

# **MCRA Documentation**

*Release 9*

# **Biometris, Wageningen University and Research**

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# **CONTENTS**













Reference and user manual for MCRA 9.1.

## **ABOUT THE TOOLBOX**

<span id="page-10-0"></span>Humans are exposed to a mixture of multiple chemicals via food intake, inhalation and dermal contact. The risk to health that may result from this depends on the effects of different chemicals in the mixture and how they combine.

MCRA 9 is the model and data toolbox developed in the EuroMix project (http://www.euromixproject.eu). It implements methods for exposure, hazard and risk assessment, following guidelines from a.o. the Joint Research Centre (JRC) of the European Commission and the European Food Safety Authority (EFSA). The toolbox should provide computational tools for future risk management decisions on the safety of chemicals in mixtures to be taken by the European Commission and the Codex Alimentarius.

MCRA 9 is a collection of data and models. The system consists of modules that are arranged in eight categories according to a *modular design*. See *Modules overview*.

Each module represents a certain type of data, which can be computed from data provided by other (sub)modules, or the data may be obtained from a dataset selected from the *data repository*. Likewise, each module may be of interest by its own merit, or may just be required as a sub-part of larger calculations. The modular design of the toolbox reveals a netw[ork of data and](#page-10-2) mod[els, and shows how](#page-24-0) data of types and from various sources can be combined in overarching modules. The most overarching module is *health impact estimates*. The toolbox allows the user to start in any of the modules in the modular design for performi[ng calculations.](#page-13-2)

For each module, an *action* can be created to configure and run the module. For data modules, such as the concentrations module, such an action comprises specifying the dataset, specifying the scope (i.e., foods of interest, substances of interest, etc.), and perhaps specifying specific selecti[ons or model settings for](#page-259-1) data manipulations (e.g., imputation of water concentrations in the concentrations module). For calculation modules, when calculating the data of the module based on ot[her dat](#page-15-1)a, configuration of an action comprises specification of the model settings and selection of the calculation inputs, which is data provided by other (sub-)modules. When running an action in the toolbox, the module produces output of its associated data type (which can be used as input for other modules), and a report will be generated of the selected data, the selection and model settings, and the module and all intermediate (i.e., sub-modules) results.

### **1.1 Data and calculation model**

### <span id="page-10-1"></span>**1.1.1 Modular design**

<span id="page-10-2"></span>The modular design distinguishes between three types of modules: primary entity modules, data modules, and calculation modules. For an overview see *Modules*.

- The primary entity modules are data modules determining the scope of the assessments in the toolbox. That is, in each assessment, the scope specifies the *foods*, *substances*, *effects*, *populations*, *responses*, and/or *test systems* that are of interest.
- The data modules give summari[es of the](#page-24-0) available data which depend on (some of) the primary entities. For example *consumptions* data.
- The calculation modules perform calculati[ons on](#page-27-3)i[nput data t](#page-44-1)[o produ](#page-25-1)[ce data on an](#page-41-0)[other type](#page-42-2), as spe[cified by the](#page-46-0) module name. E.g. the *dietary-exposures* calculation module calculates dietary exposures from consumption and occurrence data. Some calculation modules can also act as a data module, in which case the data are directly [specified rathe](#page-48-0)r than calculated. Examples are, the *relative potency factors* module: relative potency

factors can be supplied as such (*Data*) or computed based on hazard characterizations (*Compute*); the *single value consumptions* module: Large Portions can be supplied as such (*Data*) or computed based on consumption distribution data of a population (*Compute*).

### **1.1.2 [Nominal run](#page-56-0) and uncertainty analysis**

Within the toolbox two types of simulation runs are distinguished: the nominal run and the uncertainty analysis loop.

<span id="page-11-0"></span>The nominal run represents a single simulation which is aimed to compute the most likely, unbiased estimates for the specified model. E.g., when a *dietary exposure assessment* is requested, in the nominal run a single exposure distribution is estimated using nominal values for all data and parameters.

In the *uncertainty analysis* loop, each simulation run is repeated a large number of times. Each run starts with a different scenario using data obtained with bootstrapping, parametric resampling and/or re-calculation of uncertain values. As a result, a large number [of uncertain dietary exposure](#page-116-1) distributions is estimated which are used to estimate uncertainty limits (p5, p95).

Runni[ng a nominal run firs](#page-12-0)t has the advantage that the user may evaluate these modelling results before doing the final analysis. The model specification of rather complex simulation models and the corresponding output results are evaluated to detect any errors or misspecifications. Possible errors in the data and/or model settings are identified and corrected. In the final exposure assessment, the uncertainty analysis is included and the uncertainty of estimates is assessed.

### **1.1.3 Retain & Refine and tiered approaches**

<span id="page-11-1"></span>A basic idea of Retain & Refine is that entities (e.g., substances) can be handled in different ways (more or less refined) while still being considered together in the same risk assessment (retain). We refer to such different approaches as tiers.

In the modular design, a tier is defined as a specific set of settings for a module or a group of modules. Tiers can differ in many respects, and there is no single dimension to rank tiers as low vs. high. In risk assessment, typical tiers contrast deterministic to probabilistic approaches, conservative to realistic approaches, approaches using restricted data to approaches using more extensive data, and approaches using different degrees of model complexity. For each of the modules of the toolbox, as many tiers are implemented as considered useful for the practice of risk assessment.

Each calculation in the modular design may involve multiple, nested, calculations of sub-modules. A *risk* (or health impact) assessment builds on an *exposure assessment* and a *hazard assessment*, the exposure assessment builds on a *dietary* and a *non-dietary exposure* assessment, the dietary exposure assessment builds on a *consumption assessment* and an *occurrence assessment*, etc. Tiers can be defined at each node of the assessment network. An example consists of the tiers *'IESTI'*, *'EFSA basic optimistic'* and *'EFSA basic pessimistic'* which are defined at the lev[el of](#page-259-0) a dietary exposure assessment, but includ[e the settings for th](#page-115-1)e corr[esponding tiers at](#page-202-0) the level of the concentration model [calculat](#page-116-1)or.

Each c[alculator has as a main](#page-59-1) output entities that can be specified to have different tiers (tiered entities). For example, in a *hazard [assessm](#page-197-0)ent*[, some substances m](#page-73-0)ay be [assessed using a tier 'H](#page-75-2)azard Dose from dose-response data', other substances may be assessed using a tier 'TTCx100' or 'sample from general NOAEL distribution x100' (which only requires knowledge of the Cramer class of the substance). As another example, in dietary exposure assessment some food-substance combinations may be recognised as risk drivers for which a more complex approach (e.g. probabilistic mo[delling\) is required,](#page-202-0) whereas a simpler approach (e.g. *deterministic modelling*) may be sufficient for all other foodsubstance combinations. So in this case the tiered entity is 'food-substance'. A typical risk assessment will start at a tier that is simple to perform for all tiered entities (potential risk drivers). Note that, based on data availability and ease of application, the initial assessment can already include more complex elements, such as probabilistic modelling. If the initial calculations produce risk estimates that do not exc[lude concern, refinemen](#page-194-3)t of the modelling for the perceived risk drivers is useful for checking whether this concern is real.

### **1.1.4 Uncertainty**

<span id="page-12-0"></span>Uncertainties may arise in different forms in many of the models and data of the toolbox. One may encounter uncertainty in the data values (e.g., uncertain NOAELs, uncertain RPFs, or uncertain processing factors), uncertainty due to limited data (e.g., a limited number of food samples), uncertainty due to a lack of data (e.g., missing concentration data for some foods/substances or missing processing factors), and uncertainty of the models, (e.g., due to a lack of detail). In many situations it is desirable to analyse how the model outcomes vary for the different scenarios that uncertainties give rise to. For this, the toolbox offers:

- 1) for many types of data, the possibility to provide data including quantifications of uncertainty for many types of data,
- 2) imputation methods for filling in missing data in various types of models, and
- 3) a generic uncertainty analysis method that providing uncertainty estimates of the modelling results for many of the modules, which are based on bootstrapping, parametric resampling, and/or re-calculation on all submodules for which this is possible.

### **Uncertainty due to limited sampled data**

<span id="page-12-1"></span>For some type of data, e.g., processing factors, it may be that in some cases it is possible to not only provide nominal estimates of the data values, but also to provide quantified estimates of the uncertainties of these values. In other cases, it may happen that quantifications of the uncertainties of these estimates are not available. In the toolbox, the aim is to provide the possibility to work with both quantified and unquantified uncertainties. That is, include quantified uncertainties in a quantitative uncertainty analysis when available, or to ignore their absence and only use the nominal estimates, perhaps in combination with an offline qualitative uncertainty analysis.

Uncertainties of the data values may be expressed in different forms, and it depends on the type of data which forms are available, suitable, and implemented in the toolbox. For some data values, uncertainty may be quantified by means of parametric distribution parameters (e.g., *processing factor uncertainties*, or kinetic model instance parameter uncertainties). Alternatively, uncertainty values may be provided in the form of an empirical set of uncertainty values (e.g., *relative potency factor uncertainties*, or *points of departure uncertainties*).

Whenever data include quantified uncertainties, and the data module to which they belong is included as a sub-module of a calculation module. These uncertainties ma[y be chosen to be included in](#page-106-0) an uncertainty analysis of the main mod[ule, and if this is so, the data values](#page-237-2) ar[e resampled in each](#page-234-3) *uncertainty analysis cycle* based on the uncertainty quantifications.

The basic *acute exposure* distribution is estimated in a Monte Carlo simulation by combining dietary consumption records (person-days) with sampled residue values. The resulting distribution represents a combination of variability in consumption within the population and between residues [in a food lot. Percentiles](#page-11-0) may be used for further quantification e.g. the median or 99th percentile. Due to the limited size of the underlying data, these outcomes are uncert[ain. Confidenc](#page-117-0)e (or uncertainty) intervals reflect the uncertainty of these estimates, where MCRA uses bootstrap methodology and/or, depending on the available data, parametric methods to estimate the uncertainty.

### **Empirical method, resampling**

The empirical bootstrap is an approach to estimate the accuracy of an outcome. In its most simple, non-parametric form, the bootstrap algorithm resamples a dataset of n observations to obtain a *bootstrap sample* or *resampled set* of again n observations (sampling with replacement, that is: each observation has a probability of  $1/n$  to be selected at any position in the new resampled set). By repeating this process  $B$  times, one can obtain  $B$  resampled sets, which may be considered as alternative data sets that might have been obtained during sampling from the population of interest. Any statistic that can be calculated from the original dataset (e.g. the median, the standard deviation, the 99th percentile, etc.) can also be calculated from each of the resampled sets. This generates a *uncertainty distribution* for the statistic under consideration. The uncertainty distribution characterises the uncertainty of the inference due to the sampling uncertainty of the original dataset: it shows which statistics could have been obtained if random sampling from the population would have generated another sample than the one actually observed [Efron, 1979], [Efron et al., 1993].

### **Parametric methods**

Instead of bootstrapping the observed data, inference about parameters is based on parametric methods. For processing, where factors are specified through a nominal and/or upper value this is the natural choice. For concentration data, where the lognormal model is used to represent less conservative scenario's (EFSA, 2012) [EFSA, 2012], the *parametric bootstrap* may be an alternative, especially when data are limited and the empirical bootstrap fails.

According to Cochran's theorem, sample variance  $\hat{\sigma}_y^2$  follows a scaled chi-square distribution. In the parametric bootstrap for the *lognormal* distribution, the sample variance  $\hat{\sigma}_y^2$  is replaced by a random draw from a chi-square distribution with  $n_1 - 1$  degrees of freedom; the sample mean  $\hat{\mu}_y$  is replaced by a random dr[aw from a no](#page-355-0)rmal [distribution with para](#page-75-1)meters  $\hat{\mu}_y$  and  $\hat{\sigma}_y^{*2}/n_1$ , giving a new set of parameters  $\hat{\mu}_y$  and  $\hat{\sigma}_y^{*2}$ . This is repeated B times.

For the *truncated [lognormal](#page-66-0)* and *censored lognormal*, large sample maximum likelihood theory is used to derive new parameters  $\hat{\mu}_y$  and  $\hat{\sigma}_y^{*2}$ . This is repeated B times.

The binomial fraction of non-detects for the *mixture lognormal* and *mixture truncated* distribution is sampled using the beta distribution with uniform priors  $a = b = 1$  (with the *beta* distribution as the empirical Bayes estimator for the bin[omial distribution\). T](#page-67-0)his [is repeated](#page-66-1)  $B$  times.

### **Uncertainty due to missing data**

<span id="page-13-0"></span>In some cases, it may be that data as only available for specific (primary) entities and missing for others. E.g., points of departure (such as NOAELs or BMDs) may only be available for some of the substances of interest.

### **Uncertainty due to modelling approach**

<span id="page-13-1"></span>There is also uncertainty of model outcomes that may arise by conducting different modelling approaches or applying alternative modelling assumptions.

**Note:** TODO

# **1.2 Data repository**

<span id="page-13-2"></span>Figure 1.1 shows the toolbox data repository browser. The data repository enables users to upload and organise their own datasets and to share these with other users. The data sources available in the data repository can be used directly as data sources for *modelling actions*. Each user has their own repository Ŗ and is free to upload data files and to organise files into folders and sub-folders. Users may be granted access to one or more shared repositories: shared, [maintained](#page-14-1), and used by multiple users. Shared repositories and their contents are free to use by granted users in their own calculations.

The central panel [of the repository](#page-15-1) browser shows the data sources and sub-folders of the currently opened folder/repository. The top bar of the repository browser shows the path of the currently opened repository, buttons to collapse/expand the repository folder tree-view sidebar on the left  $\equiv$  and the info-sidebar on the right  $\bullet$ , and a button to open the action menu  $\ddot{\cdot}$ . The tree-view sidebar shows the hierarchical structure of the repositories and sub-repositories to which the user has access. The info-panel shows the details of the selected data source or folder. If the selected item is a data source, then the info panel shows the types of data available in the data source and the different data source versions of the data source. If the selected item is a folder, then the info panel shows info about the owner of the repository, the *access level* of the user, and info about the other users and user groups that have access to this repository.

Users with read-write access (or higher) may upload new data source files by pressing the add button  $+$  on the bottom right and selecting the *upload new file(s)* item. A new sub-repository can be created by pressing the same add button and selecting the *create new folder* [item. A t](#page-14-0)hird option is to create an external Proast link, which can be seen as a data source repository folder in which the data sources link to datasets (outputs) available on Proast web.

<span id="page-14-1"></span>

<b>MCRA 9 - EuroMix toolbox</b> G <b>Exposure, Hazard &amp; Risk Assessment</b>						<b>the Second Exercise</b> ⊞ $\bullet$ Eı
$\equiv$ Data / EuroMix / Dose-response data						0 ∶
Folders	Name $\blacktriangledown$	Version Date		Uploader		BfR-HepG2-RGA-Mixtures-2.xlsx
$\Box$ kruisselbrink ≺	个 (EuroMix)					
Acropolis $\overline{\phantom{0}}$	BfR-HepaRG-AdipoRed-Mixtures.xlsx	$\mathfrak{D}$	25-06-2019 10:20	kruisselbrink		Data groups
<b>EuroMix</b> $\checkmark$ Combined datasets	BfR-HepaRG-AdipoRed-Single.xlsx	2	25-06-2019 10:20	kruisselbrink	:	Dose response data
<b>Concentrations</b>	BfR-HepG2-RGA-Mixtures-2.xlsx	3	25-06-2019 10:19	kruisselbrink	з	Versions
Consumptions	HepaRG-AdipoRed-one-exp-five-subst-for training no summary.xlsx 1		18-03-2019 16:23	kruisselbrink	:	BfR-HepG2-RGA-Mixtures-2.xl R v1 (28/02/2019 04:02 kruisselbrink)
Dose-response data Effects and AOP networks	HepaRG-AdipoRed-one-exp-five-subst-for training.xlsx		18-03-2019 16:23	kruisselbrink		BfR-HepG2-RGA-Mixtures-2.xl v2 (25/06/2019 10:06 kruisselbrink)
Foods and food translations	HepaRG-AdipoRed-one-exp--two-subst-for training no summary.xlsx 1		18-03-2019 16:23	kruisselbrink	÷	BfR-HepG2-RGA-Mixtures-2.xl v3 (25/06/2019 10:06 kruisselbrink)
Hazard data HumanMonitoring	HepaRG-AdipoRed-one-exp--two-subst-for training.xlsx		18-03-2019 16:23	kruisselbrink	÷	
In-silico data	RIVM-EST-CardioDiff-Mixtures.xlsx	$\overline{2}$	25-06-2019 10:19	kruisselbrink		
Kinetic models	UGent-HepaRG-Mitochondria-Mixtures.xlsx	$\overline{2}$	25-06-2019 10:20	kruisselbrink		
Non-dietary exposures						
<b>Processing</b>						
Substances						
Test-systems and responses						

Figure 1.1: The toolbox data repository browser.

### **1.2.1 Repository access levels**

Shares and access rights can be granted on the level of repositories and sub-folders. Data sources inherit the access rights of the repository/folder in which these are located. The following access rights are available:

- <span id="page-14-0"></span>• **visible:** the user can only see that the repository exists, but cannot see its contents, except for sub-folders that may also visible to the user.
- **use:** the user is only allowed to use the data sources in this repository, but is **not** allowed to download the original data of the data sources of the repository.
- **read:** the user can use data sources in this repository **and** is allowed to download the original data files of the data sources of the repository.
- **read/write:** the user can use and download data sources in this repository and is allowed to add/remove files and folders to/from this repository.
- **admin:** the is considered as an administrator of this repository and has full control over it, including the rights to add/remove files and folders to/from this repository and to add/remove user and group shares.
- **owner:** the user is considered to be the owner of this repository and therefore has full control over it.

Users with administrator or owner rights on a repository/folder are allowed to add/remove user and group access using the *edit shares dialog* (Figure 1.2) that can be opened by pressing the *edit shares* button  $\leq$ .



Figure 1.2: The edit-shares dialog of the toolbox data repository browser: user and group access rights are added and removed by repository owners and administrators.

### **1.2.2 Linking remote data repositories**

<span id="page-15-0"></span>The toolbox also offers to link external data repositories  $\bigoplus$ . These are remote websites not part of the toolbox, but containing data sources that can be used for calculations. Currently, only one remote source can be linked as external repository in the toolbox, the PROASTweb (https://proastweb.rivm.nl/). PROASTweb users may link directly the outputs of their PROAST analyses (i.e., dose response models) as an external repository to the toolbox.

Figure 1.3 shows how PROAST outputs of a PROASTweb user are linked to an external repository in the toolbox. Data sources of remote repositories have to be explicitly imported in the toolbox before they can be used in analyses. Initially, all data sources in a remote reposito[ry have a status of not-impo](https://proastweb.rivm.nl/)rted  $\bigcirc$ . Pressing the import button  $\bullet$ , the toolbox will attempt to import the data source and once that is finished, the data source is ready to be used in analyses.

[A new PR](#page-16-0)OAST remote repository link is created by pressing the add button  $+$  on the bottom right and selecting the *Create Proast link* option. A dialog (Figure 1.4) opens asking for the local name of the external repository/folder, the PROASTweb username of the user of which the outputs should be linked, and the PROASTweb access key of the user, which is required as authentication token to access the analyses of the specified user.

### **1.3 Workspaces and [actions](#page-16-1)**

<span id="page-15-1"></span>User work is organized in workspaces. A workspace is a collection of work items that are logically grouped together. A workspace has a name, description and, optionally, a number of tags. Workspaces may be shared with other users. Users are the owners of their own workspace folders and possible subfolders.

Actions are configurations of the modules of the modular design. Each action is of a certain action type, which specifies the particular module for which this action is a configuration. An action can be available in two forms: 1) a data selection action and 2) a calculation action. A data selection action comprises the selection of already available data of that action type and specification of (subset) selections on that data. A calculation action is an action in which the data of that action is calculated based on relevant input and specific calculator settings. Within a workspace, multiple actions can be created.

When running an action, a task is spawned that produces output. Output is available in the form of reports or in the form of data that can be used as input in other actions. Actions have multiple outputs when settings are changed.

<span id="page-16-0"></span>

Figure 1.3: The remote (PROASTweb) repository in the toolbox data repository browser.

<span id="page-16-1"></span>

Figure 1.4: The dialog for creating a new PROASTweb remote repository link.

Output reports are presented as screen reports or print reports. Output reports are composed of one or multiple sections.

### **1.3.1 Workspace browser**

<span id="page-17-0"></span>Figure 1.5 shows the workspace browser. Users scroll through their workspaces and select the workspace which they want to work with. Detailed information about the selected item in the browser is shown in the info panel, which can be expanded/collapsed using the info button  $\bullet$  on the right of the toolbar. The *filter text box*  $\alpha$  is used to quickly find/filter workspaces by name or tag. A workspace is opened by clicking on the workspace name or selecting the *[open work](#page-17-2)space*  $\bullet$  option of the *action menu*  $\colon$  of the workspace. Opening a workspace will redirect you to the *workspace overview page*.

A new workspace is added by pressing the add button  $+$  on the bottom right of the screen. Delete workspaces by opening the *action menu*  $\vdots$  of the workspace item in the browser and selecting the delete  $\ddot{\bullet}$  option.

<span id="page-17-2"></span>

MCRA 9 - EuroMix toolbox / E EuroMix - Exposure G <b>Exposure, Hazard &amp; Risk Assessment</b> workspace	<b>the Second Second</b> Seconds & ⊞ Eīl	2			
Workspaces		Order by		$\bullet$ Q euromix $\times$	
Name $\rightarrow$	Created	Last modified	Tags		
EuroMix - CAG memberships calculations	01-10-2018 15:51	08-07-2019 14:42	EuroMix		
EuroMix - Calculations case study R&R	15-12-2018 10:31	08-07-2019 14:37	EuroMix		
EuroMix - Dose response models	14-03-2019 09:11	08-07-2019 14:40	EuroMix		
EuroMix - Effect representations	01-05-2018 12:11	08-07-2019 14:40	EuroMix		
EuroMix - Examples hazard characterisations	23-08-2018 10:01	08-07-2019 16:07	EuroMix		
EuroMix - Examples PROAST	03-04-2018 10:41	08-07-2019 16:41	EuroMix	PROAST	
EuroMix - Exposure mixtures calculations	23-04-2019 15:21	08-07-2019 16:42	EuroMix	Mixtures	
EuroMix - Hazard characterisation calculations	18-09-2018 14:51	08-07-2019 14:36	EuroMix		
EuroMix - Hazard characterisation tests	13-07-2018 14:41	08-07-2019 14:41	EuroMix		
EuroMix - Human monitoring example	11-11-2018 14:41	08-07-2019 15:54	EuroMix		
FuroMix - Ivive	28-04-2019 11:41	08-07-2019 16:14	EuroMix	Risk <b>IVIVE</b>	
EuroMix - Kinetic model calculations	07-01-2019 09:41	08-07-2019 16:40	EuroMix	Kinetic-models PBPK	
EuroMix - RPF calculation scenarios	17-07-2018 15:51	08-07-2019 16:43	EuroMix	$\pm$ <b>RPF</b>	
EuroMix - Target exposure assessments	30-04-2019 16:31	08-07-2019 15:54	EuroMix		

Figure 1.5: The workspace browser.

### **1.3.2 Workspace overview page**

<span id="page-17-1"></span>Figure 1.6 shows the workspace overview page. This page provides an overview of the actions, data, tasks, and results of a workspace, shown as four tabs at the top of the page. The actions tab shows all actions of the workspace, and from this tab, actions are opened. The data tab shows all data sources used in this workspace. I.e., all data sources that are used by the actions of the workspace. The results tab shows all tasks and results of simulation jobs that have [been subm](#page-18-1)itted by the actions of the workspace. The properties tab shows the general information of the workspace (i.e., name, descriptions, and tags) and edit functionality.

In the actions tab, all actions of the workspace are listed. The list of actions can be filtered by action type or by filter text using the controls on the toolbar. An action is opened by clicking on the action name or by selecting the *open action* option of the action menu <sup>●</sup> of the selected action item. Opening a workspace will redirect you to the *action details pages*. A new action is added to the workspace by pressing the *add button*  $+$  at the bottom right of the page.

<span id="page-18-1"></span>

<b>Exposure, Hazard &amp; Risk Assessment</b>	MCRA 9 - EuroMix toolbox / E EuroMix - User gro workspace				<b>the Secondary Seconds</b> ₩ ET	2
<b>Actions</b>	三 Data	<b>IL</b> Results	$\blacksquare$ Properties			
Workspace actions				(44 selected)	▼ Q Type filter text here	A
Name $\blacktriangledown$		Type	Created	Last modified	Tags	
Aggregate exposure assessment		<b>Exposures</b>	07-05-2018 12:45	22-06-2018 13:57	target-exposures aggregate	÷
Dietary exposure assessment		<b>Dietary exposures</b>	07-05-2018 16:56	08-05-2018 08:58	dietary	i
<u></u>	Example hazard characterisation calculation	<b>Hazard characterisations</b>	07-05-2018 14:08	08-06-2018 14:43	example target euromix	
Example target exposures calculation		<b>Exposures</b>	07-05-2018 17:06	08-06-2018 16:56	target-exposures	
Relative potency factors		<b>Relative potency factors</b>	07-05-2018 16:04	07-05-2018 16:41	rpf	i
Relative potency factors from data		<b>Relative potency factors</b>	07-05-2018 16:45	07-05-2018 16:53	rpf	i
<b>Risk assessment example</b>		<b>Risks</b>	08-05-2018 09:19	08-05-2018 09:44	Risk	ł
						$\pm$

Figure 1.6: The workspace overview page.

### **1.3.3 Action page**

<span id="page-18-0"></span>After opening an action, the user is directed to the main panel of the action. Each action has its own specific panel. In the main action page and sub-action pages, an action is configured, simulation jobs started, and output results are evaluated. The panel in Figure 1.7 shows the following sections:

- **Scope:** Links to the scope-panels in which the scope entities of the action are set (e.g., foods or substances).
- **Inputs:** Links are shown for panels in which the calculation inputs or selection inputs are set (e.g., concentration models that are i[nputs for co](#page-19-1)mputing dietary exposures).
- **Data source:** If the action is a data action, then a form is shown in which the data source should be specified (e.g., selection of the concentration data source in a concentrations action).
- **Settings:** A form is shown in which the calculation and/or selection settings of the action are set/changed (e.g., specify the exposure type, chronic/acute, of an exposure assessment).

All modules of the toolbox have equally structured panels. In each panel, data sources and settings for the action are specified and the scope and input sub-module links that are relevant are shown. This presentation reflects the modular design and allows the user to select the data and settings required for running the action. In the summary panel ■ the main settings and data of the action are summarized. The output settings panel is used to specify general output settings. In the uncertainty settings panel  $\pm$  the number of uncertainty runs and uncertainty sources is specified. In the results panel  $\bullet$  running tasks and output results of the actions are shown. An alternative form of navigating from action to sub-action is provided by the navigation menu in the left sidebar that can be expanded/collapsed by clicking the menu button on the top left in the Action bar. In this menu, all required modules for the action are shown in one list, allowing a linear way of navigation.

An action is valid and ready to run when all scopes and inputs are valid and all required data and settings are configured. For each sub-action, the check symbol  $\checkmark$  indicates that it has been configured correctly and is ready to run. In case a sub-action has a warning symbol  $\blacktriangle$ , some user action is required. When the main action is ready to run, a simulation job is started by clicking the run button  $\blacktriangleright$  in the grey action bar on the top right. Optionally, sub-actions can be started by clicking the run button  $\triangleright$  in the green (sub)action bar on the top right. Clicking the run button will send the simulation task of this (sub)action to the job-scheduler, and the progress of the task is shown in the results panel ʸ. After completing the task, output is available in the form of a screen report, download as pdf, or download of tables in csv format.

<span id="page-19-1"></span>

Figure 1.7: The main page of an action.

#### **Scoping: entity selection**

<span id="page-19-0"></span>Each action starts with the selection of the relevant primary entities. In this context, entity selection or scoping plays an important role. Scoping of the action is defining the members for its primary entities, and, occasionally, also for other entities.

As an example, Figure 1.8 shows the substances module panel. At the top, the data source file with substances is selected containing the primary entity data of substance codes. In the selection card, a selection is made of the entities in the dataset that are relevant for the current action (3 in scope). Note that if no explicit selection is made, the scope is set to all entities by default. In the settings form, additional (selection) settings are shown, e.g., selection of the index sub[stance \(rele](#page-20-0)vant for a cumulative assessment). In this way, the scope of the action is specified by selection of the primary entities.

The panels for the data modules have a similar structure and selection is essentially the same. The only difference is that data actions always have a scope. I.e., data modules always relate to one or more primary entities.

#### **Implicit versus explicit scoping**

MCRA distinguishes between implicit and explicit selection of entities (scoping). By default, the selection is defined implicitly as 'all entities' found in all data are linked to the action. For instance, the substance scope will contain all substance codes found. That is, not only substances as specified in the substance data source, but also all other substances found in data sources that link to substances like concentration sample data or points of departure data. These are implicit selections. Explicit selections are made in the specific module panel of this data type (e.g., by selecting the substances in the substances panel). Once made explicit, selections are no longer automatically expanded when new data sources are linked to the action.

For example, the substances scope shown in Figure 1.8 is defined explicitly, having three substances in the scope, and excluding 1626 substances also present provided through substances data source and/or other linked data sources like

<span id="page-20-0"></span>

Figure 1.8: The substances module panel as an example of a primary entity module panel.

concentration samples. By pressing the *clear filter* button, the explicit scope is cleared and is made implicit again. Then, the scope contains all substances found as primary entities and found in all linked data sources, in total 1629  $(1626 + 3)$  substances.

### **Comparing new data to set scopes**

After linking a data source to an action, MCRA performs a check whether the new data links well to the current scope (selected entities) of the action and reports the results. For instance, after linking new substance concentration data to an action which already has an implicit or explicit substance scope, it should be checked whether the substance codes used in the concentration data match with the current substances in scope. Note that this check is also performed after linking a primary entity substances data source to an action which already has a set of substances in scope, i.c. substances already specified in other selected data sources.

