actions/Training Chronic Prospec-	Fixed
	DataMRL2016.mdb	tive Risk Assessment Tier II	
Effects	SecondaryInput-	Standard Actions/Training Chronic Prospec-	Fixed
	DataMRL2016.mdb	tive Risk Assessment Tier II	
FoodExtrapolations	SecondaryInput-	Standard Actions/Training Chronic Prospec-	Fixed
-	DataMRL2016.mdb	tive Risk Assessment Tier II	
FoodTranslations	SecondaryInput-	Standard Actions/Training Chronic Prospec-	Fixed
	DataMRL2016.mdb	tive Risk Assessment Tier II	
Foods	SecondaryInput-	Standard Actions/Training Chronic Prospec-	Fixed
	DataMRL2016.mdb	tive Risk Assessment Tier II	
HazardDoses	SecondaryInput-	Standard Actions/Training Chronic Prospec-	Fixed
	DataMRL2016.mdb	tive Risk Assessment Tier II	
MaximumResidueLimits	SecondaryInput-	Standard Actions/Training Chronic Prospec-	Vari-
	DataMRL2016.mdb	tive Risk Assessment Tier II	able
Processing	SecondaryInput-	Standard Actions/Training Chronic Prospec-	Fixed
0	DataMRL2016.mdb	tive Risk Assessment Tier II	
ResidueDefinitions	SecondaryInput-	Standard Actions/Training Chronic Prospec-	Fixed
	DataMRL2016.mdb	tive Risk Assessment Tier II	
MaximumResidueLimits	SecondaryInput-	Standard Actions/Training Chronic Prospec-	Vari-
	DataMRL2019.mdb	tive Risk Assessment Tier II	able
DeterministicSubstance-	SingleValueCalcula-	Standard Actions/Training Chronic Prospec-	Fixed
ConversionFactors	tions.mdb	tive Risk Assessment Tier II	
Concentrations		.rsttbndard Actions/Training Chronic Prospec-	Fixed
		tive Risk Assessment Tier II	
Survey	c_ConsumptionsDE.mdb	Standard Actions/Training Chronic Prospec-	Fixed
- J	r	tive Risk Assessment Tier II	
Survey	c_ConsumptionsNL2.mdb	Standard Actions/Training Chronic Prospec-	Fixed
	r nonor (22mildo	tive Risk Assessment Tier II	
Survey	c ConsumptionsNL3 6 md	bStandard Actions/Training Chronic Prospec-	Fixed
		tive Risk Assessment Tier II	I Incu

Table 3.8: Datasources for Training prospective risk assessment chronic Tier II.

3.10 Training substance prioritisation acute neuro

This standard action is of type: Risks

This standard action allows to prioritise substances to be included in a cumulative risk assessment by inspecting a list of (semi-)probabilistic hazard quotients, ranked from high to low. This is a risk assessment for either all neurotoxicological effects in the Dutch toddler population (level 1), or one of the effect subgroups (NAM or NAN) (level 2). The hazard quotient (termed hazard index in MCRA) compares a percentile from the dietary exposure distribution (P99 or P99.9 can be selected) to the ARfD, which a fixed value for the hazard characterisation (hence the designation as semi-probabilistic).

Table Group	Name	Repository	Туре
AssessmentGroup- Memberships	ActiveSubstances.xlsx	Standard Actions/Training Substance Prioritisa- tion Acute Neuro	Fixed
MaximumResidueLim- its	ConcentrationLimits.xlsx	Standard Actions/Training Substance Prioritisa- tion Acute Neuro	Fixed
Effects	Effects.xlsx	Standard Actions/Training Substance Prioritisa- tion Acute Neuro	Fixed
FoodExtrapolations	FoodExtrapolations.xlsx	Standard Actions/Training Substance Prioritisa- tion Acute Neuro	Fixed
FoodTranslations	FoodRecipes.xlsx	Standard Actions/Training Substance Prioritisa- tion Acute Neuro	Fixed
Foods	Foods.xlsx	Standard Actions/Training Substance Prioritisa- tion Acute Neuro	Fixed
HazardCharacterisa- tions	HazardCharacterisa- tions.xlsx	Standard Actions/Training Substance Prioritisa- tion Acute Neuro	Fixed
Processing	ProcessingFactors.xlsx	Standard Actions/Training Substance Prioritisa- tion Acute Neuro	Fixed
AuthorisedUses	SubstanceAuthorisa- tions.xlsx	Standard Actions/Training Substance Prioritisa- tion Acute Neuro	Fixed
ResidueDefinitions	SubstanceConver- sions.xlsx	Standard Actions/Training Substance Prioritisa- tion Acute Neuro	Fixed
Compounds	Substances.xlsx	Standard Actions/Training Substance Prioritisa- tion Acute Neuro	Fixed
UnitVariability	UnitVarPrimo.xlsx	Standard Actions/Training Substance Prioritisa- tion Acute Neuro	Fixed
Concentrations	a_ConcentrationsSSD_Neu	roStaddard Actions/Training Substance Prioritisa- tion Acute Neuro	Fixed
Survey	a_ConsumptionsNL2.mdb	Standard Actions/Training Substance Prioritisa- tion Acute Neuro	Fixed
Survey	a_ConsumptionsNL3_6.mc	bStandard Actions/Training Substance Prioritisa- tion Acute Neuro	Fixed

Table 3.9: Datasources for Training substance prioritisation acute neuro.

- Chronic cumulative exposure assessment PFAS
- Chronic cumulative exposure assessment PA
- Demo acute cumulative risk assessment
- EU acute cumulative exposure assessment (2018) Tier I and Tier II
- EU chronic cumulative exposure assessment (2018) Tier I and Tier II
- IPGF Acute Cumulative Risk Assessment No Background
- Risk steatosis from imazalil
- Training prospective risk assessment acute Tier II
- Training prospective risk assessment chronic Tier II
- Training substance prioritisation acute neuro

CHAPTER FOUR

EXAMPLES

Note: This section is under construction. Please contribute!

Training materials used in EuroMix training sessions:

- EuroMix dietary exposure
- RPF-exercise 1-for training-draft

There are a few exercises prepared that you could follow to get started.

4.1 Cumulative dietary exposure assessment

4.1.1 Introduction

The goal of this exercise is to perform a probabilistic cumulative dietary exposure assessment, illustrating all data needed. In Example 1 we will upload and use nine different files containing the data. In Example 2 we will upload and use a single data file for the same purpose. In the example the exposure will be characterised by upper tail percentiles, and the risk driving substances and foods can be examined. In Example 3 an uncertainty analysis is added.

4.1.2 Preparation

In the workspace browser (icon), create a new workspace *Examples*, using the + button in the bottom right corner.

4.1.3 Example 1

Calculate a cumulative chronic dietary exposure for Dutch young adults in 2003 regarding a group of eight triazole substances according to the basic optimistic model of the EFSA 2012 guidance document. Use liver steatosis as a focal effect and Cyproconazole as an index substance. The data files are already available in the data folder *Documentation-Examples / Exercise Dietary Exposure Assessment*.

Detailed steps are as follows.

- In the *Examples* workspace, create a new action using the + button in the bottom right corner.
 - Select action type Dietary exposures
 - Name it, e.g. Triazoles exposures
 - (Optional) You can also add tags (e.g. triazoles, NL, steatosis) as labels that can be used later to find similar actions
 - (Optional) You can add a description for further information
 - · Click Next

- Specify Dietary exposures settings
 - Tier: EFSA 2012 Optimistic
 - Risk type *Chronic*
 - Click Create

You are now directed to the main page of the new action. You can always return to this main page by clicking Action settings \clubsuit or the action type name (*Dietary exposures*) in the green bar.

The main page contains at least three blocks of information: Scope, Inputs and Settings. We will now first link all nine data files needed for this cumulative assessment. For most settings we will use default values in accordance with the chosen tier (*EFSA 2012 Optimistic*).

Scope of the assessment:

- Click Effects (path in the green bar changes Total Dietary exposures / Effects)
 - At *Effects data source*, click 🖍 and browse to the file *Effect Steatosis.xlsx*, then click *Select*
 - At Effect Settings for focal effect select Steatosis-liver and click D Save Changes
 - In the green navigation bar, click *Dietary exposures* to go up one level.
- Click Foods (path: Dietary exposures / Foods)
 - At *Foods data source*, click 🖍 and browse to the file *Foods.xlsx*, then click *Select*
 - In the green navigation bar, click *Dietary exposures* to go up one level
- Click Populations (optional) (path: Dietary exposures / Populations)
 - At Populations data source, click 🖍 and browse to the file Populations.xlsx, then click Select
 - This file contains two populations, only one is allowed. Click \checkmark under Populations selection, this opens a pop-up window. Deselect *NL_2006*, then click *Save*. The red warning signs \blacktriangle should now be gone. (Note: green warning signs \bigstar point at details and can usually be ignored)
 - In the green navigation bar, click *Dietary exposures* to go up one level.
- Click Substances (path: Dietary exposures / Substances)
 - At Substances data source, click 🖍 and browse to the file Substances Triazoles.xlsx, then click Select
 - At Substance settings for Index substance select Cyproconazole and click D Save Changes
 - In the green navigation bar, click Dietary exposures to go up one level

Next we choose the other input data:

- Click Consumptions by modelled foodd (path: Dietary exposures / Consumptions by modelled food)
 - Click Consumptions (path: Dietary exposures / Consumptions by modelled food / Consumptions)
 - At *Consumptions data source*, click 🖍 and browse to the file *FoodConsumptions.xlsx* and *Select*
 - At *Consumptions data selection*, with 🖍 open the food consumption surveys selection.
 - The file contains two surveys, but only one is allowed. Click ✓ under Consumptions data selection, this opens a pop-up window. Deselect *VCP-kids*, then click *Save* (the red warning ▲ should now be gone)
 - In the green navigation bar, click Consumptions by modelled food to go up one level
 - Click Food conversions (path: Dietary exposures / Consumptions by modelled food / Food conversions)
 - Click Foods as measured (path: Dietary exposures / Consumptions by modelled food / Food conversions / Foods as measured)
 - Click Concentrations (path: Dietary exposures / Consumptions by modelled food / Food conversions / Foods as measured / Concentrations)

- At *Concentrations data source*, click \checkmark and browse to the file *ConcentrationData.xlsx*, then click *Select*
- In the green navigation bar, click Food conversions to go up two levels
- Click Food recipes (path: Dietary exposures / Consumptions by modelled food / Food conversions / Food recipes)
 - At *Food recipes data source*, click \checkmark and browse to the file *FoodTranslations.xlsx*. then click *Select*
 - In the green navigation bar, click Dietary exposures to go up three levels
- Click Concentration models (path: Dietary exposures / Concentration models)
 - Click Relative potency factors (path: Dietary exposures / Concentration models / Relative potency factors)
 - At *Relative potency data source*, click 🖍 and browse to the file *RPFs.xlsx*, then click *Select*
 - In the green navigation bar, click Dietary exposures to go up two levels
- Click Processing factors (path: Dietary exposures / Processing factors)
 - At Processing factors data source, click / and browse to the file ProcessingFactors.xlsx, then click Select
 - In the green navigation bar, click *Dietary exposures* to go up one level
- Click Active substances (optional) (path: Dietary exposures / Active substances)
 - In this example we have a fixed list of relative potency factors for the eight substances, and don't need point of departure (POD) data to decide which substances are active with respect to the health effect and therefore belong to the cumulative assessment group. Deselect the setting "Derive memberships from POD presence", then click Save Changes
 - In the green navigation bar, click *Dietary exposures* to go up one level

Now run the model, either by clicking the \blacktriangleright run icon in the grey bar, or by clicking the \blacktriangleright run icon in the green bar (Note: \blacktriangleright in the green bar can also be used to run subactions on their own).

The \blacktriangleright icon is replaced by the text "Running". When the run has finished, the interface automatically changes to the Results screen. You can also click the Results icon \bigoplus to go there.

As an exercise, try find the following results:

- 1. The 99th percentile of cumulative exposure
- 2. The substance(s) with highest contribution to the total exposure
- 3. The food(s)-as-measured with the highest contribution to the upper tail of the exposure distribution

Answers:

- In the grey bar, browse to the results panel by clicking the ① icon and click on the latest output (path: *Results / Dietary exposures*)
 - In the *Dietary exposures* tab, browse in the tree (unfold by clicking > where necessary) to > *Dietary exposures* > *Distribution (OIM)* > *Percentiles*
 - In the table it states that the 99% exposure percentile is at an exposure of 0.02127 µg/kg bw/day.
 - In the *Dietary exposures* tab, browse in the tree (unfold by clicking > where necessary) to > *Dietary exposures* > *Details* > *Exposures by substance* > *Total distribution*
 - From the pie chart it is clear that Tebuconazole contributes the most to the total exposure distribution with 32.7%. In the table below the graph more details can be found.
 - In the *Dietary exposures* tab, browse in the tree (unfold by clicking > where necessary) to > *Dietary exposures* > *Details* > *Exposures by food and substance* > *Risk drivers upper tail*
 - From the pie chart it is clear that Flusilazole in grapefruit contributes the most (16.7%) to the upper tail exposure distribution

4.1.4 Example 2

We will create a new action to demonstrate uploading all the data at once. All data is now contained within one file, *MCRA-Documentation Example Dietary exposures.xlsx*.

Detailed steps are as follows.

- In the *Examples* workspace, create a new action (using +)
 - Select action type Dietary exposures
 - Name it, e.g. Triazoles exposures from one data file
 - Click Next
- Specify Dietary exposures settings
 - Tier: EFSA 2012 Optimistic
 - Risk type Chronic
 - Click Create
- Then go to the actions settings 🌣 of this action (path: *Dietary exposures*)
 - Click Effects (path: Dietary exposures / Effects)
 - At *Effects data source*, click ✓ and browse to the file *MCRA-Documentation Example Dietary exposures.xlsx.* Click *Toggle all*, then *Select.* This will load all available data tables for all subactions of *Dietary exposures.*

You still need to specify the focal effect (under *Effects*), index substance (under *Substances*), and food surveys (under *Consumptions by modelled food / Consumptions*). You also need to deselect the "Derive memberships from POD presence" setting under *Active substances*. Navigate to the subaction where these changes have to be made using the green bar.

You now have achieved the same as in Example 1, only with the upload of one single file. You can now run the model, and inspect the results, which should be the same as for Example 1.

4.1.5 Example 3

Repeat the run of the previous task, but in addition to the nominal run, perform an uncertainty analysis as well.

- Click on the \neq icon (in the grey bar) to open the uncertainty settings panel
 - At Uncertainty settings, check ✓ Perform uncertainty analysis
 - For Monte Carlo iterations per uncertainty run choose 100, and press D Save Changes
- Now run the model, by pressing the 🕨 run icon in the grey bar. Note that the run will take much more time.

Compare with the previous results, to find:

- 1. 95% uncertainty bounds for the 99% exposure percentile
- 2. 95% uncertainty bounds for the highest contribution from a substance to the total exposure distribution
- 3. 95% uncertainty bounds for the highest contribution from a food to the total exposure distribution

4.2 Aggregate exposure assessment

4.2.1 Introduction

The goal of this exercise is to assess aggregate exposure assessment.

4.2.2 Preparation

If you haven't done so, in the workspace browser (use the \square icon), create a new workspace named *Examples*, using the +.

The data files used in the example(s) in this section, are located in the data folder *Documentation-Examples / Exercise* Aggregate Exposure Assessment.

4.2.3 Example 1

- In the *Examples* workspace, create a new action (using +)
 - Then select ✓ *Show all action types*, select *Exposures*
 - Name it exposures
 - At Exposure settings choose:
 - As Risk type Chronic
 - Check ✓ Include dietary and non-dietary routes of exposure
 - Press Create
- Then go to the Actions settings 🌣 of this action (path: *Exposures*)
 - At Scope, click Effects (path: Exposures / Effects)
 - At *Effects data source* with 🖍 browse to the file *Effect Steatosis.xlsx* and *Select*
 - At Effect settings for Focal effect select Steatosis-liver and press Save Changes
 - In the green navigation bar, click *Exposures* to go up one level
 - At Scope, click Foods (path: Exposures / Foods)
 - At *Foods data source* with *I* browse to the file *Foods.xlsx* and *Select*
 - In the green navigation bar, click *Exposures* to go up one level
 - At Scope, click Substances (path: Exposures / Substances)
 - At Substances data source with 🖍 browse to the file Substances.xlsx and Select
 - At Substance settings
 - for *Index substance* select *Cyproconazole* and press **B** *Save Changes*
 - In the green navigation bar, click *Exposures* to go up one level
 - At Inputs, click Dietary exposures (path: Exposures / Dietary exposures)
 - At Inputs, click Consumptions by modelled food (path: Exposures / Dietary exposures / Consumptions by modelled food)
 - At Inputs, click Consumptions (path: Exposures / Dietary exposures / Consumptions by modelled food / Consumptions)
 - At *Consumptions data source* with \checkmark browse to the file *Consumptions.xlsx* and *Select*
 - At *Consumptions data selection* with \checkmark open the food consumption surveys selection.

- Now select *DNFCS_2003* and press *Save* (the red warning **A** should now be gone)
- In the green navigation bar, click Consumptions by modelled food to go up one level
- At Inputs, click Food conversions (path: Exposures / Dietary exposures / Consumptions by modelled food / Food conversions)
 - At Inputs, click Foods as measured (path: Exposures / Dietary exposures / Consumptions by modelled food / Food conversions / Foods as measured)
 - At Inputs, click Concentrations (path: Exposures / Dietary exposures / Consumptions by modelled food / Food conversions / Foods as measured Concentrations)
 - At Concentration data source with ✓ browse to the file ConcentrationData.xlsx and Select
 - In the green navigation bar, click Food conversions to go up two levels
 - At Inputs, click Food recipes (path: Exposures / Dietary exposures / Consumptions by modelled food / Food conversions / Food recipes)
 - At Food recipes data source, with 🖍 browse to the file FoodRecipes.xlsx and Select
 - In the green navigation bar, click Dietary exposures to go up three levels
- At Inputs, click Concentration models (path: Exposures / Dietary exposures / Concentration models)
 - At Inputs, click Relative potency factors (path: Exposures / Dietary exposures / Concentration models / Relative potency factors)
 - At *Relative potency factors data source* with \checkmark browse to the file *RelativePotencyFactors.xlsx* and *Select*
 - In the green navigation bar, click Dietary exposures to go up two levels
- At Inputs, click Processing factors (path: Exposures / Dietary exposures / Processing factors)
 - At Processing factors data source with 🖍 browse to the file ProcessingFactors.xlsx and Select
 - In the green navigation bar, click *Dietary exposures* to go up one level
- At Inputs, click Active substances (optional) (path: Exposures / Dietary exposures / Active substances)
 - At Inputs, click Points of departure (path: Exposures / Dietary exposures / Active substances / Points of departure)
 - At Points of departure data source with 🖍 browse to the file HazardDoses Triazoles.xlsx
 - In the green navigation bar, click Dietary exposures to go up two levels
- At Dietary exposure settings, for Dietary exposure calculation tier select EFSA 2012 Optimistic, and press Save Changes
- In the green navigation bar, click *Exposures* to go up one level
- At Inputs, click Non-dietary exposures (path: Exposures / Non-dietary exposures)
 - At Non-dietary exposures data source with ✔ browse to the file NonDietaryExposures.xlsx and Select
- Now run the model, by pressing the **b** run icon in the grey bar.