After linking a data table from a new data source to an action which already has a defined scope for one of the entities in the table, there are three possible states for entity codes:

- codes included in both the scope and the data source
- codes included in the scope, but not present in the data source
- codes included in the data source, but not present in the scope

The first case represents a successful link, no further action is required. For the second and third type of mismatch, it depends on the type of data link whether this is considered a serious problem (red flag  $\triangle$ ) or merely a point of attention (green flag  $\triangle$ ). For instance, in the case of concentration data, for some substances no concentrations are available, and therefore MCRA allows missing concentration data for part of the substances in the scope: a green warning symbol is shown. The concentration data source may equally well contain codes that are not in the scope (e.g., concentrations for substances that are not specified in the primary entity data for substances). It may be desirable to extend the scope with these substances found in the concentration data. Also this situation is flagged with a green warning symbol.

Figure 1.9 shows an example of a point of departure action. The substances scope has already been defined by other data in the action (in this case points of departure data), and subsequently a substances data source is selected. Here, there are 140 substances in the current scope (explicitly defined). However, 132 of these 140 substances are not present in the substances data source (*not in table*). Hence, we are missing the definitions of these substances. This is considered a critical linking issue that should be solved by updating the substances data source to include these substances, therefore a red warning symbol is shown. On the other hand, the substances data source also contains 3 substances that are not part of the current scope (*only in table*). This is a non-critical error, normally leading to a green warning symbol, but in this case, it is overruled by the red warning symbol.

<span id="page-21-0"></span>

Figure 1.9: Checking substances data in a substances data source against an already set substances scope.

Another example is shown in Figure 1.10. The primary entities effects and substances are selected and in the scope. Then, a points of departure data source is selected containing effect and substance codes. For effects, no linking errors are observed, hence the new data source matches perfectly with the effects already in scope. For substances, we see that there are 7 substances that are in the points of departure data source but not in the substances scope (*new*) and for 3 substances in the sc[ope no poin](#page-22-0)ts of departure are available (*not in table*). The former is fine, but it might be needed to extend the scope with these 7 substances (*add to scope*). The latter, in general, is not a problem but just a point of consideration. These substances might be removed from the scope (*remove from scope*) or not.

<span id="page-22-0"></span>

Points of departure	be168c0b		
Scope Effects (1 in scope)			A
Substances (140 in scope)			A
Points of departure data source CAG_steatose_PESTICIDES_april 2017.mdb $\checkmark$		✓◢	
$\vee$ Hazard doses:			
Effects: no linking errors			
Substances: 7 new (add to scope)   3 not in table (remove from scope)			

Figure 1.10: Checking substances data in a POD data source against an already set substances scope.

# **MODULES**

<span id="page-24-0"></span>MCRA is a modular system. The diagram of Figure 2.1 shows the modules and their relations. Each module is associated with its own type of data, and is linked to one or more other modules. Note that not all details can be fully shown in the scheme, for details consult the table below, which specifies all relations between the modules in MCRA.



Figure 2.1: Diagram of the modular design of MCRA.

# **2.1 Primary entity modules**

<span id="page-25-0"></span>The MCRA modular system is based on six primary entities, defining who (*Populations*) is to be protected against what impact (*Effects*) caused by what agent (*Substances*) originating from where (*Foods*), with an indication how the effects are quantified (*Responses* in *Test systems*).

### **2.1.1 Eff[ects](#page-25-1)**

Effects are biological [or toxicolo](#page-42-2)gi[cal conseque](#page-46-0)nces for human health, that may result from chemical exposure and are the focus of hazard or risk assessment.

<span id="page-25-1"></span>Output of this module is used by: *Concentration models High exposure food-substance combinations Dietary exposures Exposure mixtures QSAR membership models Molecular docking models Active substances Relative potency factors Hazard characterisations Points of departure Effect representations Inter-species conversions Intra species factors AOP networks Risks Single value risks*

### **[Effects data formats](#page-221-1)**

### **[Effects](#page-206-2)**

<span id="page-25-2"></span>Effects are primary entities of the data model. Health effects are defined as (critical) changes relative to a treatment or exposure.

### **Effects**

Effects are uniquely identified by a code (idEffect). Optionally, a name and description can be added. Health effects are commonly distinguished in two types, acute and chronic. Further properties may be specified, e.g. in relation to decision schemes such as the use of thresholds of toxicological concern (TTCs).

<b>Name</b>	<b>Type</b>	Description	<b>Aliases</b>	Required
idEffect	AlphaNumeric $(50)$	Unique identification code of	idEffect,	Yes
		the effect.	EffectId, Code-	
			FocalEffect, Id,	
			Code, KeyEvent,	
			idKeyEvent	
CodeSystem	AlphaNumeric(100)	Identifier of the coding	CodeSystem	No
		system of the effect code.		
Name	AlphaNumeric(100)	Name of the effect.	Name	N <sub>o</sub>
Description	AlphaNumeric(200)	Additional description or label	Description	$\overline{No}$
		of the effect.		
Biological-	AlphaNumeric(100)	Biological organisation of the	Biological-	$\overline{No}$
Organisation		effect: Molecular, Cellular,	Organisation	
		Tissue, Organ, Individual.		
		This is in line with AOP wiki		
		terminology and can be used		
		for grouping.		
KeyEvent-	AlphaNumeric(100)	Description of AOP Key	Process	$\overline{No}$
Process		event component process.		
		E.g., receptor signalling.		
KeyEvent-	AlphaNumeric(100)	Description of AOP Key	Object	$\overline{No}$
Object		event component object. E.g.,		
		PPAR-alpha.		
KeyEvent-	AlphaNumeric(100)	Description of AOP Key	Action	$\overline{No}$
Action		event component action. E.g.,		
		decreased.		
<b>KeyEventOrgan</b>	AlphaNumeric(100)	Description of AOP Key	Organ	$\overline{No}$
		event organ. E.g., liver.		
<b>KeyEventCell</b>	AlphaNumberic(100)	Description of AOP Key	$\overline{\text{Cell}}$	$\overline{No}$
		event organ. E.g., hepatocyte.		
<b>AOPwikiKE</b>	AlphaNumeric(200)	Key event ID number in AOP	AOPWikiIds,	$\overline{No}$
		wiki	<b>AOPwikiKE</b>	
		https://aopwiki.org/events		
		Several ID possible Some		
		effects might not be in the		
		wiki, and this field will be		
		empty.		
Reference	AlphaNumeric(200)	External reference(s) to	References	$\overline{No}$
		sources containing more		
		information about the AOP		
		key event. E.g., the AOP		
		wiki, and the associated AOP		
		wiki Ids.		

Table 2.1: Table definition for Effects.

Table aliases: Effects, Effect, KeyEvents, KeyEvent.

#### **Effects calculation**

#### NOTE CHANGE THIS WHEN READY

<span id="page-27-0"></span>Option *Multiple effects analysis* selects multiple selects. Effects are selected using **Effects selection**. Press **Clear filter** and/or **change selection** and select multiple effects in the scroll down box. If both this option and *Include related effects of AOP network* is unchecked, it is obligatory to select one and only one effect. If *Include related effects of AOP network* is checked, a **Focal effect** is specified and all related effects in the *AOP network* are selected as well.

### **[Effects settings](#page-27-5)**

### <span id="page-27-1"></span>**[Selection set](#page-27-5)tings**

<span id="page-27-5"></span>



#### **Effects as data**

Effect definitions are provided as lists/catalogues of effect definitions.

- <span id="page-27-2"></span>• *Effects data formats*
- *Effects calculation*

### **2.1.2 [Foods](#page-25-2)**

<span id="page-27-3"></span>Foods [are uniquely define](#page-27-0)d sources of dietary exposure to chemical substances. Foods may refer to 1) foods as eaten, foods as coded in food consumption data (e.g. pizza); 2) foods as measured, foods as coded in concentration data (e.g. wheat, tomato); 3) any other type of food (e.g. ingredients like flour, tomato sauce).

Output of this module is used by: *Consumptions Single value consumptions Market shares Food recipes Concentrations Concentration distributions Single value concentrations Processing factors Unit variability factors Occurrence patterns Occurrence frequencies Substance authorisations Deterministic substance conversion factors Concentration limits Concentration models Modelled foods Focal food concentrations Total diet study sample compositions Food extrapolations Food conversions Consumptions b[y modelled food](#page-48-0) [High exposure food-subst](#page-56-0)[ance combinati](#page-54-1)ons [Dietary e](#page-53-2)[xposures](#page-76-1) Single [value dietary exposures](#page-59-2) Exposures [Exposure mixtures](#page-106-2)*

#### **[Foods data fo](#page-176-1)[rmats](#page-115-2)**

#### **[Foods](#page-194-3)**

<span id="page-27-4"></span>Foods are of interest in (dietary) consumption assessments and the sources of exposure within exposure assessments. The foods table is the main table of the food definitions. Relevant food related data, such as processing types, additional properties (e.g., unit weight and brand loyalty), facets, and hierarchies, can be described in the food properties, food hierarchies, and faces and facet descriptors tables.

### **Foods**

Each food is identified by a unique code (idFood) in a code system of choice, a name, and a description. In the EuroMix data collection, FoodEx1 codes are used for both foods in consumption surveys (foods as eaten) and for raw agricultural commodities (foods-as-measured). Example: 'A.19.01.002.002' is pizza and pizza-like pies, cheese, and vegetables and 'A.01.02.001' is wheat grain. Food codes can have a hierarchical structure (as in the FoodEx1 and FoodEx2 coding systems), using '.' or '\$' as separator between adjacent hierarchical levels, e.g. 'A.05' is fruits and fruit products, 'A.05.01' is citrus fruits, and 'A.05.01.001' is grapefruit (citrus paradisi). Additional forms of foods, such as foods in processed form, can be specified via food facets according to the FoodEx2 system of EFSA.



Table 2.3: Table definition for Foods.

Table aliases: Foods, Food.

### **Food properties**

Additional food properties, such as portion sizes can be attached using the food properties table.

Name	<b>vpe</b>	Description	Aliases	Required
idFood	AlphaNumeric $(50)$	The code of the food to which	idFood, FoodId,	Yes
		the property is attached. The	Food,	
		provided food code should	FoodCode,	
		match with a code of the	Code	
		foods table.		

Table 2.4: Table definition for FoodProperties.

Table aliases: FoodProperties, FoodProperty.

### **Food unit weights**

Food unit weights as specified for a food, and possibly a location.





Table aliases: FoodUnitWeights, UnitWeights.

### **Food hierarchies**

Food items are commonly categorised in hierarchies, e.g. oranges and mandarins are citrus fruits. For example FoodEx is a food description and food classification (FDFC) system consisting of a large number of individual food items aggregated into food groups and broader food categories in a hierarchical structure of parent-child relationships.





Table aliases: FoodHierarchies, FoodHierarchy, FoodsHierarchy.

### **Facets**

Food codes can be linked to facets, as e.g. in FoodEx.





Table aliases: Facets, Facet, FoodFacets, FoodFacet.

### **Facet descriptors**





<span id="page-30-0"></span>Table aliases: FacetDescriptors, FacetDescriptor, FoodFacetDescriptors, FoodFacetDescriptor.





Table aliases: ProcessingTypes, ProcessingType.

### **Foods as data**

<span id="page-31-0"></span>Food definitions are provided as lists/catalogues of food definitions, optionally with encompassing processing type definitions, facet definitions, hierarchy definitions, and additional food property information.

• *Foods data formats*

#### **Food coding systems**

<span id="page-32-0"></span>MCRA is intended to retain complete transparence of the results of risk assessment in terms of the foods that were actually consumed (foods-as-eaten). In many cases measurements of substances have not been made on the **foodas-eaten**, e.g. pizza, but on a raw agricultural commodity (RAC), e.g. tomato, onion etc. The food on which the concentration measurements have been made is termed the **food-as-measured**. MCRA implements a *recursive search algorithm* to link foods-as-eaten to foods-as-measured. This means that there can be intermediate steps, e.g. if unpeeled *apple* and *grapes* are the foods-as-measured, the food-as-eaten *apple pie* contains *peeled apple* and *raisins*, *peeled apple* is linked to unpeeled *apple*, and *raisins* are dried *grapes*. *Peeled* and *dried* are the *processing types*.

### **[Food classific](#page-176-2)ation: FoodEx1**

### **Food code definition**

In MCRA, a food code is a string consisting of symbols. Some special symbols  $(., \, \$, \, -, \#)$  are reserved for special use (see below), and can not be used freely in own codes.

Codes can be hierarchical. Any code can be followed by \$ or . plus a subtype code. This can be repeated any number of times, e.g. A\$B\$C\$D, or A.B.C.D.

Codes can specify the food processing type (e.g. peeling). Any code can be followed by a hyphen ('-') plus a processing type code (e.g. FP0226-2). More than one level of processing code is allowed (e.g. FP0226-2-13). Subtype codes should precede processing codes (e.g. NL005\$123\$456-2).

### **Food codes in consumption surveys**

Any coding system for foods-as-eaten can be used in MCRA. For example, in Europe EFSA develops a Food Classification and Description System for exposure assessment named FoodEx 2 ([EFSA, 2011a], [EFSA, 2011b]), featuring a hierarchical system of a core list of foods, an extended list, and domain-specific hierarchies.

#### **Food codes in concentration data**

Any coding system for foods-as-measured can be used in MCRA.

#### **Food processing**

Concentrations of substances in foods may change when foods are processed. Examples of *processing types* are peeling (e.g. of apples), cooking (e.g. of spinach), drying (e.g. of grapes), juicing (e.g. of oranges). In MCRA a processing factor can be specified for any food. Processing factors specify the ratio of concentrations in the processed and unprocessed food. The food code of the processed food (e.g. FP0226-2) will be converted to the food code of the unprocessed food (FP0226) and a processing type (2). In an exposure assessment, the [concentration in](#page-30-0) the unprocessed food will then be multiplied by the processing factor. Special attention is needed when food processing also includes changes of the weight of the food. Traditionally, processing factors combine the effects of chemical alteration and weight change, so the weight change should not be double-counted. The *processing correction factor* is introduced to correct processing factors that combine both effects, e.g. when 100g *raisins* (dried grapes) are translated to 300g *grape* (modelled food) and the processing factor for drying combines both effects, the processing correction factor is 3.

### **Recipes and food translation**

*Recipes* specify the composition of composite foods, e.g. *pizza*, in terms of relevant ingredients, e.g. 100g pizza contains 10g *tomato*, 5g *cheese* and 50g *flour*. Recipes are also used to specify weight changes, e.g. to obtain 100g *raisins* (dried grapes) 300g of the modelled food *grape* is needed, see also *processing correction*.

A special use of recipes and food translation is found in *Total Diet Studies*. Here, the composition of a Total Diet Study [food is s](#page-53-2)pecified, e.g. TDS-food *FruitMix* is composed of *apple*, *orange* and *pear* with a default translation proportion of 100%. So in MCRA, the food-as-eaten *apple* is converted to *FruitMix* (100%) and *FruitMix* is considered as the modelled food (TDS-food). A conversion from *apple-pie* (food-as-eaten) to *FruitMix* [\(modelled](#page-131-0) food) is based on a recipe for apple-pie and a TDS composition for Fruit[Mix.](#page-129-0)

Another use of converting foods (as-eaten or as an intermediate step), is through the specification of so-called food extrapolations (read across translations), e.g. for *pineapple* no measurements are found but by specifying that *pineapple* is converted to *FruitMix* (with a default proportion of 100%), the TDS sample concentration value of *FruitMix* will be used for *pineapple* (as-eaten or as ingredient).

### **Market shares and brand loyalty**

Sometimes measurements of substances in food are available at a more detailed food coding level than consumption data. For example, measurements may have been made for specific brands of a food whereas the consumption survey did not record the brand. MCRA allows to specify market share data for subtypes of a food (e.g. A\$1, A\$2, A\$3 are three brands of food A), and to calculate acute exposure based on such *market shares*.

### **Supertypes**

Sometimes measurements of substances on food are available at a less [detailed food c](#page-55-1)oding level than consumption data. MCRA allows to use the concentration data of a supertype for all underlying food codes. However, this is not the default, and an explicit permission should be given to allow this feature.

#### **Maximum Residue Levels**

Maximum residue levels are the upper legal levels of a concentration for substance residues in a food, e.g. pesticide, or feed based on good agricultural practices and to ensure the lowest possible consumer exposure.

### **MCRA food code conversion algorithm**

The conversion algorithm links food as eaten codes to modelled food codes using a *7-step procedure*.

### **Food classification: FoodEx2**

'The collection and evaluation of data on levels of chemical occurrence or presence of biological agents in food and feed are important tasks of EFSA. By combining the data with information on food consumption allows for detailed intake and exposure estimates crucial to any food and feed safety risk assessment or nutrient adequacy analysis. The EU Member States provide an increasing volume of data to EFSA and other European bodies. To provide a common link to all the diverse food and feed databases, a system for the unique and universal identification and characterisation of food and feed items is essential. EFSA has developed a preliminary standardised food classification and description system called FoodEx2 (version 2 of the EFSA Food Classification and Description System [FCDC] for exposure assessment). The system consists of descriptions of a large number of individual food items aggregated into food groups and broader food categories in a hierarchical parent-child relationship. Central to the system is a common 'core list' of food items or generic food descriptions that represent the minimum level of detail needed for intake or exposure assessments. More detailed terms may exist in addition to the core list and these are identified as the 'extended list'. A parent-child relationship exists between a core list food item and its related extended list food items. The terms of the core and extended list may be aggregated in different ways according to the needs of the different food safety domains. In the present version four hierarchies are proposed: three domain-specific and a general purpose one. Facets are used to add further detail to the information provided by the food list term. Facets are collections of additional terms describing properties and aspects of foods from various perspectives'. For more information visit: http://www.efsa.europa.eu/en/datex/datexfoodclass.htm.

For MCRA, having a different set of food codes is in itself not a problem. That is, for MCRA, it does not matter how foods are coded, as long as they can be linked to consumptions and concentrations within an exposure assessment. What makes FoodEx2 different from other food coding systems is that it provides additional food hierarchies, food facets, and a co[mbined food/facet coding system. Below follows a brief](http://www.efsa.europa.eu/en/datex/datexfoodclass.htm) summary of these main features of the FoodEx 2 coding system from the perspective of exposure assessment using MCRA.

### **Foods and food hierarchies**

FoodEx 2 contains different food hierarchy definitions and allows for creation of custom food hierarchy definitions. These hierarchies could, for exposure assessment, allow to assess intake or consumption data based on the groups defined by these hierarchies.

Code	Level	Name	ParentCode	Scopenotes
A000J		Grains and grain-based products	<b>ROOT</b>	The category covers all
A000K	$\mathcal{D}_{\mathcal{L}}$	Cereals and similar	A000J	$\cdots$
A0001	3	Cereal and cereal-like grains	A000K	$\cdots$
A000M	4	Amaranth grain	A000L	$\cdots$
A000N	5	Buckwheat grain	A000L	$\cdots$
A000P	6	Barley grain	A000L	$\cdots$
$\cdots$	$\cdots$	$\cdots$	$\cdots$	$\cdots$

Table 2.10: Food hierarchy export from FOODEX 2.0 Browser version 0.1.3

### **Facets and facet descriptors**

FoodEx 2 allows to provide supplementary details on specific aspects of foods by means of so-called facets and facet descriptors. Facets are collections of terms defining specific characteristics of food from particular points of view and facet descriptors describe specific characteristics foods. For example, *processing technology* is a facet, and *baking* is a facet descriptor belonging to this facet. Currently, 26 facets are defined, containing in total 2172 descriptors (EFSA 2011b) [EFSA, 2011b]. Facets are also defined in a hierarchical system. For instance, *cooking in fat (A07GR)* and *baking (A07GX)* are sub-items of the descriptor *cooking and similar thermal preparation processes (A0BA1)*. Facets are coded as small strings that consist of a facet code and a facet descriptor code separated by a '.'-character. For example, the facet code *F28.A07GX* holds

- 1. th[e facet code](#page-355-1) *F28*, which is the facet code for *process technology*, and
- 2. *A07GX* , which is the descriptor code for *baking*.

Code	Level	Name	ParentCode	Scopenotes
A04SF		Animals	<b>ROOT</b>	$\cdots$
A056H	2	Mammals (food source animal)	A04SF	$\cdots$
A056Z	-3	Farmed / non-game mammals (food source animal)	A056H	$\cdots$
A057A	$\overline{4}$	African buffalo (food source animal)	A056Z	$\cdots$
A057B	$\overline{4}$	American buffalo (food source animal)	A056Z	$\cdots$
A057C	$\overline{4}$	Buffalo (food source animal)	A056Z	$\cdots$
A057D	$\overline{4}$	Cape buffalo (food source animal)	A056Z	$\cdots$
A057E	$\overline{4}$	Cattle (food source animal)	A056Z	$\cdots$
$\cdot\cdot\cdot$	$\cdots$	.	$\cdots$	.

Table 2.11: Part of the FoodEx 2 facet descriptor codes of the source facet  $(F01)$ 

### **Implicit facets**

Implicit facets are facets of a product that are already implied by the food product itself. Consider, for example, *potato boiled (A011P)*, where *boiling (A011P)* is an implicit facet, because boiling is already implied by the product. According to EFSA [EFSA, 2011a] 'inclusion of implicit facets in the string recorded for each food database record is not encouraged' and it is suggested to identify and record the implicit facet descriptors in a separate table.

### **Foods as facets**

Foods and facet descriptors share the same unique alphanumerical coding system; in some cases, like *characterising ingredient or sweetening agent* food list elements may be used as facet descriptors.

### **The FoodEx 2 coding system**

In the coding system, facets can be added to the primary food codes to provide supplementary detailed information of particular data records. The structure of the FoodEx 2 codes is:

*idFood#idFacet.idFacetDescriptor\$idFacet.idFacetDescriptor\$….*

The code starts with the primary FoodEx2 food code. Then, when there are supplementary facets, the food code is followed by a '**#'**-character and the facets string. The facets string is constructed as a concatenation of the individual facets strings, separated by means of the '**\$**' character. As an example, consider the string *A011P#F28.A07GL\$F28.A07KQ* which is composed of:

- Food: *A011P Potato boiled*
- Facet 1: *F28.A07GL Process technology Boiling*
- Facet 2: *F28.A07KQ Process technology Freezing*

### **FoodEx2**

For MCRA, FoodEx 2 introduces the following points of attention:

- <span id="page-35-0"></span>• Reading and dealing with FoodEx 2 coded data sets
- Reading and dealing with food facets
- Reading and exploiting food hierarchy data

### **Reading and dealing with FoodEx 2 codes**

All data entities that contain foods data are potentially affected by the introduction of FoodEx 2. In MCRA, the following data tables are adapted to allow for input of full FoodEx 2 food codes:

- Foods
- Consumptions
- Concentrations

For these tables, the food code is allowed to be the complete FoodEx 2 food code and automatically recognized as such. As an example, Table 2.12 shows how the FoodEx 2 coded consumptions should be provided to the system. On important note: the maximum field length of the food code is 50. This means that there is a maximum of five facets that can be specified for a food.




## **Reading and dealing with facets data**

Within MCRA, the following facets related aspects are accounted for:

- Reading facets data
- Dealing with facets
- Facets in concentration data
- Facets in food conversion
- Using facets as processing factors
- Using hierarchy data in the output

## **Reading facets data**

To incorporate input of facets data in MCRA, two tables Facets and FacetDescriptors are introduced as optional tables of the Foods data group. The *table for Facets* and *table for FacetDescriptors*.

Within MCRA, the facets of FoodEx 2 coded foods, consumptions, and concentrations are automatically linked to the provided facets and facet descriptors. Also, the facet descriptor names are added automatically to the foods containing these facets.

#### **Dealing with facets**

The introduction of food facets allows for much more detailed specifications of consumption and concentration data. However, it introduces the problem of deciding on which level of detail the exposure assessment should be performed. That is, should concentration models be generated on the level of foods-without-facets or on the level of foods-with-facets? E.g., should the concentrations of *clementine peeled (A01CE#F28.A07LC)* and *clementine unprocessed (A01CE#F28.A0C0S)* be modelled separately or should one model be constructed for *clementine (A01CE)*? Treating all clementine's as equal may yield over-simplified conversions, whereas treating all separately may lead to many concentration models based on only few measurements. In MCRA, no implicit grouping of concentrations of equal foods with different facets is applied. If concentrations are provided for both *clementine peeled (A01CE#F28.A07LC)* and *clementine unprocessed (A01CE#F28.A0C0S)*, then these are modelled separately. Another question is whether the order of the facets is relevant or not. E.g., is *A0BYV#F02.A06GF\$F03.A06HY* the same as *A0BYV#F03.A06HY\$F02.A06GF*? Regarding this matter, MCRA considers the facet order to be important. I.e., *A0BYV#F02.A06GF\$F03.A06HY* is not the same as *A0BYV#F03.A06HY\$F02.A06GF*.

## **Facets in food conversion**

For conversion of foods-as-eaten to foods-as-measured, MCRA considers foods with different facet strings as different foods. I.e., there is no implicit conversion of foods-with-facets to foods-without-facets and also the order of the facets is important. However, as it is realistic to convert food-with-facets to the base food without facets, an additional (explicit) conversion step remove-all-facets is added that converts foods with facets to the base foods. I.e., the action is "remove all". There is no conversion step for "stripping off one facet at a time". The reason for this is that there is no good way of deciding which facet to strip off first. This new conversion step is somewhat equivalent to the already existing default processing conversion step (step 6), and is therefore implemented as step 6b of the conversion algorithm. Particular rules followed by this step:

• Conversion of food-with-facets to food-without-facets.

## **Using facets that reveal processing data**

Facets containing processing information, such as *part-consumed-analysed (F20)* and *processing technology (F28)* could be integrated with processing data. As an example, consider *clementine peeled (A01CE#F28.A07LC)*. This could be linked to *clementine (A01CE)*, with processing type *removal of external layer (A07LC)*. Linking to processing data could be achieved by entering processing data using the facet codes. As an alternative to the current processing factor tables, a facet-based processing factors table is defined for processing facets. That is, the codes for food processed and unprocessed are implicitly defined for FoodEx 2.

Table 2.13: Example of a MCRA processing factors table using FoodEx 2 foods and facets codes.

FacetCode	Substance	FoodCode	<b>ProcNom</b>	ProcUpp	Proc-	Proc-
					NomUnc-	UppUnc-
					Upp	$J$ pp
A07LC	SubstanceX	A01CE	0.5	0.6	0.05	0.06
F <sub>28</sub> .A07GV	SubstanceX	A0BY	0.2	0.1	0.03	0.04

Note that in the example, the facet code could be specified as the full facet code, or just the code of the facet descriptor. As a more elaborate example consider

*French fries from cut potato (A0BYV#F02.A06GF\$F03.A06HY\$F04.A00ZT\$F28.A07GR)*

For this food code, the substring of the processing facet is extracted from the list of facets.

- *A0BYV#F02.A06GF\$F03.A06HY\$F28.A07GR\$F04.A00ZT* with processing facet link A07GR
- *A0BYV#F02.A06GF\$F03.A06HY\$F04.A00ZT*

In MCRA, a table FacetProcessingFactors is introduced that allows for specification of processing factors by means of facets. This table has the following structure:





The integration with the food conversion algorithm is as follows: Conversion step 2 (*processing*) is extended with a step 2c (*processing facet*) that attempts to match facets of a food code to processing data provided in the processing facets table. The following important rules are followed:

- Processing factors can be defined for base-food-code/facet-code combinations and translate as food-withprocessing-facet to food-without-processing-facet.
- If multiple processing facets are present in the food-as-eaten code, then the last processing facet is used first for conversion.
- Facet processing factors can be specified using the full facet code (i.e., facet-code plus facet-descriptor-code) or just the facet descriptor code. If both are specified for the same food, the full facet code is used.
- Facet processing factors can be defined substance-specific, and non-substance-specific. Processing factors that are defined substance-specific always precede non-substance specific processing factors.
- Processing factors defined by a food-processed/food-unprocessed combination precede processing factors defined through facets.

Weight reduction factors for processing factors defined for facets should be included in the food translation table and should match exactly.

## **Food hierarchies**

## **Reading and dealing with food hierarchy data**

Within MCRA, the following hierarchy related aspects are accounted for:

- Reading food hierarchy data
- Using hierarchical data for conversion of foods
- Using hierarchy data in the output

## **Reading food hierarchy data**

A new data group named *Foods* is added. In this group, a new *table for FoodHierarchies* is used for input of food hierarchies. This table contains food hierarchy node-definition records that reflect a hierarchical structure. For foods that are not in this list as idFood, it is implicitly assumed that these foods are root items.

Note: It is common practice to describe hierarchies using tree structures. Here, the elements of the tree are named *nodes*, the lines connectin[g the no](#page-27-0)des are named *branches*, and n[odes without children are](#page-29-0) *leaf nodes/end-nodes*. This terminology is also used throughout the remainder of this document.

#### **Using food hierarchies for food conversion**

The introduction of the hierarchy structure allows for integration with step 4 and step 5 of the food conversion algorithm; the *subtype* and *supertype* linking steps. That is, when no concentration data is found for a certain product, the concentration data of a (according to the hierarchy) related product could be used. In MCRA, the *supertype* conversion step also contains a *hierarchy-supertype* step based on the food hierarchy.

#### **Supertype link (step 5)**:

- a) **Supertype**: Try to find supertypes base on '\$'-coded strings, e.g., 'xxx\$yyy' is converted to 'xxx'
- b) **Hierarchy-supertype**: try to find the supertype of the current food based on the food hierarchy (i.e., convert the current food to its parent).
- Note 1: the *supertype* conversion step is optional and should be specified in the conversion settings panel.

Note 2: the *hierarchy-supertype* step only applies for foods-without-facets. The reason for this is that for the conversion, the base type of a food-with-facets can be considered as a better conversion candidate than the parent food with the same facets.

#### **Using hierarchy data in the output**

Food hierarchy information could be used in presentation of various tables of the output of MCRA. That is, in the tables in which foods data is presented, these records could be grouped based on the hierarchy and/or a tree-like display can be built for the presentation of this data. Tables that are candidate for being extended are, for example, the input data tables foods-as-eaten/foods-as-measured and the exposure by food-as-eaten/food-as-measured output tables.