Try to find the following results:

- 1. Exposure percentiles daily intakes with uncertainty bounds
- 2. Substance with highest contribution to the total exposure distribution
- 3. The modelled food measured with the highest contribution to the upper tail of the exposure distribution

4.2.4 Example 2

In this example we will elaborate on the previous one with kinetic models.

- Go to the Actions settings 🍄 of this action (path: *Exposures*)
 - At Inputs, click Kinetic models (default) (path: Exposures / Kinetic models)
 - At Kinetic models data source with browse to the file UserGroupDemo-KineticModelsArtificial.xlsx and Select
 - At Kinetic model settings for Kinetic model select Cosmos Version 5
- Now run the model, by pressing the ▶ run icon in the green bar.

4.3 Hazard characterisations from PoDs

4.3.1 Introduction

The goal of this exercise is to try to establish hazard characterisations from PoDs (NOAELs).

4.3.2 Preparation

If you haven't done so, in the workspace browser (use the \square icon), create a new workspace named *Examples*, using the +.

The data files used in the example(s) in this section, are located in the data folder *Documentation-Examples / Exercise Hazard characterisations*.

4.3.3 Example 1

In this example, Imazalil target dose from NOAEL will be calculated.

- In the *Examples* workspace, create a new action (using +)
 - Then select ✓ Show all action types, and select Hazard characterisations
 - Name it TargetDoseImazalil
 - Use as Hazard characterization settings
 - Risk type: Chronic
 - Target level: External
 - Press Create
- Then go to the Actions settings \clubsuit of this action.
 - At Scope, click Effects (path: Hazard characterisations / Effects)
 - At Effects data source with ✓ browse to the file Effects and AOP Network Steatosis.xlsx and Select
 - At *Effects selection* with 🖍
 - Deselect everything by clicking \checkmark on the first line, next to the word *Code*
 - On the second page, select only *Steatosis-liver*, and **B** Save
 - At *Effect Settings* for *focal effect* select *Steatosis-liver* and press **B** Save Changes.
 - In the green navigation bar, click Hazard characterisations to go up one level
 - At Scope, click Substances (path: Hazard characterisations / Substances)

- At Substances data source with ✓ browse to the file TargetDosescalculation-Substances.xlsx and Select
- At Substances selection with ▲
 - Deselect everything, by clicking the \checkmark on the first line, next to the word *Code*
 - Select only *Imazalil*, and **B** Save
 - In the green navigation bar, click Hazard characterisations to go up one level
- At Inputs, click Points of departure (path: Hazard characterisations / Points of departure)
 - At *Points of departure data source* with \checkmark browse to the file *TargetDosesCalculation-HazardDoses.xlsx* and *Select*
 - In the green navigation bar, click Hazard characterisations to go up one level
- At Hazard characterisations settings, for Expression type select NOAEL (convert all hazard characterisations as NOAELs)
- At Hazard characterisations settings, Select ✓ Use inter-species conversions
- At *Hazard characterisations settings*, Select \checkmark Use intra-species factors, and press 🖻 Save Changes
- Now run the model, by pressing the **b** run icon in the grey bar.

Try to find the following results:

- 1. The NOAEL for Imazalil used as point of departure.
- 2. The target hazard dose based on the default assessment factors 1/10 and 1/10 for inter-species and within-species conversion.

Answers:

- In the grey bar, browse to the results panel by clicking the \bigcirc icon and click on the latest output (path: *Results / TargetDoseImazalil*)
 - In the *Hazard characterisations* tab, browse in the tree (unfold by clicking > where necessary) to > *Available hazard characterisations*
 - The NOAEL for Imazalil is 40 µg/kg bw/day.

4.4 Health impact estimates

4.4.1 Introduction

The goal of this exercise is to assess a health impact estimate.

4.4.2 Preparation

If you haven't done so, in the workspace browser (use the \Box icon), create a new workspace named *Examples*, using the +.

The data files used in the example(s) in this section, are located in the data folder *Documentation-Examples / Exercise Health Impact*.

4.4.3 Example 1

- In the *Examples* workspace, create a new action (using +)
 - Then select ✓ Show all action types, select Risks
 - Name it Risks
 - Press Create
- Then go to the Actions settings 🌣 of this action (path: *Risks*)
 - At Scope, click Effects (path: Risks / Effects)
 - At Effects data source with 🖍 browse to the file Effects and AOP Network Steatosis.xlsx and Select
 - At * Effect settings*, for *focal effect* select *Steatosis-liver* and press **B** *Save Changes*
 - In the green navigation bar, click Risks to go up one level
 - At Scope, click Foods (path: Risks / Foods)
 - At *Foods data source* with 🖍 browse to the file *Foods.xlsx* and *Select*
 - In the green navigation bar, click Risks to go up one level
 - At Scope, click Substances (path: Risks / Substances)
 - At Substances data source with ✓ browse to the file Substances.xlsx and Select
 - At Substance settings, for index substance select Cyproconazole, and press D Save Changes
 - In the green navigation bar, click Risks to go up one level
 - At Inputs, click Exposures (path: Risks / Exposures)
 - At Inputs, click Dietary exposures (path: Risks /Exposures / Dietary exposures)
 - At Inputs, click Consumptions by food measured (path: Risks /Exposures / Dietary exposures / Consumptions by modelled food)
 - At Inputs, click Consumptions (path: Risks /Exposures / Dietary exposures / Consumptions by modelled food / Consumptions)
 - At Consumptions data source with 🖍 browse to the file Consumptions.xlsx and Select
 - At *Consumptions data selection* with \checkmark open the food consumption surveys selection.
 - The file contains two surveys, but only one is allowed. So deselect everything by clicking ✓ on the first line, next to the word *Code*
 - Now select *DNFCS_2003* and press *Save* (the red warning **A** should now be gone)
 - At Consumptions settings for Food survey select DNFCS_2003 and press Save Changes
 - In the green navigation bar, click Consumptions by modelled food to go up one level
 - At Inputs, click Food conversions (path: Risks /Exposures / Dietary exposures / Consumptions by modelled food / Food conversions)
 - At Inputs click Food as measured (path: Risks /Exposures / Dietary exposures / Consumptions by modelled food / Food conversions / Food as measured)
 - At Inputs, click Concentrations (path: Risks /Exposures / Dietary exposures / Consumptions by modelled food / Food conversions / Food as measured / Concentrations)
 - At *Concentrations data source* with ✓ browse to the file *UserGroupDemo-ConcentrationData.xlsx* and *Select*
 - In the green navigation bar, click Food conversions to go up two levels

- At Inputs, click Food recipes (path: Risks /Exposures / Dietary exposures / Consumptions by modelled food / Food conversions / Food recipes)
 - At *Food recipes data source* with \checkmark browse to the file *UserGroupDemo-FoodRecipes.xlsx* and *Select*
 - In the green navigation bar, click *Dietary exposures* to go up three levels
- At Inputs, click Processing factors (path: Risks /Exposures / Dietary exposures / Processing factors)
 - At *Processing factors data source* with \checkmark browse to the file *UserGroupDemo-ProcessingFactors.xlsx* and *Select*
 - In the green navigation bar, click *Risks* to go up three levels
- At Inputs, click Hazard characterisations (path: Risks / Hazard characterisations)
 - At Inputs, click Active substances (path: Risks / Hazard characterisations / Active substances)
 - At Inputs, click Points of departure (path: Risks / Hazard characterisations / Active substances / Points of departure)
 - At Points of departure data source with \checkmark browse to the file UserGroupDemo-HazardDoses.xlsx and Select
 - In the green navigation bar, click Active substances to go up one level
 - At 'Active substances' click ^{*} Compute

4.5 Assessment group membership probabilities

4.5.1 Introduction

The goal of this exercise is to assess group membership probabilities.

4.5.2 Preparation

If you haven't done so, in the workspace browser (use the \square icon), create a new workspace named *Examples*, using the +.

The data files used in the example(s) in this section, are located in the data folder *Documentation-Examples / Exercise Dietary Exposure Assessment*.

4.5.3 Example 1

- In the *Examples* workspace, create a new action (using +)
 - Then select Dietary exposures
 - Name it Dietary exposures
 - Use as Dietary exposures settings
 - Tier: EFSA Guidance Optimistic
 - Risk type Chronic
 - Select \checkmark *Cumulative*
 - Press Create
- Then go to the actions settings 🌣 of this action (path: *Dietary exposures*)

- At Scope, click Foods (path: Dietary exposures / Foods)
 - At Foods data source with 🖍 browse to the file UserGroupDemo-Foods.xlsx and Select
 - In the green navigation bar, click Dietary exposures to go up one level
- At Scope, click Substances (path: Dietary exposures / Substances)
 - At Substances data source with ✓ browse to the file UserGroupDemo-Substances.xlsx and Select
 - At Substance settings for Index substance select Cyproconazole and press B Save Changes
 - In the green navigation bar, click Dietary exposures to go up one level
- At Scope, click Effects (path: Dietary exposures / Effects)
 - At *Effects data source* with 🖍 browse to the file *Effect Steatosis.xlsx* and *Select*
 - At Effect Settings for focal effect select Steatosis-liver and press D Save Changes
 - In the green navigation bar, click *Dietary exposures* to go up one level.
- At Inputs, click Consumptions by modelled food (path: Dietary exposures / Consumptions by modelled food)
 - At Inputs, click Consumptions (path: Dietary exposures / Consumptions by modelled food / Consumptions)
 - At *Consumptions data source* with \checkmark browse to the file *UserGroupDemo-Consumptions.xlsx* and *Select*
 - At Consumption settings for Food survey select DNFCS_2003 and press Save Changes
 - In the green navigation bar, click *Consumptions by modelled food* to go up one level
 - At Inputs, click Food conversions (path: Dietary exposures / Consumptions by modelled food / Food conversions)
 - At Inputs, click Foods as measured (path: Dietary exposures / Consumptions by modelled food / Food conversions / Foods as measured)
 - At Inputs, click Concentrations (path: Dietary exposures / Consumptions by modelled food / Food conversions / Foods as measured / Concentrations)
 - At *Concentrations data source* with *I* browse to the file *UserGroupDemo-ConcentrationData.xlsx* and *Select*
 - In the green navigation bar, click Food conversions to go up two levels
 - At Inputs, click Food recipes (path: Dietary exposures / Consumptions by modelled food / Food Food recipes)
 - At Food recipes data source, with 🖍 browse to the file UserGroupDemo-FoodRecipes.xlsx
 - In the green navigation bar, click *Dietary exposures* to go up three levels
- At Inputs, click Concentration models (path: Dietary exposures / Concentration models)
 - At Inputs, click Relative potency factors (path: Dietary exposures / Concentration models / Relative potency factors)
 - At *Relative potency data source* with *I* browse to the file *UserGroupDemo-RelativePotencyFactors.xlsx* and *Select*
 - In the green navigation bar, click *Dietary exposures* to go up two levels
- At Inputs, click Processing factors (path: Dietary exposures / Processing factors)
 - At *Processing factors data source* with ✓ browse to the file *UserGroupDemo-ProcessingFactors.xlsx* and *Select*
 - In the green navigation bar, click Dietary exposures to go up one level
- At Inputs, click Active substances (optional) (path: Dietary exposures / Active substances)

- At Inputs, click Points of departure (path: Dietary exposures / Active substances / Points of departure)
 - At Points of departure data source, with 🖍 browse to the file HazardDoses Triazoles.xlsx
 - In the green navigation bar, click *Dietary exposures* to go up two levels
- Now run the model, by pressing the **>** run icon in the grey bar.

Try to find the following results:

- 1. Exposure percentiles daily intakes
- 2. Substance with highest contribution to the total exposure distribution
- 3. The food-as-measured with the highest contribution to the upper tail of the exposure distribution

4.5.4 Example 2

Repeat the run of the previous task, but instead of the nominal run, now do an uncertainty analysis loop.

- Click on the [≇] icon (in the grey bar) to open the uncertainty settings panel, and check ✓ *Perform uncertainty analysis*
 - For Monte Carlo iterations per uncertainty run choose 100, and press Save Changes
- Now run the model, by pressing the **>** run icon in the grey bar.

Compare with the previous results, to find:

- 1. Exposure percentiles daily intakes with uncertainty bounds
- 2. Substance with highest contribution to the total exposure distribution
- 3. The food-as-measured with the highest contribution to the upper tail of the exposure distribution

PUBLICATIONS USING MCRA

- European Food Safety Authority (EFSA), P.S. Craig, B. Dujardin, A. Hart, A.F. Hernandez-Jerez, S. Hougaard Bennekou, C. Kneuer, B. Ossendorp, R. Pedersen, G. Wolterink, and L. Mohimont. Cumulative dietary risk characterisation of pesticides that have chronic effects on the thyroid. *EFSA Journal*, 18(4):e06088, 2020. URL: https://efsa.onlinelibrary.wiley.com/doi/abs/10.2903/j.efsa.2020.6088.
- European Food Safety Authority (EFSA), P.S. Craig, B. Dujardin, A. Hart, A.F. Hernández-Jerez, S. Hougaard Bennekou, C. Kneuer, B. Ossendorp, R. Pedersen, G. Wolterink, and L. Mohimont. Cumulative dietary risk characterisation of pesticides that have acute effects on the nervous system. *EFSA Journal*, 18(4):e06087, 2020. URL: https://efsa.onlinelibrary.wiley.com/doi/abs/10.2903/j.efsa.2020.6087.
- A. Beronius, J. Zilliacus, A. Hanberg, M. Luijten, van der Voet, H, and J. van Klaveren. Methodology for health risk assessment of combined exposures to multiple chemicals. *Food and Chemical Toxicology*, pages 111520, July 2020. URL: https://doi.org/10.1016/j.fct.2020.111520.
- J. Cotterill, N. Price, E. Rorije, and A. Peijnenburg. Development of a QSAR model to predict hepatic steatosis using freely available machine learning tools. *Food and Chemical Toxicology*, 142:111494, August 2020. URL: https://doi.org/10.1016/j.fct.2020.111494.
- B.C. Fischer, S. Rotter, J. Schubert, P. Marx-Stoelting, and R. Solecki. Recommendations for international harmonisation, implementation and further development of suitable scientific approaches regarding the assessment of mixture effects. *Food and Chemical Toxicology*, 141:111388, July 2020. URL: https: //doi.org/10.1016/j.fct.2020.111388.
- C. Karrer, M. Andreassen, N. von Goetz, F. Sonnet, A.K. Sakhi, K. Hungerbühler, H. Dirven, and T. Husøy. The EuroMix human biomonitoring study: source-to-dose modeling of cumulative and aggregate exposure for the bisphenols BPA, BPS, and BPF and comparison with measured urinary levels. *Environment International*, 136:105397, March 2020. URL: https://doi.org/10.1016/j.envint.2019.105397.
- M.C. Kennedy, A.D.M. Hart, J.W. Kruisselbrink, M. van Lenthe, W.J. de Boer, H. van der Voet, E. Rorije, C. Sprong, and J. van Klaveren. A retain and refine approach to cumulative risk assessment. *Food and Chemical Toxicology*, 138:111223, April 2020. URL: https://doi.org/10.1016/j.fct.2020.111223.
- C. Sprong, A. Crépet, F. Metruccio, U. Blaznik, C. Anagnostopoulos, D.L. Christodoulou, B.H. Jensen, M. Kennedy, N. González, I. Rehurkova, J. Ruprich, J.D. te Biesebeek, M. Vanacker, A. Moretto, and J. van Klaveren. Cumulative dietary risk assessment overarching different regulatory silos using a margin of exposure approach: a case study with three chemical silos. *Food and Chemical Toxicology*, 142:111416, August 2020. URL: https://doi.org/10.1016/j.fct.2020.111416.
- C. Tebby, H. van der Voet, G. de Sousa, E. Rorije, V. Kumar, W. de Boer, J.W. Kruisselbrink, F.Y. Bois, M. Faniband, A. Moretto, and C. Brochot. A generic PBTK model implemented in the MCRA platform: predictive performance and uses in risk assessment of chemicals. *Food and Chemical Toxicology*, 142:111440, August 2020. URL: https://doi.org/10.1016/j.fct.2020.111440.
- A.D. van den Brand, M. Beukers, M. Niekerk, G. van Donkersgoed, M. van der Aa, B. van de Ven, A. Bulder, H. van der Voet, and C.R. Sprong. Assessment of the combined nitrate and nitrite exposure from food and drinking water: application of uncertainty around the nitrate to nitrite conversion factor. *Food Additives & Contaminants: Part A*, 37(4):568–582, January 2020. URL: https://doi.org/10.1080/19440049.2019.1707294.

- H. van der Voet, J.W. Kruisselbrink, W.J. de Boer, M.S. van Lenthe, J.J.B. van den Heuvel, A. Crépet, M.C. Kennedy, J. Zilliacus, A. Beronius, C. Tebby, C. Brochot, C. Luckert, A. Lampen, E. Rorije, C. Sprong, and J.D. van Klaveren. The MCRA toolbox of models and data to support chemical mixture risk assessment. *Food and Chemical Toxicology*, 138:111185, April 2020. URL: https://doi.org/10.1016/j.fct.2020.111185.
- M. Vanacker, P. Quindroit, K. Angeli, C. Mandin, P. Glorennec, C. Brochot, and A. Crépet. Aggregate and cumulative chronic risk assessment for pyrethroids in the French adult population. *Food and Chemical Toxicology*, 143:111519, September 2020. URL: https://doi.org/10.1016/j.fct.2020.111519.
- C. Vlachou, D. Hofstädter, E. Rauscher-Gabernig, A. Griesbacher, K. Fuchs, and J. König. Risk assessment of nitrites for the Austrian adult population with probabilistic modelling of the dietary exposure. *Food and Chemical Toxicology*, 143:111480, September 2020. URL: https://doi.org/10.1016/j.fct.2020.111480.

- F.Y. Bois, C. Tebby, and C. Brochot. EuroMix PBPK model for combined exposures. 2019. URL: https://zenodo.org/record/2532334.
- A. Boobis. Report of EuroMix workshops on international harmonisation on the risk assessment of combined exposure to multiple chemicals. 2019. URL: https://zenodo.org/record/3479150.
- P.E. Boon, M. Van Der Aa, A. Dusseldorp, P. Janssen, M.J. Zeilmaker, and S. Schulpen. Loodinname via kraanwater: blootstellingsschatting en risicobeoordeling voor diverse risicogroepen. RIVM Letter report 2019-0090, 2019. URL: https://rivm.openrepository.com/handle/10029/623516.
- P.E. Boon, G. Van Donkersgoed, W. Van Der Vossen, M. Sam, M.Y. Noordam, and H. Van Der Schee. Tussenevaluatie van de nota 'gezonde groei, duurzame oogst'. RIVM Letter report 2018-0127, 2019. URL: https://rivm.openrepository.com/handle/10029/623125.
- P.E. Boon, M.J. Zeilmaker, and M.J.B. Mengelers. Risicobeoordeling van GenX en PFOA in moestuingewassen in helmond. RIVM Letter report 2019-0024, 2019. URL: https://rivm.openrepository.com/handle/ 10029/622988.
- A. Crépet, M. Vanacker, C. Sprong, W. de Boer, U. Blaznik, M. Kennedy, C. Anagnostopoulos, D.L. Christodoulou, J. Ruprich, I. Rehurkova, J.L. Domingo, B.H. Jensen, F. Metruccio, A. Moretto, L. Jacxsens, P. Spanoghe, D. Senaeve, H. van der Voet, and J. van Klaveren. Selecting mixtures on the basis of dietary exposure and hazard data: application to pesticide exposure in the European population in relation to steatosis. *International Journal of Hygiene and Environmental Health*, 222(2):291–306, March 2019. URL: https://doi.org/10.1016/j.ijheh.2018.12.002.
- J. de Rop, D. Senaeve, L. Jacxsens, M. Houbraken, J. van Klaveren, and P. Spanoghe. Cumulative probabilistic risk assessment of triazole pesticides in Belgium from 2011-2014. *Food Additives & Contaminants: Part A*, 36(6):911–921, April 2019. URL: https://doi.org/10.1080/19440049.2019.1606943.
- B. Fischer, J. Schubert, S. Rotter, and R. Solecki. Specific recommendations regarding implementation of mechanism-based test strategy for harmonised cumulative risk assessment according oecd, who, efsa and EuroMix guidance. 2019. URL: https://zenodo.org/record/3490547.
- G. Heinemeyer, M. Jantunen, and P. Hakkinen. *The Practice of Consumer Exposure Assessment*. Springer International Publishing, 2019. URL: https://doi.org/10.1007/978-3-319-96148-4.
- C. Karrer, W. de Boer, C. Delmaar, Y. Cai, A. Crépet, K. Hungerbühler, and N. von Goetz. Linking probabilistic exposure and pharmacokinetic modeling to assess the cumulative risk from the bisphenols BPA, BPS, BPF, and BPAF for Europeans. *Environmental Science & Technology*, 53(15):9181–9191, July 2019. URL: https://doi.org/10.1021/acs.est.9b01749.
- M. Kennedy, A. Hart, J.W. Kruisselbrink, M. van Lenthe, W. de Boer, H. van der Voet, E. Rorije, C. Sprong, and J. van Klaveren. Methodology and results of the retain and refine approach. 2019. URL: https://zenodo. org/record/3465690.
- M.C. Kennedy, D.G. Garthwaite, W.J. de Boer, and J.W. Kruisselbrink. Modelling aggregate exposure to pesticides from dietary and crop spray sources in UK residents. *Environmental Science and Pollution Research*, 26(10):9892–9907, February 2019. URL: https://doi.org/10.1007/s11356-019-04440-7.

- A.E. Kolbaum, K. Berg, F. Müller, O. Kappenstein, and O. Lindtner. Dietary exposure to elements from the German pilot total diet study (TDS). *Food Additives & Contaminants: Part A*, 36(12):1822–1836, October 2019. URL: https://doi.org/10.1080/19440049.2019.1668967.
- B. Sachse, A.E. Kolbaum, R. Ziegenhagen, S. Andres, K. Berg, B. Dusemund, K.I. Hirsch-Ernst, O. Kappenstein, F. Müller, C. Röhl, O. Lindtner, A. Lampen, and B. Schäfer. Dietary manganese exposure in the adult population in Germany—what does it mean in relation to health risks? *Molecular Nutrition & Food Research*, 63(16):1900065, July 2019. URL: https://doi.org/10.1002/mnfr.201900065.
- T. Tietz, A. Lenzner, A.E. Kolbaum, S. Zellmer, C. Riebeling, R. Gürtler, C. Jung, O. Kappenstein, J. Tentschert, M. Giulbudagian, S. Merkel, R. Pirow, O. Lindtner, T. Tralau, B. Schäfer, P. Laux, M. Greiner, A. Lampen, A. Luch, R. Wittkowski, and A. Hensel. Aggregated aluminium exposure: risk assessment for the general population. *Archives of Toxicology*, 93(12):3503–3521, October 2019. URL: https://doi.org/10. 1007/s00204-019-02599-z.
- H. van der Voet, J.W. Kruisselbrink, W.J. de Boer, M.S. van Lenthe, J.J.B. van den Heuvel, A. Crépet, M.C. Kennedy, J. Zilliacus, A. Beronius, E. Rorije, C. Sprong, and J.D. van Klaveren. The EuroMix model toolbox MCRA 9. 2019. URL: https://zenodo.org/record/3462181.
- H. van der Voet, J.W. Kruisselbrink, W.J. de Boer, M.S. van Lenthe, J.J.B. van den Heuvel, A. Crépet, M.C. Kennedy, J. Zilliacus, A. Beronius, C. Tebby, C. Brochot, E. Rorije, C. Sprong, and J.D. van Klaveren. Draft paper on the EuroMix toolbox of models and data to support chemical mixture risk assessment. 2019. URL: https://zenodo.org/record/3474943.
- J.D. van Klaveren, J.W. Kruisselbrink, W.J. de Boer, G. van Donkersgoed, J.D. te Biesebeek, M. Sam, and H. van der Voet. Cumulative dietary exposure assessment of pesticides that have acute effects on the nervous system using MCRA software. *EFSA Supporting Publications*, 16(9):1708E, 2019. URL: https: //efsa.onlinelibrary.wiley.com/doi/abs/10.2903/sp.efsa.2019.EN-1708.
- J.D. van Klaveren, J.W. Kruisselbrink, W.J. de Boer, G. van Donkersgoed, J.D. te Biesebeek, M. Sam, and H. van der Voet. Cumulative dietary exposure assessment of pesticides that have chronic effects on the thyroid using MCRA software. *EFSA Supporting Publications*, 16(9):1707E, 2019. URL: https://efsa.onlinelibrary. wiley.com/doi/abs/10.2903/sp.efsa.2019.EN-1707.
- M.S. van Lenthe, W.J. de Boer, J.W. Kruisselbrink, H. van der Voet, A. Crépet, M. Vanacker, and L. Trocellier. Validation of the EuroMix model toolbox and comparison with us software. 2019. URL: https://zenodo.org/ record/3467409.
- J. Zilliacus, A. Beronius, A. Hanberg, M. Luijten, J. van Klaveren, and H. van der Voet. EuroMix handbook for mixture risk assessment. 2019. URL: https://zenodo.org/record/3560719.
- J. Zilliacus, E. Rorije, M. Kennedy, and J. van Klaveren. Proceedings and training material from second training session for stakeholders. 2019. URL: https://zenodo.org/record/3560731.

- P.E. Boon, J.D. Te Biesebeek, H. Brants, M.C. Bouwmeester, and E.V.S. Hessel. Dietary sources of exposure to bisphenol A in the Netherlands. RIVM Letter report 2017-0187, 2018. URL: http://rivm.openrepository. com/rivm/handle/10029/621792.
- P.E. Boon, G. Van Donkersgoed, J.D. Te Biesebeek, G. Wolterink, and A.G. Rietveld. Cumulative exposure to residues of plant protection products via food in the Netherlands. RIVM Letter report 2017-0018, 2018. URL: http://rivm.openrepository.com/rivm/handle/10029/622169.
- Jardim, A.N.O, D.C. Mello, A.P. Brito, H. van der Voet, P.E. Boon, and E.D. Caldas. Probabilistic dietary risk assessment of triazole and dithiocarbamate fungicides for the Brazilian population. *Food and Chemical Toxicology*, 118:317–327, August 2018. URL: https://doi.org/10.1016/j.fct.2018.05.002.
- Jardim, A.N.O, D.C. Mello, A.P. Brito, G. van Donkersgoed, P.E. Boon, and E.D. Caldas. Dietary cumulative acute risk assessment of organophosphorus, carbamates and pyrethroids insecticides for the Brazilian population. *Food and Chemical Toxicology*, 112:108–117, February 2018. URL: https://doi.org/10.1016/j.fct.2017. 12.010.

- M. Mengelers, J.D. Te Biesebeek, M. Schipper, W. Slob, and P.E. Boon. Risicobeoordeling van GenX en PFOA in moestuingewassen in Dordrecht, Papendrecht en Sliedrecht. RIVM Letter report 2017-0017, 2018. URL: http://rivm.openrepository.com/rivm/handle/10029/621785.
- S. Rotter, A. Beronius, A.R. Boobis, A. Hanberg, J. van Klaveren, M. Luijten, K. Machera, D. Nikolopoulou, H. van der Voet, J. Zilliacus, and R. Solecki. Overview on legislation and scientific approaches for risk assessment of combined exposure to multiple chemicals: the potential EuroMix contribution. *Critical Reviews in Toxicology*, 48(9):796–814, October 2018. URL: https://doi.org/10.1080/10408444.2018.1541964.
- J. Suomi, P. Tuominen, S. Niinistö, S.M. Virtanen, and K. Savela. Dietary heavy metal exposure of Finnish children of 3 to 6 years. *Food Additives & Contaminants: Part A*, 35(7):1305–1315, June 2018. URL: https://doi.org/10.1080/19440049.2018.1480065.
- B.M. van De Ven, S. Fragki, J.D. te Biesebeek, A.G. Rietveld, and P.E. Boon. Mineral oils in food; a review of toxicological data and an assessment of the dietary exposure in the Netherlands. RIVM Letter report 2017-0018, 2018. URL: http://rivm.openrepository.com/rivm/handle/10029/622044.

- P.E. Boon, J.D. te Biesebeek, and G. van Donkersgoed. Dietary exposure to lead in the Netherlands. RIVM Letter report 2016-0206, 2017. URL: https://www.rivm.nl/bibliotheek/rapporten/2016-0206.pdf.
- K. Presser, C. Zoom, J. Szymanek, and G. Zappa. Development of a pilot service for the electronic infrastructure of METROFOOD-RI. In *Proceedings of 3rd IMEKOFOODS Conference: Metrology Promoting Harmonization and Standardization in Food and Nutrition*. International Measurement Confederation, 2017. URL: https://imeko.org/publications/tc23-2017/IMEKO-TC23-2017-045.pdf.
- C. Sieke, B. Michalski, and T. Kuhl. Probabilistic dietary risk assessment of pesticide residues in foods for the German population based on food monitoring data from 2009 to 2014. *Journal of Exposure Science & Environmental Epidemiology*, 28(1):46–54, July 2017. URL: https://doi.org/10.1038/jes.2017.7.
- R.C. Sprong, E.M. Niekerk, and M.H. Beukers. Intake assessment of the food additives nitrites (e 249 and e 250) and nitrates (e 251 and e 252). RIVM Letter report 2016-0208, 2017. URL: https://www.rivm.nl/ bibliotheek/rapporten/2016-0208.pdf.

- P.E. Boon and J.D. te Biesebeek. Preliminary assessment of dietary exposure to 3-MCPD in the Netherlands. RIVM Letter report 2015-0199, 2016. URL: https://www.rivm.nl/bibliotheek/rapporten/2015-0199.pdf.
- P.E. Boon, J.D. te Biesebeek, S.P.J. van Leeuwen, M.J. Zeilmaker, and L.A.P. Hoogenboom. Dietary exposure to polybrominated diphenyl ethers in the Netherlands. RIVM Letter report 2016-0037, 2016. URL: https://www.rivm.nl/bibliotheek/rapporten/2016-0037.pdf.
- C Rompelberg, M.B. Heringa, G. van Donkersgoed, J. Drijvers, A. Roos, S. Westenbrink, R. Peters, G. van Bemmel, W. Brand, and A.G. Oomen. Oral intake of added titanium dioxide and its nanofraction from food products, food supplements and toothpaste by the Dutch population. *Nanotoxicology*, 10(10):1404–1414, September 2016. URL: https://doi.org/10.1080/17435390.2016.1222457.
- R.C. Sprong, L. de Wit-Bos, J.D. te Biesebeek, M. Alewijn, P. Lopez, and M.J.B. Mengelers. A mycotoxindedicated total diet study in the Netherlands in 2013: part III – exposure and risk assessment. *World Mycotoxin Journal*, 9(1):109–128, February 2016. URL: https://doi.org/10.3920/wmj2015.1905.
- C.L. Stephenson and C.A. Harris. An assessment of dietary exposure to glyphosate using refined deterministic and probabilistic methods. *Food and Chemical Toxicology*, 95:28–41, September 2016. URL: https://doi.org/ 10.1016/j.fct.2016.06.026.
- H. van der Voet, W.J. de Boer, J.W. Kruisselbrink, G. van Donkersgoed, and J.D. van Klaveren. MCRA made scalable for large cumulative assessment groups. *EFSA Supporting Publications*, 13(1):910E, 2016. URL: https://efsa.onlinelibrary.wiley.com/doi/abs/10.2903/sp.efsa.2016.EN-910.

- Y. Akhandaf, J. van Klaveren, S. de Henauw, G. van Donkersgoed, T. van Gorcum, A. Papadopoulos, V. Sirot, M. Kennedy, H. Pinchen, J. Ruprich, I. Rehurkova, G. Perelló, and I. Sioen. Exposure assessment within a total diet study: a comparison of the use of the pan-European classification system FoodEx-1 with national food classification systems. *Food and Chemical Toxicology*, 78:221–229, April 2015. URL: https://doi.org/ 10.1016/j.fct.2015.01.019.
- U. Blaznik, A. Yngve, I. Eržen, and C.H. Ribič. Consumption of fruits and vegetables and probabilistic assessment of the cumulative acute exposure to organophosphorus and carbamate pesticides of schoolchildren in Slovenia. *Public Health Nutrition*, 19(3):557–563, May 2015. URL: https://doi.org/10.1017/ s1368980015001494.
- P.E. Boon and H. Van der Voet. Probabilistic dietary exposure models relevant for acute and chronic exposure assessment of adverse chemicals via food. RIVM Letter report 2015-0191, 2015. URL: https://www.rivm.nl/ bibliotheek/rapporten/2015-0191.pdf.
- P.E. Boon, G. van Donkersgoed, D. Christodoulou, A. Crépet, L. D'Addezio, V. Desvignes, B. Ericsson, F. Galimberti, E. Ioannou-Kakouri, B.H. Jensen, I. Rehurkova, J. Rety, J. Ruprich, S. Sand, C. Stephenson, A. Strömberg, A. Turrini, H. van der Voet, P. Ziegler, P. Hamey, and J.D. van Klaveren. Cumulative dietary exposure to a selected group of pesticides of the triazole group in different European countries according to the EFSA guidance on probabilistic modelling. *Food and Chemical Toxicology*, 79:13–31, May 2015. URL: https://doi.org/10.1016/j.fct.2014.08.004.
- D. He, X. Ye, Y. Xiao, N. Zhao, J. Long, P. Zhang, Y. Fan, S. Ding, X. Jin, C. Tian, S. Xu, and C. Ying. Dietary exposure to endocrine disrupting chemicals in metropolitan population from China: a risk assessment based on probabilistic approach. *Chemosphere*, 139:2–8, November 2015. URL: https://doi.org/10.1016/j. chemosphere.2015.05.036.
- R. Jacobs, H. van der Voet, and C.J.F. ter Braak. Integrated probabilistic risk assessment for nanoparticles: the case of nanosilica in food. *Journal of Nanoparticle Research*, June 2015. URL: https://doi.org/10.1007/s11051-015-2911-y.
- M.C. Kennedy, C.R. Glass, B. Bokkers, A.D.M. Hart, P.Y. Hamey, J.W. Kruisselbrink, W.J. de Boer, H. van der Voet, D.G. Garthwaite, and J.D. van Klaveren. A European model and case studies for aggregate exposure assessment of pesticides. *Food and Chemical Toxicology*, 79:32–44, May 2015. URL: https://doi.org/10.1016/j.fct.2014.09.009.
- M.C. Kennedy, C.R. Glass, S. Fustinoni, A. Moretto, S. Mandic-Rajcevic, P. Riso, A. Turrini, H. van der Voet, M.T. Hetmanski, R.J. Fussell, and J.D. van Klaveren. Testing a cumulative and aggregate exposure model using biomonitoring studies and dietary records for Italian vineyard spray operators. *Food and Chemical Toxicology*, 79:45–53, May 2015. URL: https://doi.org/10.1016/j.fct.2014.12.012.
- M.C. Kennedy, H. van der Voet, V.J. Roelofs, W. Roelofs, C.R. Glass, W.J. de Boer, J.W. Kruisselbrink, and A.D.M. Hart. New approaches to uncertainty analysis for use in aggregate and cumulative risk assessment of pesticides. *Food and Chemical Toxicology*, 79:54–64, May 2015. URL: https://doi.org/10.1016/j.fct.2015. 02.008.
- F.R. Mancini, V. Sirot, L. Busani, J.L. Volatier, and M. Hulin. Use and impact of usual intake models on dietary exposure estimate and risk assessment of chemical substances: a practical example for cadmium, acrylamide and sulphites. *Food Additives & Contaminants: Part A*, 32(7):1065–1074, May 2015. URL: https://doi.org/10.1080/19440049.2015.1041428.
- R.C. Sprong and P.E. Boon. Dietary exposure to cadmium in the Netherlands. RIVM Letter report 2015-0085, 2015. URL: https://www.rivm.nl/bibliotheek/rapporten/2015-0085.pdf.
- J. Suomi, J. Ranta, P. Tuominen, T. Putkonen, C. Bäckman, M.L. Ovaskainen, S.M. Virtanen, and K. Savela. Quantitative risk assessment on the dietary exposure of Finnish children and adults to nitrite. *Food Additives & Contaminants: Part A*, 33(1):41–53, November 2015. URL: https://doi.org/10.1080/19440049.2015. 1117145.
- H. van der Voet, W.J. de Boer, J.W. Kruisselbrink, P.W. Goedhart, G.W.A.M. van der Heijden, M.C. Kennedy, P.E. Boon, and J.D. van Klaveren. The MCRA model for probabilistic single-compound and

cumulative risk assessment of pesticides. *Food and Chemical Toxicology*, 79:5–12, May 2015. URL: https://doi.org/10.1016/j.fct.2014.10.014.

• J.D. van Klaveren, M.C. Kennedy, A. Moretto, W. Verbeke, H. van der Voet, and P.E. Boon. The ACROP-OLIS project: its aims, achievements, and way forward. *Food and Chemical Toxicology*, 79:1–4, May 2015. URL: https://doi.org/10.1016/j.fct.2015.03.006.

2014

- P.E. Boon. Estimation of the acute dietary exposure to pesticides using the probabilistic approach and the point estimate methodology. *European Journal of Nutrition & Food Safety*, 4(1):1–3, January 2014. URL: https://doi.org/10.9734/ejnfs/2014/6899.
- P.E. Boon, J.D. te Biesebeek, L. de Wit, and G. van Donkersgoed. Dietary exposure to dioxins in the Netherlands. RIVM Letter report 2014-0001, 2014. URL: https://www.rivm.nl/bibliotheek/rapporten/2014-0001. pdf.
- P.E. Boon, H. van der Voet, J. Ruprich, A. Turrini, S. Sand, and J.D. van Klaveren. Computational tool for usual intake modelling workable at the European level. *Food and Chemical Toxicology*, 74:279–288, December 2014. URL: https://doi.org/10.1016/j.fct.2014.10.019.
- H. van der Voet, J.W. Kruisselbrink, W.J. Boer, and P.E. Boon. Model-then-add: usual intake modelling of multimodal intake distributions. RIVM Letter report 090133001/2014, 2014. URL: http://hdl.handle.net/ 10029/314361.