Summarizing over the food hierarchy is many cases not a straightforward task. Consider, for instance, the statistic *number of consumption days* given the artificial hierarchy of *Citrus Fruits* containing two child-nodes *Mandarin* and *King Mandarin*: the number of consumption of *Citrus Fruits* is not "just" the sum of the consumption day of *Mandarin* and *King Mandarin*. A difficulty for summarizing based on a hierarchy arises when a node contains both data and child-nodes with data. E.g., concentrations are defined on the level of *Citrus Fruits* and on the level of *Mandarin*. In this case, the hierarchy view should ideally summarize for both *Citrus Fruits* as data record and *Citrus Fruits* as summary node. An additional complication is the status of facet-coded foods within the hierarchy. In a hierarchical view, foods-with-facets should ideally be added to their base-foods for visualization.

In MCRA, an alternative view (treetable) is added that can display hierarchical data. This alternative view is used to present a hierarchical view based on the foods hierarchy for the consumption input summary tables food as eaten and modelled food. The data summary methods for these tables are updated such that the data is also summarized per hierarchy-node.

鹽 ຨ							
Food name	Food code	Mean (q)	Mean consumption consumption days davs(a)	<b>Consumption Percentage</b>	consumption weights davs	Total consumption weights davs	Percentage total consumption days
$\Box$ Fruit and fruit products	A01BS	167	200	5.	83.3%	5.0	83.3%
$F$ Fresh fruit	A04RK	167	200	5	83.3%	5.0	83.3%
$\Box$ Starchy roots or tubers and products thereof, sugar plants	A00ZR	100	600	٦	16.7%	1.0	16.7%
$\Box$ Starchy root and tuber products	A011B	66.7	400	٦	16.7%	1.0	16.7%
$\Box$ Processed root and tuber products	A04MI	66.7	400	٦	16.7%	1.0	16.7%
$\Box$ Potato boiled	A011P	66.7	400	٦	16.7%	1.0	16.7%
Potato boiled Tuber (as part-nature)	A011P#F02.A067V	16.7	100	ı	16.7%	1.0	16.7%
Potato boiled Tuber (as part-nature), Potatoes, <b>Boiling</b>	A011P#F02.A067VSF27.A00ZTSF28.A07GL	16.7	100		16.7%	1.0	16.7%
Potato boiled Tuber (as part-nature), Potatoes, <b>Boiling</b>	A011P#F02.A067VSF28.A07GLSF27.A00ZT	16.7	100		16.7%	1.0	16.7%
Potato boiled Tuber (as part-nature), Potatoes, Boiling, Baking	A011P#F02.A067VSF27.A00ZT\$F28.A07GL\$F28.A07GX	16.7	100		16.7%	1.0	16.7%
□ Starchy roots and tubers	A0075	33.3	200	٦	16.7%	1.0	16.7%
$\Box$ Tubers	A04MC	33.3	200	ı	16.7%	1.0	16.7%
$\Box$ Potatoes	A00ZT	33.3	200	٦	16.7%	1.0	16.7%
Potatoes Potatoes (food source plant). Tuber (as part-nature)	A00ZT#F01.A05KG\$F02.A067V	16.7	100		16.7%	1.0	16.7%
Potatoes Potatoes (food source plant), Tuber (as part-nature), Baking	A00ZT#F01.A05KG\$F02.A067V\$F28.A07GX	16.7	100		16.7%	1.0	16.7%

Figure 2.2: Hierarchy view for the foods as eaten input summary table.

If a node contains both data and a child record, then this node is split-up in two nodes: a summary node that summarizes the data of the node and all of its child nodes, and a data record with the string "(unspecified)" added as a child of this summary node. See Figure 2.2 for an example (*Citrus Fruits* versus *Citrus Fruits (unspecified)*). In MCRA, foods-with-facets are added as child nodes of the foods-without-facets.

## **Food unit weights**

Food unit weights specify the standard weights of food units. E.g., the standard weight of an apple. This unit weight may be specified as the weight of the whole food (raw agricultural commodity/RAC) or the weight of the edible portion (EP), e.g., without peel. Unit weights are specified in the table *table for FoodUnitWeights* and used in combination with *unit variability factors* to account for unit-to-unit variation in concentrations between single units of the same food in *single value dietary exposures assessments* and *(individual) dietary exposures assessments*.

Food unit weights can be location specific or specified as overall (default) unit weights. For some models, e.g., the *IESTI model*, location specific unit weights are preferred over o[verall unit weights. The ove](#page-29-1)rall unit weights are then used [when no location spec](#page-113-0)ific uses are available. For other methods, only overall unit weights are used. If, for a food, a[n overall unit weight is not available, but th](#page-194-0)ere [are location specific unit weights available,](#page-116-0) then the overall unit weight is computed as the average weight of the location specific unit weights (similar to EFSA PRIMo revision 3 [\[EFSA, 2018](#page-197-0)]).

**Note:** Note that in earlier versions of the software, food unit weights were specified in the *table for FoodProperties*. Although this is still possible, the recommended way of specifying unit weights is in the *table for FoodUnitWeights*. [If, for a food,](#page-355-0) unit weights are specified in both tables, then the unit weights specified in the *table for FoodUnitWeights* have priority. The unit weights specified in the *table for FoodProperties* are then only used as fallbacks for the overall unit weight when no overall unit weight is specified in the *table for FoodUnitWeights*.

# **2.1.3 Populations**

Populations are groups of human individuals that are the scope of exposure or risk assessments. Optional descriptors of populations are location (e.g. a country), time period (start date, end date), age range and gender. Example: the French population in 2005-2007 of women of child-bearing age (18-45 yr).

<span id="page-41-1"></span>Output of this module is used by: *Consumptions Single value consumptions Consumptions by modelled food Dietary exposures Single value dietary exposures Non-dietary exposures Exposures Human monitoring analysis Risks Single value risks*

## **[Populati](#page-116-0)[ons data formats](#page-194-0)**

## **[Populatio](#page-263-0)ns**

<span id="page-41-0"></span>Populations are primary entities of the data model.

## **Populations**

Populations identify human groups, and e.g. dietary, nondietary and human monitoring surveys. Optionally, a name and description can be added. Population can be restricted to a certain time period. AgeMin, AgeMax and Gender are optional properties of a population.

Name	<b>Type</b>	Description	<b>Aliases</b>	Required
idPopulation	AlphaNumeric(50)	Unique identification code of the population.	IdPopulation, PopulationId, Code, Id	Yes
Name	AlphaNumeric(100)	The name of the population.	Name, <b>PopulationName</b>	No
Description	AlphaNumeric(200)	Description of of the population.	Description	N <sub>o</sub>
Location	AlphaNumeric(50)	Location.		No
<b>StartDate</b>	<b>DateTime</b>	Starting date of the specific time window marking this population.	<b>StartDate</b>	N <sub>o</sub>
EndDate	<b>DateTime</b>	End date of the specific time window marking this population.	EndDate	No
AgeMin	Integer	Inclusive minimum bound (in years) of the specific age group of this population.	AgeMinimum	N <sub>o</sub>
AgeMax	Integer	Inclusive maximum bound (in years) of the specific age group of this population.	AgeMaximum	N <sub>o</sub>
Gender	AlphaNumeric(50)	Gender levels of this population.	<b>Sex</b>	No
NominalBody- Weight	Numeric	Nominal body weight (in kg) of the individuals of this population.	NominalBody- Weight, BodyWeight	N <sub>o</sub>

Table 2.15: Table definition for Populations.

Table aliases: Populations, Population.

## **Selection settings**

Table 2.10: Selection settings for module Populations.		
Name	Description	
Population	Specifies which population is selected.	

 $Table 2.16:$  Selection settings for modul

#### **Populations as data**

Populations are provided as data.

• *Populations data formats*

## **2.1.4 Responses**

Respo[nses are measurable entiti](#page-41-0)es in test systems. Responses are used to represent effects (see effect representations) and their measured values are collected in dose response data.

This module has as primary entities: *Test systems*

Output of this module is used by: *Dose response models Dose response data Effect representations*

#### **Responses data formats**

#### **Responses**

<span id="page-42-0"></span>A response is a measurable endpoint on in a test system. E.g., in a rat test system a response may be the percentage of fatty hepatocytes observed after 90 days. Responses are defined in the responses table.

#### **Responses**

Each response is identified by a unique code (idResponse) in a code system of choice, a name, and a description. Also, each response should be linked to a test system (idTestSystem) on which the response is measured. Responses can be of various types (ResponseType), e.g., ContinuousMultiplicative (= non-negative real values using a ratio scale), ContinuousAdditive (= real values using an interval scale), Ordinal, Quantal, or Binary. For continuous variables, the response unit (ResponseUnit) is also relevant. Additionally, also a reference to the test method guideline, e.g., standardised assay kit may also be specified (GuidelineMethod).

Table 2.17: Table definition for Responses.

<b>Name</b>	Type	Description	<b>Aliases</b>	Required
idResponse	AlphaNumeric(50)	Unique identification code of	idResponse,	Yes
		the response. In the EuroMix	ResponseId,	
		data collection, a EuroMix	Response, Id	
		coding system has been set up		
		in which the id of the test		
		system prefixes the id of the		
		response. E.g.,		
		'HepaRG-PCR-PPARA',		
		'RatWEC-PCR-CYP26a1'		
		and 'MouseDevelopmental-		
		FacialPrimordia-malformed-		
		E9'.		
CodeSystem	AlphaNumeric(100)	Identifier of the coding	CodeSystem	N <sub>o</sub>
		system of the response code.		
Name	AlphaNumeric(100)	Name of the response.	Name	N <sub>o</sub>
Description	AlphaNumeric(200)	Additional description or label	Description	$\overline{No}$
		of the response.		
idTestSystem	AlphaNumeric $(50)$	Unique identification code of	idTestSystem,	Yes
		the test system.	idSystem,	
			SystemId,	
			TestSystem	
Guideline-	AlphaNumeric(200)	Reference to the test method	Guideline-	N <sub>o</sub>
Method		guideline, e.g., standardised	Method	
		assay kit.		
ResponseType	ResponseTypes	The data type of the response	ResponseType	Yes
		measurements (e.g.,		
		continuous multiplicative,		
		ordinal, categorical).		
ResponseUnit	AlphaNumeric(100)	If the response type is	ResponseUnit	N <sub>o</sub>
		Continuous, then this should		
		be the unit of the response,		
		e.g., kg.		

Table aliases: Responses, Response.

## **Responses settings**

## **Selection settings**





#### **Responses as data**

A response is a measurable endpoint defined in a test system. It has a unit and a measurement type (e.g., continuous non-negative, quantal).

• *Responses data formats*

# **2.1.5 Substances**

<span id="page-44-0"></span>Subst[ances are chemical entiti](#page-42-0)es that can refer to: 1) active substances such as investigated in toxicology; 2) measured substances such as defined in specific analytical methods. MCRA assessments can have one or more substances as the scope. When more than one substance is specified, there is an option to perform a cumulative assessment. In that case one of the substances has to be indicated as the index/reference substance, and results will be expressed in equivalents of the index substance.

Output of this module is used by: *Concentrations Concentration distributions Single value concentrations Processing factors Unit variability factors Occurrence patterns Occurrence frequencies Substance authorisations Substance conversions Deterministic substance conversion factors Concentration limits Concentration models Modelled foods Focal food concentrations Food conversions Consumptions by modelled food High exposure food-substance combinations Dietary exposures Single value diet[ary exposures](#page-76-0) [Non-dietary exposures](#page-59-0) Exposures [Exposure mixtures](#page-106-0) Hu[man moni](#page-102-0)[toring d](#page-102-0)ata [Human monitoring](#page-113-0) analysis [QSAR mem](#page-97-0)[bership models](#page-95-0) Molecul[ar docking models](#page-108-0) Kineti[c models](#page-109-0) Active [substanc](#page-109-0)es [Relative potency factors](#page-91-0) Hazard characterisations [Points of de](#page-60-0)parture [Dose response](#page-62-0) models [Dose res](#page-94-0)[ponse](#page-92-0) data [Inter-species co](#page-92-0)[nversions](#page-176-0) Intra species factors Risks [Single value ris](#page-115-0)[ks](#page-147-0)*

## **[Substan](#page-202-0)[ces data formats](#page-237-0)**

#### **[Sub](#page-209-0)[stances](#page-231-0)**

Substances are primary entities of the data model. Substance intakes are of main interest in exposure assessments and the effect of intake on human health is of interest in risk assessments. In the substances table, the substance entities and other relevant substance properties that are relevant for the assessment at hand should be defined.

#### **Substances**

Each substance should have a unique identification code (idSubstance), and optionally, a name and description may be used for a more detailed description of the entity. Additional properties, such as the molecular mass (MolecularMass) and Cramer class (CramerClass) may also be specified. Example: Captan (idSubstance RF-0061-001-PPP) has MolecularMass 300.5922 and CramerClass 3.

Table 2.19: Table definition for Compounds.

Name	Type	<b>Description</b>	<b>Aliases</b>	Required
idSubstance	AlphaNumeric(50)	The unique identification code	idSubstance,	<b>Yes</b>
		of the substance. This code	SubstanceId,	
		may be from an existing	Substance,	
		coding system, such as	Code, Id	
		CAS-codes or Param codes of		
		EFSA, or it may be a		
		used-defined code.		
Name	AlphaNumeric(100)	The substance name.	Name,	No
			SubstanceName,	
			PesticideName	
Description	AlphaNumeric(200)	Substance description.	Description	N <sub>o</sub>
<b>ARFD</b>	Numeric	The acute reference dose of	<b>ARFD</b>	N <sub>o</sub>
		the critical effect. Note that		
		this is always specified in		
		mg/kg bw/day (exposure).		
ADI	Numeric	The acceptable daily intake.	<b>ADI</b>	N <sub>o</sub>
		Note that this is always		
		specified in mg/kg bw/person		
		(exposure).		
<b>SF</b>	Numeric	The safety factor belonging to	<b>SF</b>	N <sub>o</sub>
		the ADI/ARFD.		
<b>CramerClass</b>	Integer	The Cramer class of the	<b>CramerClass</b>	No
		substance.		
MolecularMass	Numeric	The molecular (molar) mass.	MolecularMass,	N <sub>o</sub>
			Mass,	
			MolarMass,	
			Molecular-	
			Weight,	
			MolarWeight	

Table aliases: Substances, Substance.

## **Substances settings**

## **Selection settings**





## **Substances as data**

Substances are provided as data (code, name).

• *Substances data formats*

# **2.1.6 Test systems**

Test systems are biological or artificial systems used for assessing hazard in relation to chemical exposure from substances in varying doses. Test systems may refer to 1) in-vivo test systems (e.g. a rat 90-day study, a human biomonitoring study); 2) in-vitro test systems (e.g. HepaRG cells).

Output of this module is used by: *Responses Dose response models Dose response data*

### **Test systems data formats**

#### **Test Systems**

<span id="page-46-0"></span>Test systems are the biological systems (e.g., animals) or in-vitro systems on which responses related to health effects can be measured.

## **Test Systems**

Each test system should have a unique identification code (idSystem), and (optionally) a name and a description. The test system's type (TestSystemType) indicates the type whether the test system is an in-vivo test system (in which case it is a model for external exposure) or any of a range of other, in-vitro, options (cell-line, etc., which all will be interpreted as models for internal exposure). Additionally, if applicable, the organ (e.g., liver) of the test system and the route of exposure (RouteExposure) for in-vivo test systems (oral, dermal or inhalation) may be specified.

Table 2.21: Table definition for TestSystems.

<b>Name</b>	<b>Type</b>	<b>Description</b>	<b>Aliases</b>	Required
idTestSystem	AlphaNumeric $(50)$	Unique identification code of	idTestSystem,	Yes
		the test system.	idSystem, Id,	
			Code	
CodeSystem	AlphaNumeric(50)	Identifier of the code system	CodeSystem	$\overline{No}$
		of the test systems.		
Name	AlphaNumeric(100)	Name of the test system.	Name	No
Description	AlphaNumeric(200)	Additional description or label of the test system.	Description	N <sub>o</sub>
TestSystem-	<b>TestSystemTypes</b>	The type of the test system,	TestSystem-	$\overline{No}$
Type		i.e., in-vivo, cell-line, etc.	Type,	
			SystemType	
Organ	AlphaNumeric(100)	If applicable, the organ that	Organ	$\overline{No}$
		the cells originate from		
		associated with the in vitro		
		test-system.		
Species	AlphaNumeric(100)	If applicable, the species	<b>Species</b>	No
		associated with the		
		test-system.		
Strain	AlphaNumeric(100)	If applicable, the strain of the	Strain	N <sub>o</sub>
		species associated with the		
		test-system.		
RouteExposure	ExposureRouteTypes	If applicable, the route of	ExposureRoute-	$\overline{No}$
		exposure associated with the in vivo test-system, oral,	Type, ExposureRoute,	
		dermal, inhalation, s.c., i.v.	RouteExposure	
Guideline-	AlphaNumeric(200)	Reference to test guideline.	GuidelineStudy	No
Method				
Reference	AlphaNumeric(200)	External reference(s) to other	Reference	N <sub>o</sub>
		sources containing more		
		information about the test		
		system. E.g., publications,		
		website, documents.		

Table aliases: TestSystems, TestSystem, Systems, System.

## **Test systems as data**

Test systems are provided as data.

• *Test systems data formats*

# **2.2 [Consumption](#page-46-0) modules**

Consumption modules specify the *consumptions* or *single value consumptions* of *foods* by surveyed individuals in *populations*. Foods can be related to each other using *food recipes*.

# **2.2.1 Consumptions**

Consumptions data are the amounts of foods consumed on specific days by individuals in a food consumption survey. For acute exposure assessments, the interest is in a population of person-days, so one day per individual may be sufficient. For chronic exposure assessments, the interest is in a population of persons, so preferably two or more days per individual are needed.

This module has as primary entities: *Populations Foods*

Output of this module is used by: *Food conversions Consumptions by modelled food*

## **Consumptions data formats**

<span id="page-48-0"></span>Consumption data is often collect[ed in 24-hour die](#page-176-0)[tary recall studies and contains t](#page-115-0)he food consumptions and consumption amounts for a number of individuals on a number of days. For each of the individuals, the bodyweight should be specified, and optionally also age, sex, and other properties may be recorded. If applicable, sampling weights may also be specified that can be used to correct the sample of individuals in the survey to a more representative sample of the targeted population. The consumption amounts are usually expressed in grams, but may also be expressed in alternative units of plates, cups, or spoons. Optionally, the uncertainty of food consumption quantifications can be specified, see [Souverein et al., 2011].

## **Consumptions**

Consumption surveys are described using three tables: FoodSurveys, Individuals, and Consumptions. Individuals are linked to food surveys using the survey code (idFoodSurvey), and consumptions are linked to individuals using the individual codes (idIndividual). The food codes used to identify the consumed foods should match with the codes provided by the foods entity definitions.

## **Food consumption surveys**

The records of the food consumption surveys table contain the ids, names, descriptions, and other relevant metadata of consumption surveys.

Table 2.22: Table definition for FoodSurveys.

<b>Name</b>	<b>Type</b>	Description	<b>Aliases</b>	Required
idSurvey	AlphaNumeric(50)	Unique identification code of	idSurvey,	Yes
		the food consumption survey.	idFoodSurvey,	
			Survey,	
			FoodSurvey,	
			SurveyId,	
			FoodSurveyId,	
			Name, Code, Id	
Description	AlphaNumeric(200)	Description of the food	Description	$\overline{No}$
		consumption survey.		
Location	AlphaNumeric $(50)$	The location or country where	Location,	$\overline{No}$
		survey is held. It is	Country	
		recommended to use ISO		
		Alpha-2 country codes.		
BodyWeight-	<b>Body WeightUnits</b>	The unit of bodyweight of the	BodyWeight-	No
Unit		individuals of the survey: kg	Unit.	
		(default) or g.	UnitBody-	
			Weight,	
			WeightIn	
AgeUnit	AgeUnit	The unit of age, i.e., year or	UnitAge, agein,	N <sub>0</sub>
		month.	AgeUnit	
Consumption-	Consumption Units	The unit of the	AmountUnit,	$\overline{No}$
Unit		use/consumption amounts of	UnitAmount,	
		the consumptions of the	AmountUnit,	
		survey: g (default) or kg or	Consumption-	
		CustomUnit (see table food	Unit	
		consumption quantifications		
		table).		
<b>StartDate</b>	<b>DateTime</b>	The start date of the survey.	<b>StartDate</b>	$\overline{No}$
EndDate	<b>DateTime</b>	The end date of the survey.	EndDate	N <sub>o</sub>
NumberOf-	Integer	The number of days each	NumberOf-	Yes
SurveyDays		individual participated in the	SurveyDays,	
		survey.	NDaysInSurvey	
idPopulation	AlphaNumeric(50)	Unique identification code of	IdPopulation,	N <sub>o</sub>
		the population.	PopulationId	

Table aliases: FoodSurvey, FoodSurveys, Survey, Surveys.

## **Individuals**

The individuals of a survey are recorded in the individuals table.

<b>Name</b>	Type	Description	<b>Aliases</b>	Required
idIndividual	AlphaNumeric(50)	Unique identification code of	idIndividual,	Yes
		the individual.	IndividualId,	
			Individual, Id	
idFoodSurvey	AlphaNumeric(50)	The identification code / short	idSurvey,	Yes
		name of survey.	idFoodSurvey,	
			Survey,	
			FoodSurvey,	
			SurveyId,	
			FoodSurveyId,	
			SurveyCode	
BodyWeight	Numeric	The body weight of the	BodyWeight,	Yes
		individual.	Weight	
Sampling-	Numeric	The sampling weight for an	<b>SamplingWeight</b>	$\overline{No}$
Weight		individual (default $= 1$ ).		
NumberOf-	Integer	The number of days the	NumberOf-	N <sub>o</sub>
SurveyDays		individual participated in the	SurveyDays,	
		survey.	NumberOfDays-	
			InSurvey,	
			DaysInSurvey,	
			NDaysInSurvey	
Age	Numeric	The age of the individual.	Age	N <sub>o</sub>
Gender	AlphaNumeric(50)	The gender of the individual.	Gender	N <sub>0</sub>
		It is recommended to use the		
		codes Male/Female for coding		
		the gender.		
Other individual		Other individual properties		No
properties		can be added just like the		
		fields age and gender. These		
		properties are automatically		
		parsed as co-factors or		
		co-variables.		

Table 2.23: Table definition for Individuals.

Table aliases: Individuals, Individual.

## **Individual properties**

Individual properties, additional columns that can also be specified as additional columns in the Individuals table





Table aliases: IndividualProperties, IndividualProperty.

## **Individual property values**

Individual property values, additional columns that can also be specified as additional columns in the Individuals table

Name	vpe	Description	Aliases	Required
idIndividual	AlphaNumeric(50)	The identification number of	Id	Yes
		the Individual.		
PropertyName	AlphaNumeric $(50)$	The name of the property.	Name	Yes
<b>TextValue</b>	AlphaNumeric $(50)$	The value of the property as		N <sub>0</sub>
		text value.		
DoubleValue	Numeric	The value of the property as		N <sub>0</sub>
		number.		

Table 2.25: Table definition for IndividualPropertyValues.

Table aliases: IndividualPropertyValues, IndividualPropertyValue.

## **Consumptions**

The individual consumptions are recorded in the consumptions table.

Name	Type	<b>Description</b>	Aliases	Required
idIndividual	AlphaNumeric(50)	The unique identification code	idIndividual,	Yes
		of the consumer (individual).	IndividualId,	
			Individual	
idFood	AlphaNumeric(50)	The food code (food as eaten	idFood, Food,	Yes
		code).	FoodId,	
			FoodConsumed,	
			FoodAsEaten	
idUnit	AlphaNumeric(50)	Identification code of the unit	idUnit, Unit,	N <sub>o</sub>
		in which the food is consumed	UnitId	
		(e.g. plate, cup, spoon).		
idDay	AlphaNumeric $(50)$	Identification code of the day	idDay, DayId,	Yes
		of consumption, sequential	Day,	
		number	DayOfSurvey	
idMeal	AlphaNumeric(50)	Identification code of the meal	idMeal, MealId,	N <sub>o</sub>
		(eating occasion within a day).	Meal	
Amount	Numeric	The consumed portion of food	Amount,	Yes
		in g (default) or kg or quantity	Amount-	
		of a plate, cup, spoon. Days	Consumed	
		without consumptions are not		
		recorded.		
DateConsumed	<b>DateTime</b>	The date of the consumption.	DateConsumed,	N <sub>o</sub>
			Consumption-	
			Date	

Table 2.26: Table definition for Consumptions.

Table aliases: FoodConsumptions, FoodConsumption, Consumptions, Consumption.

## **Food consumption quantifications**

Food consumption quantifications record information about food consumption quantities that are associated with unit-consumptions of foods.





Table aliases: FoodConsumptionQuantifications, FoodConsumptionQuantification.

## **Consumptions settings**

## **Selection settings**

Name	Description
Food survey	The food consumption representative for the population of
	interest.
Restrict population to	Specifies whether the population should be restricted to the
consumers or consumer days	individuals (chronic) or individual days (acute) that have non-zero
only	consumption.
Restrict population to	Specifies whether the population should be restricted to the
consumers or consumer days	individuals (chronic) or individual days (acute) consuming any of
with consumptions of specific	the foods of the specified subset.
foods	
Selected foods-as-eaten	Set of consumed foods that are of particular interest for restricting
	the consumers / consumption days.
Consumption subset: restrict to	If checked, then the consumptions are restricted to those of the
consumptions of specific foods	specified food-as-eaten subset.
Selected foods-as-eaten	Set of consumed foods that are of particular interest.
Ignore sampling weights	If checked, individual sampling weights are not used (sampling
	weight $= 1$ ). If unchecked, the specified sampling weights are
	used.

Table 2.28: Selection settings for module Consumptions.





#### **Consumptions uncertainty**

In MCRA, in an *acute exposure* assessments, individual consumption day data are *resampled*, thus preserving the multivariate consumption patterns and associated weights and/or other individual characteristics. In MCRA we resample the set of individuals x number of survey days. We think that this implementation better reflects the notion of acute exposure which is expressed as the normalized intake per day. For *chronic exposure* assessments the resampling algorithm remain[ed unchanged a](#page-117-0)nd the set of individuals (with corresponding days) is *[resample](#page-12-0)d*.

## **Consumptions as data**

Consumptions data are the amounts of foods consumed on specific days by individuals [in a food co](#page-12-0)nsumption survey.

• *Consumptions data formats*

# **2.2.2 Food recipes**

Food [recipes data specify the com](#page-48-0)position of specific foods (typically: foods-as-eaten) in terms of other foods (intermediate foods or foods-as-measured) by specifying proportions in the form of a percentage.

This module has as primary entities: *Foods*

Output of this module is used by: *Food conversions*

## **Food recipes data formats**

## **Food recipes**

<span id="page-53-0"></span>Recipe data to specify the ingredients of foods. Food recipes can be used to describe the ingredients of a composite food (e.g., of apple pie), or to specify the amount of a primary ingredient needed to obtain 100g of the food (e.g., grapes to raisins). Recipe is commonly used recursively (e.g., apple pie contains apple and flour, flour contains wheat).





Table aliases: FoodTranslations, FoodTranslation, FoodCompositions, FoodComposition.

## **Food recipes as data**

Food recipes are provided as data in the form of simple composition tables.

• *Food recipes data formats*

# **2.2.3 Market shares**

Mark[et shares data specify for a](#page-53-0) given food, percentages of more specific foods (subfoods, e.g. brands) representing their share in a market. Market shares are used when consumption data are available at a more generalised level than concentration data.

This module has as primary entities: *Foods*

Output of this module is used by: *Food conversions*

## **Market shares data formats**

## **MarketShares**

<span id="page-54-0"></span>Describes the shares (proportions) in a market.

#### **Market shares**

Market shares main table.





Table aliases: MarketShares, MarketShare, FoodMarketShares, FoodMarketShare.

## **Market shares as data**

Market shares are provided as data in the form of percentages.

• *Market shares data formats*

#### **Market shares and brand loyalty**

Somet[imes measurements of subs](#page-54-0)tances in food are available at a more detailed food coding level than consumption data. For example, measurements may have been made for specific brands of a food whereas the consumption survey did not record the brand. MCRA allows to specify market share data for subtypes of a food (e.g. A\$1, A\$2, A\$3 are three brands of food A), and to calculate acute exposure based on such market shares.

For chronic assessments **brand loyalty** should be specified according to a simple Dirichlet model [Goodhardt et al., 1984]. Technically, the Dirichlet model for brand choice needs nbrand parameters  $\alpha_i$  (which should be positive real numbers). The average brand choice probability for each brand is

 $\alpha_i/S$ 

[where](#page-355-1)

$$
S=\sum \alpha_i
$$

By definition, the market shares  $m_i$  should be proportional to the brand choice probabilities, and thus to the parameters  $\alpha_i$ . Thus means that S, the sum of the alphas, is the only additional parameter that should be specified, and indeed this is the parameter that determines brand loyalty.  $S = 0$  corresponds to absolute brand loyalty, and brand loyalty decreases with increasing S. We define  $L = (1 + S)^{-1}$  as an interpretable brand loyalty parameter, where now  $L = 0$  and  $L = 1$  correspond to the situations of no brand loyalty and absolute brand loyalty, respectively. Given empirical or parametric distributions of consumption and concentration values, the algorithm for chronic exposure assessment now operates as follows:

- 1. Simulate consumptions for a large number  $n$  of individuals.
- 2. Simulate  $n$  selection probabilities from the Dirichlet distribution
- 3. For each individual, simulate  $d$  brand choices from a multinomial distribution using the individual specific selection probabilities from step 2.
- 4. For all individuals and days simulate values from the appropriate concentration distribution.
- 5. Multiply consumption with concentration to obtain exposure.

# **2.2.4 Single value consumptions**

Single value consumption data are the single value amounts (Large Portion, Mean Consumption, p97.5Consumption) of modelled foods (foods-as-measured) consumed in a population.

This module has as primary entities: *Populations Foods*

Output of this module is used by: *Single value dietary exposures*

#### **Single value consumptions da[ta formats](#page-41-1)**

Single value consumptions data pr[ovides a single per-individual-d](#page-194-0)ay and per-food consumption amount for a population. Also the bodyweight should be specified, and optionally also age, sex, and other properties may be recorded. The consumption amounts are usually expressed in grams, but may also be expressed in alternative units of plates, cups, or spoons. Optionally, the uncertainty of food consumption quantifications can be specified, see [Souverein et al., 2011].