2013

- A.J.C. Roodenburg, A.J. van Ballegooijen, M. Dötsch-Klerk, H. van der Voet, and J.C. Seidell. Modelling of usual nutrient intakes: potential impact of the Choices programme on nutrient intakes in young Dutch adults. *PLoS ONE*, 8(8):e72378, August 2013. URL: https://doi.org/10.1371/journal.pone.0072378.
- E.H.M. Temme, H. van der Voet, J.T.N.M. Thissen, J. Verkaik-Kloosterman, G. van Donkersgoed, and S. Nonhebel. Replacement of meat and dairy by plant-derived foods: estimated effects on land use, iron and SFA intakes in young Dutch adult females. *Public Health Nutrition*, 16(10):1900–1907, February 2013. URL: https://doi.org/10.1017/s1368980013000232.

- P.E. Boon, J.D. te Biesebeek, I. Sioen, I. Huybrechts, J. Moschandreas, J. Ruprich, A. Turrini, M. Azpiri, L. Busk, T. Christensen, M. Kersting, L. Lafay, K.-H. Liukkonen, S. Papoutsou, L. Serra-Majem, I. Traczyk, S. de Henauw, and J.D. van Klaveren. Long-term dietary exposure to lead in young European children: comparing a pan-European approach with a national exposure assessment. *Food Additives & Contaminants: Part A*, 29(11):1701–1715, November 2012. URL: https://doi.org/10.1080/19440049.2012.709544.
- P.W. Goedhart, H. van der Voet, S. Knüppel, A.L.M. Dekkers, K.W. Dodd, H. Boeing, and J. van Klaveren. A comparison by simulation of different methods to estimate the usual intake distribution for episodically consumed foods. *EFSA Supporting Publications*, 9(6):299E, 2012. URL: https://efsa.onlinelibrary.wiley.com/ doi/abs/10.2903/sp.efsa.2012.EN-299.
- I. Sioen, T. Fierens, M. van Holderbeke, L. Geerts, M. Bellemans, M. de Maeyer, K. Servaes, G. Vanermen, P.E. Boon, and S. de Henauw. Phthalates dietary exposure and food sources for Belgian preschool children and adults. *Environment International*, 48:102–108, November 2012. URL: https://doi.org/10.1016/j.envint. 2012.07.004.
- J.D. van Klaveren, P.W. Goedhart, D. Wapperom, and H. van der Voet. A European tool for usual intake distribution estimation in relation to data collection by EFSA. *EFSA Supporting Publications*, 9(6):300E, 2012. URL: https://efsa.onlinelibrary.wiley.com/doi/abs/10.2903/sp.efsa.2012.EN-300.

• EFSA Panel on Plant Protection Products and their Residues (PPR). Guidance on the use of probabilistic methodology for modelling dietary exposure to pesticide residues. *EFSA Journal*, 10(10):2839, 2012. URL: https://efsa.onlinelibrary.wiley.com/doi/abs/10.2903/j.efsa.2012.2839.

2011

- P.E. Boon, M. Bonthuis, H. van der Voet, and J.D. van Klaveren. Comparison of different exposure assessment methods to estimate the long-term dietary exposure to dioxins and ochratoxin A. *Food and Chemical Toxicology*, 49(9):1979–1988, September 2011. URL: https://doi.org/10.1016/j.fct.2011.05.009.
- C.W. Noorlander, S.P.J. van Leeuwen, J.D. te Biesebeek, M.J.B. Mengelers, and M.J. Zeilmaker. Levels of perfluorinated compounds in food and dietary intake of PFOS and PFOA in the Netherlands. *Journal of Agricultural and Food Chemistry*, 59(13):7496–7505, July 2011. URL: https://doi.org/10.1021/jf104943p.
- O.W. Souverein, W.J. de Boer, A. Geelen, H. van der Voet, J.H. de Vries, M. Feinberg, and P. van 't Veer. Uncertainty in intake due to portion size estimation in 24-hour recalls varies between food groups. *The Journal of Nutrition*, 141(7):1396–1401, May 2011. URL: https://doi.org/10.3945/jn.111.139220.

- P.E. Boon, I. Sioen, H. van der Voet, I. Huybrechts, M. de Neve, P. Amiano, M. Azpiri, L. Busk, T. Christensen, A. Hilbig, T. Hirvonen, S. Koulouridaki, L. Lafay, K.-H. Liukkonen, J. Moschandreas, S. Papoutsou, L. Ribas-Barba, J. Ruprich, L. Serra-Majem, M. Tornaritis, A. Turrini, M. Urtizberea, E. Verger, A. Westerlund, M. Kersting, S. de Henauw, and J.D. van Klaveren. Long-term dietary exposure to lead in young children living in different European countries. *EFSA Supporting Publications*, 7(5):51E, 2010. URL: https://efsa.onlinelibrary.wiley.com/doi/abs/10.2903/sp.efsa.2010.EN-51.
- P.E. Boon, J.D. te Biesebeek, I. Sioen, I. Huybrechts, M. de Neve, P. Amiano, C. Arganini, M. Azpiri, L. Busk, T. Christensen, A. Hilbig, T. Hirvonen, S. Koulouridaki, L. Lafay, K.-H. Liukkonen, J. Moschandreas, S. Papoutsouk, L. Ribas-Barba, J. Ruprich, L. Serra-Majem, M. Tornaritis, A. Turrini, M. Urtizberea, E. Verger, A. Westerlund, M. Kersting, S. de Henauw, and J.D. van Klaveren. Long-term dietary exposure to chromium in young children living in different European countries. *EFSA Supporting Publications*, 7(5):54E, 2010. URL: https://efsa.onlinelibrary.wiley.com/doi/abs/10.2903/sp.efsa.2010.EN-54.
- I. Huybrechts, I. Sioen, P.E. Boon, M. de Neve, P. Amiano, C. Arganini, E. Bower, L. Busk, T. Christensen, A. Hilbig, T. Hirvonen, A. Kafatos, S. Koulouridaki, L. Lafay, K.-H. Liukkonen, S. Papoutsou, L. Ribas-Barba, J. Ruprich, I. Rehurkova, M. Kersting, L. Serra-Majem, A. Turrini, E. Verger, A. Westerlund, M. Tornaritis, J.D. van Klaveren, and S. de Henauw. Long-term dietary exposure to different food colours in young children living in different European countries. *EFSA Supporting Publications*, 7(5):53E, 2010. URL: https://efsa. onlinelibrary.wiley.com/doi/abs/10.2903/sp.efsa.2010.EN-53.
- A. König, A.H. Kuiper, H.J.P. Marvin, P.E. Boon, L. Busk, F. Cnudde, S. Cope, H.V. Davies, M. Dreyer, L.J. Frewer, M. Kaiser, G.A. Kleter, I. Knudsen, G. Pascal, A. Prandini, O. Renn, M.R. Smith, B.W. Traill, H. van der Voet, H. van Trijp, E. Vos, and M.T.A. Wentholt. The SAFE FOODS framework for improved risk analysis of foods. *Food Control*, 21(12):1566–1587, December 2010. URL: https://doi.org/10.1016/j. foodcont.2010.02.012.
- EFSA Panel on Contaminants in the Food Chain (CONTAM). Scientific opinion on lead in food. *EFSA Journal*, 8(4):1570, 2010. URL: https://efsa.onlinelibrary.wiley.com/doi/abs/10.2903/j.efsa.2010.1570.
- I. Sioen, P.E. Boon, I. Huybrechts, M. de Neve, P. Amiano, C. Arganini, L. Busk, C. Chadjigeorgiou, T. Christensen, A. Hilbig, T. Hirvonen, S. Koulouridaki, L. Lafay, K.-H. Liukkonen, J. Moschandreas, S. Papoutsou, L. Ribas-Barba, J. Ruprich, L. Serra-Majem, A. Turrini, M. Urtizberea, M. Kersting, E. Verger, A. Westerlund, J.D. van Klaveren, and S. de Henauw. Long-term dietary exposure to selenium in young children living in different European countries. *EFSA Supporting Publications*, 7(5):56E, 2010. URL: https://efsa.onlinelibrary.wiley.com/doi/abs/10.2903/sp.efsa.2010.EN-56.
- W. Slob, W.J. de Boer, and H. van der Voet. Can current dietary exposure models handle aggregated intake from different foods? a simulation study for the case of two foods. *Food and Chemical Toxicology*, 48(1):178–186, January 2010. URL: https://doi.org/10.1016/j.fct.2009.09.035.

- E.H.M. Temme, H. van der Voet, A.J.C. Roodenburg, A. Bulder, G. van Donkersgoed, and J. van Klaveren. Impact of foods with health logo on saturated fat, sodium and sugar intake of young Dutch adults. *Public Health Nutrition*, 14(4):635–644, September 2010. URL: https://doi.org/10.1017/s1368980010002089.
- J.D. van Klaveren, G. van Donkersgoed, H. van der Voet, C. Stephenson, and P.E. Boon. Cumulative exposure assessment of triazole pesticides. *EFSA Supporting Publications*, 7(2):40E, 2010. URL: https://efsa.onlinelibrary.wiley.com/doi/abs/10.2903/sp.efsa.2010.EN-40.

- B.G.H. Bokkers, M.I. Bakker, P.E. Boon, S. Bosgra, G.W.A.M. van der Heijden, G. Janer, W. Slob, and H. van der Voet. The practicability of the integrated probabilistic risk assessment (IPRA) approach for substances in food. RIVM Report 320121001/2009, 2009. URL: http://hdl.handle.net/10029/260367.
- P.E. Boon, M.I. Bakker, J.D. van Klaveren, and C.T.M. van Rossum. Risk assessment of the dietary exposure to contaminants and pesticide residues in young children in the Netherlands. RIVM report 35007000, 2009. URL: http://www.rivm.nl/bibliotheek/rapporten/350070002.pdf.
- P.E. Boon, K. Svensson, S. Moussavian, H. van der Voet, A. Petersen, J. Ruprich, F. Debegnach, W.J. de Boer, G. van Donkersgoed, C. Brera, J.D. van Klaveren, and L. Busk. Probabilistic acute dietary exposure assessments to captan and tolylfluanid using several European food consumption and pesticide concentration databases. *Food and Chemical Toxicology*, 47(12):2890–2898, December 2009. URL: https://doi.org/10.1016/j.fct.2009.01.040.
- P.E. Boon, E.D. van Asselt, M.I. Bakker, A.G. Kruizinga, and M.C.J.F. Jansen. Trends in diet and exposure to chemicals in Dutch children. Report 2009.002, RIKILT, Wageningen, 2009. URL: http://edepot.wur.nl/7507.
- P.M.J. Bos, P.E. Boon, H. van der Voet, G. Janer, A.H. Piersma, B.J. Brüschweiler, E. Nielsen, and W. Slob. A semi-quantitative model for risk appreciation and risk weighing. *Food and Chemical Toxicology*, 47(12):2941–2950, December 2009. URL: https://doi.org/10.1016/j.fct.2009.03.009.
- S. Bosgra, H. van der Voet, P.E. Boon, and W. Slob. An integrated probabilistic framework for cumulative risk assessment of common mechanism chemicals in food: an example with organophosphorus pesticides. *Regulatory Toxicology and Pharmacology*, 54(2):124–133, July 2009. URL: https://doi.org/10.1016/j.yrtph. 2009.03.004.
- W.J. de Boer, H. van der Voet, B.G.H. Bokkers, M.I. Bakker, and P.E. Boon. Comparison of two models for the estimation of usual intake addressing zero consumption and non-normality. *Food Additives & Contaminants: Part A*, 26(11):1433–1449, November 2009. URL: https://doi.org/10.1080/02652030903161606.
- B.H. Jensen, A. Petersen, and T. Christensen. Probabilistic assessment of the cumulative dietary acute exposure of the population of Denmark to organophosphorus and carbamate pesticides. *Food Additives & Contaminants: Part A*, 26(7):1038–1048, July 2009. URL: https://doi.org/10.1080/02652030902859754.
- A.K. Müller, S. Bosgra, P.E. Boon, H. van der Voet, E. Nielsen, and O. Ladefoged. Probabilistic cumulative risk assessment of anti-androgenic pesticides in food. *Food and Chemical Toxicology*, 47(12):2951–2962, December 2009. URL: https://doi.org/10.1016/j.fct.2009.07.039.
- S.D. Muri, H. van der Voet, P.E. Boon, J.D. van Klaveren, and B.J. Brüschweiler. Comparison of human health risks resulting from exposure to fungicides and mycotoxins via food. *Food and Chemical Toxicology*, 47(12):2963–2974, December 2009. URL: https://doi.org/10.1016/j.fct.2009.03.035.
- EFSA Panel on Contaminants in the Food Chain (CONTAM). Scientific opinion on arsenic in food. *EFSA Journal*, 7(10):1351, 2009. URL: https://efsa.onlinelibrary.wiley.com/doi/abs/10.2903/j.efsa.2009.1351.
- A. J. C. Roodenburg, E. H. M. Temme, O. Howell Davies, and J. C. Seidell. Potential impact of the Choices programme on nutrient intakes in the Dutch population. *Nutrition Bulletin*, 34(3):318–323, September 2009. URL: https://doi.org/10.1111/j.1467-3010.2009.01767.x.
- J. Ruprich, I. Rehurkova, P.E. Boon, K. Svensson, S. Moussavian, H. van der Voet, S. Bosgra, J.D. van Klaveren, and L. Busk. Probabilistic modelling of exposure doses and implications for health risk characterization: glycoalkaloids from potatoes. *Food and Chemical Toxicology*, 47(12):2899–2905, December 2009. URL: https://doi.org/10.1016/j.fct.2009.03.008.

- H. van der Voet, G.W.A.M. van der Heijden, P.M.J. Bos, S. Bosgra, P.E. Boon, S.D. Muri, and B.J. Brüschweiler. A model for probabilistic health impact assessment of exposure to food chemicals. *Food and Chemical Toxicology*, 47(12):2926–2940, December 2009. URL: https://doi.org/10.1016/j.fct.2008.12.027.
- H.J. van Ooijen, H. van der Voet, and M.I. Bakker. Identification and handling of uncertainties in dietary exposure assessment. RIVM Report 320103004, 2009. URL: http://hdl.handle.net/10029/261706.
- EFSA Panel on Plant Protection Products and their Residues (PPR Panel). Scientific opinion on risk assessment for a selected group of pesticides from the triazole group to test possible methodologies to assess cumulative effects from exposure through food from these pesticides on human health. *EFSA Journal*, 7(9):1167, 2009. URL: https://efsa.onlinelibrary.wiley.com/doi/abs/10.2903/j.efsa.2009.1167.

- P.E. Boon, H. Van der Voet, M.T.M. Van Raaij, and J.D. Van Klaveren. Cumulative risk assessment of the exposure to organophosphorus and carbamate insecticides in the Dutch diet. *Food and Chemical Toxicology*, 46(9):3090–3098, September 2008. URL: https://doi.org/10.1016/j.fct.2008.06.083.
- A.L. Brantsæter, M. Haugen, A. de Mul, T. Bjellaas, G. Becher, J. van Klaveren, J. Alexander, and H.M. Meltzer. Exploration of different methods to assess dietary acrylamide exposure in pregnant women participating in the Norwegian mother and child cohort study (MoBa). *Food and Chemical Toxicology*, 46(8):2808–2814, August 2008. URL: https://doi.org/10.1016/j.fct.2008.05.020.
- A. de Mul, M.I. Bakker, M.J. Zeilmaker, W.A. Traag, S.P.J. van Leeuwen, R.L.A.P. Hoogenboom, P.E. Boon, and J.D. van Klaveren. Dietary exposure to dioxins and dioxin-like PCBs in the Netherlands anno 2004. *Regulatory Toxicology and Pharmacology*, 51(3):278–287, August 2008. URL: https://doi.org/10.1016/j.yrtph.2008.04.010.
- B.H. Jensen, J.H. Andersen, A. Petersen, and T. Christensen. Dietary exposure assessment of Danish consumers to dithiocarbamate residues in food: a comparison of the deterministic and probabilistic approach. *Food Additives & Contaminants: Part A*, 25(6):714–721, June 2008. URL: https://doi.org/10.1080/02652030701858262.
- C.J. Seal, A. de Mul, G. Eisenbrand, A.J. Haverkort, K. Franke, S.P.D. Lalljie, H. Mykkänen, E. Reimerdes, G. Scholz, V. Somoza, S. Tuijtelaars, M. van Boekel, J. van Klaveren, S.J. Wilcockson, and L. Wilms. Riskbenefit considerations of mitigation measures on acrylamide content of foods a case study on potatoes, cereals and coffee. *British Journal of Nutrition*, 99(S2):S1–S46, April 2008. URL: https://doi.org/10.1017/s0007114508965314.

- European Food Safety Authority (EFSA). Opinion of the scientific panel on plant protection products and their residues on acute dietary intake assessment of pesticide residues in fruit and vegetables. *EFSA Journal*, 5(8):538, 2007. URL: https://efsa.onlinelibrary.wiley.com/doi/abs/10.2903/j.efsa.2007.538.
- M.I. Bakker, R. de Winter-Sorkina, A. de Mul, P.E. Boon, G. van Donkersgoed, J.D. van Klaveren, B.A. Baumann, W.C. Hijman, S.P.J. van Leeuwen, W. de Boer, and M.J. Zeilmaker. Dietary intake and risk evaluation of polybrominated diphenyl ethers in the Netherlands. *Molecular Nutrition & Food Research*, 52(2):204–216, December 2007. URL: https://doi.org/10.1002/mnfr.200700112.
- P.E. Boon, A.M.J. Ragas., and J.D. van Klaveren. Exploration of aggregate exposure to compounds present in food. Report 2007.016, RIKILT, Wageningen, 2007. URL: http://www.rikilt.wur.nl/NL/publicaties/ Rapporten.
- R. de Winter-Sorkina, M.I. Bakker, G. Wolterink, and M.J. Zeilmaker. Brominated flame retardants: occurrence, dietary intake and risk assessment. RIVM report 320100002/2006, 2007. URL: http://rivm. openrepository.com/rivm/handle/10029/7303.
- H. van der Voet and W. Slob. Integration of probabilistic exposure assessment and probabilistic hazard characterization. *Risk Analysis*, 27(2):351–371, April 2007. URL: https://doi.org/10.1111/j.1539-6924.2007. 00887.x.

- E.D. Caldas, P.E. Boon, and J. Tressou. Probabilistic assessment of the cumulative acute exposure to organophosphorus and carbamate insecticides in the Brazilian diet. *Toxicology*, 222(1-2):132–142, May 2006. URL: https://doi.org/10.1016/j.tox.2006.02.006.
- E.D. Caldas, J. Tressou, and P.E. Boon. Dietary exposure of Brazilian consumers to dithiocarbamate pesticides—a probabilistic approach. *Food and Chemical Toxicology*, 44(9):1562–1571, September 2006. URL: https://doi.org/10.1016/j.fct.2006.04.014.
- J.D. van Klaveren, M.Y. Noordam, P.E. Boon, G. van Donkersgoed, B.C. Ossendorp, M.T.M. van Raaij, and J.G. van der Roest. Trends in normoverschrijdigen, overschrijdingen van de acute referentiewaarde en gesommeerde blootstelling - tussenevaluatie nota duurzame gewasbescherming - deelrapport voedselveiligheid. Report 2006.011, RIKILT, Wageningen, 2006. URL: http://edepot.wur.nl/24544.

2005

- P.E. Boon, A. de Mul, H. van der Voet, G. van Donkersgoed, M. Brette, and J.D. van Klaveren. Calculations of dietary exposure to acrylamide. *Mutation Research/Genetic Toxicology and Environmental Mutagenesis*, 580(1-2):143–155, February 2005. URL: https://doi.org/10.1016/j.mrgentox.2004.10.014.
- A. de Mul, R. de Winter-Sorkina, P.E. Boon, G. van Donkersgoed, M.I. Bakker, and J.D. van Klaveren. Dietary intake of brominated diphenyl ether congeners by the Dutch population. Report 2005.006, RIKILT, Wageningen, 2005. URL: http://edepot.wur.nl/26982.
- M.J. Paulo, H. van der Voet, M.J.W. Jansen, C.J.F. ter Braak, and J.D. van Klaveren. Risk assessment of dietary exposure to pesticides using a Bayesian method. *Pest Management Science*, 61(8):759–766, 2005. URL: https://doi.org/10.1002/ps.1060.
- R.C. Schothorst, H.P. van Egmond, A. de Mul, P.E. Boon, J.D. van Klaveren, and G.J.A. Speijers. Trichothecenes in baby food. RIVM Report 310301002, 2005. URL: http://www.rivm.nl/bibliotheek/rapporten/ 310301002.pdf.

- P.E. Boon, S. Lignell, J.D. van Klaveren, and E.I.M. Tjoe Nij. Estimation of the acute dietary exposure to
 pesticides using the probabilistic approach and the point estimate methodology the generation of work examples using food consumption data from the Netherlands and Sweden. Report 2004.008, RIKILT, Wageningen,
 2004. URL: http://edepot.wur.nl/28647.
- P.E. Boon, E.I.M. Tjoe Nij, N. Koopman, and J.D. van Klaveren. Dietary habits and exposure to pesticides in Dutch infants. Report 2004.017, RIKILT, Wageningen, 2004. URL: http://edepot.wur.nl/44408.
- P.E. Boon, E.I.M. Tjoe Nij, G. van Donkersgoed, and J.D. van Klaveren. Probabilistic intake calculations performed for the codex committee on pesticide residues. Report 2004.005, RIKILT, Wageningen, 2004. URL: http://edepot.wur.nl/36066.
- H. van der Voet and M.J. Paulo. Some explorations into Bayesian modelling of risks due to pesticide intake from food. In M.A.J.S. van Boekel, A. Stein, and A.H.C. van Bruggen, editors, *Bayesian statistics and quality modelling in the agro-food production chain*, pages 145–162. Kluwer, Dordrecht, 2004. URL: http://library.wur.nl/frontis/bayes/13_van_der_voet.pdf.

- P.E. Boon, H. van der Voet, and J.D. van Klaveren. Validation of a probabilistic model of dietary exposure to selected pesticides in Dutch infants. *Food Additives and Contaminants*, 20(sup001):S36–S49, October 2003. URL: https://doi.org/10.1080/0265203031000134956.
- P.E. Boon and J.D. van Klaveren. Cumulative exposure to acetylcholineterase inhibiting compounds in the Dutch population and young childeren. Report 2003.003, RIKILT, Wageningen, 2003. URL: http://edepot. wur.nl/30057.
- P.E. Boon and J.D. van Klaveren. Dietary exposure to pesticides relevant variables and probabilistic modelling. Report 2003.008, RIKILT, Wageningen, 2003. URL: http://edepot.wur.nl/23045.
- R. de Winter-Sorkina, M.I. Bakker, G. van Donkersgoed, and J.D. van Klaveren. Dietary intake of brominated flame retardants by the Dutch population. Report 2003.019, RIKILT, Wageningen, 2003. URL: http://hdl. handle.net/10029/7303.
- R. de Winter-Sorkina, G. van Donkersgoed, M.I. Bakker, and J.D. van Klaveren. Dietary intake of heavy metals (cadmium, lead and mercury) by the Dutch population. Report 2003.016, RIKILT, Wageningen, 2003. URL: http://edepot.wur.nl/41597.
- M.J. Gibney and H. van der Voet. Introduction to the Monte Carlo project and the approach to the validation of probabilistic models of dietary exposure to selected food chemicals. *Food Additives and Contaminants*, 20(sup001):S1–S7, October 2003. URL: https://doi.org/10.1080/0265203031000134947.
- H. van der Voet, P.E. Boon, and J.D. van Klaveren. Validation of Monte Carlo models for estimating pesticide intake of Dutch infants. Report 2003.002, RIKILT, Wageningen, 2003. URL: http://edepot.wur.nl/39363.

2002

 P.E. Boon, G. van Donkersgoed, and J.D. van Klaveren. Human acute exposure assessment of pesticides in fruits and vegetables. Report 2002.002, RIKILT, Wageningen, 2002. URL: https://library.wur.nl/WebQuery/ wurpubs/reports/320297.

LIST OF SYMBOLS

HC	Hazard characterisation	
HI	Hazard Index	
MCR	Maximum Cumulative Ratio	
MOE	Margin of Exposure	
RPF	Relative Potency Factors	
μ	Mean	
s	Standard deviation	
σ^2	Variance	

CHAPTER

SEVEN

APPENDICES

7.1 Api Documentation

Note: This section is under construction.

GET /Api/Actions/GetAll/{idWorkspace} Gets all actions of the workspace with the specified id.

Parameters

• idWorkspace (integer:int32, required) - The id of the workspace.

Status Codes

- 200 OK The actions of the workspace with the specified id.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

GET /Api/Actions/Get/{id} Gets the action with the specified id.

Parameters

• id (integer:int32, required) - The id of the action.

Status Codes

- 200 OK The action with the specified id.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

POST /Api/Actions/Create

Creates a new action action.

Status Codes

- 200 OK A record of the created action.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

GET /Api/Actions/{id}/Settings Get the action with the provided id

Parameters

• id (integer:int32, required) -

- 200 OK The SettingsDto of the requested action id
- 401 Unauthorized Authorization error.
- 404 Not Found Not found.
- 500 Internal Server Error Internal server error.

PATCH /Api/Actions/{id}/Settings Patch the action settings for action with provided id.

Parameters

• id (integer:int32, required) - Action id

Status Codes

- 200 OK The patched action settings
- 401 Unauthorized Authorization error.
- 404 Not Found Not found.
- 400 Bad Request Bad request.
- 500 Internal Server Error Internal server error.

POST /Api/Actions/ImportSettingsConfigFile/{idAction} Sets the action settings of the provided action settings xml file.

Parameters

• idAction (*integer:int32*, *required*) – Id of the action that should be cloned.

Status Codes

- 200 OK –
- 401 Unauthorized Authorization error.
- 400 Bad Request Bad request.
- 500 Internal Server Error Internal server error.

POST /Api/Actions/ImportDataSourceConfigFile/{idAction} Sets the action settings of the provided action data source config xml file.

Parameters

- **idAction** (*integer:int32*, *required*) Id of the action that should be cloned.
- **Status Codes**
 - 200 OK –
 - 401 Unauthorized Authorization error.
 - 400 Bad Request Bad request.
 - 500 Internal Server Error Internal server error.

POST /Api/Actions/Clone

Duplicates an action and all of its settings and output.

Status Codes

- 200 OK A record of the cloned action.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

POST /Api/Actions/UploadActionZipFile/{idWorkspace}

Creates an upload task that creates an action from an action+data zip file. Returns the task id for monitoring the progress.

Parameters

• idWorkspace (integer:int32, required) - Id of the action that should be cloned.

Status Codes

- 200 OK Task id of the upload task.
- 401 Unauthorized Authorization error.
- 400 Bad Request Bad request.
- 500 Internal Server Error Internal server error.

DELETE /Api/Actions/Delete/{id} Deletes the action with the specified id.

Parameters

```
• id (integer:int32, required) – The id of the action that should be deleted.
```

Status Codes

- 200 OK -
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

PUT /Api/Actions/UpdateMetaData/{id} Updates the action meta data.

Parameters

• **id** (*integer:int32*, *required*) – The id of the action.

Status Codes

- 200 OK A record of the updated action meta data.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

PUT /Api/Actions/ConvertToCustomAction/{id} Converts the (standard)action to a custom action.

Parameters

• id (integer: int 32, required) - The id of the action.

Status Codes

- 200 OK A record of the updated action.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

PUT /Api/Actions/UpdateStandardActionVersion/{id} Updates the (standard)action to use the latest standard action version.

Parameters

• id (integer:int32, required) – The id of the action.

- 200 OK A record of the updated action.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

PUT /Api/Actions/SetIsCompute/{id}

```
Specifies whether the data of a (sub)module of the action should be computed or obtained from data.
```

Parameters

```
• id (integer: int 32, required) - The id of the action.
```

Status Codes

- 200 OK A record of the updated action.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

PUT /Api/Actions/SetActionDataSource/{id} Adds a data source to the data source configuration of the action.

Parameters

```
• id (integer:int32, required) – The id of the action.
```

Status Codes

- 200 OK A record of the updated action.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

PUT /Api/Actions/ReplaceActionDataSource/{id} Replaces an action data source with another data source.

Parameters

```
• id (integer: int 32, required) - The id of the action.
```

Status Codes

- 200 OK A record of the updated action.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

GET /Api/Actions/GetSummary/{id}

Returns the summary of the action with the specified id.

Parameters

• id (integer:int32, required) – The id of the action.

Status Codes

- 200 OK Summary section of the specified action.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

GET /Api/Actions/GetDataReadingSummary/{id} Returns the data reading summary of the specified (sub-)module for the action with the specified id.

Parameters

• id (integer:int32, required) - The id of the action.

Query Parameters

• actionType (required) - The id of the (sub-)module

Status Codes

• 200 OK – The data reading summary of the specified (sub-)module for the action with the specified id.

- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.
- GET /Api/Actions/GetDataSelection/{id} Retrieves a data entities page for the provided table.

Parameters

• id (integer:int32, required) -

Query Parameters

• entityType (required) -

Status Codes

- 200 OK A record of the updated action.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.
- GET /Api/Actions/GetDataReadingEntityRecords/{id} Retrieves a data entities page for the provided table.

Parameters

- id (integer:int32, required) -
- **Query Parameters**
 - tableGroupId (required) -
 - scopingType (required) -
 - filteredStatusTypes (array) -
 - page (integer:int32) -
 - pageSize (integer: int 32) -
 - **sort** (*string*) -
 - order (string) -

Status Codes

- 200 OK A collection of data reading summary records.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

GET /Api/Actions/GetDataLinkingEntityRecords/{id} Get action data linking records.

Parameters

• id (integer:int32, required) -

Query Parameters

- tableGroupId (required) -
- scopingType (required) -
- referencedScopingType (required) -
- filteredStatusTypes (array) –
- page (integer:int32) -
- pageSize (integer:int32) -
- **sort** (*string*) -

• order (string) -

Status Codes

- 200 OK A collection of data linking summary records.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

PUT /Api/Actions/SetScope/{id}

Sets the action scope. I.e., specify which data entities (specified by their codes) should be considered in the action.

Parameters

• id (integer: int 32, required) - The id of the action.

Status Codes

- 200 OK A record of the updated action.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

PUT /Api/Actions/ClearScope/{id} Clears the action scope/data selection.

Parameters

• id (integer:int32, required) - The id of the action.

Status Codes

- 200 OK A record of the updated action.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

PUT /Api/Actions/ExtendScope/{id} Extends the action scope.

Parameters

• id (integer: int 32, required) - The id of the action.

Status Codes

- 200 OK A record of the updated action.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

PUT /Api/Actions/ReduceScope/{id} Reduces the action scope.

Parameters

• id (integer: int 32, required) - The id of the action.

Status Codes

- 200 OK A record of the updated action.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

GET /Api/Actions/ExportActionSettings/{id} Returns the settings xml of the action with the specified id.

Parameters

• id (integer: int 32, required) - The id of the action.

Status Codes

- 200 OK File response.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.
- GET /Api/Actions/ExportActionDataSources/{id} Returns the data sources xml of the action with the specified id.

Parameters

• id (integer: int 32, required) - The id of the action.

Status Codes

- 200 OK File response.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

GET /Api/Actions/DownloadActionZip/{id}

Returns all settings and data of the action as a zip-file. Project data is included in the form of csv files (zipped csv format).

Parameters

• **id** (*integer:int32*, *required*) – The id of the action.

Status Codes

- 200 OK File response.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

GET /Api/Actions/DownloadActionZipOriginalData/{id} Returns all settings and data of the action as a zip-file. Project data source files are included in their original forms.

Parameters

• id (integer:int32, required) - The id of the action.

Status Codes

- 200 OK File response.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

GET /Api/Actions/DownloadActionZipNoData/{id} Returns all settings and data source configuration of the action as a zip-file.

Parameters

• **id** (*integer:int32*, *required*) – The id of the action.

Status Codes

- 200 OK File response.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

GET /api/Actions/FileUploadProgress

Retrieves the task progress of the upload task with the specified task id.

Query Parameters

• **taskId** (*string*, *required*) – The task id of the upload task.

Status Codes

- 200 OK Task report of the specified task.
- 400 Bad Request Bad request.
- 500 Internal Server Error Internal server error.

GET /Api/DataSources/Get/{id}

Returns a data source description record of the data source with the specified id.

Parameters

• id (integer:int32, required) – The id of the data source.

Status Codes

- 200 OK Data source description records.
- 401 Unauthorized Authorization error.
- 400 Bad Request Bad request.
- 404 Not Found Record not found.
- 500 Internal Server Error Internal server error.

GET /Api/DataSources/GetAll

Gets all data sources available to the user.

Status Codes

- 200 OK Collection of data source description records.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

GET /Api/DataSources/GetRepositoryDataSources/{id} Gets the data sources of the repository with the specified id.

Parameters

• id (integer:int32, required) -

Status Codes

- 200 OK Collection of data source description records.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

GET /Api/DataSources/GetRemoteRepositoryDataSources/{id} Gets the data sources available in the remote repository with the specified id.

Parameters

• id (integer:int32, required) -

Status Codes

- 200 OK Collection of data source description records.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

POST /Api/DataSources/ImportRemoteDataSource Imports the remote data source specified by the import settings.

- 200 OK Data source description record of the imported data source.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

PUT /Api/DataSources/Move/{id}

Moves the data source with the specified id to another repository.

Parameters

• **id**(*integer:int32*, *required*) – The id of the data source that should be moved.

Status Codes

- 200 OK Data source description record of the moved data source.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

DELETE /Api/DataSources/Delete/{id}

Deletes the data source with the specified id.

Parameters

• **id**(*integer:int32*, *required*) – The id of the data source that should be deleted.

Status Codes

- 200 OK –
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

PUT /Api/DataSources/Rename/{id}

Renames the data source with the specified id.

Parameters

• id (integer: int 32, required) – The id of the data source that is to be renamed.

Status Codes

- 200 OK Data source description record of the moved data source.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

POST /Api/DataSources/UploadNewDataSource/{idRepository}

Creates an upload task that adds a new data source to the repository with the specified id. Returns the task id for monitoring the progress.

Parameters

• **idRepository** (*integer:int32*, *required*) – The id of repository in which the data source should be created.

Status Codes

- 200 OK Task id of the upload task.
- 401 Unauthorized Authorization error.
- 400 Bad Request Bad request.
- 500 Internal Server Error Internal server error.

POST /Api/DataSources/UploadNewDataSourceVersion/{idDataSource}

Creates an upload task that adds a new version to the specified data source. Returns the task id for monitoring the progress.

Parameters

• idDataSource (*integer:int32*, *required*) – The id of data source for which this is a new version.

Status Codes

- 200 OK Task id of the upload task.
- 401 Unauthorized Authorization error.
- 400 Bad Request Bad request.
- 500 Internal Server Error Internal server error.
- GET /Api/DataSources/GetVersion/{idVersion} Gets the data source version with the specified id.

Parameters

• idVersion (integer:int32, required) – The id of the data source version.

Status Codes

- 200 OK Data source version record.
- 401 Unauthorized Authorization error.
- 400 Bad Request Bad request.
- 500 Internal Server Error Internal server error.

GET /Api/DataSources/GetVersions/{id} Gets all versions of the data source with the specified id.

Parameters

• id (integer:int32, required) – The id of the data source.

Status Codes

- 200 OK Data source version records.
- 401 Unauthorized Authorization error.
- 400 Bad Request Bad request.
- 500 Internal Server Error Internal server error.

GET /Api/DataSources/GetWorkspaceDataSourceVersions/{idWorkspace} Returns all data source versions used in the workspace with the specified id.

Parameters

• idWorkspace (integer:int32, required) - The id of the workspace.

Status Codes

- 200 OK Data source version records.
- 401 Unauthorized Authorization error.
- 400 Bad Request Bad request.
- 500 Internal Server Error Internal server error.

GET /Api/DataSources/GetDataSourceVersionUsage/{idVersion} Returns the data source version's usage in actions.

Parameters

```
• idVersion (integer:int32, required) – The id of the version.
```

- 200 OK Data source version usage.
- 401 Unauthorized Authorization error.

- 400 Bad Request Bad request.
- 500 Internal Server Error Internal server error.
- GET /Api/DataSources/GetDataSourceUsage/{idDataSource} Returns the data source's usage in actions.

Parameters

• idDataSource (integer: int 32, required) – The id of the data source.

Status Codes

- 200 OK Data source usage.
- 401 Unauthorized Authorization error.
- 400 Bad Request Bad request.
- 500 Internal Server Error Internal server error.

GET /Api/DataSources/DownloadVersion/{idVersion}

Downloads the data source version dataset file of the version with the specified id.

Parameters

• idVersion (integer:int32, required) -

Status Codes

- 200 OK File response.
- 401 Unauthorized Authorization error.
- 400 Bad Request Bad request.
- 404 Not Found File not found.
- 500 Internal Server Error Internal server error.

GET /Api/DataSources/DownloadVersionCsv/{idVersion}

Downloads the data source version raw data (as imported by MCRA) of the version as zipped csv file collection.

Parameters

• idVersion (integer:int32, required) -

Status Codes

- 200 OK File response.
- 401 Unauthorized Authorization error.
- 400 Bad Request Bad request.
- 404 Not Found File not found.
- 500 Internal Server Error Internal server error.

POST /Api/DataSources/UploadActionZipFile

Creates an upload task that creates an action from an action+data zip file. Returns the task id for monitoring the progress.

- 200 OK Task id of the upload task.
- 401 Unauthorized Authorization error.
- 400 Bad Request Bad request.
- 500 Internal Server Error Internal server error.

GET /api/DataSources/FileUploadProgress

Retrieves the task progress of the upload task with the specified task id.

Query Parameters

• **taskId** (*string*, *required*) – The task id of the upload task.

Status Codes

- 200 OK Task report of the specified task.
- 400 Bad Request Bad request.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

GET /api/Repositories/GetAll

Status Codes

• 200 OK –

GET /Api/Repositories/Get/{id}

Parameters

• id (integer: int 32, required) -

Status Codes

- 200 OK –
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

GET /Api/Repositories/GetDetails/{id}

Parameters

• id (integer: int 32, required) -

Status Codes

- 200 OK –
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

POST /api/Repositories/Create

Status Codes

• 200 OK –

POST /api/Repositories/Update

Status Codes

• 200 OK –

POST /Api/Repositories/Delete/{id}

Parameters

• id (integer:int32, required) -

- 200 OK –
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

POST /Api/Repositories/ForceDelete/{id}

Parameters

• id (integer: int 32, required) -

Status Codes

- 200 OK –
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

POST /api/Repositories/Move

Status Codes

• 200 OK –

POST /api/Repositories/ChangeOwner

Status Codes

• 200 OK –

POST /api/Repositories/AddUserShare

Status Codes

```
• 200 OK –
```

POST /api/Repositories/UpdateUserShare

Status Codes

• 200 OK –

POST /api/Repositories/RemoveUserShare

Status Codes

• 200 OK –

POST /api/Repositories/AddGroupShare

Status Codes

• 200 OK –

POST /api/Repositories/UpdateGroupShare

Status Codes

• 200 OK –

POST /api/Repositories/RemoveGroupShare

Status Codes

• 200 OK –

GET /Api/Repositories/GetRepositoryDataSourcesInUse/{id} Gets the data sources that are in use in the repository with the specified id.