## **Single value consumptions**

<span id="page-56-0"></span>Single value consumptions are described using one table: PopulationConsumptionSingleValues.

## **Population consumption single values**

Population consumption single values describe population food consumptions in the form of single value statistics.

Name	Type	Description	<b>Aliases</b>	Required
idPopulation	AlphaNumeric(50)	Unique identification code of	IdPopulation,	Yes
		the population.	PopulationId	
idFood	AlphaNumeric(50)	The unique identification code	idFood,	Yes
		of the consumed food.	FoodCode, Food	
Value type of	Consumption Value-	The value type of this	Consumption-	Yes
the single value	Types	consumption value.	Type,	
consumption			ValueType,	
amount.			Consumption-	
			ValueType,	
			Consumption-	
			SingleValue-	
			Type	
Percentile	Numeric	The percentile (if	Percentile	No
		consumption value type is a percentile).		
Consumption-	Numeric	The consumed amount.	Amount,	Yes
Amount			Consumption,	
			Consumption-	
			Amount.	
			Amount-	
			Consumed	
Consumption-	ConsumptionIntake-	The unit of the consumption	AmountUnit,	N <sub>o</sub>
Unit	<b>Units</b>	amount.	UnitAmount,	
			Consumption-	
			Unit	
Reference	AlphaNumeric(200)	Reference to the source from	Reference,	N <sub>o</sub>
		which this value is obtained.	References.	
			Source, Sources	

Table 2.32: Table definition for PopulationConsumptionSingleValues.

Table aliases: ConsumptionSingleValues, SingleValueConsumptions, PopulationConsumptionSingleValues, PopulationConsumptionValues.

#### **Single value consumptions calculation**

<span id="page-57-0"></span>Single value consumptions can be supplied *as data* or computed. When single value consumptions are computed from *consumptions by modelled food*, then the mean, median and large portion (p97.5 percentile) are computed for all modelled food consumption distributions. Besides these statistics, also the mean bodyweight of the population is computed. The following options are relevant in this calculation:

- Set the *risk type* option to *acute* if the [consum](#page-56-0)ptions should be based on the individual-day distributions. Oth[erwise, choose](#page-115-0) *chronic* to base them on the distributions aggregated by individual.
- Checking the *apply processing factors* option will compute the single value consumptions for the processed foods. When using this option, the output will also show a reverse yield factor, that is the ratio of the quantity of the [raw comm](#page-58-0)odity required to to obtain the processed commodity. Note, when no processing factors are available, the single-value consumption amounts of processed foods are expressed in terms of the processed commodities. [The yield factor, i.c. the](#page-58-0) factor for translating the processed amount to the unprocessed amount, is not applied. In the IESTI calculations (and also chronic single-value calculations), calculations are done using the processed amounts, this is on the level of raw processed foods.
- Check the *restrict population to consumers or consumer days only (modelled-food)* option to compute the single value consumption statistics for each food based on the food consumers only. Note that checking this option

will also affect the computed bodyweight, which is then computed by food based on the food-consumers only and can be different for each food.

- There is also an option to *ignore sampling weights* in the calculation.
- Check the *standardise consumption with body weight before calculation of single values or afterwards (with mean bodyweight)* option to compute the single value consumptions from the per bodyweight distribution. If unchecked, the per-person distribution will be used for computing the statistics. Note that although the results are reported per-d[ay, the statistics are estab](#page-58-0)lished by multiplying the statistics obtained from the per bw distributio[n by the bodyweight.](#page-58-0)

#### **Single value consumptions settings**

#### **Calculation settings**

<span id="page-58-0"></span>

Name	Description
Risk type	The type of exposure considered in the assessment; acute (short
	term) or chronic (long-term).
Restrict population to	Specifies whether the population should be restricted to the
consumers or consumer days	individuals (chronic) or individual days (acute) that have non-zero
only	consumption.
Ignore sampling weights	If checked, individual sampling weights are not used (sampling
	weight $= 1$ ). If unchecked, the specified sampling weights are
	used.
Use standardised consumption	Specifies whether single values are calculated on individual
distributions before calculation	consumptions standardised with body weight and then multiplied
of single values	by the mean body weight. Otherwise, single values are calculated
	on the original consumptions (per day). Note that both methods
	lead to different estimates for the single value.
Apply processing factors	Specified in table ProcessingFactor. If checked, processing factors
	are applied. Concentrations in the consumed food may be
	different from concentrations in the food as measured in
	monitoring programs (typically raw food) due to processing, such
	as peeling, washing, cooking etc. If unchecked, no processing
	information is used. This is in most (though not all) cases a
	worst-case assumption
Restrict population to	Specifies whether the population should be restricted to the
consumers or consumer days	individuals (chronic) or individual days (acute) with consumptions
only (food-as-measured)	containing any of the foods-as-measured.

Table 2.33: Calculation settings for module Single value consumptions.

#### **Single value consumptions as data**

Single value consumption data are the single value amounts of modelled foods (foods-as-measured) consumed in a population.

• *Single value consumptions data formats*

#### **Calculation of single value consumptions**

Single value consumptions are calculated as a percentile (p97.5 or p99) or mean of the modelled food consumption distribution. For an acute single value dietary exposure assessment, this is the individual day consumption distribution, for chronic single value dietary exposure assessment, the individual consumption distribution is used.

• *Single value consumptions calculation*

Inputs used: *Consumptions by modelled food*

Settings used

• *[Calculation Settings](#page-57-0)*

# **2.3 [Occurrence](#page-58-0) modules**

The basic occurrence data are *concentrations* for *substances* in *foods*, sometimes specified separately for a focal food as *focal food concentrations*. In some cases *concentration limits* are used as a stand-in when data are missing.

Concentration data are recalculated (if needed) as *active substance concentrations* in *modelled-foods*. If substance concentrations are not specified directly for the *active substances*, then they are converted using *substance conversions* and/or specified authorised *oc[currence patter](#page-76-0)ns*. [The compo](#page-44-0)sit[ion of](#page-27-1) mixed samples in total diet studies is described in *[total diet study sample com](#page-92-0)positions*. *Food [extrapolation rules](#page-60-0)* specify if insufficient data for a food can be suppleted with data from another food. From these basic data the list of *[modelled-foods](#page-76-0)* is deriv[ed.](#page-94-0)

*Active substance concentrations* in *modelled-foods* [are modelled](#page-202-0) in *concentration models*, optio[nally allowing for](#page-109-0) *occurrence pattern models*. In addition, *[processin](#page-97-0)[g factors](#page-93-0)* and *unit variability factors* can be provided for further use in *di[etary exposure assessment](#page-112-0)*.

# **2.3.1 [Concentra](#page-97-0)[tion d](#page-76-0)i[str](#page-94-0)[ibutions](#page-102-0)**

[Concentration distributions](#page-116-0) describe substance concentrations on foods in the form of summary statistics.

<span id="page-59-0"></span>This module has as primary entities: *Foods Substances*

Output of this module is used by: *Concentration models Dietary exposures*

## **Concentration distributions d[ata fo](#page-27-1)[rmats](#page-44-0)**

## **Concentration distributions**

<span id="page-59-1"></span>Concentration distributions describe substance concentrations on foods in the form of summary statistics. These distributions can be characterised by a mean and a dispersion factor, the standard deviation or, preferably, a percentile point e.g. p95.

## **Concentration distributions**

Substance concentrations on foods specified in the form of summary statistics.

Name	Type	Description	<b>Aliases</b>	Required
idFood	AlphaNumeric(50)	Food code, the raw agricultural commodity.	idFood	<b>Yes</b>
idSubstance	AlphaNumeric(50)	The code of the substance.	idSubstance, SubstanceId, SubstanceCode, Substance	Yes
Mean	Numeric	The mean of (monitoring) samples, on the original scale $(in \, mg/kg).$	Mean	<b>Yes</b>
<b>CV</b>	Numeric	Coefficient of variation, for samples of the size of the TDS pooled amount.	<b>CV</b>	N <sub>o</sub>
Percentile	Numeric	The percentile at the point specified by the percentage.	Percentile	No
Percentage	Numeric	The percentage that belongs to the given the percentile, e.g., 95 (in mg/kg).	Percentage	No
Limit	Numeric	The specified norm value or limit value (in mg/kg).	Limit	N <sub>0</sub>
Concentration- Unit	<b>Concentration Units</b>	The unit of the limit value (default mg/kg).	Concentration- Unit, Unit	N <sub>0</sub>

Table 2.34: Table definition for ConcentrationDistributions.

Table aliases: Concent[rationDistributions,](#page-341-1) ConcentrationDistribution.

## **Concentration distributions as data**

Concentration distributions describe substance concentrations on foods in the form of summary statistics.

• *Concentration distributions data formats*

# **2.3.2 Concentration limits**

<span id="page-60-0"></span>Conce[ntration limits specify \(legal\) limit valu](#page-59-1)es for substance concentrations on foods and are sometimes used as conservative values for concentration data. In the framework of pesticides the legal Maximum Residue Limit (MRL) is the best known example.

This module has as primary entities: *Foods Substances*

Output of this module is used by: *Concentrations Single value concentrations Concentration models Modelled foods*

## **Concentration limits data for[mats](#page-27-1)**

<span id="page-60-1"></span>The concentration limits table de[scribes limit va](#page-76-0)[lues \(e.g., MRLs\) for spec](#page-106-0)i[fic food/substance co](#page-62-0)[mbinations. Thi](#page-94-0)s data may be used, for instance, for the food/substance combinations for which no concentration data is available. The food codes (idFood) and substance codes (idSubstance) should match the codes of the foods and substances table respectively.

## **Concentration limits**

Concentration limits are concentration limit values for specific food and substance combinations originating from regulations (e.g., MRLs). This data may be used, for instance, for the food/substance combinations for which no concentration data is available.

## **Concentration limits**

The food codes (idFood) and substance codes (idSubstance) should match the codes of the foods and substances table respectively.





Table aliases: ResidueLimits, ResidueLimit, MaximumResidueLimits, MaximumResidueLimit, MRLs, MRL.

#### **Concentration limits as data**

Maximum Residue Limits (MRL) are provided as data.

• *Concentration limits data formats*

## **2.3.3 Concentration models**

<span id="page-62-0"></span>Conce[ntration models are distributional](#page-60-1) models of substance concentrations on foods. They describe both the substance presence (yes/no, with no representing an absolute zero concentration) and the substance concentrations. Concentration models are specified per food/substance combination.

This module has as primary entities: *Foods Substances Effects*

Output of this module is used by: *High exposure food-substance combinations Dietary exposures*

#### **Concentration models calcula[tion](#page-27-1)**

There are a number of *concentration model types* [are available. A basic disti](#page-147-0)[nction is between](#page-116-0) using the empirical concentration data (empirical model), fitting a statistical model to the concentration data (parametric model), or to construct a model from (conservative) limit values. Settings relevant for some of these model types as well as other settings are described under *concentration model settings*.

Concentration data from *[Total Diet Studies](#page-62-1)* does not include variability information. In order to include variability in the concentration models created for TDS samples, the variance of the TDS concentrations may be *estimated using variances of the foods maki[ng up the composite TDS food](#page-70-0)*.

#### **Concentration mode[l types](#page-129-0)**

#### <span id="page-62-1"></span>**Empirical model**

Data points are sampled at random from the available set. Non-detects are handled by imputation. If *occurrence patterns* are used, a proportion  $p_0/p_{ND}$  of non-detects is set as 0. See also *concentration models*.

#### **Non-detect spike lognormal model**

A binomial model is used to estimate the proportion  $p$  of positive values (detects). This is just the proportion observed in the data (unless *agricultural use* data have been used to set a proportion of true zeroes). A lognormal model is fitted to the positive data. This provides estimates of  $\mu$  and  $\sigma$ , which are the mean and standard deviation of the natural logarithm of the concentration. Simulated concentrations are a non-detect with probability  $p_{ND} = 1 - p$  or a value sampled from the fitted lognormal distribution with probability  $p$ . Non-detects are handled by imputation. If occurrence pattern[s are used, a pro](#page-97-0)portion  $p_0/p_{ND}$  of non-detects is set as 0. Minimum requirements: at least two positive concentration values. See also *concentration models*.

#### **Non-Detect-Spike Truncated lognormal model**

A binomial model is used to estimate the proportion  $p$  of positive values (detects). This is just the proportion observed in the data (unless agricultural use data have been used to set a proportion of true zeroes in which case  $p$  is calculated on the remaining proportion). A truncated lognormal model, with LOR as the truncation limit, is fitted to the positive data, leading to estimates of  $\mu$  and  $\sigma$ , which are the mean and standard deviation of the natural logarithm of the concentration. Simulated concentrations are a non-detect with probability  $p_{ND} = 1 - p$  or a value sampled from the fitted lognormal distribution with probability  $p$ . Non-detects are handled by imputation. If occurrence patterns are used, a proportion  $p_0/p_{ND}$  of non-detects is set as 0. Minimum requirements: at least two positive concentration values, all non-detects must have one LOR value. See also *concentration models*.









## **NDSpike-TruncLogN**



Figure 2.5: Nondetect Spike Truncated Lognormal distribution

## **Censored Lognormal model**

A censored lognormal model, with LOR as the censoring limit, is fitted to the data, both positives and non-detects. This provides estimates of  $\mu$  and  $\sigma$ , which are the mean and standard deviation of the natural logarithm of the concentration. If agricultural use data are being used, then a proportion  $p_0/p_{ND}$  of non-detects will be excluded, where  $p_0$  will be lowered to  $p_{ND}$  if it would be higher. Simulated concentrations are sampled from the fitted lognormal distribution. If agricultural use data have been used, simulated concentrations are 0 with probability  $p_0$  or are sampled from the fitted lognormal distribution with probability  $1 - p_0$ . Minimum requirements: at least one positive concentration value. See also *concentration models*.

## **Zero-spike censored lognormal model**

A mixture distribution of a spike of true zeroes and a censored lognormal model, with LOR as the censoring limit, is fitted to the data (non-detects and positives. This provides estimates of  $p_0$ , which is the proportion of true zeroes, and of  $\mu$  and  $\sigma$ , which are the mean and standard deviation of the natural logarithm of the concentration. Simulated concentrations are 0 with probability  $p_0$  and are sampled from the fitted lognormal distribution with probability  $1-p_0$ . Minimum requirements: at least one positive concentration value, no agricultural use data for the food-substance combination (which directly specify  $p_0$ , therefore it should not be estimated from the data). See also *concentration models*.



Figure 2.6: Censored Lognormal distribution





#### **Non-detect spike MRL model**

This model simply takes values specified in an input table as Maximum Residue Level (MRL) to be used for the proportion of positive values in the concentration dataset, and can be used to force the use of a pessimistic value.

#### **Summary statistics model**

For this model, no individual measurements on raw agricultural commodities are needed. The final estimates of  $\mu$  and  $\sigma$  are simply provided or pooled or estimated using e.g. a coefficient of variation. Specific use of this model is found in *Total Diet Study* assessments. In general, each TDS food sample is prepared only once, yielding one measurement for a TDS food sample. The variability of the underlying distribution is unknown. However, a rough guess can be made using the e.g. coefficient of variation of the subsamples (in general raw agricultural commodities) that compose the TDS food sample. The estimated standard deviation is *calculated as a pooled estimate using the coefficient of va[riation and the c](#page-129-0)ount of each subsample in the TDS food*.

#### **Concentration models**

Let x denote a random variable from a lognormal distribution. Then, the log transformed variable  $y = ln(x)$  is normally distributed with  $\mu$  and variance  $\sigma$ . The probability density function (p.d.f.) of y may be expressed as:

$$
f_y(y,p_0,\mu_y,\sigma_y^2) = p_0 I(y;0) + (1-p_0)(1-I(y;0)) \cdot \frac{1}{\sqrt{2\pi \sigma_y}} \exp{\frac{(y-\mu_y)^2}{2\sigma_y^2}}
$$

where  $p_0 = Pr(y < log(X_{tor}))$ ,  $x_{tor}$  is the limit of reporting and  $I(y; 0)$  is an indicator function for  $y < log(X_{tor})$ . For  $p_0 = 0$  the p.d.f. of y reduces to the usual lognormal density. The left truncated density for  $y \geq \log(X_{tor})$  may be expressed as:

$$
f_y(y;\mu_y,\sigma_y^2)=\frac{1}{\sqrt{2\pi\sigma_y}}\exp{\frac{(y-\mu_y)^2}{2\sigma_y^2}}/(1-\Phi(z))
$$

with  $\Phi(\cdot)$  the standard normal c.d.f. and  $z = (\log(x_{tor}) - \mu_y)/\sigma_z$ . Model parameters are estimated using maximum likelihood estimation based on the loglikelihood functions specified below. The loglikelihood functions are evaluated in R, using the **optim** algorithm to find estimates for  $\mu_y$ ,  $\sigma_y^2$  and  $p_0$ .

#### **Mixture zero spike and censored lognormal**

The loglikelihood may be expressed as:

$$
\log L(p_0, \mu_y, \sigma_y^2) = \sum_{i=1}^{n_0} \log(p_0 + (1-p_0) \Phi(z_i)) + n_1 \log(\frac{1-p_0}{\sqrt{2\pi\sigma_y}}) - \sum_{i=n_0+1}^{n} \frac{(y_i - \mu_y)^2}{2\sigma_y^2}
$$

where  $y_i = \log(x_i)$ ,  $\Phi(\cdot)$  is the standard normal c.d.f.,  $z = (\log(x_{i,lor}) - \mu_y)/\sigma_y$ ,  $z_{lor} = (\log(lor) - \mu_y)/\sigma_y$  with  $n_0$  number of censored values  $(x_i < x_{i,lor})$ ,  $n_1$  number of uncensored values  $(x_i \ge x_{i,lor})$  and  $x_i$ ,  $i = 1 \cdots n$ .

Multiple values for LOR are allowed.

#### **Censored lognormal**

When  $p_0 = 0$  the loglikelihood reduces to:

$$
\log L(\mu_y,\sigma_y^2) = \sum_{i=1}^{n_0} \log(\Phi(z)) + n_1 \log(\frac{1}{\sqrt{2\pi \sigma_y}}) - \sum_{i=n_0+1}^{n} \frac{(y_i-\mu_y)^2}{2\sigma_y^2}
$$

Multiple values for LOR are allowed.

#### **Mixture non-detect spike and truncated lognormal**

Ignoring the  $n_0$  values below  $x_{lor}$ , the loglikelihood may be expressed as:

$$
\log L(\mu_y, \sigma_y^2) = -n_1\log(1-\Phi(z)) + n_1\log(\frac{1}{\sqrt{2\pi\sigma_y}}) - \sum_{i=n_0+1}^n \frac{(y_i-\mu_y)^2}{2\sigma_y^2}
$$

Only one value for LOR is allowed.

#### **Mixture non-detect spike and lognormal**

Ignoring the  $n_0$  values below  $x_{lor}$ , the loglikelihood may be expressed as:

$$
\log L(\mu_y,\sigma_y^2) = n_1 \log(\frac{1}{\sqrt{2\pi \sigma_y}}) - \sum_{i=n_0+1}^n \frac{(y_i-\mu_y)^2}{2\sigma_y^2}
$$

Only one value for LOR is allowed.

#### **Imputation**

A complication in concentration modelling occurs if results are reported as being below a limit. Different names may be used for such a limit, e.g. limit of detection or limit of quantification. For the purpose of exposure assessment it is only relevant whether results are reported as a positive value or as a non-detect, therefore we refer to any limit as the **Limit Of Reporting** (LOR), and any result reported as '<LOR' is termed a **nondetect**. The value of LOR should always be known for the particular analytical method used.

Non-detects are a very common phenomenon for some classes of substances like pesticides. Non-detects can be handled by replacing them with a given value (**imputation**), or by incorporating them in a parametric model. In the imputation approach, non-detects (values reported less than LOR) can be replaced in simulations by any value between 0 and LOR \* *constant*.

Imputation may be also dependent on the authorisation status of a substance i.c. whether the use of a substance on a agricultural crop is allowed or not.

In Figure 2.8 to Figure 2.11, the various scenarios are displayed. Two substances, Fenamidine and Hexythiazox are indicated with a brown box, these substances are authorized.

#### **N[o imputat](#page-68-0)io[n](#page-69-0)**

#### **Impute all nondetects**

#### **Impute nondetects based on authorized uses**

#### **No imputation except for authorized uses**

#### **Deriving the variance of TDS samples from monitoring**

Variability of TDS food sample concentrations can be derived using *concentration distributions* for the sub-foods of the TDS food samples (defined by the *TDS compositions*). For each sub-food, e.g. *apple* (sub-food of TDS food *FruitMix*), a coefficient of variation (CV) is specified that is derived using the available monitoring samples. Note that monitoring samples may be composite samples. For *apple*, composite food samples are measured and each sample contains, for instance, 12 apples with unit weight 200 g. So monitoring [concentrations,](#page-59-0)  $c_{mi}$ , are based on composite samples with a total weight  $w_{mi} = 2400$  [g each.](#page-112-0)

<span id="page-68-0"></span>

Figure 2.8: Tier 1: Non-detects are not replaced. For Fenamidine and Hexythiazox (brown boxes) authorized use is assumed.



Figure 2.9: All non-detects are replaced by a constant factor x LOR. For Fenamidine and Hexythiazox (brown boxes) authorized use is assumed.



Figure 2.10: Nondetects are replaced by a constant factor x LOR for authorized uses. For Fenamidine and Hexythiazox (brown boxes) authorized use is assumed.

<span id="page-69-0"></span>

Figure 2.11: Tier 1: Nondetects are not replaced except for authorized uses (replaced by a constant factor x LOR). For Fenamidine and Hexythiazox (brown boxes) authorized use is assumed.

A TDS food sample is composed of  $w_i$  g of food i with  $i = 1...k$ ,  $w_i$  represents the *PooledAmount* in *TDS food sample compositions table*. Then, the concentration of a TDS food sample may be represented as:

$$
c_{\text{TDS}} = \sum_{i=1}^k (w_i \cdot c_i) / \sum_{i=1}^k w_i
$$

with variance:

$$
var(c_{\textit{TDS}}) = \sum_{i=1}^k (w_i \cdot var(c_i)) / \sum_{i=1}^k w_i
$$

and  $var(c_i)$  is the variance of concentrations  $c_i$  of food *i* with portion sample size  $w_i$ .

It is expected that increasing the number of units in a composite sample will have a reverse effect on the variation between concentrations. Suppose TDS food *FruitMix* is composed of 2 x 200 = 400 g *apple*. The expected variation between portion sizes of 400 g will be larger than between portion sizes of 2400 g:

$$
var(c_i) = var(c_{mi}) \cdot w_{mi}/w_i
$$

The variance of the monitoring samples are corrected as follows, calculate:

1. 
$$
var(c_{mi}) = \log(CV_{mi}^2 + 1)
$$
  
2.  $var(c_i) = var(c_{mi}) \cdot w_{mi}/w_i$ 

3. 
$$
CV_i = \sqrt{\exp(var(c+i) - 1)}
$$

#### <span id="page-70-0"></span>**Concentration models settings**

# **Calculation settings**

<b>Name</b>	Description
Concentration model tier	Custom model, or set according to EFSA Guidance 2012. Note:
	you may need to set the tier separately in sub-modules.
Default concentration model	The concentration model type that will be used as default for all
	food/substance combinations. If this model type cannot be fitted,
	e.g., due to a lack of data, a simpler model will be chosen
	automatically as a fall-back.
Include MRL fallback model	Use the MRL as fallback model in case the occurrence data is
	insufficient for other concentration modelling options.
Restrict LOR imputation to	Specifies whether imputation of factor x LOR should be limited to
authorised uses	authorised uses only.
Non-detects replacement	How to replace non-detects (when not co-modelled, as in
	censored models).
Factor f (f x LOR)	Replace non-detects by Limit Of Reporting (LOR) times this
	factor. Constant (f), e.g. 0.5.
MRL Factor (f x MRL)	Use f x MRL as concentration estimate of the MRL models.
Sample based	Include co-occurrence of substances in samples in simulations. If
	checked, substance residue concentrations are sampled using the
	correlations between values on the same sample. If unchecked,
	any correlation between substances is ignored, substance residue
	concentrations are sampled ignoring the correlations between
	values on the same sample.
Imputation of missing values	If checked, in procedure of EFSA Guidance 2012, Appendix 1,
	impute missing values using substance based concentration
	models. If unchecked, missing values are not imputed (set to 0).
Correlate imputed values with	If checked, in procedure of EFSA Guidance 2012, Appendix 1,
sample potency	correlate high imputed values with high cumulative potency
	samples. If unchecked, random imputation.
Use occurrence frequencies for	Use of occurrence frequencies (e.g., agricultural use frequencies)
imputation	is relevant for imputation of non-detects in the concentration data.
	Part of the observed non-detects and missing values may be
	imputed with zero when the occurrence frequency is smaller than
	100%. If checked, occurrence frequencies are expected as input
	of this action, otherwise 100% potential presence is assumed for
	all substances on all foods.

Table 2.36: Calculation settings for module Concentration models.

## **Uncertainty settings**

Table 2.37: Uncertainty settings for module Concentration models.

Name	Description
Parametric uncertainty	For resample concentrations: specifies whether the uncertainty
	assessment is based on a parametric approach.
### **Concentration models tiers**

In addition to the possibility for users to work with their own choices for all settings, MCRA implements four tiers from two documents:

- The optimistic and pessimistic basic assessments from the *EFSA 2012 Guidance on the Use of Probabilistic Methodology for Modelling Dietary Exposure to Pesticide Residues* [EFSA, 2012].
- Tier 1 and 2 from the *European Commission working document SANTE-2015-10216 rev. 7 (2018)* on risk management aspects related to the assessment of cumulative exposure [EC, 2018].

## **Overview**

Name	<b>EFSA</b>	<b>EFSA</b>	<b>EFSA</b>	EC 2018	<b>EC 2018</b>
	2012 Op-	2012	2012	Tier 1	Tier <sub>2</sub>
	timistic	Pes-	Pes-		
		simistic -	simistic -		
		Acute	Chronic		
Default concentration	Empirical	NonDe-	NonDe-	Empirical	Empirical
model		tect-	tect-		
		SpikeLog-	SpikeLog-		
		Normal	Normal		
<b>Include MRL fallback</b>	false	true	true	false	false
model					
<b>Restrict LOR</b>		false	false	false	false
imputation to					
authorised uses					
Non-detects	Replace-	Replace-	Replace-	Replace-	Replace-
replacement	<b>ByZero</b>	<b>ByLOR</b>	<b>ByLOR</b>	<b>ByLOR</b>	<b>ByLOR</b>
Factor f (f x LOR)		1	1	0.5	$\overline{0.5}$
<b>MRL</b> Factor (f x		1	1		
MRL)					
Sample based	true	true	true	true	true
Imputation of missing	false	true	true	true	true
values					
Correlate imputed	false	true	true	true	false
values with sample					
potency					
Use occurrence	false			true	true
frequencies for					
imputation					
Parametric uncertainty	false	true	false	false	false

Table 2.38: Tier overview for module Concentration models.

The sections below describe the settings specified by each tier in detail.

# **EFSA 2012 Optimistic**

Use the optimistic model settings according to the EFSA Guidance 2012. Non-detects and missing values are replaced by zero.





## **EFSA 2012 Pessimistic - Acute**

Concentration model settings for acute pessimistic dietary exposure assessments according to the EFSA Guidance 2012. A non-detect spike lognormal model is fitted to the positive residue values and non-detects are replaced by the LOR. When the number of positives is smaller than 2, the maximum residue limit (if available) is used instead. Missing values are imputed.

Name	Setting
Default concentration model	NonDetectSpikeLogNormal
Include MRL fallback model	true
Restrict LOR imputation to authorised uses	false
Non-detects replacement	ReplaceByLOR
Factor f (f x LOR)	
MRL Factor (f x MRL)	
Sample based	true
Imputation of missing values	true
Correlate imputed values with sample potency	true
Parametric uncertainty	true

Table 2.40: Tier definition for EFSA 2012 Pessimistic - Acute.

# **EFSA 2012 Pessimistic - Chronic**

Concentration model settings for acute pessimistic dietary exposure assessments according to the EFSA Guidance 2012. A non-detect spike lognormal model is fitted to the positive residue values and non-detects are replaced by the LOR. When the number of positives is smaller than 2, the maximum residue limit (if available) is used instead. Missing values are imputed.

Table 2.41: Tier definition for EFSA 2012 Pessimistic - Chronic.

Name	Setting
Default concentration model	NonDetectSpikeLogNormal
Include MRL fallback model	true
Restrict LOR imputation to authorised uses	false
Non-detects replacement	ReplaceByLOR
Factor f (f x LOR)	
MRL Factor (f x MRL)	
Sample based	true
Imputation of missing values	true
Correlate imputed values with sample potency	true
Parametric uncertainty	false

# **EC 2018 Tier 1**

# Table 2.42: Tier definition for EC 2018 Tier 1.



# **Input tiers**





# **EC 2018 Tier 2**

Name	Setting
Default concentration model	Empirical
Include MRL fallback model	false
Restrict LOR imputation to authorised uses	false
Non-detects replacement	ReplaceByLOR
Factor f (f x LOR)	0.5
Sample based	true
Imputation of missing values	true
Correlate imputed values with sample potency	false
Use occurrence frequencies for imputation	true
Parametric uncertainty	false

Table 2.44: Tier definition for EC 2018 Tier 2.



# **EFSA 2012 Pessimistic**

**Note:** This tier is deprecated and has been replaced by separate acute/chronic tiers.

Concentration model settings for pessimistic dietary exposure assessments according to the EFSA Guidance 2012. A non-detect spike lognormal model is fitted to the positive residue values and non-detects are replaced by the LOR. When the number of positives is smaller than 2, the maximum residue limit (if available) is used instead. Missing values are imputed.

Name	Setting
Default concentration model	NonDetectSpikeLogNormal
Include MRL fallback model	true
Restrict LOR imputation to authorised uses	false
Non-detects replacement	ReplaceByLOR
Factor f (f x LOR)	
MRL Factor (f x MRL)	
Sample based	true
Imputation of missing values	true
Correlate imputed values with sample potency	true
Parametric uncertainty	true

Table 2.46: Tier definition for EFSA 2012 Pessimistic.

### **Concentration models uncertainty**

When using empirical distributions, concentration model uncertainty is covered by the the inputs. I.e., concentration models can be recomputed from *resampled/bootstrapped* concentration data. This happens for both the univariate concentration models, being recomputed from the bootstrapped residue collections for each food and substance, and also for the samples of the sample-based approach that are re-generated from the bootstrapped samples (including the necessary steps of missing value imputation and imputation of non-detects).

When parametric uncertainty is pr[eferred over empirical bo](#page-12-0)otstrapping, the parameters of the univariate concentration models fitted as a parametric distributions can be *resampled parametrically*.

Let x denote a random variable from the specified distribution. The log transformed variable  $y = ln(x)$  is normally distributed with mean  $\mu_y$  and variance  $\sigma_y$ . The maximum likelihood estimates are  $\hat{\mu}_y$  and  $\hat{\sigma}_y$ . In each bootstrap sample, values are drawn from a normal distribut[ion where the maximum l](#page-13-0)ikelihood estimates are replaced by (  $\hat{\mu}_y^*$ ,  $\hat{\sigma}_y^*$ ).

### **Calculation of concentration models**

Concentration models can be computed from concentration data.

• *Concentration models calculation*

Inputs used: *Concentrations Concentration limits Modelled foods Substance authorisations Occurrence frequencies Relative potency factors Concentration distributions Total diet study sample compositions*

Settin[gs used](#page-62-0)

• *Calcul[ation Settings](#page-76-0)*

# **2.3.4 Concentrations**

<span id="page-76-0"></span>Conce[ntrations data are an](#page-70-0)alytical measurements of chemical substances occurring in food samples. In their simplest form, concentration data can just be used as provided by datasets. Optionally, concentrations data can be manipulated for active substances, extrapolated to other foods, and/or default values can be added for water.

This module has as primary entities: *Foods Substances*

Output of this module is used by: *Single value concentrations Occurrence patterns Concentration models Modelled foods*

### **Concentrations data formats**

[Three](#page-94-0) schemes for data are implemented:

- <span id="page-76-1"></span>1. MCRA scheme: relational tables that can hold all information about Food samples (e.g. sampling date and location), Analytical methods, Analytical method properties for substances (e.g. LOR), Analysis samples (e.g. analysis date) and Concentrations;
- 2. SSD scheme: data according to the EFSA Standard Sample Description (SSD) guideline; SSD data are converted automatically to the MCRA scheme;
- 3. Tabulated data scheme: simplified data format, where samples and analytical methods are not explicitly specified. Tabulated concentration data are converted automatically to the MCRA scheme.

### **Concentration data**

In this group all tables are collected that store information related to concentration or concentration related entities.

### **Sample-based concentration data**

This sub-group contains five tables to specify food samples, analytical methods, their properties for given substances, analyses and concentrations.

### **Analytical methods**

The analytical methods used for analyzing the samples are recorded in the analytical methods table. Each analytical method should have a unique identification code (idAnalyticalMethod). The description field may be used for a more detailed description of the analytical method. The records of this table should be linked to one or more analytical-method-substance records, which record the substances that are measured by this method (and their limits of reporting).





Table aliases: AnalyticalMethod, AnalyticalMethods.

# **Analytical method properties for substances**





Table aliases: AnalyticalMethodSubstances, AnalyticalMethodSubstance.

# **Food samples**

Food sample for analysis of concentrations. May be characterised by location and/or date of sampling. A sample can be analysed multiple times, the results per analysis are stored as analysis samples.





Table aliases: FoodSamples, FoodSample, Samples, Sample, PrimarySample, PrimarySamples.

# **Sample properties**

Food sample properties, additional columns that can also be specified as additional columns in the food samples table

Table 2.50: Table definition for SampleProperties.

Name	<b>vpe</b>	Description	Aliases	Required
Name	AlphaNumeric(50)	The name of the property.	Id	Y es

Table aliases: SampleProperties, SampleProperty.

# **Sample property values**

Food sample property values, additional columns that can also be specified as additional columns in the food samples table

Name	vpe	Description	Aliases	Required
idSample	AlphaNumeric $(50)$	The identification number of	Id	Yes
		the food sample.		
PropertyName	AlphaNumeric(50)	The name of the property.	Name	Yes
<b>TextValue</b>	AlphaNumeric $(50)$	The value of the property as		N <sub>0</sub>
		text value.		
DoubleValue	Numeric	The value of the property as		N <sub>0</sub>
		number.		

Table 2.51: Table definition for SamplePropertyValues.

Table aliases: SamplePropertyValues, SamplePropertyValue.

# **Sample Analyses**

An analysis sample specifies the analysis of a sample by an analytical method. A sample can be analysed multiple times, the results per analysis are stored as analysis samples.





Table aliases: AnalysisSamples, AnalysisSample, SampleAnalysis, SampleAnalyses.

# **Sample concentrations**

The positive concentration values for substances from analysis in the unit specified in table AnalysisSamples. Nondetects (i.e. results 'less than LOR') are not included, their existence can be inferred from the tables AnalysisSamples and AnalyticalMethodSubstances, and the LOR itself from the table AnalyticalMethods.

Name	Type	Description	Aliases	Required
idSample-	AlphaNumeric(50)	The identification number of	idSample-	<b>Yes</b>
Analysis		the analysed sample.	Analysis,	
			SampleAnalysis,	
			idAnalysis-	
			Sample,	
			AnalysisSample-	
			Id	
idSubstance	AlphaNumeric $(50)$	The substance code.	idSubstance.	<b>Yes</b>
			SubstanceId,	
			Substance	
Concentration	Numeric	The measured concentration.	Concentration	<b>Yes</b>

Table 2.53: Table definition for ConcentrationsPerSample.

Table aliases: SampleConcentrations, ConcentrationsPerSample, ConcentrationPerSample.

## **Tabulated concentration data**

Tabulated concentration data provide a simplified concentration data format, where samples and analytical methods are not explicitly specified and analysis results can be tabulated for repeats of the same outcome. This is a convenient data format for single-substance analyses, but it should be noted that it is not possible to use this data in sample-based methods of multiple substances, because it does not record co-occurrence information of substances in samples. Tabulated concentrations data is converted to the internal, relational data format of MCRA.

# **Tabulated concentrations**

In the tabulated concentration data table, each record represents one or multiple samples, and each sample contains a concentration value for a food/substance combination. Non-detects (i.e. concentrations less than LOR) are specified as negative values, i.e. 'less than LOR' should be specified as minus the LOR value. MCRA uses the concept of samples analysed by analytical methods, where the analytical method is characterised by the substances analysed and the LORs for these substances. However, the tabulated data do not provide this information explicitly. Samples are reconstructed from the tabulated records using the NumberOfSamples field to create that number of single substance samples. Analytical methods are reconstructed from the data, with each analytical method having only one analysed substance with a LOR and concentration unit. When a negative concentration value is given (i.e., it is a non-detect measurement), this value is recorded as the LOR (negated). All non-detect measurements of the same substance with the same LOR and concentration unit are linked to the same analytical method. When a positive concentration value is given, this value is recorded as the measured concentration of the sample. All positive measurements of the same substance are linked to the same analytical method that has an artificial LOR that is smaller than the lowest positive concentration. When a concentration of 0 (zero) is given, the measurement is considered to be a non-detect measurement and the LOR is set as default to the value 1E-08.

Name	Type	Description	<b>Aliases</b>	Required
<b>GUID</b>	AlphaNumeric $(50)$	Unique identifier of the	idAnalysis-	N <sub>o</sub>
		analysis sample of this	Sample,	
		tabulated concentration	SampleId,	
		record.	SampleCode,	
			Code, Id	
idSubstance	AlphaNumeric(50)	The code of the substance of	idSubstance,	Yes
		this concentration value.	SubstanceId.	
			Substance	
idFood	AlphaNumeric(50)	The food code.	idFood, FoodId,	Yes
			FoodMeasured,	
			Food	
DateSampling	AlphaNumeric(10)	The date of sampling.	DateSampling	No
SamplingType	AlphaNumeric(50)	The type of sampling	SamplingType	No
		(monitoring).		
Location	AlphaNumeric $(50)$	The location or country of	Location.	No
		sampling.	Country	
NumberOf-	Integer	The count of the number of	NumberOf-	Yes
Samples		times the specified	<b>Samples</b>	
		concentration or limit of		
		reporting (LOR) occurs.		
Concentration	Numeric	The concentration or LOR.	Concentration,	Yes
		LORs are specified using a	Value	
		$minus (-)$ sign.		
Concentration-	<b>Concentration Units</b>	The unit of the specified	Concentration-	No
Unit		concentrations/LORs (default	Unit,	
		$mg/kg$ ).	Unit	

Table 2.54: Table definition for ConcentrationTabulated.

Table aliases: Concent[rationTabulated, Co](#page-341-0)ncentrationValues, TabulatedConcentrations, TabulatedConcentration.

## **EFSA SSD concentration data**

MCRA provides an option to upload concentration data that is formatted according to the EFSA Standard Sample Description (SSD) guideline. SSD formatted concentrations data is converted to the internal, relational data format of MCRA.

### **SSD concentrations**

MCRA uses the concept of samples analysed by analytical methods, where the analytical method is characterised by the substances analysed and the LORs for these substances. However, the SSD data do not provide information on the analytical methods at this level of detail. Therefore, from the provided SSD records, analytical methods are reconstructed and samples are linked to these analytical methods. All SSD records with the same labSampCode and labSubSampCode are considered to be from the same sample. All SSD samples that have records for the same substances, with the same LOQ/LOD values and resUnit are considered to originate from the same reconstructed analytical method. If both LOQ and LOD are provided, LOQ is used as LOR of the reconstructed analytical method. It is highly recommended to supply LOQ/LOD values, even for positive measurement, because this reduces the number of reconstructed analytical methods.





Table aliases: ConcentrationsSSD, SSDConcentrations.

# **Concentrations calculation**

Occasionally, concentrations of substances measured in food samples are exceeding a specified concentration limit e.g. the *Maximum Residue Limits* (MRL). An MRL is the highest level of a substance that is legally tolerated in or on food or feed when substances are applied correctly. *Filter samples* exceeding the concentration limits filter out all samples where one of the substances measured is exceeding the *MRL*.

*Substance conversions* rules may be used to *convert* concentration data at the level of measured substances to concentrati[on data at the level of pot](#page-61-0)entially active substances. These rules (provided as data) may be applicable, for example, when a measured substance represents multi[ple substances](#page-83-0) and these measurements should be converted into measurement values for these substances. This conversion [may d](#page-61-0)epend on *substance authorisations* which pro[vides information on](#page-109-0) the likelihood of certa[in trans](#page-83-1)lations to occur. *points of departure* or *relative potency factors* might be needed when the substance conversion should select the most toxic candidate in case a measured substance translates to multiple active substances.

If there are only a few measurements in the concentration data, then *extrapolation of concentration data* may be desired. In that case, *food extrapolation rules* may be provided to specify per food the alternative foods from which extrapolation is allowed. The extrapolation of concentrations will then be performed within this module and the results are included in the resulting active substance concentrations data. *Substance authorisations* and/or *concentration limits* may be used to further restrict the to-food/from-food combinations per [substance for which extrapolation is p](#page-85-0)ossible.

Concentration data fo[r water are often not ava](#page-93-0)ilable in the concentration data, but it may be desirable to include them in the assessments. For this, *imputation* of low-tier, deterministic estimates of water concentratio[ns of the most toxic](#page-60-0) substances may be used to include (typically conservative) estim[ates in the calculations.](#page-108-0)

In some scenarios it may be desired to perform a prospective analysis in which anticipated (or foreground) *focal commodity concentration data* for a particular focal commodity food (and substance) is added to, or replaces part of the background concentr[ation data t](#page-85-1)hat is used for the null-scenario. The concentrations module offers various options to perform such *focal commodity scenario analyses*.

# **[Sample filtering](#page-92-0)**

<span id="page-83-0"></span>If the option **Filter samples exceeding the concentration limits** is checked, all samples with one or more substance concentrations exceeding the *MRL* are filtered out. Then a **concentration limit filter exceedance factor** (factor)isspecified, which filtersoutsamples with at least one substance concentration higher than :  $math:$ 

If the option **Filter samples exceeding t[he con](#page-61-0)centration limits** remains unchecked all samples are retained in the analysis.

# **Substance conversion**

<span id="page-83-1"></span>When concentration data at the level of measured substances have to be converted to concentration data at the level of *active substances* (or perhaps also inactive substances), then *substance conversion rules* can be specified to provide the rules. This section first describes the basic substance conversion, and then the refinements using available *substance authorisations*.

For each measured substance in the concentration data, there may be zero or more conversion rules (records in the [substance conver](#page-202-0)sion rules data source), each linking to an [active or inactive substanc](#page-109-0)e. Substance conversion rules may specify a link to an exclusive substance or not. For an exclusive conversion it is assumed that only one [substance](#page-108-0) [is present in th](#page-108-0)e sample, therefore the measured substance is considered to be just one of the linked substances. It can also be that measured substances link to one or more exclusive substances plus one (non-exclusive) substance that is considered a metabolite of the other exclusive substances. The metabolite can occur together with any of the exclusive substances. It is assumed that either all conversion rules linked to a measured substance are marked as exclusive (case 1), or precisely one rule is marked as exclusive and the other rules are marked as not exclusive (case 2). If this is not the case for any set of rules linked to a measured substance, then this is regarded as erroneous data.

Four methods are implemented for substance conversion:

**1. Allocate most potent (EC 2018 Tier 1):** For each measured substance, the linked substances are restricted to the active substances of interest. The concentration of the measured substance is assigned to the most potent active substance in this set. Potency is specified by the *relative potency factors*. All other candidate active substances are assigned a zero concentration. I.e., the measured substance concentration is allocated for 100% to the most potent substance specified by the conversion rules and for this allocation, the concentration or LOR is multiplied by the molecular weight correction factor. See *EC2018 Tier 1*.

**2. Random allocation (EC 2018 Tier 2):** One [of the conversion rules i](#page-237-0)s drawn randomly (with equal probability), including the rules of both active and other substances. This drawn rule is used as follows to generate active substance concentrations:

• **If the drawn conversion r[ule is marked](#page-145-0) as exclusive**, the concentration or LOR is allocated to the linked substance.

• **If the drawn conversion rule is marked as not exclusive**, a proportion *p*, specified by the drawn conversion rule, of the concentration or LOR is allocated to the linked substance. The remaining proportion *(1-p)* is allocated to one other substance, which is the substance that is linked to the measured substance in a conversion rule marked as exclusive (in this case it is assumed that precisely one record per measured substance is marked as exclusive).

All assigned concentrations are multiplied by the molecular weight correction factor. All unselected candidate substances are assigned a zero concentration. See *EC2018 Tier 2*.

**3. Nominal estimate:** The substances specified through the conversion rules are allocated with a nominal value based on all possible conversion rules. This may be regarded as the nominal or average allocation value of the random sampling method.

- **All conversion rules are marked as exclusive:** The meas[ured substance](#page-145-1) concentration is divided over all *n* active substances specified with equal proportions *1/n*, accounting for the molecular weight correction factor for all substances.
- **Precisely one conversion rule is marked exclusive and n conversion rules are marked as not exclusive:** The measured substance concentration is divided over all active substances specified, with a proportion *1/2 + 1/n* for the substance belonging to the exclusive conversion rule, and equal proportions *1/n* for the other substances, accounting for the molecular weight correction factor for all substances.

**4. Allocate all:** The concentration of a measured substance is allocated to each active substance associated with the measured substance as if it were the most potent substance. I.e., the same measured substance is allocated to all associated active substances simultaneously. This method is not sensible when using it in a cumulative assessment, but it is of use in substance screening assessments, where in a combined analysis of multiple substances all active substances are considered independently.

### **Use of substance authorisations in substance conversion**

When *substance authorisations* are available, then these can be used to exclude conversions of measured substances to unauthorised substances on a given food. The information is used as follows in the substance conversion procedures:

**1. Allocate most potent:** The set of candidate active substances from which the most potent active substance is to be drawn is reduced to only the substances with authorised uses. However, if none of the [candidate active substanc](#page-108-0)es is authorised, then the most potent of the unauthorised substances is selected for active substance allocation.

**2. Random allocation:** The set of conversion rules from which to draw is reduced to the rules linking to authorised substances or the non-exclusive substance (thus allowing the selection of a possibly unauthorised metabolite of an authorised substance). If none of the conversion rules links to an authorised substance, then one rule is drawn from the full set of all (unauthorised) conversion rules.

**3. Nominal estimate:** The set of conversion rules is reduced in the same way as in *Tier 2*. Nominal calculation is performed on the resulting set of conversion rules.

**4. Allocate all:** For this method, the same rules apply as for *allocate most potent*. The set of candidate active substances that are to be allocated is reduced to only the substances with authorised uses. Hence, a substance is not allocated when it is not authorised and there is at least one other [candid](#page-145-1)ate active substance that is authorised. However, if none of the candidate active substances is authorised, then the most potent of the unauthorised substances is selected for active substance allocation.

## **Food extrapolation**

If the *food extrapolation* setting has been checked, extrapolation of concentrations is performed for all food/*active substance* combinations for which:

- <span id="page-85-0"></span>1. the number of measurements in the analytical scope is smaller than a given threshold for extrapolation (default 10), and
- 2. [ther](#page-202-0)e is an *extrapolation rule* allowing extrapolation of concentrations from one or more other foods (the [from](#page-202-0)food(s)) to the given food (the to-food), and
- 3. (optional criterion:) the substance is associated with *authorised use* for both foods, and
- 4. (optionalc[riterion:\)](#page-93-0) *concentration limits (e.g. MRLs)* on the from-food and to-food exist and are equal. Note: if the **active substance** is not a **measured substance**, then the MRL check has to be made per measurement at the level of the measured substance which provid[ed the concentr](#page-97-0)ations assigned to the active substance.

Food extrapolation is performed by one of the following procedures: 1) Substance-specific imputation of missing values by extrapolated me[asurements, or 2\) Extrapolation o](#page-60-0)f complete samples for multiple substances.

# **1. Substance-specific imputation of missing values by extrapolated measurements**

The missing values in the active substance concentrations of the to-food are *imputed* in a random order by active substance concentrations (positive, nondetect or zero) from a randomised list obtained from the fromfood(s). By matching the randomised lists, each from-food measurement is assigned at most once, so after extrapolation there may still be missing values left, or not all measurements of the from-food(s) may have been used for extrapolation.

Note: In this method, it is assumed that the to-food has a sufficient number of [samples](#page-67-0). No extrapolation is applied for foods with no samples at all, and data gaps will also remain for foods with fewer than *n* samples, because no new samples are added.

Note: the resulting *occurrence patterns* will be random with respect to the extrapolated substances, i.e., observed occurrence patterns for the from-food are not extrapolated to the to-food.

# **2. Extrapolation [of complete sam](#page-97-0)ples for multiple substances**

### (not yet implemented)

All samples of the from-food(s), i.e., complete samples with data for all active substances, are copied as samples for the to-food and added to the existing to-food samples. For example, extrapolate all apple sample records to the available pear sample records. However, measurements for substances that do not fulfil the (optional) criteria 3 and 4 above are non-valid extrapolations and are replaced by missing values. The status of the extrapolated samples is stored to distinguish between extrapolated and non-extrapolated sample records. Note that this method maintains correlations in the occurrence patterns and postpones imputation of MVs until the concentration models step.

### **Water imputation**

<span id="page-85-1"></span>If water has been selected as an additional source of exposure, but concentration data is missing, then, fixed concentration values can be assigned to water for the five most toxic *active substances*, with the toxicity ranking being based on the *relative potency factors*. For all other substances, zero concentrations are *imputed*. The default imputation value is 0.05 µg/L, but this value can be chosen as a setting. If specified, *substance authorisations* may be used to restrict to the set of active substances for which water concentrations are imputed to only those for which concentrations may be expected from *authorised use*.

### **Focal commodity scenario analysis**

There are different methods for modifying the (background) concentration data for specific (prospective) focal commodity scenario analyses. In the front end, these focal commodity scenario analysis method are accessible through the option *include focal commodity concentrations*. Checking this option will open the focal commodity scenario analysis form (see Figure 2.12 where the method and the focal commodity food/substance can be selected, and accompanying other settings can be configured.



Figure 2.12: Focal commodity scenario analysis form of the front end. This form is a sub-form of the concentrations module panel.

### **Replace samples with focal commodity samples**

This method will replace all samples for the selected focal commodity food by samples from the *focal commodity concentration data*. It works substance independent, and will therefore replace all substance concentrations of the focal commodity food in the background concentration data.

# **[Append focal co](#page-92-0)mmodity samples**

This method adds the *focal commodity samples* of the selected focal commodity food to the background concentration data. This method is also substance-independent and may be a useful approach when the substances measured in the field trial do not overlap with the substances of the (background) concentration data. In this case, the focal commodity substance concentrations will be missing for the background concentration data and (also the other way around) the substance concentrati[ons of all other substance](#page-92-0)s will be considered missing for the focal commodity samples. These missing values may be imputed at a later stage following the "normal procedures".

### **Replace measurements of focal food/substance combinations**

This method replaces, for the selected (focal) combination of food and substance, all substance concentrations with focal concentrations. This method knows two variants:

- **Replace by focal commodity samples:** The focal food/substance measurements are obtained from *focal commodity samples*. Here, substance measurements of the focal commodity food in the background concentration data set are replaced by randomly assigned substance measurements of the focal commodity samples.
- **Replace by concentration limits:** The focal food/substance measurements are obtained from *focal commodity samples*. Here, substance measurements of the focal commodity food in the background concentrat[ion data set](#page-92-0) [are replaced by](#page-92-0) the concentration limit value (e.g., an MRL) obtained from the provided *concentration limits data*.

Using the *focal commodity substance occurrence percentage*, it is possible to specify an occurren[ce percentage for](#page-92-0) the co[mbinatio](#page-92-0)n of focal food and substance. When this percentage is less than 100%, this will partly (i.e., for the selected percentage) replace the concentrations of the focal commodity food and substance wit[h the focal concen](#page-60-0)tratio[ns, an](#page-60-0)d for the other part replace the concentrations with zero concentrations. E.g., when aiming to replace backgroun[d concentrations of the substance fluopyram on po](#page-88-0)tatoes with an MRL value, then specifying a focal commodity substance occurrence percentage of 40% will replace 40% of the measurements with the MRL, and 60% of the measurements with zero concentrations. Note that, because the allocation is random (i.e., each substance measurement has a probability of being assigned a focal concentration or a zero defined by the percentage), the realized replacement percentage may differ from the specified percentage. This option can, for example, be used to simulate a percentage of agricultural use.

Using the *adjustment factor for the focal food/substance concentration*, it is possible to adjust the (positive) concentrations of the focal food and substance measurements. This factor can be used when the focal commodity concentrations (e.g., from field trials) are assumed to be higher than what may be reasonably expected in practice. In this case, this factor could be set for instance, to the expected ratio of mean monitoring concentration and mean field trial concentratio[n. Note that for replacement by focal commodity measure](#page-88-0)ments, this factor will only adjust the positive concentrations and not the LORs.

By default, the focal commodity substance measurements are replaced before the optional step of *converting the concentrations from measured to active substance concentrations*. This also means that for these replaced measurements, the same rules apply, and the measurements may be converted to active substance measurements after replacement. Alternatively, it is possible to replace substance measurements after having done the allocation, and to use *deterministic substance conversions factors* for the focal commodity food and substance to convert these m[easurements to the](#page-83-1) level of *[active substances](#page-83-1)*.

Note that when also using using *substance authorisations*, the focal food and substance combination will be treated as authorised, even if there is no authorisation supplied for the combination. The approved authorisatio[n status is](#page-91-0) [considered to be part of this scena](#page-91-0)rio analysis.

### **Remove measurements of focal food/substance combinations**

This method will simply remove all background concentrations for the selected focal commodity food and substance combination, and will not replace them with other values. This method may be useful when a separate analysis is desired for the background and foreground concentrations.

# <span id="page-88-0"></span>**Concentrations settings**

# **Selection settings**

Concentrations tier Specifies the concentration data should be treated according to a pre-defined tier or custom.
If checked, samples with at least one substance concentration Filter samples exceeding the
concentration limits higher than some factor (concentration limit filter exceedance
factor) times the MRL are filtered out.
Concentration limit filter The multiplication factor for the concentration limit exceedance
exceedance factor filter.
If checked, concentrations are modelled in terms of active Use substance conversion rules
substances (using substance conversion).
Allocation method for assigning active substance concentrations Substance conversion method
from measured substance concentrations based on substance
translations.
Retain all allocated substances If checked, all allocated substances kept after substance
after active substance allocation conversion. Otherwise, the concentration data is restricted to the
active substances of the assessment group.
Account for substance authorisations when allocating measured Account for substance
authorisations in substance substances to active substance using substance conversions.
conversions
Use extrapolation rules Use extrapolation rules to extrapolate food samples for foods with
a limited amount of samples (data poor foods) from other foods
(data rich foods).
Threshold for extrapolation. Threshold for extrapolation
Restrict extrapolations to equal Restrict extrapolations to equal MRLs.
<b>MRLs</b>
Only extrapolate if substance use is authorised. Restrict extrapolations to
authorised uses
Impute constant concentration values on the selected (water) Impute water concentrations
commodity.
The commodity for which constant concentration values should be Water commodity
added. Water concentration value Constant concentration value that should be used for water (in
$\mu$ g/kg). $(\mu g/kg)$ Restrict water imputation to the five most toxic substances. Restrict water imputation to
the five most toxic substances
Restrict water imputation to authorised uses. Restrict water imputation to
authorised uses
Include focal commodity Specifies whether there is monitoring data that should replace part
concentrations of the consumption data for the specified focal commodities.
Replacement method to be used for replacing base concentration Focal commodity
data with concentration data of the focal commodity/commodities concentrations replacement
method concentrations.
Focal commodity substance Anticipated occurrence percentage / agricultural use percentage
occurrence percentage of the focal commodity.
Adjustment factor for the focal Optional adjustment factor for the focal food/substance
food/substance concentration concentration. E.g., the expected ratio of mean monitoring
concentration and mean field trial concentration.
Use deterministic substance Convert measured substance concentrations of focal commodity
conversions for focal
to active substance concentrations using deterministic substance conversion factors.
commodity Modelled-foods subset: restrict
If checked, then the assessment is restricted to the specified modelled foods.
to specific modelled-foods The locations for which samples are filtered. Sample locations

Table 2.56: Selection settings for module Concentrations.

# **Uncertainty settings**





# **Concentrations tiers**

In addition to the possibility for users to work with their own choices for all settings, MCRA implements Tier 1 and 2 from the European Commission working document SANTE-2015-10216 rev. 7 (2018) on risk management aspects related to the assessment of cumulative exposure.

# **Overview**



# Table 2.58: Tier overview for module Concentrations.

# **EC 2018 Tier 1**







### Table 2.60: Tier definition for EC 2018 Tier 2.

### **Concentrations uncertainty**

Uncertainty due to a limited number of samples can be accounted for by resampling/bootstrapping. Resampling is done on a sample-based basis preserving co-occurrence of substance residue values on the same sample for multiplesubstance analyses.

### **Concentrations as data**

Concentration data can be entered using the internal, relational data format or using the EFSA SSD format. Depending on the settings, the entered concentration data can be pre-processed for conversion to active substances, extrapolation to other foods, and/or default values can be added for water.

- *Concentrations data formats*
- *Concentrations calculation*

Inputs used: *Focal food concentrations Food extrapolations Substance conversions Deterministic substance conversion factors [Relative potency factors](#page-76-1) Substance authorisations Active substances Concentration limits*

# **2.3.5 De[terministic substa](#page-92-0)[nce conversio](#page-93-0)[n factors](#page-109-0)**

[Determ](#page-91-0)[inistic substance conver](#page-237-0)[sion factors.](#page-108-0)

<span id="page-91-0"></span>This module has as primary entities: *Substances Foods*

Output of this module is used by: *Concentrations Single value concentrations*

## **Deterministic substance conv[ersion fa](#page-44-0)[ctors d](#page-27-0)ata formats**

### **Deterministic substance co[nversion fact](#page-76-0)[ors](#page-106-0)**

<span id="page-91-1"></span>Deterministic substance conversion factors. Foods are optional.

### **Deterministic substance conversion factors**

Deterministic substance conversion factors for translating measured substance concentrations to active substance concentrations.



Table 2.61: Table definition for DeterministicSubstanceConversionFactors.

Table aliases: SingleValueSubstanceConversionFactors, SingleValueConversionFactors, SingleValueConversions, SubstanceConversionsFixed, DeterministicSubstanceConversionFactors.

### **Deterministic substance conversion factors as data**

Deterministic substance conversion factors.

• *Deterministic substance conversion factors data formats*

# **2.3.6 Focal food concentrations**

<span id="page-92-0"></span>In so[me cases the attention in an assessment is to evaluate co](#page-91-1)ncentrations (e.g., from specific field trials) for a specific food (and substance), in combination with a background of concentration data for other foods. Focal food concentrations can be included to provide these separate (foreground) concentration data for one or more focal food commodities that should replace measurements in the (background) *concentration data* in *focal commodity scenario analyses*.

This module has as primary entities: *Foods Substances*

Output of this module is used by: *Concentrations*

### **Focal food concentrations data formats**

See *concentration data formats*.

# <span id="page-93-1"></span>**Focal food concentrations settings**

# **Sel[ection settings](#page-76-1)**





# **Calculation settings**





# **Focal food concentrations as data**

Focal food concentrations are concentration data and specified in the exact same manner. The difference is that this data will be used to replace part of the concentration data in order to combine specific concentration data with a background of ordinary concentration data.

• *Focal food concentrations data formats*

# **2.3.7 Food extrapolations**

<span id="page-93-0"></span>Food [extrapolations data specify which foods](#page-93-1) (data rich foods) can be used to impute concentration data for other foods with insufficient data (data poor foods).