Parameters

• id (integer:int32, required) -

- 200 OK –
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

GET /Api/Repositories/GetRepositoryDataSourceUsage/{id}

Gets all project information still associated with any datasources in the repository or any subrepositories

Parameters

• id (integer:int32, required) – Id of the main repository to search

Status Codes

- 200 OK -
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

GET /Api/Workspace/GetAll

Gets all workspaces available to the user.

Status Codes

- 200 OK Collection of workspace description records.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

GET /Api/Workspace/Get/{id} Gets the workspace with the specified id.

Parameters

• id (integer:int32, required) - The id of the workspace.

Status Codes

- 200 OK The workspace with the specified id.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

POST /Api/Workspace/Create

Creates a workspace based on the provided form data.

Status Codes

- 200 OK A description record of the created workspace.
- 500 Internal Server Error Internal server error.

POST /Api/Workspace/Update/{id}

Updates the workspace meta data.

Parameters

• id (integer:int32, required) – The id of the workspace that is to be updated.

Status Codes

• 200 OK – A description record of the created workspace.

DELETE /Api/Workspace/Delete/{id}

Deletes the workspace with the specified id.

Parameters

• id (integer: int 32, required) – The id of the workspace that should be deleted.

Status Codes

• 200 OK -

GET /Api/Workspace/GetDataSources/{id} Returns all data sources used in the workspace with the specified id.

Parameters

• id (integer: int 32, required) - The id of the workspace.

Status Codes

- 200 OK A description record of the created workspace.
- 401 Unauthorized Authorization error.
- 400 Bad Request Bad request.
- 500 Internal Server Error Internal server error.

POST /Api/Workspace/AddDataSource/{id} Adds a data source version to the workspace.

Parameters

• id (integer:int32, required) - The id of the workspace.

Status Codes

- 200 OK On success.
- 401 Unauthorized Authorization error.
- 400 Bad Request Bad request.
- 500 Internal Server Error Internal server error.

POST /Api/Workspace/RemoveDataSource/{id} Removes a data source version from the workspace.

Parameters

• id (integer: int 32, required) - The id of the workspace.

Status Codes

- 200 OK On success.
- 401 Unauthorized Authorization error.
- 400 Bad Request Bad request.
- 500 Internal Server Error Internal server error.

GET /Api/About/VersionInfo

Returns a version info record containing version number and build date.

Status Codes

• 200 OK – About info record.

GET /Api/About

Returns a version info record containing version number and build date.

Status Codes

• 200 OK – About info record.

GET /api/ActionTypes/Get

Returns a collection of the action type definitions available in the toolbox.

Status Codes

- 200 OK Collection of action type definition records.
- 500 Internal Server Error Internal server error.

GET /api/DataFormats/Get

Returns a collection of the data format definitions available in the toolbox.

- 200 OK Collection of data format definition records.
- 500 Internal Server Error Internal server error.

GET /Api/Outputs/Get/{id} Gets the output with the specified id.

Parameters

• id (integer: int 32, required) - The id of the output.

Status Codes

- 200 OK Output record.
- 400 Bad Request Bad request.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

GET /Api/Outputs/GetFromAction/{idAction}

Gets the outputs of the action with the specified id. Returns a list of OutputInfoViewModel records.

Parameters

• idAction (integer:int32, required) - Id of the action.

Status Codes

- 200 OK List of output records.
- 400 Bad Request Bad request.
- 500 Internal Server Error Internal server error.

GET /Api/Outputs/GetOutputReportToc/{id}

Get the output summary descrialized into the object hierarchy.

Parameters

• **id** (*integer:int32*, *required*) – Id of the output.

Status Codes

- 200 OK Output hierarchy object.
- 400 Bad Request Bad request.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

GET /Api/Outputs/GetOutputReportTocs Gets the hierarchical output report summary toc objects of the outputs with the specified ids.

Query Parameters

• ids (array) – Ids of the outputs.

Status Codes

- 200 OK Array of summary hierarchical toc objects.
- 400 Bad Request Bad request.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.
- GET /Api/Outputs/GetOutputReportSection Gets output report section with the specified id of the output with the specified id.

Query Parameters

• idOutput (integer:int32, required) - Id of the output

• idSection (string:guid, required) - Id of the section.

Status Codes

- 200 OK Html string of the section content.
- 400 Bad Request Bad request.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

GET /Api/Outputs/GetShortOutputSummary Gets short output report summary with the specified id of the output

Query Parameters

• **idOutput** (*integer:int32*, *required*) – Id of the output

Status Codes

- 200 OK Html string of the short output summary content.
- 400 Bad Request Bad request.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

GET /Api/Outputs/GetOutputReportTableData Gets output table data section with the specified id of the output with the specified id.

Query Parameters

- idOutput (integer:int32, required) Id of the output.
- **idSection** (*string:guid*, *required*) Id of the section.
- maxRecords (integer:int32, required) Maximum number of records.
- **isTree** (*boolean*, *required*) if true, the hierarchy defining data should be incorporated

Status Codes

- 200 OK Html string of the table data.
- 400 Bad Request Bad request.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

GET /Api/Outputs/GetOutputReportTableContent

Gets output table data section with the specified id of the output with the specified id.

Query Parameters

- idOutput (integer:int32, required) Id of the output.
- idSection (*string:guid*, *required*) Id of the section.
- caption (string, required) Table caption
- maxRecords (integer:int32, required) Maximum number of records.
- **columnOrder** (*string*, *required*) Comma separated list of column indices determining the shown columns
- **isTree** (*boolean*, *required*) if true, the hierarchy defining data should be incorporated
- **rotate** (*boolean*, *required*) if true, the rows and columns are switched, effectively rotating the table

Status Codes

- 200 OK Html string of the full table data.
- 400 Bad Request Bad request.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

GET /Api/Outputs/GetOutputReportChart

Gets output table data section with the specified id of the output with the specified id.

Query Parameters

- idOutput (integer:int32, required) Id of the output.
- **idSection** (*string:guid*, *required*) Id of the section.

Status Codes

- 200 OK Html string of the table data.
- 400 Bad Request Bad request.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

GET /Api/Outputs/GetTaskSettingsSection

Gets the settings section content of the task with the specified id.

Query Parameters

• **idTask** (*integer:int32*, *required*) – Id of the task.

Status Codes

- 200 OK Html string of the section.
- 400 Bad Request Bad request.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

GET /Api/Outputs/DownloadFullPdf

Returns pdf file of the output report with the specified id.

Query Parameters

• idOutput (integer:int32, required) - Id of the output.

Status Codes

- 200 OK Report pdf file.
- 400 Bad Request Bad request.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

GET /Api/Outputs/DownloadShortReportPdf Returns pdf file of the output report with the specified id.

Query Parameters

• idOutput (integer:int32, required) - Id of the output.

- 200 OK Report pdf file.
- 400 Bad Request Bad request.

- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

GET /Api/Outputs/DownloadSectionPdf

Returns pdf file of the report section with the specified id of the output with the specified id.

Query Parameters

- idOutput (integer:int32, required) Id of the output.
- **idSection** (*string:guid*, *required*) Id of the section.

Status Codes

- 200 OK Report pdf file.
- 400 Bad Request Bad request.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

GET /Api/Outputs/GetSectionHtml

Returns pdf file of the report section with the specified id of the output with the specified id.

Query Parameters

- idOutput (integer:int32, required) Id of the output.
- **idSection** (*string:guid*, *required*) Id of the section.

Status Codes

- 200 OK Report section HTML file.
- 400 Bad Request Bad request.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

GET /Api/Outputs/GetShortReportHtml

Returns pdf file of the report section with the specified id of the output with the specified id.

Query Parameters

• idOutput (integer:int32, required) - Id of the output.

Status Codes

- 200 OK Short report HTML.
- 400 Bad Request Bad request.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

GET /Api/Outputs/GetFullReportHtml

Returns pdf file of the report section with the specified id of the output with the specified id.

Query Parameters

• **idOutput** (*integer:int32*, *required*) – Id of the output.

- 200 OK Short report HTML.
- 400 Bad Request Bad request.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

GET /Api/Outputs/DownloadReportCsvZip

Returns zip file with csv files of the tables of the output report of the output with the specified id.

Query Parameters

• idOutput (integer:int32, required) - Id of the output.

Status Codes

- 200 OK Zip file.
- 400 Bad Request Bad request.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

GET /Api/Outputs/DownloadTableCsv

Returns csv file of the report table with the specified id of the output with the specified id.

Query Parameters

- **idOutput** (*integer:int32*, *required*) Id of the output.
- idSection (string:guid, required) Id of the section.

Status Codes

- 200 OK Csv file.
- 400 Bad Request Bad request.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

GET /Api/Outputs/DownloadSectionCsvZip

Returns zip file with csv files of the tables of the section with the specified id of the output with the specified id.

Query Parameters

- idOutput (integer:int32, required) Id of the output.
- idSection (string:guid, required) Id of the section.

Status Codes

- 200 OK Zip file.
- 400 Bad Request Bad request.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

GET /Api/Outputs/DownloadRawData

Returns a zip file of the generated raw data of the output with the specified id.

Query Parameters

• idOutput (integer:int32, required) - Id of the output.

Status Codes

- 200 OK Zip file.
- 400 Bad Request Bad request.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

GET /api/ScopingTypes/Get

Returns a collection of the scoping type definitions available in the toolbox.

Status Codes

- 200 OK Collection of workspace description records.
- 500 Internal Server Error Internal server error.

GET /api/UnitDefinitions/Get

Returns a collection of the unit definitions available in the toolbox.

Status Codes

- 200 OK Collection of unit definition records.
- 500 Internal Server Error Internal server error.

GET /Api/Tasks/Get/{id} Returns the task with the specified id.

Parameters

• id (integer:int32, required) – The id of the task.

Status Codes

• 200 OK – Task record.

GET /Api/Tasks/GetWorkspaceTasks/{idWorkspace} Gets all tasks of a workspace.

Parameters

• idWorkspace (integer:int32, required) - The id of the workspace.

Status Codes

• 200 OK – Task record.

PUT /Api/Tasks/UpdateDescription/{id}

Updates the task description.

Parameters

• **id** (*integer:int32*, *required*) – The id of the task.

Status Codes

- 200 OK Task record.
- 401 Unauthorized Authorization error.
- 400 Bad Request Bad request.
- 500 Internal Server Error Internal server error.

GET /Api/Tasks/GetTaskLog/{id}

Retrieves the latest tasklog of the task with the specified id.

Parameters

• id (integer: int 32, required) - The id of the task.

Status Codes

- 200 OK Task log string.
- 401 Unauthorized Authorization error.
- 400 Bad Request Bad request.
- 500 Internal Server Error Internal server error.

POST /Api/Tasks/ScheduleTask/{idAction}

Starts a calculation task for the given action and returns the id of this task.

Parameters

• idAction (*integer:int32*, *required*) – The id of the action from which the task should be spawned.

Status Codes

- 200 OK Task record.
- 401 Unauthorized Authorization error.
- 400 Bad Request Bad request.
- 500 Internal Server Error Internal server error.

POST /Api/Tasks/ScheduleSubTask/{idAction} Starts a concentration model calculation task for the given project and returns the id of this task.

Parameters

• idAction (integer:int32, required) – The id of the action.

Query Parameters

• **actionType** (*required*) – The type of the sub-module of the action.

Status Codes

- 200 OK Task record.
- 401 Unauthorized Authorization error.
- 400 Bad Request Bad request.
- 500 Internal Server Error Internal server error.

GET /Api/Tasks/GetProgress/{taskId} Returns the progress of the specified task.

Parameters

• **taskId** (*integer:int32*, *required*) – Id of the task.

Status Codes

- 200 OK Task progress record.
- 401 Unauthorized Authorization error.
- 400 Bad Request Bad request.
- 500 Internal Server Error Internal server error.

GET /Api/Tasks/GetActiveTaskStatuses/{idWorkspace} Gets the statuses of all active tasks of a workspace.

Parameters

• idWorkspace (integer:int32, required) - Id of the workspace.

Status Codes

- 200 OK Task progress record.
- 401 Unauthorized Authorization error.
- 400 Bad Request Bad request.
- 500 Internal Server Error Internal server error.

PUT /Api/Tasks/Abort/{id} Aborts the execution of the task with the specified id.

Parameters

```
• id (integer:int32, required) - The id of the task.
```

Status Codes

- 200 OK –
- 401 Unauthorized Authorization error.
- 400 Bad Request Bad request.
- 500 Internal Server Error Internal server error.

DELETE /Api/Tasks/Delete/{id}

Deletes the task with the specified id.

Parameters

• id (integer:int32, required) -

Status Codes

- 200 OK –
- 401 Unauthorized Authorization error.
- 400 Bad Request Bad request.
- 500 Internal Server Error Internal server error.

DELETE /Api/Tasks/BatchDelete

Deletes the tasks with the specified ids.

Query Parameters

• **ids** (*array*) – The ids of the tasks that should be deleted.

Status Codes

- 200 OK –
- 401 Unauthorized Authorization error.
- 400 Bad Request Bad request.
- 500 Internal Server Error Internal server error.

GET /Api/Tasks/DownloadTaskZip/{id}

Download a zip file containing the task's settings and data at the time of the task's creation. The task's settings and data sources are deserialized from the ProjectSettings and DataSourceSettings fields

Parameters

• id (integer:int32, required) - The ID of the task

Status Codes

- 200 OK File result (zip file).
- 401 Unauthorized Authorization error.
- 400 Bad Request Bad request.
- 500 Internal Server Error Internal server error.

GET /api/StandardActionTypes/Get Returns a collection of the standard action type definitions available for the user.

Status Codes

- 200 OK Collection of standard action type definition records.
- 500 Internal Server Error Internal server error.

GET /api/ActionClasses/Get

Returns a collection of the action class definitions available in the toolbox.

Status Codes

• 200 OK – Collection of action class definition records.

• 500 Internal Server Error – Internal server error.

7.2 Munro collection

This collection can be downloaded here.

7.3 Unit definitions

7.3.1 Benchmark response types

Accepted benchmark response types.

Name	Short name	Aliases	Description
Fraction change	Fraction change	Fraction- Change, FactorChange	The benchmark response is defined as a fraction change of the background response (i.e., defined for both increase and decrease). E.g., for a factor of 0.1, the benchmark response is at +/- 10% of background response.
Percentage change	Percentage change	Percent- ageChange	The benchmark response is defined as a percentage change of the background response (i.e., defined for both increase and decrease). E.g., for a percentage of 10, the benchmark response is at +/- 10% of background response.
Fraction of background response	Fraction of background	Factor, Facto- rOfBackground	The benchmark response is defined as a fraction of the background response. E.g., for a factor of 0.9, the benchmark response is at 0.9 times the background response (i.e., a decrease).
Percentage of background response	Percentage of background	Percentage, PercentageOf- Background	The benchmark response is defined as a percentage of the background response. E.g., for a percentage of 90, the benchmark response is at 90% of the background response (i.e., a decrease).
Extra risk	ER	ExtraRisk	For quantal response types. The benchmark dose is defined as the dose that corresponding with an extra risk of a factor times the background risk. A factor of 0.05 corresponds with 5% extra risk.
Additional risk	AR	AdditionalRisk	For quantal response types. The benchmark dose is defined as the dose that corresponding with an additional risk of a factor times the background risk. A factor of 0.05 corresponds with 5% additional risk.
ED50	ED50	ED50	For quantal response types. The benchmark dose is defined as the dose that corresponds with an estimated risk of 50% (ED50).
Absolute threshold value	Threshold value	Absolute	The benchmark dose is defined as an absolute threshold value.
Absolute difference	Absolute difference	Difference	The benchmark dose is defined an absolute difference with the background risk.

Table 7.1: Unit definition for Benchmark response types.

7.3.2 Body weight units

Units for describing person body weights.

T 11 70 II '	1 C '''	C D 1	• • •
Table 7.2: Unit	definition	for Body	weight units.

Name	Short name	Aliases
Kilogram	kg	kg, kilograms, kilogr, 3, G167A
Gram	g	g, grams, gr, 0, G148A

7.3.3 Concentration units

Units for describing substance concentrations.

Name	Short name	Aliases
kilogram/kilogram	kg/kg	kg/kg, kilogram/kilogram, kilogram/kg, 0, G063A
gram/kilogram	g/kg	g/kg, gram/kilogram, gram/kg, gr/kg, -3, G015A, G060A, G191A
milligram/kilogram	mg/kg	mg/kg, milligram/kilogram, milligram/kg, milligr/kg, -6, G049A, G061A
micro- gram/kilogram	µg/kg	μg/kg, ug/kg, microgram/kilogram, microgram/kg, microgr/kg, -9, G050A, G076A
nanogram/kilogram	ng/kg	ng/kg, nanogram/kilogram, nanogram/kg, nanogr/kg, -12, G077A, G080A
picogram/kilogram	pg/kg	pg/kg, picogram/kilogram, picogram/kg, picogr/kg, -15, G081A
kilogram/liter	kg/L	kg/l, kg/L, kilogram/liter, kilogram/litre, G017A
gram/liter	g/L	g/l, g/L, gram/liter, gram/litre, gr/l, gr/L, G016A
milligram/liter	mg/L	mg/l, mg/L, milligram/liter, milligram/litre, milligr/l, milligr/L, G052A, G062A
microgram/liter	µg/L	μg/l, ug/L, microgram/liter, microgram/litre, microgr/l, microgr/L, G051A, G079A
nanogram/liter	ng/L	ng/l, ng/L, nanogram/liter, nanogram/litre, nanogr/l, nanogr/L, G078A
picogram/liter	pg/L	pg/l, pg/L, picogram/liter, picogram/litre, picogr/l, picogr/L
micro- gram/milliliter	µg/mL	μg/ml, ug/mL, microgram/milliliter, microgram/millilitre, microgr/ml, microgr/mL
nanogram/milliliter	ng/mL	ng/ml, ng/mL, nanogram/milliliter, nanogram/millilitre, nanogr/ml, nanogr/mL

	** * * * * * *	
Table 7.3:	Unit definition	for Concentration units.

7.3.4 Concentration value types

Concentration value types.

Table 7.4. One definition for Concentration value types.			
Name	Short name	Aliases	Description
Mean	MC	MeanConcen-	Mean value from the residue trials.
concentration		tration,	
		Concentration-	
		Mean, MC	
Median	MR	MedianConcen-	Median concentration / residue value of the
concentration		tration, MR,	positive measurements of the residue trials.
		STMR, Super-	
		visedTrialMedi-	
		anResidue	
Highest	HR	HighestConcen-	Highest measured residue / concentration
concentration		tration,	value.
		HighestResidue,	
		HR	
Concentration	СР	Percentile	
percentile			
Limit of	LOQ	LOQ	
quantification			
Maximum	MRL	MRL	
residue limit			

Table 7.4: Unit definition for Concentration value types.

7.3.5 Consumption intake units

Units for consumption intakes amounts.

Table 7.5:	Unit definition for	· Consumption	intake units.
1 uoie 7.5.	Onit dominition for	consumption	mune unus.