This module has as primary entities: *Foods*

Output of this module is used by: *Concentrations Food conversions*

# **Food extrapolations data form[ats](#page-27-0)**

# **Food extrapolation rules**

<span id="page-93-2"></span>Food extrapolations (or read-across food translations) can be used to specify whether data (e.g, occurrence data) on a food for which this is missing (a data poor food) may be extrapolated from another food for which data is available (read-across food).

## **Food extrapolations**

Food extrapolations are simply specified as combinations of two food codes. One code for the food for the data poor food, and one for the data rich food (or read-across food).

Name	Type	Description	Aliases	Required
DataPoorFood	AlphaNumeric(50)	The code of the data poor	IdFoodData-	Yes
		food. I.e., the food for which	Poor,	
		missing data is allowed to be	FoodDataPoor,	
		extrapolated.	idFromFood,	
			FromFoodId,	
			FromFood,	
			FoodFrom,	
			Food, IdFood	
CodeDataRich-	AlphaNumeric $(50)$	The code of the read-across	IdFoodData-	Yes
Food		food (or data rich food). I.e.,	Rich,	
		the food from which data is	FoodDataRich,	
		used for extrapolation.	IdFoodRead-	
			Across,	
			FoodRead-	
			Across.	
			IdReadAcross-	
			Food,	
			ReadAcross-	
			Food, idToFood,	
			ToFoodId,	
			ToFood, FoodTo	

Table 2.64: Table definition for FoodExtrapolations.

Table aliases: ReadAcrossFoodTranslations, ReadAcrossFoodTranslation, ReadAcrossTranslations, ReadAcrossTranslation, FoodExtrapolations, FoodExtrapolation.

### **Food extrapolations as data**

Food extrapolations are specified as data in the form of simple tuples of data rich food and data poor food for which extrapolation is allowed/reasonable.

• *Food extrapolations data formats*

# **2.3.8 Modelled foods**

Mode[lled foods are foods within the fo](#page-93-2)ods scope for which concentration data or MRLs of substances are available (or expected).

<span id="page-94-0"></span>This module has as primary entities: *Foods Substances*

Output of this module is used by: *Concentration models Food conversions*

### **Modelled foods calculation**

<span id="page-95-0"></span>Modelled foods are the foods within the foods scope for which concentration data or MRLs of substances are available (or expected). Modelled foods are derived primarily from *concentration data*. That is, all foods for which food samples are available in the concentration data or MRL data are considered to be modelled foods. In addition, this set may be extended when *concentration limits* such as MRLs are available (see *calculation settings*) and/or when *food extrapolation rules* are used. Foods for which such data is available are considered to be modelled foods. The set of foods can also be restricted by omitting foods with only non-[detect measuremen](#page-76-0)ts (see *calculation settings*).

# **[Modelled foods](#page-93-0) setti[ngs](#page-60-0)**

# **Calculation settings**

<span id="page-95-1"></span>

### Table 2.65: Calculation settings for module Modelled foods.

# **Calculation of modelled foods**

Modelled foods are computed from concentration data (which may also be in the form of single-value concentrations) and/or derived from available maximum residue limits.

• *Modelled foods calculation*

Inputs used: *Concentrations Single value concentrations Concentration limits*

Settings used

• *[Calculation Settings](#page-95-0)*

# **2.3.9 Occurrence frequencies**

Occur[rence frequencies sp](#page-95-1)ecify how often substances occur on foods. Frequencies are expressed as percentages.

This module has as primary entities: *Foods Substances*

Output of this module is used by: *Concentration models Single value dietary exposures*

## **Occurrence frequencies data formats**

## **Occurrence frequencies**

<span id="page-96-0"></span>Occurrence frequencies are described by one simple table, specifying for pairs of food and substance, the associated occurrence frequencies as percentages.

### **Occurrence frequencies**

Occurrence frequencies are specified as percentages for pairs of food and substance. Optionally, a reference can be included in each record to specify the source (e.g., from literature) from which the percentage was obtained.

Name	Type	Description	Aliases	Required
idFood	AlphaNumeric(50)	The food code.	idFood.	Yes
			CodeFood,	
			FoodId.	
			FoodCode, Food	
idSubstance	AlphaNumeric(50)	Code of the substance.	idSubstance,	Yes
			CodeSubstance.	
			SubstanceId,	
			SubstanceCode.	
			Substance	
Percentage	Numeric	The occurrence frequency	Percentage,	Yes
		percentage.	Frequency-	
			Percentage	
Reference	AlphaNumeric(200)	Reference to the source from	Reference,	N <sub>o</sub>
		which this use frequency	References,	
		value is obtained.	Source, Sources	

Table 2.66: Table definition for OccurrenceFrequencies.

Table aliases: OccurrenceFrequencies.

### **Occurrence frequencies calculation**

<span id="page-96-1"></span>Occurrence frequencies can be provided as data or computed from *occurrence patterns*. Occurrence frequencies for a food and substance are computed by collecting all occurrence patterns of this food and summing up the frequencies of the occurrence patterns containing the substance. In the unlikely case that the total frequency of the occurrence patterns of a food exceeds 100%, then a rescaling is applied first. For foods for which the sum of the frequencies of the occurrence patterns does not sum up to 100%, the interpretati[on of the remaining](#page-97-0) unspecified percentage can be "no use", assuming that none of the substances occur on this remaining percentage, or (more conservatively) "all use", assuming all of the substances occur on this remaining percentage. This choice is available as the setting *associate the unspecified percentage with no-occurrence for foods with at least one specified occurrence pattern*.

Depending on the setting *apply occurrence pattern percentages*, occurrence frequencies can be computed in a crisp form in which the occurrence frequency is either 0% or 100% or as percentages ranging from 0% to 100%[.](#page-97-1)

### **Selection settings**

<span id="page-97-1"></span>

Name	Description		
Associate the unspecified	If checked, for foods with at least one specified occurrence		
percentage with no-occurrence	pattern, unspecified occurrence patterns for the same food are		
for foods with at least one	assumed to be associated with no use. If unchecked, all		
specified occurrence pattern	substances are considered to be authorised (potentially present in		
	samples). Note that this setting cannot be used for foods that have		
	no specified AUs. These foods have 100% potential presence of		
	all substances. To declare all AUs on such a food un-authorised,		
	include an empty AU with percentage 100% in the AU data table		
	(i.e., use an AU for this food, without specifying substances in the		
	AU Substances table)		
Apply occurrence pattern	If checked, use the percentages of potential presence as specified		
percentages	by the occurrence patterns. If unchecked, 100% potential		
	presence in samples is assumed for all substances identified by the		
	occurrence patterns.		

Table 2.67: Selection settings for module Occurrence frequencies.

### **Occurrence frequencies as data**

Occurrence frequencies are described by one table, specifying for a food and substance the associated occurrence frequency as percentage.

• *Occurrence frequencies data formats*

Inputs used: *Active substances*

### **Calc[ulation of occurrence frequenc](#page-96-0)ies**

Occurrence [frequencies for a f](#page-202-0)ood and substance are computed according to the model that is part of the EC 2018 Tier II defintion (see van Klaveren et al. 2019)

• *Occurrence frequencies calculation*

Inputs used: *Occurrence patterns*

# **2.3.10 [Occurrence patterns](#page-96-1)**

<span id="page-97-0"></span>Occurrence [patterns \(OPs\) are t](#page-97-0)he combinations (or mixtures) of substances that occur together on foods and the frequencies of these mixtures occurring per food, expressed in percentages. In the context of pesticides, occurrence patterns are associated with agricultural use percentages. Occurrence patterns are relevant to account for co-occurrence of active substances in exposed individuals. Occurrence patterns may be specified as data or modelled based on observed patterns of positive concentrations.

This module has as primary entities: *Foods Substances*

Output of this module is used by: *Occurrence frequencies Dietary exposures*

# <span id="page-98-0"></span>**Occurrence patterns data formats**

# **Agricultural uses**

Agricultural use percentages for plant protection products (PPPs) may be of use for concentration modelling, as they provide information about what substance mixtures are expected to be present simultaneously on food samples. Especially for non-detect concentration measurements, this information may aid to determine whether the non-detect measurement originated from a true zero or may be a positive measurement below the limit of detection. Agricultural use percentages are specified using the agricultural uses and agricultural use substances table. This data format expects agricultural use percentages to be specified for mixtures of substances. Each mixture has an id (idAgriculturalUse) and a list of substances that are part of this mixture (agricultural use substances). These agricultural uses are assumed to be exclusive (i.e., only one mixture or PPP is used per sample). Hence, the sum of the agricultural uses for one food should not exceed 100%.

# **Agricultural uses**

The AgriculturalUses contains the definitions of the agricultural use mixtures, or PPPs and the specification of the percentage of the products treated with this mixture. Optionally also the time period of the use percentage may be specified.

Name	Type	Description	Aliases	Required
idAgricultural-	AlphaNumeric(50)	The unique identification code	idAgricultural-	<b>Yes</b>
Use		of the agricultural use group /	Use,	
		plant protection product	AgriculturalUse-	
		(PPP).	Id,	
			Id	
idFood	AlphaNumeric(50)	The food code.	idFood, FoodId,	Yes
			Food	
Location	AlphaNumeric(50)	The location or country code,	Country,	N <sub>o</sub>
		agricultural use location.	Location	
<b>StartDate</b>	<b>DateTime</b>		<b>StartDate</b>	No
EndDate	DateTime		EndDate	N <sub>o</sub>
Percentage-	Numeric	The percentage agricultural	PercentageCrop-	Yes
CropTreated		use $(\%).$	Treated,	
			Percentage,	
			PercCrop-	
			Treated,	
			PercentageUse	

Table 2.68: Table definition for AgriculturalUses.

Table aliases: AgriculturalUses, AgriculturalUse.

### **Agricultural use substances**

The agricultural use substances table records the substances that are part of the agricultural use mixtures (PPPs).

Name	Type	Description	Aliases	Required
idAgricultural-	AlphaNumeric(50)	The agricultural use code,	idAgricultural-	<b>Yes</b>
Use		normally a code for a	Use.	
		combination of authorised	AgriculturalUse-	
		substances.	Id	
idSubstance	AlphaNumeric $(50)$	The code of the substance.	idSubstance,	<b>Yes</b>
			SubstanceId,	
			SubstanceCode,	
			Substance	

Table 2.69: Table definition for AgriculturalUsesHasCompounds.

Table aliases: AgriculturalUseHasSubstances, AgriculturalUsesHasSubstances, AgriculturalUseSubstances, AgriculturalUseGroups, AgriculturalUseGroup.

### **Occurrence patterns calculation**

Assumptions can be made for each food on the basis of findings in concentration data.

**Tier 1:** 0% occurrence is assumed for all substances with no positive concentrations at all; 100% occurrence is assumed for all substances with at least one positive concentration;

**Tier 2:** 0% occurrence is assumed for all substances with no positive concentrations at all; for substancefood combinations with at least one positive (finding), use findings patterns to implement a specific interpretation of Option 5 in the SANTE document, as described below.

Therefore in both tiers, substance-food combinations without any positive finding are handled in the optimistic way by assuming absolute zeroes for any non-detect observation.

If Tier 2 is selected, then for each of the modelled foods a tabulation is made of the observed frequencies of positives for all substance combinations (including the empty set), based on the *active substance concentrations*. For an OP consisting of just one substance, the basic frequency is the number of samples with a positive concentration divided by the number of samples where the substance has been measured (i.e., is not a MV). For an OP consisting of multiple substances, the basic frequency is the number of samples with all concentrations positive for the members divided by the number of samples where all members of the set have been measure[d.](#page-76-0)

After calculation of the basic frequencies for all occurrence patterns, these frequencies are rescaled such that the overall sum of frequencies is 100%. When *substance authorisations* are available, then patterns involving unauthorised substances are not rescaled and only those patterns for which all substances are authorised are rescaled such that the sum of all frequencies is 100%.

Note: the Tier 2 procedure is not what is literally written in the SANTE document, but is an interpretation agreed upon by EFSA and RIVM. An alternative [model, not yet implemen](#page-108-0)ted, but perhaps more in line with the text of the SANTE document, would be to double the basic frequencies to modelled occurrence pattern frequencies. Only if the sum of all frequencies becomes larger than 100%, the set of frequencies would be normalised to 100% sum.

# **Occurrence patterns settings**

# **Selection settings**



Table 2.70: Selection settings for module Occurrence patterns.

# **Uncertainty settings**





# **Occurrence patterns tiers**

# **Overview**





Table 2.73: Tier definition for EC 2018 Tier 1.

Name	
Apply occurrence pattern percentages	false

### **Input tiers**

Table 2.74: Input tiers for EC 2018 Tier 1.

Module	Input tier	
<i>Concentrations</i>	<i>EC</i> 2018 Tier 1	

### **EC 2018 Tier 2**





## **Input tiers**

Table 2.76: Input tiers for EC 2018 Tier 2.

Module	Input tier	
<i>Concentrations</i>	<i>EC</i> 2018 Tier 2	

### **Occurrence patterns as data**

Occurrence patterns are provided as data by specification of the occurrence mixtures and their associated occurrence/agricultural use percentages.

• *Occurrence patterns data formats*

Inputs used: *Substance authorisations Active substances*

### **Calc[ulation of occurrence patter](#page-98-0)ns**

Occurrence [patterns are computed fro](#page-108-0)[m the observed p](#page-202-0)atterns of positive concentrations in the concentration data.

• *Occurrence patterns calculation*

Inputs used: *Concentrations*

# **2.3.11 Processing factors**

Processing factors are multiplication factors to derive the concentration in a processed food from the concentration in an unprocessed food and can be specified for identified processing types (e.g., cooking, washing, drying). Processing factors are primarily used in dietary exposure assessments to correct for the effect of processing on substance concentrations in dietary exposure calculations.

This module has as primary entities: *Foods Substances*

Output of this module is used by: *Food conversions Dietary exposures Single value dietary exposures*

# **Processing factors data forma[ts](#page-27-0)**

Processing factors connect two fo[od codes, one for](#page-176-0) [the processed foo](#page-116-0)[d and one for the unprocessed](#page-194-0) food. There are two schemes to make this connection:

- <span id="page-102-0"></span>1) specify the two food codes and the processing type, or
- 2) use food facets, i.e. specify only the code of the unprocessed food and the processing type (facet), the code of the processed food is defined by the other two.

# **Processing factors**

Processing factors are defined for triplets of processing type, food, and substance. The processing types are defined in the processing types table and the processing factors are defined in the processing factors table.

# **Processing factors**

Processing factor records should be linked to processing types using the processing type code (idProcessingType) and for the foods and substances. The codes of the processing factor records should match the codes of the foods, substances, and processing type definitions.

<span id="page-103-0"></span>



Table aliases: ProcessingFactors, ProcessingFactor, Processing.

# **Food facet processing factors**

This table can be used to define processing factors for (FoodEx2) food/food-facet combinations.





Table aliases: FoodFacetProcessingFactors, FoodFacetProcessingFactor, FacetProcessingFactors,

FacetProcessingFactor, FacetProcessing.

# **Processing factors calculation**

### **Processing factors fixed or distribution based**

<span id="page-105-2"></span>Processing factors can be specified as fixed factors (nominal) or as statistical distributions for the variability across samples.

- The distribution is either *the logistic-normal distribution* for processing types with factors restricted between 0 and 1 (e.g. washing),
- or the lognormal distribution *the lognormal distribution* for processing types with non-negative factors (e.g. drying).

Variability distribution paramet[ers are specified indirectly via th](#page-105-0)e 50th and 95th percentile. Uncertainty for processing factors can be specified using uncertainty distributions of the same form as for variability. Uncertainty distribution parameters are specified indirectlyv[ia the 95th uncertainty per](#page-105-1)centiles on the 50th and 95th variability distribution percentiles.

For distribution based processing factors specify  $f_{k,nominal}$  and  $f_{k, upper}$  (*Nominal* and *Upper* in table **Processing-Factors**). Two situations are distinguished depending on the type of transformation.

### **Nonnegative processing factors**

<span id="page-105-1"></span>Equate the logarithms of  $f_{k,nominal}$  and  $f_{k,upper}$  to the mean and the 95% one-sided upper confidence limit of a normal distribution. This normal distribution is specified by a mean

$$
ln(f_{k,nominal})
$$

and a standard deviation

$$
\ln(f_{k,upper}) - \ln(f_{k,nominal})/1.645
$$

### **Processing factors between 0 and 1**

<span id="page-105-0"></span>Equate the logits of  $f_{k,nominal}$  and  $f_{k, upper}$  to the mean and the 95% one-sided upper confidence limit of a normal distribution. This normal distribution is specified by a mean

$$
logit(f_{k,nominal})
$$

and a standard deviation

$$
logit(f_{k,upper})-logit(f_{k,nominal})/1.645.
$$

See also *processing correction*

### **Processing factors settings**

### **Uncert[ainty settings](#page-131-0)**

Table 2.79: Uncertainty settings for module Processing factors.

Name	Description	
Resample processing factors	Specifies whether processing factors are resampled from a	
	parametric uncertainty distribution.	

#### **Processing factors uncertainty**

Processing effects are modelled either by a fixed processing factor, or by a lognormal or logistic-normal distribution (depending on the distribution type of the *processing type*). In case of a fixed factor, the uncertainty distribution is lognormal or logistic-normal with the same mean  $\mu$  as the fixed value, and with a standard deviation  $\sigma_{unc}$  which is calculated from the specified central value  $\mu$  (or nominal) and an estimate of the p95 of the *uncertainty distribution* (set *NominalUncertaintyUpper* in the *table for ProcessingFactors*).

The calculation is:

$$
\sigma_{unc} = \frac{f(NominalUncertaintyUpper) - f(\mu)}{1.645}
$$

with  $f() = logit$  for the logistic-normal distribution (distribution type 1) and  $f() = ln$  for the lognormal distribution (distribution type 2). Values lower than 0.01 or higher than 0.99 (distribution type 1 only) are replaced by default values (0.01 and 0.99); this is useful computationally to avoid problems. In each iteration of the uncertainty analysis a new value is drawn from this distribution to be used as a fixed factor in the Monte Carlo calculation. In case of distribution based processing factors (describing the variability of processing factors) two uncertainties can be specified. For  $\sigma_{unc}$ , specification and calculation is as before (set *NominalUncertaintyUpper* in the *table for ProcessingFactors*).

The uncertainty about the variability standard deviation

$$
\sigma_{var} = \frac{f(Upper) - f(\mu)}{1.645}
$$

can be specified by the *UpperUncertaintyUpper* value. This value is specified as the p95 upper limit on *Upper*. The specified value is used to derive in a iterative search the number of degrees of freedom *df* (van der Voet et al. 2009) [van der Voet et al., 2009]. In the uncertainty analysis, a modified chi-square distribution with *df* degrees of freedom is used to generate new values of  $\sigma_{var}$ . A very high value of *df* means little uncertainty and  $\sigma_{var}$  will be almost equal in all iterations of the uncertainty analysis. A  $df$  close to 0 means a large uncertainty and very different values of  $\sigma_{var}$ will be obtained in the iterations of the uncertainty analysis. The p95 upper limit on Upper is set through parameter *[UpperUncertaintyUpper](#page-357-0)*.

#### **Processing factors as data**

Specify for a combination of processing type, food and substance the processing factor (nominal, upper).

- *Processing factors data formats*
- *Processing factors calculation*

# **2.3.12 [Single value conce](#page-102-0)ntrations**

<span id="page-106-0"></span>Single [value concentrations data are](#page-105-2) the single value estimates (High Residue, Maximum Residue Limit, Supervised Trials Median Residue) of residue concentrations on foods as measured.

This module has as primary entities: *Foods Substances*

Output of this module is used by: *Modelled foods Single value dietary exposures*

#### **Single value concentrations d[ata fo](#page-27-0)[rmats](#page-44-0)**

Single value concentrations data p[rovides a single](#page-94-0) [value concentration for a subst](#page-194-0)ance.

# **Single value concentration data**

# **Concentration single values**

The food codes (idFood) and substance codes (idSubstance) should match the codes of the foods and substances table respectively.

Name	Type	Description	<b>Aliases</b>	Required
idFood	AlphaNumeric(50)	Code of the food of this	idFood, FoodId,	Yes
		concentration single value.	Food	
idSubstance	AlphaNumeric(50)	Code of the substance of this	idSubstance,	Yes
		concentration single value.	SubstanceId,	
			Substance,	
			idCompound,	
			CompoundId,	
			Compound	
Value	Numeric	Concentration single value.	Value,	Yes
			Concentration,	
			Concentration-	
			Value	
ValueType	Concentration Value-	Value type of the	Concentration-	Yes
	<b>Types</b>	concentration value.	SingleValue-	
			Type,	
			Concentration-	
			ValueType,	
			SingleValue-	
			Type,	
			Concentration-	
			Type,	
			ValueType,	
			Type	
Percentile	Numeric	Percentile.	Percentile	N <sub>o</sub>
Concentration-	<b>Concentration Units</b>	The unit of the concentration	Concentration-	N <sub>o</sub>
Unit		single value (default mg/kg).	Unit,	
			Unit	
Reference	AlphaNumeric(200)	Reference to the source from	Reference,	N <sub>o</sub>
		which this concentration	References,	
		single value is obtained.	Source, Sources	

Table 2.80: Table definition for ConcentrationSingleValues.

Table aliases: ConcentrationSingleValues, SingleValueConcentrations.
#### **Single value concentrations calculation**

<span id="page-108-0"></span>Single value concentrations as data are supplied as mean concentrations, median concentrations, highest residues, percentiles, LOQs or maximum residue limits. Specify the 'Use data' option in the interface. In a retrospective context, the single values are computed based on the concentration distributions available for the modelled food as supplied in the *Concentrations module*. Specify option 'Compute' in the Single value concentrations action.

#### **Single value concentrations settings**

#### **Selection se[ttings](#page-76-0)**



#### Table 2.81: Selection settings for module Single value concentrations.

#### **Single value concentrations as data**

Single value concentrations data are the single value concentrations of residues on foods as measured.

• *Single value concentrations data formats*

Inputs used: *Active substances*

#### **Calc[ulation of single value concentrati](#page-106-0)ons**

Single value [concentrations ar](#page-202-0)e calculated as a percentile (p50, p97.5 or maximum residue limit) of the food as measured concentration distribution.

• *Single value concentrations calculation*

Inputs used: *Concentrations Concentration limits Deterministic substance conversion factors*

# **2.3.13 [Substance authorisation](#page-108-0)s**

Substance au[thorisations spe](#page-76-0)[cify which food/subs](#page-60-0)[tance combinations are authorised for \(agri](#page-91-0)cultural) use. If substance authorisations are used, then only the food/substance combinations that are specified in the data are assumed to be authorised and all other combinations are assumed to be not authorised. This information may, for instance, be used to determine whether concentration measurements below the LOR could be assumed true zeros. I.e., if a food/substance combinations is assumed to be unauthorised, then the LOR may be assumed to be a zero.

This module has as primary entities: *Foods Substances*

Output of this module is used by: *Concentrations Occurrence patterns Concentration models*

## **Substance authorisations data formats**

#### **Substance authorisations**

<span id="page-109-0"></span>Authorised uses data provides information about whether substance use is allowed for specified foods. For cumulative exposure assessments, this information is used for imputation of non-detects/missing values.

#### **Authorised uses**

The authorised uses table





Table aliases: AuthorisedUses, AuthorisedUse.

## **Substance authorisations as data**

Substance authorisations are specified as data in the form of a list of authorised food/substance combinations, with combinations not on the list associated with no authorised use.

• *Substance authorisations data formats*

# **2.3.14 Substance conversions**

Subst[ance conversions specify how measured](#page-109-0) substances are converted to active substances, which are the substances assumed to cause health effects. In the pesticide legislation such measured substances and the substance conversion rules are known as residue definitions.

This module has as primary entities: *Substances*

Output of this module is used by: *Concentrations*

## **Substance conversions dataf[ormats](#page-44-0)**

<span id="page-109-1"></span>Two types of substance conversio[ns are implemen](#page-76-0)ted, with two subtypes for the first type:

1a) The measured substance is one or more of a set of possible substances (e.g. isomers or metabolites), and the toxicity of all substances in this set is assumed to be the same and is expressed in one active substance. Example: The measured substance Parathion-methyl(RD) is either Parathion-methyl or paraoxon-methyl, but both are expressed as the active substance Parathion-methyl.

1b) The measured substance is one or more of a set of possible substances (e.g. isomers or metabolites), and the toxicity of all substances in this set is assumed to relate with equal probability to one of a subset of active substances. Example: The measured substance Dithiocarbamates includes the active substances maneb, mancozeb, metiram, propineb, thiram and ziram, one of which will be assumed to be the active substance present with equal probability.

2) If *∶n active substances* all metabolise to the same active substance (the metabolite), it is assumed that all  $:n+1$  substances have equal probability of being the source of the measured concentration. The measured substance then is either one active substance (the metabolite) or a mixture of two active substances, one being the metabolite and the other one of the possible parent substances. Example: The measured substan[ce Carbofuran\(RD](#page-204-0)) is either the active substance Carbufuran or a mixture of Carbofuran and one of the possible active parent substances Benfuracarb or Carbosulfan.

## **Substance conversion rules**

Substance conversions are described by a single substance conversions table.

## **Substance conversion rules**

The records of the substance translations definitions table specify which active substances (idActiveSubstance) link to a measured substance (idMeasuredSubstance). Each record contains a conversion factor that specifies how a concentration of the measured substance translates to a concentration of the active substance, a flag that states whether the residue definition should be assumed to translate exclusively to one of its active substances, and a proportion. The proportion specifies the proportion of the samples that should translate to this specific active substance in case the translation is exclusive, otherwise it specifies the proportion of the concentration that is assumed to be attributed to the active substance.





Table aliases: ResidueDefinitions, ResidueDefinition.

#### **Substance conversions as data**

Substance conversions are provided as data.

• *Substance conversions data formats*

Inputs used: *Active substances*

# **2.3.15 [Total diet study samp](#page-109-1)le compositions**

Total diet st[udy sample comp](#page-202-0)ositions specify the composition of mixed food samples, such as used in a total diet study (TDS), in terms of their constituting foods.

<span id="page-112-1"></span>This module has as primary entities: *Foods*

Output of this module is used by: *Concentration models Food conversions*

## **Total diet study sample comp[ositio](#page-27-0)ns data formats**

## **Total diet study data**

<span id="page-112-0"></span>Total diet studies (TDS) complement traditional monitoring of substance concentrations on raw commodities by measuring substance occurrence in main foods prepared as consumed and pooled into representative food groups. To include occurrence data from TDS for exposure assessment, the composition of the TDS samples is needed in order to link the composite samples to the consumed foods (either directly or indirectly). TDS composition data describes the composition of TDS samples by specifying the foods (and the amounts) of TDS samples.

## **TDS food sample compositions**

The TDS food sample compositions table contains the descriptions of the TDS samples and specifications of the foods (with amounts) included in the TDS samples.

Name	Type	Description	Aliases	Required
idTDSFood	AlphaNumeric(50)	The code of the TDS food.	idTDSFood	Yes
idFood	AlphaNumeric(50)	Sub-food of the TDS food.	idFood	Yes
PooledAmount	Numeric	Total weight (in g) or volume	PooledAmount,	Yes
		(in ml) of the food.	Weight	
Description	AlphaNumeric $(200)$	Additional description of the	Description	N <sub>0</sub>
		TDS sample (e.g. number of		
		subsamples).		
Regionality	AlphaNumeric	Regionality information.	Regionality	N <sub>0</sub>
Seasonality	AlphaNumeric	Seasonality information.	Seasonality	N <sub>0</sub>

Table 2.84: Table definition for TDSFoodSampleCompositions.

Table aliases: TDSFoodSampleCompositions, TDSFoodSampleComposition, CompositionTDSFoodSamples, CompositionTDSFoodSample.

## **Total diet study sample compositions as data**

Total diet study sample compositions are provided as data.

• *Total diet study sample compositions data formats*

# **2.3.16 Unit variability factors**

Unitv[ariability factors specify the variation in concentrat](#page-112-0)ions between single units of the same food, which have been put together in a mixture sample on which the concentration measurements have been made. Unit variability factors are used for *modelling unit variability* in acute *(individual) dietary exposures calculations* to account for the fact that concentration data often relate to composite samples, whereas an acute risk may result from consumption of single food units. For the same purpose, they are also used in the *IESTI model* for *single value dietary exposures calculations*.

This module [has as primary entities:](#page-117-0) *Foods Su[bstances](#page-116-0)*

Output of this module is used by: *Dietary exposures Single value dietary exposures*

## **Unit variability factors data fo[rmats](#page-27-0)**

## **Unit variability factors**

<span id="page-113-0"></span>Unit variability factors specify the unit-to-unit variation of substance concentrations on foods. Unit variability factors are described by a single unit variability factors table.

## **Unit variability factors**

Unit variability factors are defined for a food, and may possibly also be specified for a specific substance and/or processing type. The unit variability factors are linked to the foods by means of the food code (idFood). Unit variability factors can be specified as unit variability factors (P97.5/mean) or as coefficients of variation of a statistical distribution.





Table aliases: UnitVariabilityFactors, UnitVariabilityFactor, VariabilityFactor, VariabilityFactors, VariabilityProcCompProd, UnitVariability.

# **IESTI special cases**

IESTI special cases for specified combinations of food, substance. The application type (post-harvest or pre-harvest) determines whether Case 1 or Case 3 should be used.

Name	Type	Description	<b>Aliases</b>	Required
idFood	AlphaNumeric(50)	The unique identification code	idFood, Code,	<b>Yes</b>
		of the food.	FoodId.	
			FoodCode,	
			Food, Id	
idSubstance	AlphaNumeric(50)	The unique identification code	idSubstance,	Yes
		of the substance. This code	SubstanceId,	
		may be from an existing	Substance,	
		coding system, such as	Code, Id	
		CAS-codes or Param codes of		
		EFSA, or it may be a		
		used-defined code.		
Application-	HarvestApplication-	Harvest application type	Application-	Yes
Type	Types	(pre-harvest or post-harvest).	Type,	
			Harvest-	
			ApplicationType	
Reference	AlphaNumeric(200)	External reference(s) to	Reference	N <sub>0</sub>
		pre-harvest use.		

Table 2.86: Table definition for IestiSpecialCases.

Table aliases: IestiSpecialCases.

## **Unit variability factors as data**

Unit variability factors are provided as data.

• *Unit variability factors data formats*

# **2.4 [Exposure modules](#page-113-0)**

*Exposures* are, in the simplest applications, *dietary exposures*, which combine consumption and occurrence data, either for single or for multiple *substances* causing the same adverse *effect*. Links between the foods-as-eaten and the *modelled foods* are made using *food conversions*, and the consumptions are expressed as *consumptions per modelled food*. For large assessment groups, the use of *dietary exposures screening* may be used to reduce the complexity of [the calcula](#page-150-0)tions and only focus calculations o[n the risk drivers.](#page-116-0)

In aggregate exposure assessments, *[expos](#page-44-0)[ures](#page-176-0)* combine *dietary exp[osures](#page-25-0)* with *non-dietary exposures*[, which have to](#page-115-0) [be entered as pr](#page-94-0)e-calculated data.

*Human monitoring data* can be compared to *exposures* using *[human monit](#page-147-0)oring analysis*.

In cumulative assessments, importa[nt mixture](#page-150-0)s of *substances* [can be ident](#page-116-0)ified using *[exposure mixture](#page-189-0)s*.

# **2.4.1 [Consumpti](#page-182-0)ons by model[led food](#page-150-0)**

<span id="page-115-0"></span>Consumptions by modelled food are consumption[s of individ](#page-44-0)uals expressed on thel[evel of the foods f](#page-166-0)or which concentration data are available (i.e., the modelled-foods). These are calculated from consumptions of foods-as-eaten and food conversions that link the foods-as-eaten amounts to modelled-foods amounts.

This module has as primary entities: *Populations Foods Substances*

Output of this module is used by: *Single value consumptions High exposure food-substance combinations Dietary exposures*

## **Consumptions by modelledf[ood calculation](#page-56-0)**

<span id="page-115-1"></span>[Consump](#page-116-0)tions by modelled food are calculated from *consumptions* of *modelled foods* and *food conversions* that link the foods-as-eaten amounts to modelled-foods amounts. Given that the food conversion is already available, the procedure for computing the consumptions by modelled-food is straightforward. For each consumption of each individual, a modelled-food consumption record is created for each modelled-food that is linked to the consumed foods through the food conversion, with the amount b[eing the total c](#page-48-0)on[sumption amou](#page-94-0)nt m[ultiplied by the pr](#page-176-0)oportion indicated by the food conversion. Also, if in the *food conversion algorithm* one or more *processing types* are found, then these types are recorded in the consumptionby modelled food record.

## <span id="page-115-2"></span>**Consumptions by modelled food**



Table 2.87: Calculation settings for module Consumptions by modelled food.

## **Calculation of consumptions by modelled food**

Consumptions by modelled food are calculated from consumptions of foods-as-eaten and food conversions that link the foods-as-eaten amounts to modelled-foods amounts.

• *Consumptions by modelled food calculation*

Inputs used: *Consumptions Food conversions*

#### Settings used

• *[Calculation Settings](#page-115-1)*

# **2.4.2 Dietary exposures**

<span id="page-116-0"></span>Dietar[y exposures are the](#page-115-2) amounts of substances, expressed per kg bodyweight or per individual, to which individuals in a population are exposed from their diet per day. Depending on the exposure type, dietary exposures can be shortterm/acute exposures and then contain exposures for individual-days, or they can be long-term/chronic exposures, in which case they represent the average exposure per day over an unspecified longer time period.

This module has as primary entities: *Populations Foods Substances Effects*

Output of this module is used by: *Exposures*

## **Dietary exposures calculation**

In probabilistic exposure assessm[ent we con](#page-150-0)sider a population of individuals. Exposure assessment with MCRA can address *acute exposure* or *chronic exposure*. Acute exposure is relevant when the short-term effect on individuals is relevant, chronic exposure when the long-term effects on the individuals matter. In MCRA short-term is operationalised as one day, so effectively acute exposure assessment is concerned with a population of person-days, whereas chronic exposure assessment is concerned with a population of persons.

The basic o[peration in expo](#page-117-1)su[re assessment is in](#page-122-0)tegrating consumptions and concentrations per food. With multiple foods, consumptions are typically correlated, therefore MCRA works with the multivariate distribution of a consumption vector, as represented by the consumption data of individuals in a consumption survey. In contrast, the distributions of concentration for each food are typically considered to be independent between foods. E.g., eating an apple with an accidentally high residue concentration does not predict that another food eaten on the same day will also have a high residue concentration. As a consequence of this assumption, concentrations of substances are modelled for each food independently.

For large assessment groups, the use of *dietary exposures screening* may be used to reduce the complexity of dietary exposures calculations and only focus calculations on the risk drivers. In this case, only detailed information is recorded for the risk drivers. With or without screening MCRA produces the same estimated cumulative exposure distribution summarized by percentiles and exceedance percentages, the same contributions of all substances and all foods-as-measured. After screening, co[ntributions related to food-a](#page-147-0)s-eaten are available for the risk drivers only.

In cumulative exposure calculations two simple approaches are used to identify and select mixtures contributing to the exposure of a target population:

1. qualitative approach: *counting of co-exposure*. To which combinations of substances are individuals exposed?

Co-exposure of substances is a qualitative approach where the number of combinations of substances to which an individual is exposed is recorded. There is no cut-off level, the only criterion is the presence of a substance in the simulated daily diet or not. For an *acute* or short term exposure assessment, a simulated individual day is the smallest entityt[o determine co-exposure](#page-158-0). For a *chronic* or long term exposure assessment, co-exposures are summarized at the individual level, e.g. co-exposure is determined combining all consumption days of an individual. For more information see *co-exposure of substances*.

2. quantitative approach: *maximum cumula[tive ra](#page-117-1)tio (M[CR\)](#page-122-0)*. To what degree are mixtures more important than single substances?

A quantitative approach is available in the *[exposures mixtures m](#page-158-0)odule*.

## **Acute exposure assessment**

<span id="page-117-1"></span>In an acute exposure assessment, the short term exposure to a substance or group of substances is estimated. The interest is in the distribution of individual day exposures and derived statistics like the fraction of days that exceed an intake limit or point of departure (*PoD*). The PoD is calculated as the acute reference dose (ARfD) \* safety factor (SF). The basic model for the exposure to a substance in an acute exposure assessment is:

$$
y_{ij} = \frac{\sum_{k=1}^{p} x_{ijk} c_{ijk}}{bw_i}
$$

where  $y_{ij}$  is the intake by individual i on day j (in microgram substance per kg body weight),  $x_{ijk}$  is the consumption by individual *i* on day *j* of food *k* (in g),  $c_{ijk}$  is the *(simulated) concentration* of that substance in food *k* eaten by individual *i* on day *j* (in mg/kg), and  $bw_i$  is the body weight of individual *i* (in kg). Finally, *p* is the number of foods accounted for in the model. Within parenthesis, the default unit definitions are assumed, but decimal multiples or submultiples of units are easily specified using the relevant tables.

In the exposure assessment, individual days enter the [Monte Carlo sample usin](#page-130-0)g the inverse of the sampling weights when the number of MC iterations is > 0 (see *table for Individuals*, field *SamplingWeight*).

## **Modelling unit-to-unit variation**

<span id="page-117-0"></span>The basic model for an acute exposure assessmen[t assumes that the co](#page-50-0)ncentration of the substance displays the variation of residues between units in the marketplace. In general, both monitoring data and controlled field trial data are obtained using composite samples. As a result some of the unit-to-unit variation is averaged out. The model for unit variability aims to adjust the composite sample mean such that sampled concentrations represent the originally unit-to-unit variation of the units in the composite sample.

MCRA offers three distributions to sample from:

- 1. the *beta distribution*,
- 2. the *lognormal distribution*,
- 3. and the *bernoulli distribution*.

The beta [distribution simu](#page-121-0)lates values for a unit in the composite sample. It requires knowledge of the number of units in a [composite sample and](#page-121-1) of the variability between units.

The lognorma[l distribution simulate](#page-122-1)s values for a new unit in the batch. It requires only knowledge of the variability between units.



# **Contribution to total exposure distribution for foods as measured**

Figure 2.13: Example MCRA dietary exposure contributions foods as measured.

# **Contribution to total exposure distribution for foods as eaten**







# **Contribution to total exposure distribution for substances**

Figure 2.15: Example MCRA dietary exposure contributions substances

# **Contribution to total exposure distribution for foods as measured x substances (MSCC)**



Figure 2.16: Example MCRA dietary exposure contributions foods as measured x substances

The bernoulli distribution is considered as a limiting case of the beta distribution when knowledge of the variability between units is lacking and only the number of units in the composite sample is known. For the beta and lognormal distribution, estimates of unit variability are either realistic (no censoring at the value of the monitoring residue) or conservative (unit values are left-censored at the value of the monitoring residue). For the lognormal distribution sampled concentrations have no upper limit. Whereas for the beta distribution, sampled concentration values for a unit are never higher than the monitoring residue times the number of units in the composite sample.

Variability between units is specified using a variability factor  $v$  (defined as 97.5th percentile divided by mean) or a coefficient of variation  $c_V$  (standard deviation divided by mean). Following FAO/WHO recommendations, the default variability factor  $v = 1$  for small crops (unit weight < 25 g). For large crops (unit weight  $\ge 25$  g)  $v = 5$ . For foods which are processed in large batches, e.g. *juicing, marmalade/jam, sauce/puree, bulking/blending* the variability factor  $v =$ 1 is proposed.

#### **Estimation of intake values using the concept of unit variability**

<span id="page-120-0"></span>A composite sample for food  $k$  is composed of  $nu_k$  units with *nominal (whole food/RAC) unit weight*  $wu_k$ . The weight of a composite sample is  $wm_k = nu_k \cdot w_k$  with mean residue value  $cm_k$ .

- For each iteration i in the MC-simulation, obtain for each food k a simulated intake  $x_{ik}$ , and a simulated composite sample concentration  $cm_{ik}$ .
- Calculate the number of unit intakes  $nux_{ik}$  in  $x_{ik}$  (round [upwards\) and set weights](#page-40-0)  $w_{ikl}$  equal to unit weight  $wu_k$ , except for the last partial intake, which has weight  $w_{ikl} = x_{ik} - (nu x_{ik} - 1)wu_k$ .
- For the beta or bernoulli distribution: draw  $nux_{ik}$  simulated values  $bc_{ikl}$  from a beta or bernoulli distribution. Calculate concentration values as  $c_{ikl} = bc_{ikl} \cdot cm_{ik,max} = bc_{ikl} \cdot cm_{ik} \cdot nu_k = s v f_{ikl} \cdot cm_{ik}$ , where  $nu_k$  is the number of units in a composite sample of food k, and  $svf_{ikl}$  is the stochastic variability factor for this simulated unit, i.e. the ratio between simulated concentration  $c_{ikl}$  and the simulated composite sample concentration  $cm_{ik}$ . Sum to obtain the simulated concentration in the consumed portion:

$$
c_{ik}=\sum_{l=1}^{nux_{ik}} w_{ikl}c_{ikl}/x_{ik}
$$

• For the lognormal distribution: draw  $nux_{ik}$  simulated logconcentration values  $lc_{ikl}$  from a normal distribution with (optional) a biased mean  $\mu = ln(cm_{ik})$  or (default) unbiased mean  $\mu = ln(cm_{ik}) - 1/2\sigma^2$  and standard deviation  $\sigma$ . Calculate concentration values as

$$
c_{ikl} = \exp(lc_{ikl}) = s v f_{ikl} * c m_{ik}
$$

where  $svf_{ikl}$  is the stochastic variability factor for this simulated unit, i.e. the ratio between simulated concentration  $c_{ikl}$  and the simulated composite sample concentration  $cm_{ik}$ . Back transform and sum to obtain the simulated concentration in the consumed portion:

$$
c_{ik} = \sum_{l=1}^{nux_{ik}} w_{ikl}c_{ikl}/x_{ik}
$$

For cumulative exposure assessments, a sensitivity analysis may be performed by specifying a full correlation between concentrations from different substances on the same unit. As a result, high (or low) concentrations from different substances occur together on the same unit. In MCRA, for each unit the random sequence is repeatedly used to generate concentration values for all substances.

#### **Beta distribution**

<span id="page-121-0"></span>Under the beta model simulated unit values are drawn from a bounded distribution on the interval  $(0, c_{max})$  with  $c_{max} = nu_k \cdot cm_k$ . The standard beta distribution is defined on the interval (0, 1) and is usually characterised by two parameters a and b, with  $a > 0$ ,  $b > 0$  (see e.g. Mood et al. 1974) [Mood et al., 1974]. Alternatively, it can be parameterised by the mean

$$
\mu = a/(a+b)
$$

and the variance

$$
\sigma^2 = ab/(a+b+1)^{-1}(a+b)^{-2}
$$

or, as applied in MCRA, by the mean  $\mu$  and the squared coefficient of variation

$$
c_{\nabla}^2 = ba^{-1}(a+b+1)^{-1}
$$

For the simulated unit values in *each iteration of the program* we require an expected value  $cm_k$ . This scales down to a mean value  $\mu = cm_k/c_{max} = 1/nu_k$  in the (standard) beta distribution. From this value for  $\mu$  and an externally specified value for  $cv_k$  the parameters  $a$  and  $b$  of the beta distribution are calculated as:

$$
a = b(nu_k - 1)^{-1}
$$

and

$$
b = \frac{(nu_k - 1)(nu_k - 1 - cv_k^2)}{nu_kcv_k^2}
$$

From the second formula it can be seen that  $cv_k$  should not be larger than  $\sqrt{nu_k - 1}$  in order to avoid negative values for  $b$ . When the unit variability is specified by a variability factor

$$
v_k = \frac{p97.5_k}{cm_k}
$$

instead of a coefficient of variation  $cv_k$  then MCRA applies a bisection algorithm to find a such that the cumulative probability

$$
P[Beta(a, b)] = 0.975
$$

for  $b = a(nu_k - 1)$ .

Sampled values from the beta distribution are rescaled by multiplication with  $cm_{max}$  to unit concentrations  $c_{ijk}$  on the interval  $(0, cm_{max})$ .

#### **Lognormal distribution**

<span id="page-121-1"></span>The lognormal distribution is characterised by  $\mu$  and  $\sigma$ , which are the mean and standard deviation of the logtransformed concentrations. The unit log-concentrations are drawn from a normal distribution with mean  $\mu =$  $ln(cm_{ik}) - 1/2\sigma^2$ . The coefficient of variation cv is turned into the standard deviation  $\sigma$  on the log-transformed scale with:

$$
\sigma = \sqrt{\ln(cv^2 + 1)}
$$

The variability factor is defined as the 97.5th percentile of the concentration in the individual measurements divided by the corresponding mean concentration seen in the composite sample. A variability factor  $v$  is converted into the standard deviation  $\sigma$  as follows:

$$
v = \frac{p97.5}{mean} = \frac{e^{\mu + 1.96\sigma}}{e^{\mu + 1/2\sigma^2}} = e^{1.96\sigma - 1/2\sigma^2}
$$

with  $\mu$  and  $\sigma$  representing the mean and standard deviation of the log-transformed concentrations. So

$$
ln(v) = 1.96\sigma - 1/2\sigma^2
$$

Solving for  $\sigma$  gives:

$$
\sigma^2 - 2 \cdot 1.96\sigma + 2\log(v) = 0
$$

with roots for  $\sigma$  according to:

$$
\sigma = 1.96 \pm \sqrt{(1.96^2 - 2log(v))}
$$

The smallest positive root is taken as an estimate for  $\sigma$ .

#### **Bernoulli distribution**

<span id="page-122-1"></span>The bernoulli model is a limiting case of the beta model, which can be used if no information on unit variability is available, but only the number of units in a composite sample is known (see van der Voet et al. 2001). As a worst case approach we may take the coefficient of variation  $cv$  as large as possible. When  $cv$  is equal to the maximum possible value  $\sqrt{nu_k} - 1$ , the (unstandardised) beta distribution simplifies to a bernoulli distribution with probability

$$
\left(nu_k-1\right)/nu_k
$$

or

$$
(v_k-1)/v_k \\
$$

 $1/nu_k$ 

for the value 0 and probability

or

 $1/v_k$ 

for the value  $c_{max} = nu_k \cdot cm_k$ .

In MCRA values 0 are actually replaced by  $cm_k$ , to keep all values on the conservative side. For example, with  $nu_k =$ 5, there will be 80% probability at  $c_{ijk} = cm_k$  and 20% probability at  $c_{ijk} = c_{max}$ . When the number of units  $nu_k$ in the composite sample is missing, the nominal unit weight  $wu_k$  is used to calculate the parameter for unit variability.

# **Chronic exposure assessment**

<span id="page-122-0"></span>In a chronic exposure assessment, usual exposure is defined as the long-run average of daily exposure to a substance or group of substances by an individual. The interest is in the distribution of individual exposures and derived statistics like the fraction of individuals that exceed an intake limit or point of departure *PoD*). The PoD is calculated as the average daily intake (ADI) \* safety factor (SF). Usually, for an individual, dietary recall data are available on 2 (or more) consecutive days. We assume an equal number of days for each individual, unless specified differently in *table for Individuals*.

For a chronic exposure assessment the available data are used to calculate exposu[res p](#page-347-0)er person-day (daily exposure):

$$
y_{ij} = \frac{\sum_{k=1}^{p} x_{ijk} c_{ijk}}{b w_i}
$$

where  $y_{ij}$ ,  $x_{ijk}$  and  $bw_i$  are defined as before but now concentrations of the substance found in food k enter the model as the *estimated mean substance concentration value*  $c_k$ . Using the person-day exposures MCRA, provides a number of *exposure models* to calculate the distribution of usual exposure at the person level.

## **Chronic exposure models**

Using the person-day exposures MCRA uses one of the following models to calculate the distribution of usual exposure at the person level:

- 1. The observed individual means *observed individual means* (OIM) model;
- 2. The *logisticnormal-normal (LNN) model*, in a full version that includes the estimation of correlation between exposure frequency and amount, and in a simpler version without this estimation;
- 3. The *betabinomial-normal* (BB[N\) model;](#page-123-0)
- 4. The *[discrete/semi-parametric](#page-124-0)* model known as the Iowa State University Foods (ISUF) model. For this model, an equal number of days per individual is assumed.

In modelli[ng usual exposure, tw](#page-124-1)o situations can be distinguished. Foods are consumed on a *daily basis* or foods are *episodically consumed*. For the logisticnormal-normal model and the betabinomial-normal model, the latter requires fitting of a [two-part model,](#page-124-2)

- 1. a model for the frequency of consumption, and
- 2. [a model for the](#page-135-0) exposure amount on consumption days.

In the final step, both models are integrated in order to obtain the usual exposure distribution. For daily consumed foods, fitting of the frequency of consumption is skipped and modelling resorts to fitting the model to daily exposure amounts only. Note that the distinction between BNN and LNN disappears and modelling will give equivalent results.

## **Observed individual means (OIM)**

<span id="page-123-0"></span>The usual exposure distribution for a population is estimated with the empirical distribution of individual means. Each mean is the average of all single-day exposures for an individual. The mean value for an individual still contains a considerable amount of within-individual variation. As a consequence, the distribution of within-individual means has larger variance than the true usual exposure distribution and estimates using the OIM-method are biased, leading to a too high estimate of the fraction of the population with a usual exposure above some standard. Despite its known tendency to over-estimate high-tail exposures, the OIM method is the method to be used in EFSA (2012) [EFSA, 2012] basic assessments.

## **Model based and model assisted**

Following Kipnis et al. [Kipnis et al., 2009], some of the models available in MCRA are extended to predict individual usual exposures. This model assisted approach has been added to BBN and LNN when used without correlation) and may be a useful extension in evaluating the relationship between health outcomes and individual usual exposures of foods. In contrast, the estimation of the usual exposure distribution in the general population is called the model based approach. Summ[arizing, we get](#page-356-0) Table 2.88:



Table 2.88: Model based and assisted approach available for chronic exposure models

The model assisted approach builds on the proposal of Kipnis et al. [Kipnis et al., 2009], but is modified to ensure that the population mean and variance are better represented. The method is based on shrinkage of the observed individual means (modified BLUP estimates) and shrinkage of the observed exposure frequencies. The model-assisted usual exposure distribution applies to the population for which the consumption data are representative, and automatically integrates over any covariates present in the model. Model-assisted [exposures are not ye](#page-356-0)t available for LNN, and when a covariable is modelled by a spline function of degree higher than 1. In case of a model with covariates the usual exposure is presented in graphs and tables as a *function of the covariates* (conditional usual exposure distributions).

#### **Betabinomial-Normal model (BBN)**

<span id="page-124-1"></span>The *Betabinomial-Normal (BBN)* model for chronic risk assessment is described in [de Boer et al., 2009], including its near-identity to the STEM-II model presented in [Slob, 2006]. The BBN model combines a betabinomial model for the exposure frequencies with a normal model for transformed positive exposures.

## **Lo[gisticnormal-Normal mod](#page-135-1)el (LNN with an[d without](#page-356-1) correlation)**

<span id="page-124-0"></span>In the logisticnormal-normal (LNN) model, exposure frequencies are modelled by a logistic normal distribution. In notation, for probability  $p$ :

$$
\text{logit}(p) = \log(p/1-p) = \mu - i + \underline{c}_i
$$

where  $\mu_i$  represents the person specific fixed effect model and  $c_i$  represent person specific random effects with estimated variance component  $\sigma_{between}^2$ . Similarly as in the BBN model, the positive exposure amounts are modelled, after transformation (logarithmic or Box-Cox), with a normal distribution. This model is referred to as the *LogisticNormal-Normal (LNN)* model. The full *LNN model* model includes the estimation of a correlation between exposure frequency and exposure amount. This is similar to the NCI model described in Tooze et al. [Tooze et al., 2006]. A simple and computationally less demanding version of the LNN method does not estimate the correlation between frequency and amount. The models are fitted by maximum likelihood, employing *Gauss-Hermite inte[gration](#page-136-0)*.

[For chronic models](#page-356-2) amounts are usually transformed before the statistical model is fit. The power transformation, given by  $y^p$ , has been replaced by the equivalent Box-Cox transformation. The Box-Cox transformation is a linear function of the power transformation, given by  $(y^p - 1)/p$ , and has a better numerical stability. *[Gauss-Hermite](#page-349-0) [integration](#page-349-0)* is used for back-transformation (see also *Box Cox power transformation*).

## **Discrete/semi-parametric model (ISUF)**

<span id="page-124-2"></span>Nusser et al. [Nusser et al., 1996] described how to assess chronic risks for data sets with positive exposures (a small fraction of zero exposures was allowed, but then replaced by a small positive value). The modelling allowed for heterogeneity of variance, e.g. the concept that some people are more variable than others with respect to their consumption habits. However, a disadvantage of the method was the restricted use to contaminated foods which were consumed on [an almost daily basi](#page-356-3)s, e.g. dioxin in fish, meat or diary products. The estimation of usual exposure from data sets with a substantial amount of zero exposures became feasible by modelling separately zero exposure on part or all of the days via the estimation of exposure probabilities as detailed in Nusser et al. [Nusser et al., 1997] and Dodd [Dodd, 1996]. In MCRA, a discrete/semi-parametric model is implemented allowing for zero exposure and heterogeneity of variance following the basic ideas of Nusser et al. and Dodd ([Nusser et al., 1996], [Nusser et al., 1997], [Dodd, 1996]). This implementation of the ISUF model for chronic risk assessment is fully described in de Boer et al. [de Boer et al., 2009].

## **[Model-The](#page-354-0)n-Add**

The traditional approach can be termed the Add-Then-Model approach, because adding over foods precedes the statistical modelling of usual exposure. MCRA offers, as an advanced option, an alternative approach termed Model-Then-Add (van der Voet et al. 2014). In this approach the statistical model is applied to subsets of the diet (single foods or food groups), and then the resulting usual exposure distributions are added to obtain an overall usual exposure distribution. The advantage of such an approach is that separate foods or food groups may show a better fit to the normal distribution model as assumed in all common models for usual exposure (including MCRA's *betabinomialnormal* (BBN) model and *logisticnormal-normal* model (LNN)). That this principle can work in practice was shown in previous work (de Boer et al. 2009 [de Boer et al., 2009], Slob et al. 2010 [Slob et al., 2010], Goedhart et al. 2012) [Goedhart et al., 2012], and a simulation model was developed and implemented in MCRA 7.1 to show how multimodal distributions can arise from adding unimodal distributions of foods that are not always consumed (Slob et al. 2010 [Slob et al., 2010], de Boer and van der Voet 2011, [de Boer et al., 2011]). For specific cases involving separate modelling of dietary supplements and the rest of the diet, proposals have been made (Verkaik-Kloosterman et al. 2011) [Verkaik-Kloosterman et al., 2011]. However, a practical approach to apply the Model-Then-Add approach to general cases of usual exposure estimation was still missing. Therefore a module in MCRA was developed to implements[uch an approach b](#page-356-4)ased on a visual inspection of a pre[liminary estimate of](#page-354-1) the usual exposure distribution using the *Observed Individual Means* (OIM) method.

# **The Model step**

At this stage of development the division of foods into a number of food groups is performed in an interactive process, where the MCRA user is presented with a visual display (see example in Figure 2.17) which shows:

- 1. The OIM distribution represented as a histogram, where each bar shows the frequency of exposures (summed over foods) of individuals in a certain exposure interval; each bar is subdivided according to the contributions of the individual foods contributing to those exposures (left panel [Figure 2.17\)](#page-125-0).
- 2. The contributions graph, where each of the bars in the OIM histogram is expanded to 100%. This graph allows a better view of the lower bars in the OIM histogram.

The visual display identifies the nine foods that contribute most to the tot[al exposure; t](#page-125-0)he remaining foods are grouped in a rest category to avoid identification problems because of too many colours (right panel Figure 2.17).

<span id="page-125-0"></span>

Figure 2.17: Left panel: OIM usual exposure distribution to smoke flavours via the different foods (excluding the zero exposures) in young children; right panel: Contribution of foods to exposures within each bar of the OIM distribution histogram.

The user has now the possibility to select one or more foods and to split these from the main exposure histogram. A separate graph shows the OIM distribution for the split-off food or food group. The graphs for the main group (now called the rest group) are adapted to show the OIM distribution and the contributions for the remaining foods only (see Figure 2.18 upper two panels). This splitting-off can be repeated several times for other foods or food groups. In this way the user can try to obtain foods or food groups that show unimodal OIM distributions. If the result is not what is intended, a food or food group can be added again to the rest group. Per split-off food or food group the usual exposure can be modelled using either BBN or LNN, with a logarithmic or power transformation. The rest group will [always be m](#page-126-0)odelled as OIM. It is possible that the rest group is empty, when the total exposure via the different split-off foods and /or food groups is modelled with BBN or LNN.

After a split-off selection has been made, the OIM distribution is summarised in terms of the defined grouping (Figure 2.19), and the usual exposure distribution per split-off food or food group is fitted according to the chosen modelling settings.

<span id="page-126-0"></span>

Figure 2.18: Result of a selection into two split-off groups and a rest group. The graph bottom left represents the exposure via a food group containing *'Sausage, frankfurter'* and *'Sausage, smoked cooked'*. The graph bottom right represents the exposure via a food group containing *'Sausage, luncheon meat', Herbs, mixed, main brands, not prepared', 'Soup, pea', 'Ham', and 'Bacon'*. The top graph represents the exposure via the rest group.

#### Usual exposures per model



Figure 2.19: OIM usual exposure distribution showing the contributions from the three food groups as constructed in Figure 2.18.

#### **The Add step**

C[onsumptions](#page-126-0) of foods may be correlated. In the traditional Add-Then-Model approach the Add step automatically reflects any correlations that are apparent in the consumptions at the individual-day or individual level. In the Model-Then-Add approach the estimated usual exposure distributions for different foods or food groups have to be combined to assess the total usual exposure. Two approaches are available for this:

- 1. *Model-based approach*: adds independent samples from the usual exposure distribution per food or food group, ignoring any correlations in consumption;
- 2. *Model-assisted approach*: adds the model-assisted, person-specific usual exposure estimates per food or food [group, taking correlatio](#page-131-0)ns in consumptions into account.

#### See also, *episodically consumed foods, model-based, model-assisted*.

Befor[e the addition is made, in](#page-133-0) the model-based approach, model-based estimates of the usual exposure amounts distribution per food or food group are back-transformed values from the normal distribution assumed for transformed amounts per food or food group, and the *model-based frequency* distribution is sampled to decide if a simulated individua[l has exposure via the food or food group or not. Model-ass](#page-135-0)isted estimates of the usual exposure distribution are back-transformed values from a shrunken version of the transformed OIM distribution, also done per food or food group, where the shrinkage factor is based on the variance components estimated using the linear mixed model for amounts at the transformed scale (van Kla[veren et al. 2012\). For](#page-136-1) individuals with no observed exposure (OIM=0) no model-assisted estimate of usual exposure can be made and a model-based replacement is used.

The model-based approach was investigated in Slob et al. (2010) [Slob et al., 2010] and performed surprisingly well, even if correlations in consumptions of foods were present. The model-assisted approach adds exposures at the individual level, and therefore retains effects of correlations between foods in the usual exposure distribution.

MCRA calculates both the model-based and model-assisted usual [intake distributio](#page-356-4)ns.

#### **Chronic exposure as a function of covariates**

The intake frequency and transformed intake amounts may be modelled as a function of covariates. MCRA allows one covariable and/or one cofactor.



Figure 2.20: Model-assisted estimated usual exposure distributions (excluding the zero exposures).



Figure 2.21: Model-based estimated usual exposure distributions (excluding the zero exposures).





Here  $l = 1...L$  and L is the number of levels of the cofactor,  $y_{ij}$ , the intake amount,  $x_1$  is the covariable, f is a polynomial function with the degrees of freedom  $df$ ,  $c_i$  and  $u_{ij}$  are the individual effect and interaction effect, respectively. These effects are assumed to be normally distributed  $N(0, \sigma_{between}^2)$  resp.  $N(0, \sigma_{within}^2)$ . The degree of the function is determined by backward or forward selection. In the output, the usual intake is displayed for a specified number of values of the covariable and/or the levels of the cofactor.