Name	Short name	Aliases
gram/kilogram	g/kg bw/day	g/kg bw, gram/kg bw, g/kg bw/day, gram/kg bw/day, gr/kg
bodyweight/day		bw/day, G212A
gram/day	g/day	gram, grams, g/day, g/day, gram/day, gr/day

7.3.6 Consumption units

Units for consumption amounts.

Table 7.6: Unit definition for Consumption units.

Name	Short name	Aliases
kilogram	kg	kg, kilograms, kilogr, 3, G167A
Gram	g	g, grams, gr, 0, G148A

7.3.7 Consumption value types

Consumption value types.

Table 7.7: Unit definition for Consumption value types.

Name	Short name	Aliases
Large portion	LP	LP, LargePortion
Mean consumption	MC	MC, MeanConsumption
Percentile	Percentile	Percentile, P

7.3.8 Dose response model types

Known dose response model types.

Name	Short name	Aliases	Description
Exp-m1	Exp-m1	Expm1	
Exp-m2	Exp-m2	Expm2	
Exp-m3	Exp-m3	Expm3	
Exp-m4	Exp-m4	Expm4	
Exp-m5	Exp-m5	Expm5	
Hill-m1	Hill-m1	Hillm1	
Hill-m2	Hill-m2	Hillm2	
Hill-m3	Hill-m3	Hillm3	
Hill-m4	Hill-m4	Hillm4	
Hill-m5	Hill-m5	Hillm5	
TwoStage	TwoStage	TwoStage	
LogLogist	LogLogist	LogLogist	
Weibull	Weibull	Weibull	
LogProb	LogProb	LogProb	
Gamma	Gamma	Gamma	
Logistic	Logistic	Logistic	
Probit	Probit	Probit	
LVM Exp m2	LVM Exp m2	LVM Exp m2	
LVM Exp m3	LVM Exp m3	LVM_Exp_M3	
LVM Exp m4	LVM Exp m4	LVM_Exp_M4	
LVM Exp m5	LVM Exp m5	LVM_Exp_M5	
LVM Hill m2	LVM Hill m2	LVM Hill m2	
LVM Hill m3	LVM Hill m3	LVM_Hill_M3	
LVM Hill m4	LVM Hill m4	LVM_Hill_M4	
LVM Hill m5	LVM Hill m5	LVM Hill m5	

Table 7.8.	Unit definition	for Dose response	model types
1 abic 7.0.	Onit dominion		model types.

7.3.9 Dose units

Units for describing substance doses.

hort name	Aliases
/kg bw/day	g/kg bw/day, gram/kg bw/day, gr/kg bw/day
ng/kg bw/day	mg/kg bw/day, milligram/kg bw/day, milligr/kg bw/day, G211A
/	kg bw/day

Table 7.9: Unit definition for Dose units.

continues on next page

Name	Short name	- continued from previous page Aliases	
micro-	µg/kg bw/day	μg/kg bw/day, microgram/kg bw/day, microgr/kg bw/day	
gram/kilogram	µg/ng om/auj		
bodyweight/day			
nanogram/kilogram	ng/kg bw/day	ng/kg bw/day, nanogram/kg bw/day, nanogr/kg bw/day	
bodyweight/day	8 8 9 9 9		
picogram/kilogram	pg/kg bw/day	pg/kg bw/day, picogram/kg bw/day, picogr/kg bw/day	
bodyweight/day	10 0 0	18 8	
fem-	fg/kg bw/day	fg/kg bw/day, femtogram/kg bw/day, femtogr/kg bw/day	
togram/kilogram			
bodyweight/day			
gram/gram	g/g bw/day	g/g bw/day, gram/g bw/day, gr/g bw/day	
bodyweight/day			
milligram/gram	mg/g bw/day	mg/g bw/day, milligram/g bw/day, milligr/g bw/day	
bodyweight/day			
microgram/gram	µg/g bw/day	µg/g bw/day, microgram/g bw/day, microgr/g bw/day	
bodyweight/day	·		
nanogram/gram	ng/g bw/day	ng/g bw/day, nanogram/g bw/day, nanogr/g bw/day	
bodyweight/day			
picogram/gram	pg/g bw/day	pg/g bw/day, picogram/g bw/day, picogr/g bw/day	
bodyweight/day			
femtogram/gram	fg/g bw/day	fg/g bw/day, femtogram/g bw/day, femtogr/g bw/day	
bodyweight/day			
kilogram/day	kg/day	kg/day, kilogram/day, kilogr/day	
gram/day	g/day	g/day, gram/day, gr/day	
milligram/day	mg/day	mg/day, milligram/day, milligr/day	
microgram/day	µg/day	µg/day, microgram/day, microgr/day	
nanogram/day	ng/day	ng/day, nanogram/day, nanogr/day	
picogram/day	pg/day	pg/day, picogram/day, picogr/day	
femtogram/day	fg/day	fg/day, femtogram/day, femtogr/day	
kilogram/kilogram	kg/kg	kg/kg, kilogram/kilogram, kilogram/kg, kg/kg bw	
gram/kilogram	g/kg	g/kg, gram/kilogram, gram/kg, gr/kg, g/kg bw	
milligram/kilogram	mg/kg	mg/kg, milligram/kilogram, milligram/kg, milligr/kg,	
		mg/kg bw, G225A	
micro-	µg/kg	µg/kg, microgram/kilogram, microgram/kg, microgr/kg,	
gram/kilogram		μg/kg bw	
nanogram/kilogram	ng/kg	ng/kg, nanogram/kilogram, nanogram/kg, nanogr/kg,	
		ng/kg bw	
picogram/kilogram	pg/kg	pg/kg, picogram/kilogram, picogram/kg, picogr/kg, pg/kg	
		bw	
Molar	М	M, mol/L	
millimolar	mM	mM, mmol/L	
micromolar	μΜ	uM, µM, umol/L	
nanomolar	nM	nM, nmol/L	
moles	moles	moles, Moles	
millimoles	mmoles	mmoles, mMoles	
micromoles	µmoles	umoles, uMoles	
nanomoles	nmoles	nmoles, nMoles	

 Table 7.9 – continued from previous page

7.3.10 Exposure route types

The different routes in which an individual is exposed to substance concentrations.

	Tuble 7.10. Only definition for Exposure route types.				
Name	Short name	Aliases	Description		
Dietary	Dietary	Dietary	Dietary exposure.		
exposure					
Non-dietary	Oral	Oral	Non-dietary oral exposure.		
oral exposure					
Non-dietary	Dermal	Dermal	Non-dietary dermal exposure.		
dermal					
exposure					
Non-dietary	Inhalation	Inhalation	Non-dietary inhalation exposure.		
inhalation					
exposure					
At target	At target	AtTarget	Exposures directly at the target (organ).		

Table 7.10: Unit definition for Exposure route types.

7.3.11 Exposure types

The different types of exposure. I.e., acute or chronic.

Table 7.11: Unit definition for Exposure types.

Name	Short name	Aliases	Description
Acute	Acute	Acute	Acute exposure.
Chronic	Chronic	Chronic	Chronic exposure.

7.3.12 Exposure units

Units for describing substance exposures.

Table 7.12: Unit definition for Exposure units.

Name	Short name	Aliases
gram/kilogram	g/kg bw/day	g/kg bw/day, g/kg/day, gram/kg bw/day, gr/kg bw/day,
bodyweight/day		G212A
milligram/kilogram	mg/kg bw/day	mg/kg bw/day, mg/kg/day, milligram/kg bw/day,
bodyweight/day		milligr/kg bw/day, G211A
micro-	µg/kg bw/day	μg/kg bw/day, μg/kg/day, microgram/kg bw/day,
gram/kilogram		microgr/kg bw/day, G210A
bodyweight/day		
nanogram/kilogram	ng/kg bw/day	ng/kg bw/day, ng/kg/day, nanogram/kg bw/day, nanogr/kg
bodyweight/day		bw/day, G214A
picogram/kilogram	pg/kg bw/day	pg/kg bw/day, picogram/kg bw/day, picogr/kg bw/day
bodyweight/day		
fem-	fg/kg bw/day	fg/kg bw/day, fg/kg/day, femtogram/kg bw/day,
togram/kilogram		femtogr/kg bw/day
bodyweight/day		
gram/gram	g/g bw/day	g/g bw/day, g/g/day, gram/g bw/day, gr/g bw/day
bodyweight/day		
milligram/gram	mg/g bw/day	mg/g bw/day, mg/g/day, milligram/g bw/day, milligr/g
bodyweight/day		bw/day
microgram/gram	µg/g bw/day	μg/g bw/day, μg/g/day, microgram/g bw/day, microgr/g
bodyweight/day		bw/day

continues on next page

Name	Short name	Aliases
nanogram/gram	ng/g bw/day	ng/g bw/day, nanogram/g bw/day, nanogr/g bw/day
bodyweight/day		
picogram/gram	pg/g bw/day	pg/g bw/day, pg/g/day, picogram/g bw/day, picogr/g
bodyweight/day		bw/day
femtogram/gram	fg/g bw/day	fg/g bw/day, fg/g/day, femtogram/g bw/day, femtogr/g
bodyweight/day		bw/day
kilogram/day	kg/day	kg/day, kilogram/day, kilogr/day
gram/day	g/day	g/day, gram/day, gr/day
milligram/day	mg/day	mg/day, milligram/day, milligr/day
microgram/day	µg/day	μg/day, microgram/day, microgr/day
nanogram/day	ng/day	ng/day, nanogram/day, nanogr/day
picogram/day	pg/day	pg/day, picogram/day, picogr/day
femtogram/day	fg/day	fg/day, femtogram/day, femtogr/day
gram/kilogram	g/kg	g/kg, gram/kg, gr/kg, G015A
milligram/kilogram	mg/kg	mg/kg, milligram/kg, milligr/kg, G061A
micro-	µg/kg	μg/kg, microgram/kg, microgr/kg, G050A
gram/kilogram		
nanogram/kilogram	ng/kg	ng/kg, nanogram/kg, nanogr/kg, G077A
picogram/kilogram	pg/kg	pg/kg, picogram/kg, picogr/kg, G081A
fem-	fg/kg	fg/kg, femtogram/kg, femtogr/kg
togram/kilogram		
gram	g	g, gram, gr, G148A
milligram	mg	mg, milligram, milligr, G155A
microgram	μg	μg, microgram, microgr
nanogram	ng	ng, nanogram, nanogr, G120A
picogram	pg	pg, picogram, picogr, G125A
femtogram	fg	fg, femtogram, femtogr

Table 7.12 - continued from previous page

7.3.13 Harvest application types

Available harvest application types.

Name	Short name	Aliases	Description
Pre-harvest application	Pre-harvest	PreHarvest	Pre-harvest application
Post-harvest application	Post-harvest	PostHarvest	Post-harvest application

7.3.14 Hazard characterisation types

Known hazard characterisation types.

Table 7.14:	Unit definition	for Hazard	characterisation types.
14010 /.111.	onn aonnition	i i oi i i iuzui u	characterisation types.

Name	Short name	Aliases
Benchmark dose	BMD	BMD
No observed	NOAEL	NOAEL
adverse effect level		
Lowest observed	LOAEL	LOAEL
adverse effect level		
Acceptable daily	ADI	ADI
intake		
Acute reference	ARfD	ARfD
dose		
No observed effect	NOEL	NOEL
level		

7.3.15 Point of departure types

Known point of departure types.

Name	Short name	Aliases	Description	
Benchmark	BMD	BMD		
dose				
No observed	NOAEL	NOAEL		
adverse effect				
level				
Lowest	LOAEL	LOAEL		
observed				
adverse effect				
level				
No observed	NOEL	NOEL		
effect level				
LD50	LD50	LD50	Median lethal dose.	

Table 7.15: Unit definition for Point of departure types.

7.3.16 Response types

Available response types.

	Tuble 7.10. Chit demitten for Response (Spes.				
Name	Short name	Aliases	Description		
Continuous	CM	Continuous-	Response values are positive real numbers,		
multiplicative		Multiplicative	e.g., weight, size.		
Continuous	CA	ContinuousAd-	Response values are real numbers, e.g.,		
additive		ditive	weight change, temperature.		
Binary	В	Binary	Response values have binary outcomes		
			(yes/no, true/false, success/failure, 0/1, etc.).		
Quantal	Q	Quantal,	Response is measured in terms of number of		
		Binomial	successes out of N possible.		
Quantal group	QG	QuantalGroup	Individual responses are measured as binary		
			values, which may be grouped to form a		
			quantal response.		
Count	С	Count	Number of items (cells, molecules, deaths,		
			etc.) in given interval/area/volume.		
Ordinal	0	Ordinal	Relative scores (or graded scores) useable		
			only for ranking.		

Table 7.16: Unit definition for Response types.

7.3.17 Target dose level types

This unit specifies whether a dose is assumed to be an internal or external dose.

Table 7.17:	Unit definition for	Target dose	level types.

Name	Short name	Aliases	Description
External	Ext	External, Ext	External exposure.
Internal	Int	Internal, Int	Internal exposure.

7.3.18 Test system types

Available test system types.

Table 7.18: Unit definition for Test system types.

Name	Short name	Aliases	Description
In vivo	In vivo	InVivo	In vivo
Cell line	Cell line	CellLine	CellLine
Primary cells	Primary cells	PrimaryCells	PrimaryCells
Tissue	Tissue	Tissue	Tissue
Organ	Organ	Organ	Organ

7.4 Transformations

7.4.1 Box Cox power transformation

The Box-Cox power transformation is a data transformation to achieve a better normality and to stabilize the variance. In MCRA, the transformation parameter p in $(y^p - 1)/p$ is determined by maximizing the log-likelihood function

$$l(p) = -\frac{n}{s} \log \left[\frac{1}{n} \sum_{i=1}^{n} (y_i^{(p)} - \overline{y^{(p)}})^2 \right] + (p-1) \sum_{i=1}^{n} \log y_i$$

where i indexes the n observations and

$$\overline{y^{(p)}} = \frac{1}{n} \sum_{i=1}^n y_i^{(p)}$$

is the average of the $y_i^{(p)}$ (Box & Cox, 1964) [[Box et al., 1964]].

7.5 Gauss-Hermite

7.5.1 Gauss-Hermite integration

7.5.2 One-dimensional Gauss-Hermite integration

Gauss-Hermite integration approximates a specific integral as follows

$$\int\limits_{-\infty}^{\infty} f(x) \exp(-x^2) \mathrm{d}x \approx \sum_{j=1}^{N} w_j f(x_j)$$

in which w_j and x_j are weights and abscissas for N-point Gauss-Hermite integration, see Abramowitz and Stegun (1972) [[Abramowitz, 1972]]. N-point integration is exact for all polynomials f(x) of degree 2N-1, see Dahlquist and Björck (1974) [[Dahlquist et al., 1974]]. This can for instance be used to approximate the mean of a function F(Y) of a normally distributed random variable Y with mean μ and variance σ^2 :

$$\begin{split} &\int\limits_{-\infty}^{\infty}F(x)\frac{1}{\sqrt{2\pi\sigma}}\exp\left(-\frac{(y-\mu)^2}{2\sigma^2}\right)\mathrm{d}y\\ &=\int\limits_{-\infty}^{\infty}F(\mu+\sqrt{2}\sigma x)\frac{1}{\sqrt{\pi}}\exp(-x^2)\mathrm{d}x\\ &=\frac{1}{\sqrt{\pi}}\sum_{j=1}^Nw_jF(\mu+\sqrt{2}\sigma x_j) \end{split}$$

7.5.3 Two-dimensional Gauss-Hermite integration

One-dimensional Gauss-Hermite integration can readily be extended to two dimensions. The following principal result in two dimensions is more or less given in Jäckel (2005) [[Jäckel, 2005]] for the standard bivariate normal distribution $\phi(x, y; \rho)$ with correlation parameter ρ :

$$\int_{-\infty}^{\infty} \int_{-\infty}^{\infty} F(x,y)\phi(x,y;\rho)\mathrm{d}x\mathrm{d}y \approx \frac{1}{\pi} \sum_{i=1}^{N} \sum_{j=1}^{N} w_i w_j F(\sqrt{2}[ax_i + bx_j], \sqrt{2}[bx_i + ax_j])$$

in which

$$a = \frac{\sqrt{1+\rho} + \sqrt{1-\rho}}{2}$$

and

$$b = \frac{\sqrt{1+\rho} - \sqrt{1-\rho}}{2}$$

as given in Jäckel (2005) [[Jäckel, 2005]].

Jäckel (2005) discusses other Gauss-Hermite approximations to the two-dimensional integral, but found that the approximation given above generally gives the most accurate results. For the general bivariate normal distribution with means (μ_x, μ_y) and variances (σ_x^2, σ_y^2) the integral can be approximated by means of

$$\frac{1}{\pi} \sum_{i=1}^{N} \sum_{j=1}^{N} w_i w_j F(\mu_x + \sigma_x \sqrt{2} [ax_i + bx_j], \mu_y + \sigma_y \sqrt{2} [bx_i + ax_j])$$

The product $w_i w_j$ can be very small, especially when many quadrature points are used, thus wasting possibly precious calculation time. This can be remedied by pruning, i.e. by dropping combinations of (i, j) with very small values of the product $w_i w_j$.

7.5.4 Maximum likelihood for the LNN model with two-dimensional Gauss-Hermite integration

Denote non-consumption on day j for individual i as $Y_{ij} = 0$. The conditional likelihood, i.e. given random effects b_i and v_i , of a non-consumption on day j equals, with H() the inverse of the logit function

$$P(Y_{ij} = 0 | b_i, v_i) = 1 - H(\lambda + v_i).$$

The conditional likelihood of a positive intake $Y_{ij} > 0$ equals, with ϕ the density of the normal distribution

$$f(Y_{ij} = y_{ij} | y_{ij} > 0, b_i, v_i) = H(\lambda + v_i)\phi(y_{ij} - \mu - b_i; 0, \sigma_w^2)$$

The conditional likelihood contribution for individual i is the product of the individual contributions for each day. The marginal likelihood contribution for individual i is obtained by integrating over the possible values of b_i and v_i . Since the pair (b_i, v_i) follows a bivariate normal distribution, the likelihood contribution for individual i can be approximated by means of two-dimensional Gauss-Hermite integration. Individually based covariables, such as sex or age, imply that μ_i and λ_i must be used instead of μ and λ . The likelihood must be optimized by means of some general optimization routine.

CHAPTER EIGHT

COLOPHON



WUR/Biometris, Wageningen University & Research FERA, Food and Environmental Research Agency RIVM, National Institute for Public Health and the Environment

This manual was created for MCRA 9, version 9.1.32, build 9.1.32.0 for Prod-RIVM on Monday April 26, 2021, 16:16:11.

8.1 Contributors to MCRA

MCRA development team: Waldo de Boer, Johannes Kruisselbrink, Marco van Lenthe, Hans van den Heuvel, Hilko van der Voet

Many people contributed to the MCRA code over the years:

Frits van Evert, Jack van Galen, Paul Goedhart, Gerie van der Heijden, Paul Keizer, Marcel Koenders, Jaap Kokorian, Sanne Korzec, Helen Owen, Gerrit Polder, Pim Reijersen, Willem Roelofs, Gert-Jan Swinkels, Jac Thissen.

MCRA has been tested in practice by many people. Feed-back over many years from the team at RIVM is gratefully acknowledged:

Jan Dirk te Biesebeek, Polly Boon, Annick van den Brand, Gerda van Donkersgoed, Corinne Sprong, Trijntje van der Velde-Koerts, Jacob van Klaveren.

Contributors to the MCRA Reference Manual: Waldo de Boer, Paul Goedhart, Andy Hart, Marc Kennedy, Johannes Kruisselbrink, Helen Owen, Willem Roelofs, Hans van den Heuvel, Hilko van der Voet.

WUR/Biometris is the unit for Mathematical and Statistical Methods of Wageningen University & Research P.O. Box 16, 6700 AA Wageningen, Netherlands WUR Campus, Building 107 (Radix), Droevendaalsesteeg 1, 6708 PB Wageningen Telephone: +31 (0)317 476925 https://wur.nl http://www.biometris.nl

Fera Food and Environmental Research Agency

Sand Hutton, York, YO41 1LZ, United Kingdom Telephone: +44 (0)1904 462000 https://www.fera.co.uk

RIVM National Institute for Public Health and the Environment

P.O. Box 1, 3729 BA Bilthoven, Netherlands Antonie van Leeuwenhoeklaan 9, 3721 MA Bilthoven Telephone: +31 30 2749111 https://rivm.nl

BIBLIOGRAPHY

- [Abramowitz, 1972] Milton Abramowitz and Irene A. Stegun. Handbook of mathematical functions. *National Bureau of Standards Applied Mathematics Series*, 55:589–626, 1972.
- [Bopp et al., 2015] Stephanie Bopp, Elisabet Berggren, Aude Kienzler, Sander van der Linden, and Andrew Worth. Scientific methodologies for the assessment of combined effects of chemicals - a survey and literature review. *EUR - Scientific and Technical Research Reports*, 2015. doi:10.2788/093511.
- [Box et al., 1964] George EP Box and David R Cox. An analysis of transformations. *Journal of the Royal Statistical Society: Series B (Methodological)*, 26(2):211–243, 1964.
- [Butler et al., 2018] M.C. Butler Ellis, Marc C. Kennedy, C.J. Kuster, R. Alanis, and C.R. Tuck. Improvements in modelling bystander and resident exposure to pesticide spray drift: investigations into new approaches for characterizing the 'collection efficiency' of the human body. *Annals of work exposures and health*, 62(5):622–632, 2018. doi:10.1093/annweh/wxy017.
- [Béchaux et al., 2013] Camille Béchaux, Mélanie Zetlaoui, Jessica Tressou, Jean-Charles Leblanc, Fanny Héraud, and Amélie Crépet. Identification of pesticide mixtures and connection between combined exposure and diet. *Food and chemical toxicology*, 59:191–198, 2013.
- [Cramer et al., 1976] G.M. Cramer, R.A. Ford, and R.L. Hall. Estimation of toxic hazard—a decision tree approach. *Food and cosmetics toxicology*, 16(3):255–276, 1976. doi:10.1016/S0015-6264(76)80522-6.
- [Dahlquist et al., 1974] G Dahlquist and A Bjorck. Numerical methods (transl. by n. anderson). 1974.
- [de Boer et al., 2009] Waldo J de Boer, Hilko van der Voet, Bas GH Bokkers, Martine I Bakker, and Polly E Boon. Comparison of two models for the estimation of usual intake addressing zero consumption and nonnormality. *Food Additives and Contaminants*, 26(11):1433–1449, 2009.
- [de Boer et al., 2011] Waldo J de Boer and van der Voet. Mcra 7. a web-based program for monte carlo risk assessment. reference manual 2011-12-19, documenting mcra release 7.1. Technical Report, Biometris, Wageningen UR and National Institute for Public Health and the Environment (RIVM), Bilthoven, Wageningen., 2011. URL: https://mcra.rivm.nl.
- [Dodd, 1996] KW Dodd. A technical guide to c-side. Ames, Iowa: Department of Statistics and Center for Agricultural and Rural Development, Iowa State University, 1996.
- [EC, 2018] European Commission Standing Committee on Plants Animals Food and Feed. European commission working document sante-2015-10216 rev. 7. 2018.
- [Efron, 1979] B Efron. Bootstrap methods: another look at the jackknife annals of statistics 7: 1–26. *View Article PubMed/NCBI Google Scholar*, 1979.
- [Efron et al., 1993] Bradley Efron and Robert J Tibshirani. An introduction to the bootstrap chapman & hall. *New York*, 1993.
- [EFSA, 2011a] European Food Safety Authority (EFSA). Report on the development of a food classification and description system for exposure assessment and guidance on its implementation and use. *EFSA Journal*, 9(12):84, 2011. doi:doi:10.2903/j.efsa.2011.2489.
- [EFSA, 2011b] European Food Safety Authority (EFSA). The food classification and description system foodex 2 (draft-revision 1). *EFSA Journal*, pages 438, 2011.

- [EFSA, 2012] European Food Safety Authority (EFSA). Guidance on the use of probabilistic methodology for modelling dietary exposure to pesticide residues. *EFSA Journal*, 10(10):2839, 2012. doi:10.2903/j.efsa.2012.2839.
- [EFSA, 2014] European Food Safety Authority (EFSA). Guidance on the assessment of exposure of operators, workers, residents and bystanders in risk assessment for plant protection products. *EFSA Journal*, 12(10):3874, 2014. doi:10.2903/j.efsa.2014.3874.
- [EFSA, 2017a] European Food Safety Authority (EFSA), Harrie Buist, Peter Craig, Ian Dewhurst, Susanne Hougaard Bennekou, Carsten Kneuer, Kyriaki Machera, Christina Pieper, Daniele Court Marques, Gilles Guillot, Federica Ruffo, and Arianna Chiusolo. Guidance on dermal absorption. EFSA Journal, 15(6):e04873, 2017. doi:10.2903/j.efsa.2017.4873.
- [EFSA, 2017b] EFSA Panel on Contaminants in the Food Chain (CONTAM), Helle Katrine Knutsen, Jan Alexander, Lars Barregård, Margherita Bignami, Beat Brüschweiler, Sandra Ceccatelli, Bruce Cottrill, Michael Dinovi, Lutz Edler, Bettina Grasl-Kraupp, Christer Hogstrand, Laurentius (Ron) Hoogenboom, Carlo Stefano Nebbia, Isabelle P. Oswald, Annette Petersen, Martin Rose, Alain-Claude Roudot, Tanja Schwerdtle, Christiane Vleminckx, Günter Vollmer, Heather Wallace, José Angel Gomez Ruiz, and Marco Binaglia. Risks for human health related to the presence of pyrrolizidine alkaloids in honey, tea, herbal infusions and food supplements. *EFSA Journal*, 7 2017. doi:10.2903/j.efsa.2017.4908.
- [EFSA, 2018] European Food Safety Authority (EFSA), Alba Brancato, Daniela Brocca, Lucien Ferreira, Luna Greco, Samira Jarrah, Renata Leuschner, Paula Medina, Ileana Miron, Alexandre Nougadere, Ragnor Pedersen, Hermine Reich, Miguel Santos, Alois Stanek, Jose Tarazona, Anne Theobald, and Laura Villamar-Bouza. Use of efsa pesticide residue intake model (efsa primo revision 3). *EFSA Journal*, 16(1):e05147, 2018. URL: https://efsa.onlinelibrary.wiley.com/doi/abs/10.2903/j.efsa.2018.5147.
- [EFSA, 2020a] European Food Safety Authority (EFSA), Peter S Craig, Bruno Dujardin, Andy Hart, Antonio F Hernández-Jerez, Susanne Hougaard Bennekou, Carsten Kneuer, Bernadette Ossendorp, Ragnor Pedersen, Gerrit Wolterink, and Luc Mohimont. Cumulative dietary risk characterisation of pesticides that have acute effects on the nervous system. *EFSA Journal*, 18(4):e06087, 2020. URL: https://efsa.onlinelibrary. wiley.com/doi/abs/10.2903/j.efsa.2020.6087.
- [EFSA, 2020b] European Food Safety Authority (EFSA), Peter S Craig, Bruno Dujardin, Andy Hart, Antonio F Hernandez-Jerez, Susanne Hougaard Bennekou, Carsten Kneuer, Bernadette Ossendorp, Ragnor Pedersen, Gerrit Wolterink, and Luc Mohimont. Cumulative dietary risk characterisation of pesticides that have chronic effects on the thyroid. *EFSA Journal*, 18(4):e06088, 2020. URL: https://efsa.onlinelibrary. wiley.com/doi/abs/10.2903/j.efsa.2020.6088.
- [Gillis et al., 2013] Nicolas Gillis and Robert J Plemmons. Sparse nonnegative matrix underapproximation and its application to hyperspectral image analysis. *Linear Algebra and its Applications*, 438(10):3991–4007, 2013.
- [Goedhart et al., 2012] Paul W. Goedhart, Hilko van der Voet, S. Knüppel, Arnold L.M. Dekkers, Kevin W. Dodd, Hermann Boeing, and Jacob D. van Klaveren. A comparison by simulation of different methods to estimate the usual intake distribution for episodically consumed foods. Technical Report, Report: Supporting Publications 2012:EN-299, 2012. URL: http://www.efsa.europa.eu/publications.
- [Goodhardt et al., 1984] Gerald Joseph Goodhardt, Andrew SC Ehrenberg, and Christopher Chatfield. The dirichlet: a comprehensive model of buying behaviour. *Journal of the Royal Statistical Society. Series A (General)*, pages 621–655, 1984.
- [Hoyer, 2004] Patrik O Hoyer. Non-negative matrix factorization with sparseness constraints. *Journal of machine learning research*, 5(Nov):1457–1469, 2004.
- [Jäckel, 2005] Peter Jäckel. A note on multivariate gauss-hermite quadrature. London: ABN-Amro. Re, 2005.
- [Karrer et al., 2019] Karrer, Cecile, Waldo de Boer, Christiaan Delmaar, Yaping Cai, Amélie Crépet, Konrad Hungerbühler, and Natalie van Goetz. Linking probabilistic exposure and pharmacokinetic modeling to assess the cumulative risk from the bisphenols bpa, bps, bpf, and bpaf for europeans. *Environmental science & technology*, 53(15):9181–9191, 2019. doi:10.1021/acs.est.9b01749.
- [Kennedy et al., 2012] Marc C. Kennedy, Clare M.J. Butler Ellis, and Paul C.H. Miller. Bream: a probabilistic bystander and resident exposure assessment model of spray drift from an agricultural boom sprayer. *Computers and electronics in agriculture*, 88:63–71, 2012. doi:10.1016/j.compag.2012.07.004.

- [Kennedy et al., 2015a] Marc C Kennedy, C Richard Glass, Bas Bokkers, Andy DM Hart, Paul Y Hamey, Johannes W Kruisselbrink, Waldo J de Boer, Hilko van der Voet, David G Garthwaite, and Jacob D van Klaveren. A european model and case studies for aggregate exposure assessment of pesticides. *Food and Chemical Toxicology*, 79:32–44, 2015.
- [Kennedy et al., 2015b] Marc C Kennedy, Hilko van der Voet, Victoria J. Roelofs, Willem Roelofs, C. Richard Glass, Waldo J de Boer, Johannes W. Kruisselbrink, and Andy D.M. Hart. New approaches to uncertainty analysis for use in aggregate and cumulative risk assessment of pesticides. *Food and Chemical Toxicology*, 79:54–64, 2015.
- [Kennedy et al., 2017] Marc C. Kennedy and M.C. Butler Ellis. Probabilistic modelling for bystander and resident exposure to pesticides using the browse software. *Biosystems engineering*, 154:105–121, 2017. doi:10.1016/j.biosystemseng.2016.08.012.
- [Kennedy et al., 2020] Marc C. Kennedy, Andy D.M. Hart, Johannes W. Kruisselbrink, Marco van Lenthe, Waldo J. de Boer, Hilko van der Voet, Emiel Rorije, Corinne Sprong, and Jacob van Klaveren. A retain and refine approach to cumulative risk assessment. *Food and Chemical Toxicology*, April 2020. doi:10.1016/j.fct.2020.111223.
- [Kipnis et al., 2009] Victor Kipnis, Douglas Midthune, Dennis W Buckman, Kevin W Dodd, Patricia M Guenther, Susan M Krebs-Smith, Amy F Subar, Janet A Tooze, Raymond J Carroll, and Laurence S Freedman. Modeling data with excess zeros and measurement error: application to evaluating relationships between episodically consumed foods and health outcomes. *Biometrics*, 65(4):1003–1010, 2009.
- [Lee et al., 1999] Daniel D Lee and H Sebastian Seung. Learning the parts of objects by non-negative matrix factorization. *Nature*, 401(6755):788, 1999.
- [Merz et al., 2016] Karl-Heinz Merz and Dieter Schrenk. Interim relative potency factors for the toxicological risk assessment of pyrrolizidine alkaloids in food and herbal medicines. *Toxicology Letters*, 263:44–57, 2016. doi:10.1016/j.toxlet.2016.05.002.
- [Mood et al., 1974] Alexander McFarlane Mood, Franklin A Graybill, and Duane C Boes. *Introduction to the Theory* of Statistics 1974. McGraw-Hill Kogakusha, 1974.
- [Mulder et al., 2015] Patrick P.J. Mulder, Patricia López Sánchez, Anja These, Angelika Preiss-Weigert, and Massimo Castellari. Occurrence of pyrrolizidine alkaloids in food. *EFSA Supporting Publications*, 12(8):859E, 2015. doi:10.2903/sp.efsa.2015.EN-859.
- [Munro et al., 1996] Ian C. Munro, Richard A. Ford, Elke Kennepohl, and James G. Sprenger. Correlation of structural class with no-observed-effect levels: a proposal for establishing a threshold of concern. *Food and Chemical Toxicology*, 34(9):829–867, 1996. doi:10.1016/S0278-6915(96)00049-X.
- [Nusser et al., 1996] Sarah M Nusser, Alicia L Carriquiry, Kevin W Dodd, and Wayen A Fuller. A semiparametric transformation approach to estimating usual daily intake distributions. *Journal of the American Statistical Association*, 91(436):1440–1449, 1996.
- [Nusser et al., 1997] Sarah M Nusser, Wayne A Fuller, Patricia M Guenther, and others. Estimating usual dietary intake distributions: adjusting for measurement error and nonnormality in 24-hour food intake data. Technical Report, Center for Agricultural and Rural Development (CARD) at Iowa State University, 1997.
- [Price et al., 2011] Paul S Price and Xianglu Han. Maximum cumulative ratio (mcr) as a tool for assessing the value of performing a cumulative risk assessment. *International journal of environmental research and public health*, 8(6):2212–2225, 2011.
- [Saul et al., 2002] Lawrence K Saul and Daniel D Lee. Multiplicative updates for classification by mixture models. In Advances in Neural Information Processing Systems, 897–904. 2002.
- [Slob, 2006] Wout Slob. Probabilistic dietary exposure assessment taking into account variability in both amount and frequency of consumption. *Food and Chemical Toxicology*, 44(7):933–951, 2006.
- [Slob et al., 2010] Wout Slob, Waldo J de Boer, and Hilko van der Voet. Can current dietary exposure models handle aggregated intake from different foods? a simulation study for the case of two foods. *Food and chemical toxicology*, 48(1):178–186, 2010.

- [Souverein et al., 2011] Olga W. Souverein, Waldo J. de Boer, Anouk Geelen, Hilko van der Voet, Jeanne H. de Vries, Max Feinberg, and Pieter van't Veer. Uncertainty in intake due to portion size estimation in 24-hour recalls varies between food groups. *The Journal of nutrition*, 141(7):1396–1401, 2011. doi:10.3945/jn.111.139220.
- [Tebby et al., 2019] Cleo Tebby, Hilko van der Voet, Georges de Sousa, Emiel Rorije, Vikas Kumar, Waldo de Boer, Johannes H. Kruiselbrink, Frédéric Y. Bois, Moosa Faniband, Angelo Moretto, and Céline Brochot. Deliverable 6.3 - integration of in silico and in vitro data in pbpk modeling for risk assessment of foodand non-food-borne chemicals using the euromix toolbox. *Zenodo*, October 2019. URL: https://doi.org/ 10.5281/zenodo.3472609.
- [Tooze et al., 2006] Janet A Tooze, Douglas Midthune, Kevin W Dodd, Laurence S Freedman, Susan M Krebs-Smith, Amy F Subar, Patricia M Guenther, Raymond J Carroll, and Victor Kipnis. A new statistical method for estimating the usual intake of episodically consumed foods with application to their distribution. *Journal of the American Dietetic Association*, 106(10):1575–1587, 2006.
- [van den Berg et al., 2016] F. van den Berg, C.M.J. Jacobs, M.C. Butler Ellis, P. Spanoghe, K. Doan Ngoc, and G. Fragkoulis. Modelling exposure of workers, residents and bystanders to vapour of plant protection products after application to crops. *Science of the Total Environment*, 573:1010–1020, 2016. doi:10.1016/j.scitotenv.2016.08.180.
- [van der Voet et al., 2007] Hilko van der Voet and Wout Slob. Integration of probabilistic exposure assessment and probabilistic hazard characterization. *Risk Analysis: An International Journal*, 27(2):351–371, 2007. doi:10.1111/j.1539-6924.2007.00887.x.
- [van der Voet et al., 2009] Hilko van der Voet, Gerie W.A.M. van der Heijden, Peter M.J. Bos, Sieto Bosgra, Polly E. Boon, Stefan D. Muri, and Beat J. Brüschweiler. A model for probabilistic health impact assessment of exposure to food chemicals. *Food and Chemical Toxicology*, 47(12):2926–2940, 2009. doi:10.1016/j.fct.2008.12.027.
- [van Klaveren et al., 2019a] J.D. van Klaveren, J.W. Kruisselbrink, W.J. de Boer, G. van Donkersgoed, J.D. te Biesebeek, M. Sam, and H. van der Voet. Cumulative dietary exposure assessment of pesticides that have acute effects on the nervous system using mcra software. *EFSA Supporting Publications*, 16(9):1708E, 2019. URL: https://efsa.onlinelibrary.wiley.com/doi/abs/10.2903/sp.efsa.2019.EN-1708.
- [van Klaveren et al., 2019b] J.D. van Klaveren, J.W. Kruisselbrink, W.J. de Boer, G. van Donkersgoed, J.D. te Biesebeek, M. Sam, and H. van der Voet. Cumulative dietary exposure assessment of pesticides that have chronic effects on the thyroid using mcra software. *EFSA Supporting Publications*, 16(9):1707E, 2019. URL: https://efsa.onlinelibrary.wiley.com/doi/abs/10.2903/sp.efsa.2019.EN-1707.
- [Verkaik-Kloosterman et al., 2011] Janneke Verkaik-Kloosterman, Kevin W Dodd, Arnold LM Dekkers, Pieter van't Veer, and Marga C Ocké. A three-part, mixed-effects model to estimate the habitual total vitamin d intake distribution from food and dietary supplements in dutch young children. *The Journal of nutrition*, 141(11):2055–2063, 2011.
- [WHO, 2018] World Health Organization (WHO). *Guidance document on evaluating and expressing uncertainty in hazard characterization*. World Health Organization, 2018.
- [Zetlaoui et al., 2011] Mélanie Zetlaoui, Max Feinberg, Philippe Verger, and Stephan Clémençon. Extraction of food consumption systems by nonnegative matrix factorization (nmf) for the assessment of food choices. *Biometrics*, 67(4):1647–1658, 2011.