# **Total Diet Study**

In Total Diet Studies (TDS), substance occurrence data is obtained from measuring food products as consumed. TDS offers a more direct measure of substance concentrations compared to traditional monitoring and surveillance programs that are concerned with contamination of raw agricultural commodities. In a TDS, food selection is based on national consumption data in such a way that 90 to 95% of the usual diet is represented by the samples. Selected foods are collected, prepared as consumed and related foods are pooled prior to analysis. The compositions these TDS food samples are described by the *TDS food sample compositions* data module.

In MCRA, TDS concentration data can also be used in *dietary exposure assessments*, using it as an alternative type of concentration data where the foods-as-measured are not the raw primary commodities (RACs), but these are TDS food compositions. To link the concentration data to the consumed foods, the *TDS food sample composition information* is used in the *food conversion algorithm* [in a manner analo](#page-112-1)gous to the use of *food recipes* describing the composition of a composite food. The main difference [is that the translation propor](#page-116-0)tion is always 100% (default). Take, as an example, a TDS food *FruitMix* that is composed of *apple, orange* and *pear*, then a consumed food (foodas-eaten) *apple-pie* is converted to *apple*, *wheat* and *butter* (in some specific proporti[ons\) and subsequently,](#page-112-1) *apple* is [converted to](#page-112-1) food-as-measured *FruitMix* [\(100%\). N](#page-176-0)ot necessarily all foods as consume[d are represe](#page-53-0)nted in a TDS food sample. In addition to the TDS food sample compositions, there may be additional foods that are not officially part of a TDS food, but which can be extrapolated to a TDS food sample. Through the use of *food extrapolations* (read across translations), these foods may be directly linked to a TDS food sample, e.g., by specifying that *pineapple* is translated to *FruitMix*, *pineapple* or foods containing *pineapple* will also be matched to a *FruitMix* concentration.

Because TDS samples only contain one single, average measurement, TDS occurrence data can currently only be used for only applicable for chronic exposures assessments. However, when variability informa[tion is available for](#page-93-0) the raw primary foods in the TDS food samples (e.g., from monitoring), this information may be used *to approximate the variance of TDS samples*.

For more information about Total Diet Studies, visit the TDS-Exposure website http://www.tds-exposure.eu.

# **[TDS reduction to limit s](#page-67-0)cenario analysis**

The outcome of a MCRA risk assessment may be that some foods dominate the right upper tail of the exposure distribution. A scenario analysis answers the question to what extent the risk of foods with a high exposure would have been diminished by an intervention or by taking any precautions. To be able to do so, some information is needed about the concentration distributions of the raw agricultural commodities (RACs) that make up the TDS food sample. The decision to intervene or not can be based on comparison between the p95 percentile point of the concentration distribution and a concentration limit value that associated with a high risk.

- For  $p95 \le$  limit, most concentration values are below the value that is considered as a potential risk, so there is no urgency to take any precautions.
- When the opposite is true, i.c. p95 > limit, there may be an argument to intervene for this specific food.

In MCRA, limits and p95's are supplied by the *concentration distributions module*. In the MCRA interface, a scenario analysis can be checked and the scroll down list allows to select the foods that should be included in the scenario analysis. For the selected foods, concentration reduction factors are computed based on the p95 percentile and the limit value:

$$
f_{\rm reduction}= \textit{limit}/\textit{p95}
$$

These reduction factors (computed for the RACs) are applied to the simulated concentrations on the level of the TDS composite foods of which the conversion paths contain the RACs included in the scenario analysis.

$$
f_{\rm reduction} \cdot c_{TDS}
$$

Here,  $c_{TDS}$  is the concentration value of the TDS food.

#### **Substance concentrations generation**

<span id="page-130-0"></span>Both *chronic* and *acute* dietary exposure assessments rely on assigning substance concentrations to consumed foodsas-measured. For chronic exposure assessments, this concentration should be the mean concentration of the food and substance, as obtained from the concentration models. For acute, these concentrations are obtained through random sampling, for which there are two distinct approaches: sample-based and substance-based.

#### **Sample-based concentrations generation**

In the sample-based approach, the analytical samples from the concentration data form the basis for generating concentrations. For each identified food-as-measured of a consumption, substance concentrations are generated by drawing a random sample from the set of all samples available for that food-as-measured. Assuming that for the drawn sample, substance concentration values are known for all substances of interest (i.e., all missing values and non-detects are imputed with either a zero concentration or a positive concentration at or below LOR), the substance concentrations for all substances of the assessment group are set to the substance concentrations of the drawn samples. The rationale behind this approach is that it maintains correlations between substance concentrations on the same food.

As mentioned, the sample based approach relies on all samples being analysed for all substances of interest. Often, this is not the case and for a given sample, concentration may missing for one or more substances. Also, this approach requires non-detect values to be imputed with either positive concentration or a zero concentration.

For *imputation* of missing values there are two approaches:

- 1. **Imputation by zero:** all missing values are assumed zero.
- 2. **Imputation using substance-based concentration models:** all missing values are imputed by drawing a [concentr](#page-67-1)ation value from the substance-based concentration models.

For imputation of non-detects, two approaches exist:

- 1. **Replace by zero:** Non-detect values are imputated by a zero concentration value. This is an optimistic approach.
- 2. **Replace by factor times LOR:** Each non-detect value is replaced by a factor (e.g., 1 or 1/2) times its LOR.

#### **Substance-based concentrations generation**

In the substance-based approach, substance concentrations for a given food are drawn independently per substance from the food/substance concentration models.

#### **Processing correction**

Concentrations in the consumed food (food as eaten) may be different from concentrations in the modelled food in monitoring programs (typically raw food) due to processing, such as peeling, washing, cooking etc. Concentrations are therefore corrected according to

$$
c'_{jhk} = pf_{jhk} \cdot c_{jhk} = \left(\frac{PF_k}{cf_k}\right) \cdot c_{jhk}
$$

where  $c_{jhk}$  is the concentration of substance k in the food j with processing type h, and where  $pf_{jhk} = \frac{PF_{jhk}}{cf_{ihk}}$  $\frac{F f_{jhk}}{c f_{jhk}}$  is a factor indicating the mass change for a specific combination  $k$  of modelled food and processing. The processing correction factor  $cf<sub>ihk</sub>$  is used to correct for the fact that the processing factors  $PF<sub>jhk</sub>$  as commonly available from the input data describe both the effects of chemical alteration and weight change. E.g. for a dried food with a consumption of 100 gram which is translated to 300 gram raw agricultural commodity, the correction factor is 3. Note that the weight change is already included when calculating the consumption amounts of the foods-as-measured.

#### **Chronic exposure assessment, daily consumed foods**

#### <span id="page-131-0"></span>**Model based usual intake**

Foods are consumed on a daily basis.

For individual *i* on day *j* let  $Y_{ij}$  denote the 24 hour recall of a food  $(i = 1...n; j = 1...n_i)$ . In most cases withinindividual random variation is dependent on the individual mean and has a skewed distribution. It is therefore customary to define a one-way random effects model for  $Y_{ij}$  on some transformed scale

$$
Y_{ij}^* = g(Y_{ij}) = \mu_i + b_i + w_{ij}
$$

with  $b_i \sim N(0, \sigma_b^2)$  and  $w_{ij} \sim N(0, \sigma_w^2)$ 

Note that  $b_i$  represents variation between individuals and  $w_{ij}$  represents variation within individuals between days.

The mean  $\mu_i$  may depend on a set of covariate  $Z_i = (Z_{i1}, ..., Z_{in})$ :

$$
\mu_i = \beta_0 + \beta_1^t Z_i
$$

where  $\beta_0$  and  $\beta_1$  are regression coefficients.

The usual intake  $T_i$  for an individual i is defined as the mean consumption over many many days. This assumes that the untransformed intakes  $Y_{ij}$  are unbiased for true usual intake rather than the transformed intakes  $Y_{ij}^*$ . In mathematical terms  $T_i$  is the expectation of the intake for this individual where the expectation is taken over the random day effect:

$$
T_i = E_w[g^{-1}(\mu_i + b_i + w_{ij}) | b_i] = F(b_i)
$$

#### **Model based usual intake on the transformed scale**

For the model based usual intake first note that the conditional distribution

$$
(\mu_i + b_i + w_{ij}|b_i) \sim N(\mu_i + b_i, \sigma_w^2)
$$

It follows that the usual intake  $T_i$  is given by

$$
T_i = E_w[g^{-1}(\mu_i + b_i + w_{ij}|b_i)] = \int_{-\infty}^{\infty} g^{-1}(\mu_i + b_i + w_{ij}) \frac{1}{\sqrt{2\pi\sigma_w^2}} \exp\left(-\frac{w^2}{2\sigma_w^2}\right) dw
$$

#### **Model based using a logarithmic transformation**

For the logarithmic transform the usual intake  $T_i$  can be written in closed form using the formula for the mean of the lognormal distribution:

$$
T_i=\exp(\mu_i+b_i+\sigma_w^2/2)
$$

In this case  $T_i$  follows a log-normal distribution with mean  $\mu_i + \sigma_w^2/2$  and variance  $\sigma_b^2$ . This fully specifies the usual intake distribution, e.g. the mean and variance of the usual intake are given by

$$
\mu_{iT} = E[T_i] = \exp(\mu_i + \sigma_w^2/2 + \sigma_b^2/2)
$$

$$
\sigma_{iT}^2 = Var[T_i] = [\exp(\sigma_b^2) - 1] \exp(2\mu_i + \sigma_w^2 + \sigma_b^2)
$$

#### **Model based using a power transformation**

For the *power transformation* the integral can be approximated by means of N-point Gauss-Hermite integration. This results in the following usual intake

$$
T_i \approx \frac{1}{\sqrt{\pi}} \sum_{j=1}^N w_j (\mu_i + b_i + \sqrt{2} \sigma_w x_j)^p
$$

with  $p$  the inverse of the power transformation. A similar approximation can be used for the Box-Cox transformation. There can be a small problem with Gauss-Hermite integration. The summation term  $(\mu_i + b_i + \sqrt{2}\sigma_w x_j)^p$  can not be calculated when the factor between round brackets is negative and the power  $p$  is not an integer. This can happen when  $(\mu_i + b_i)$  is small relative to the between day standard error  $\sigma_w$ . In that case the corresponding term is set to zero. This is not a flaw in the numerical method but in the statistical model since the model allows negative intakes on the transformed scale which cannot be transformed back to the natural scale. The mean and variance of  $T_i$  can be approximated again by using Gauss-Hermite integration:

$$
\mu_{iT} = E[T_i] = \frac{1}{\sqrt{\pi}} \sum_{k=1}^{N} w_k \frac{1}{\sqrt{\pi}} \sum_{j=1}^{N} w_j (\mu_i + \sqrt{2}\sigma_w x_j + \sqrt{2}\sigma_b x_k)
$$

$$
\sigma_{iT} = Var[T_i] = \frac{1}{\sqrt{\pi}} \sum_{k=1}^{N} w_k \left[ \frac{1}{\sqrt{\pi}} \sum_{j=1}^{N} w_j (\mu_i + \sqrt{2}\sigma_w x_j + \sqrt{2}\sigma_b x_k) \right]^2 - \mu_T^2
$$

An alternative method for obtaining model based usual intakes for the power transformation employs a Taylor series expansion for the power, see e.g. Kipnis (2009) [Kipnis et al., 2009]. This is however less accurate than Gauss-Hermite integration. For the power transformation simulation is required to derive the usual intake distribution: simulate a random effect  $b_i$  for many individuals and then approximate  $T_i$  for these individuals. The  $T_i$  values then form a sample form the usual intake distribution.

#### **Model assisted usual intake on the transformed scale**

The model assisted approach employs a prediction for the usual intakes of every individual in the study. This requires a prediction of the individual random effect  $b_i$  for every individual.

<span id="page-133-0"></span>In the one-way random effects model the Best Linear Unbiased Prediction for  $(\mu_i + b_i)$  is given by

$$
\textit{BLUP}_i = \mu_i + (\bar{Y}^*_i - \mu_i) \left( \frac{\sigma_b^2}{\sigma_b^2 + \sigma_w^2/n_i} \right)
$$

in which  $\bar{Y}_i^*$  is the mean of the transformed intakes for individual i. BLUPs have optimal properties for some purposes, but not for the purpose of representing the variation  $\sigma_b^2$  between individuals. This can be seen by noting that

$$
\textit{Var}(\bar{Y}^*_i) = \sigma_b^2 + \sigma_w^2/n_i
$$

and thus

$$
\textit{Var}(\textit{BLUP}_i) = \left(\frac{\sigma_b^4}{\sigma_b^2 + \sigma_w^2/n_i}\right)
$$

which is smaller than the between individual variance  $\sigma_b^2$ . As an alternative a modified BLUP can be defined by means of

$$
modified B L U P_i = \mu_i + (\bar{Y}_i^* - \mu_i) \sqrt{\left(\frac{\sigma_b^2}{\sigma_b^2 + \sigma_w^2/n_i}\right)}
$$

which has the correct variance  $\sigma_b^2$  and also the correct mean  $\mu_i$ . However these optimal properties disappear when modified BLUPs are directly backtransformed to the original scale.

#### **Model assisted using a logarithmic transformation**

For the logarithmic transformation the usual intake  $T_i$  follows a log-normal distribution with mean  $\mu_i + \sigma_w^2/2$  and variance  $\sigma_b^2$ . If we can construct a BLUP like stochastic variable with the same mean and variance, then this variable be an unbiased predictor with the correct variance. It is easy to see that the following variable has the same distribution as  $T_i$ 

$$
\textit{modelassistedBLUP}_i = \mu_i + \frac{\sigma_w^2}{2} + (\bar{Y}^*_i - \mu_i) \sqrt{\left(\frac{\sigma_b^2}{\sigma_b^2 + \sigma_w^2/n_i}\right)}
$$

So the model assisted individual intake  $exp(modelassisted BLUP<sub>i</sub>)$  has the same distribution as the usual intake and is thus the best predictor for usual intake.

Kipnis et al. (2009) [Kipnis et al., 2009] employs the conditional distribution of  $b_i$  given the observations  $Y_{i1}, \cdots, Y_{in_i}$ to obtain a prediction. First note that

$$
(b_i|Y_{i1},\cdots,Y_{in_i})=(b_i|Y_{i1}^*,\cdots,Y_{in_i}^*)=(b_i|\bar{Y}_i^*)
$$

Since all distributio[ns in the one-way ra](#page-356-0)ndom effects model are normal it follows that:

$$
(b_i,\bar{Y}^*_i)\sim \textit{BivariateNormal}(0,\mu_i,\sigma_b^2,\sigma_b^2+\sigma_w^2/n_i,\sigma_b^2)
$$

where the last parameter represents the covariance between  $b_i$  and  $\bar{Y}_i^*$ . It follows that the conditional distribution

$$
(b_i|\bar{Y}^*_i) \sim N(\mu_c, \sigma_c^2)
$$

with

$$
\mu_c=\frac{\sigma_b^2}{\sigma_b^2+\sigma_w^2/n_i}(\bar{Y}_i^*-\mu_i)
$$

and

$$
\sigma_c^2 = \frac{\sigma_b^2 \sigma_w^2/n_i}{\sigma_b^2 + \sigma_w^2/n_i}
$$

A prediction for the usual intake  $T_i = F(b_i)$  is then obtained by the expectation

$$
E[F(b_i)|\bar{Y}_i^*] = \int F(b)\phi(b;\mu_c,\sigma_c^2)db
$$

For the logarithmic transform  $F(b_i) = \exp(\mu_i + b_i + \sigma_w^2/2)$  and the expectation reduces to

$$
E[F(b_i)|\bar{Y}_i^*] = \exp(\mu_i + \mu_c + \sigma_c^2/2 + \sigma_w^2/2)
$$

which is a function of  $\bar{Y}_i^*$  through  $\mu_c$ . To obtain the mean and variance of the prediction note that

$$
\mu_i + \mu_c + \sigma_c^2/2 + \sigma_w^2/2 \sim N\left(\mu_i + \frac{\sigma_b^2 \sigma_w^2/n_i}{2(\sigma_b^2 + \sigma_w^2/n_i)} + \frac{\sigma_w^2}{2}, \frac{\sigma_b^4}{\sigma_b^2 + \sigma_w^2/n_i}\right)
$$

It follows that the expectation of the prediction equals

$$
E[E[F(b_i)|\bar{Y}_i^*]] = \exp\left(\mu_i + \frac{\sigma_b^2 \sigma_w^2 / n_i}{2(\sigma_b^2 + \sigma_w^2 / n_i)} + \frac{\sigma_w^2}{2} + \frac{\sigma_b^4}{2(\sigma_b^2 + \sigma_w^2 / n_i)}\right)
$$

$$
= \exp\left(\mu_i + \frac{\sigma_b^2}{2} + \frac{\sigma_w^2}{2}\right)
$$

which equals the mean of the usual intake. However the variance of the prediction equals

$$
Var[E[F(b_i|\bar{Y}_i^*]] = \left[\exp\left(\frac{\sigma_b^4}{(\sigma_b^2 + \sigma_w^2/n_i)}\right) - 1\right] \exp(2\mu_i + \sigma_b^2 + \sigma_w^2)
$$

Which is less than the variance of the usual intake. The approach of Kipnis et al (2009) [Kipnis et al., 2009] will therefor result in too much shrinkage of the model assisted usual intake.

#### **Model assisted using a power transformation**

For the *power transformation* a model assisted BLUP with optimal properties, as derived above, cannot be constructed. The approach of Kipnis et al. (2009) [Kipnis et al., 2009] can however be used to obtain a prediction in the following way. First approximate  $T_i = F(b_i)$  by *Gauss-Hermite integration*:

$$
F(b_i) = T_i \approx \frac{1}{\sqrt{\pi}} \sum_{j=1}^{N} w_i (\mu_i + b_i + \sqrt{2} \sigma_w x_i)^p
$$

Secondly again use Gauss-Hermite to a[pproximate the expectation](#page-349-1) of the conditional distribution giving the prediction  $P_i$ .

$$
P_i = E[F(b_i)|\bar{Y}_i^*] = \int F(b_i)\phi(b;\mu_c,\sigma_c^2) \mathrm{d}b \approx \frac{1}{\pi} \sum_{k=1}^N w_k \sum_{j=1}^N w_j (\mu_i + \mu_c + \sqrt{2}\sigma_w x_j + \sqrt{2}\sigma_c x_k)^p
$$

which is a function of  $\bar{Y}_i^*$  through  $\mu_c$ . It is likely that the thus obtained predictions  $P_i$  have a variance that is too small. If we would know the mean  $\mu_{iP}$  and variance  $\sigma_{iP}^2$  of the predictions, the predictions could be linearly rescaled to have the correct mean  $\mu_{iT}$  and variance  $\frac{2}{iT}$ . The mean and variance of the prediction can be calculated using *Gauss-Hermite integration*.

$$
\mu_{iP} = \frac{1}{\sqrt{\pi}} \sum_{l=1}^{N} w_l \frac{1}{\pi} \sum_{k=1}^{N} w_k \sum_{j=1}^{N} w_j (\mu_i + \sqrt{2} \frac{\sigma_b^2}{\sigma_b^2 + \sigma_w^2/n_i} x_l + \sqrt{2} \sigma_w x_j + \sqrt{2} \sigma_c x_k)^p
$$

$$
\sigma_{iP}^2 = \frac{1}{\sqrt{\pi}} \sum_{l=1}^{N} w_l \left[ \frac{1}{\pi} \sum_{k=1}^{N} w_k \sum_{j=1}^{N} w_j (\mu_i + \sqrt{2} \frac{\sigma_b^2}{\sigma_b^2 + \sigma_w^2/n_i} x_l + \sqrt{2} \sigma_w x_j + \sqrt{2} \sigma_c x_k)^p \right]^2 - \mu_{iP}^2
$$

The proposed prediction then equals

$$
P_i^* = \mu_{iT} + \frac{\sigma_{iT}}{\sigma_{iP}}(P_i - \mu_{iP})
$$

#### **2.4. Exposure modules 127**

#### <span id="page-135-0"></span>**Chronic exposure assessment, episodically consumed foods**

For episodically consumed foods we need to take the probability of consumption into account. Define  $p_i$  as the probability that individual  $i$  consumes the food on any given day. The usual intake for this individual is then given by the product of  $p_i$  and  $T_i$  which is now defined as the usual amount on consumption days. Since individuals will vary in their probability pi, besides modelling the amounts as for daily consumed foods, it is also necessary to model the frequency of consumption. A three stage analysis of 24-hour recall data is the necessary:

- 1. A model for the frequency of consumption
- 2. A model for the intakes on consumption days
- 3. Integration of both models in order to obtain a usual intake distribution.

Step 2 uses the analysis outlined in the previous section for the positive intakes only. For step 1 two popular models which describe between-individual variation for the probability of consumption are the beta-binomial model and the logistic-normal model.

#### <span id="page-135-1"></span>**Beta-Binomial model for frequencies (BBN)**

Let  $n_i$  be the total number of recall days for individual i and  $X_i$  the number of days with a positive intake. The distribution of  $X_i$ , with  $p_i$  the probability of consumption for individual i, is given by

$$
X_i = \mathit{Binomial}(n_i, p_i)
$$

In this model the probability  $p_i$  varies among individuals according to the Beta distribution:

$$
f(p) = B^{-1}(\alpha, \beta) p^{\alpha - 1} (1 - p)^{\beta - 1}
$$

with

$$
B(\alpha, \beta) = \frac{\Gamma(\alpha)\Gamma(\beta)}{\Gamma(\alpha + \beta)}
$$

Combining the binomial and the Beta distribution results in the betabinomial distribution:

$$
P(X_i = x) = {n_i \choose r} \frac{B(\alpha + x, n_i + \beta - x)}{B(\alpha, \beta)}
$$

The mean and variance of the betabinomial distribution are given by

$$
E[X_i] = n_i \frac{\alpha}{\alpha + \beta}
$$

and

$$
\text{Var}[X_i] = n_i \frac{\alpha \beta (\alpha + \beta + n_i)}{(\alpha + \beta)^2 (\alpha + \beta + 1)}
$$

Using the reparameterization  $\pi = \alpha/(\alpha + \beta)$  and  $\phi = 1/(\alpha + \beta + 1)$ , it follows that

$$
E[X_i] = n_i \pi
$$

and

$$
\textit{Var}[X_i] = n_i \pi (1-\pi)[1+(n_i-1)\phi]
$$

This reparameterization enables to model the probability  $\pi_i$  of consumption for individual i directly as a logistic regression:

$$
\textit{logit}(\pi_i) = \gamma_0 + \gamma_1^t Z_i
$$

Note that the dispersion parameter  $\phi$ : is assumed to be equal for all individuals. The betabinomial logistic regression model can be fitted by means of maximum likelihood.

#### **Model based frequencies for usual intake**

<span id="page-136-1"></span>For the model based usual intake distribution the estimated parameters  $\pi_i$  and  $\phi$  are backtransformed using  $\alpha_i =$  $\pi_i \phi/(1 - \phi)$  and  $\beta_i = (1 - \pi_i)\phi/(1 - \phi)$ . These can then be used to draw from the Beta distribution.

#### **Model assisted frequencies for usual intake**

For the model assisted usual intake distribution a prediction of the consumption probability is required for every individual. Simple predictions are

- 1. the observed frequencies for every individual or
- 2. the fitted probability for every individual. When there are no covariables the fitted probability is the same for every individual.
- 3. Alternatively one can use the approach outlined in Kipnis et al (2009) employing the conditional expectation of the probability given the observed frequency:

$$
E(p_i|X_i = x) = \int_p pf(p|X_i = x)dp
$$
  
\n
$$
= \int_p p \frac{f(X_i = x|p)f(p)}{\int f(X_i = x|p)f(p)dp} dp
$$
  
\n
$$
= \frac{1}{P(x_i = x)} \int_p p \binom{n_i}{r} p^x (1-p)^{n_i-x} B^{-1}(\alpha_i, \beta_i) p^{\alpha_i-1} (1-p)^{\beta_i-1} dp
$$
  
\n
$$
= \frac{B^{-1}(\alpha_i, \beta_i)}{P(x_i = x)} \binom{n_i}{r} \int_p p^{\alpha_i+x} (1-p)^{n_i+\beta_i-x-1} dp
$$
  
\n
$$
= \frac{B(\alpha_i + x + 1, n_i + \beta_i - x)}{B(\alpha_i + x, n_i + \beta_i - x)}
$$
  
\n
$$
= \frac{\alpha_i + x}{\alpha_i + \beta_i - x}
$$

For individual with zero intakes on all recall days a prediction for the random individual amount effect  $b_i$  is not available. There seem to be two option for predicting the usual intake for such individuals:

- Set the individual intake to zero
- Simulate a model based prediction for the amount and combine this with the conditional expected probability given above to obtain an individual usual intake.

#### **Logistic-Normal model for frequencies (LNN0)**

<span id="page-136-0"></span>In this model the distribution of  $X_i$  is again binomial:

$$
X_i = \mathit{Binomial}(n_i, p_i)
$$

The probability  $p_i$  is now given by a logistic regression with a random effect in the linear predictor which represents the between-individual variation in the probability  $p_i$ 

$$
\textit{logit}(p_i) = \lambda_i + v_i \text{ with } v_i \sim N(0,\sigma_v^2) \text{ and the regression equation } \lambda_i = \gamma_0 + \gamma_1^t Z_i
$$

The marginal probability  $\pi_i$  is obtained by integrating over the random effect  $v_i$ , i.e. using *Gauss-Hermite integration* 

$$
\pi_i = \int H(\lambda_i + v) f(v) dv \approx \frac{1}{\sqrt{\pi}} \sum_{j=1}^N w_j H(\lambda_i + \sqrt{2} \sigma_v x_j)
$$

in which  $H()$  is the inverse of the logit transformation. Note that this is different from  $logit^{-1}(\lambda_i)$  which is the median probability. The model can be fitted by maximum likelihood using Gauss-Hermite integration. An (approximate) maximum likelihood procedure is implemented in routine glmer of the lme4 package in R. For a new vector of covariates  $Z_i^*$  the linear predictor  $\lambda_i^*$  can be calculated along with its standard error  $Se(\lambda_i^*)$ . The marginal predicted probability  $\pi_i^*$  can be calculated by means of Gauss-Hermite integration and the standard error of the predicted probability can be calculated by means of the usual Taylor series expansion:

$$
\begin{split} \label{eq:sec:reduced:expansion:expansion} \mathit{Se}(\pi_i^*) &\approx \frac{\mathit{Se}(\lambda_i^*)}{\sqrt{\pi}} \sum_{j=1}^N w_j \frac{d}{d\lambda_i^*} H(\lambda_i^* + \sqrt{2}\sigma_v x_j) \\ & = \frac{\mathit{Se}(\lambda_i^*)}{\sqrt{\pi}} \sum_{j=1}^N w_j H(\lambda_i^* + \sqrt{2}\sigma_v x_j)[1-H(\lambda_i^* + \sqrt{2}\sigma_v x_j)] \end{split}
$$

#### **Model based frequencies for usual intake**

For the model based usual intake distribution the estimated parameters  $\lambda_i$  and  $\sigma_v^2$  can be used to generate individual probabilities.

#### **Model assisted frequencies for usual intake**

For the model assisted usual intake distribution simple predictors are (a) the observed frequencies and (b) the marginal probability  $\pi_i$ . The conditional expectation (c) is given by

$$
E(p_i|X_i = x) = \int_v H(\lambda_i + v) f(v|X_i = x) \mathrm{d}v
$$
  

$$
= \int_v H(\lambda_i + v) \frac{f(X_i = x_i|v) f(v)}{\int f(X_i = x_i|v) f(v) \mathrm{d}v} \mathrm{d}v
$$
  

$$
= \frac{\int_v H(\lambda_i + v)[H(\lambda_i + v)]^{x_i} [1 - H(\lambda_i + v)]^{n_i - x_i} f(v) \mathrm{d}v}{\int_v [H(\lambda_i + v)]^{x_i} [1 - H(\lambda_i + v)]^{n_i - x_i} f(v) \mathrm{d}v}
$$

and both nominator and denominator can be approximated by means of the *Gauss-Hermite integration*. For individual with zero intakes on all recall days see above for the two options.

#### **Logistic-Normal model for frequencies correlated with amou[nts \(LNN\)](#page-349-0)**

This model is extends the LNN0 model with a correlation between the individual random effect  $b_i$  for amounts and the individual random effect  $v_i$  for frequencies. This model is also known as the NCI model and is introduced by Tooze et al (2006) [Tooze et al., 2006] with further mathematical details in Kipnis et al (2009) [Kipnis et al., 2009]. The model can be written as

$$
\textit{logit}(P(Y_{ij}>0))=\lambda_i+v_i
$$

$$
g(Y_{ij}) = \mu_i + b_i + w_{ij}
$$

and  $(v_i, b_i) \sim \text{BivariateNormal}(0, 0, \sigma_v^2, \sigma_b^2, \rho)$  and  $w_{ij} \sim N(0, \sigma_w^2)$ 

The model can be fitted by maximum likelihood employing *two-dimensional Gauss-Hermite integration*.

#### **Model based usual intake**

Model based usual intake requires generation of the pair  $(v_i, b_i)$  for many hypothetical individual. The usual intake  $U_i$  for such a hypothetical individual is then given by

$$
U_i = H(\lambda_i + \nu_i)T_i
$$
  
= 
$$
H(\lambda_i + \nu_i)E_w[g^{-1}(\mu_i + b_i + w_{ij})|b_i]
$$
  
= 
$$
H(\lambda_i + \nu_i)F(b_i)
$$

The second term can be calculated using the method outlined for daily intakes.

#### **Model assisted usual intake**

This requires simultaneous prediction of the random effect for frequency and for amount as outlined in Kipnis et al (2009) [Kipnis et al., 2009]. We have for individual *i* in the study  $(U_i|Y_{i1},...,Y_{in_i}) = (U_i|Y_{i1}^*,...,Y_{in_i}^*) =$  $(U_i|x_i, \bar{Y}_i^*)$  where  $x_i$  is the number of positive intakes and  $\bar{Y}_i^*$  is the mean of the transformed **positive** intakes. It follows that the required conditional expectation  $P_i$  equals

$$
P_i = E[U_i | x_i, \bar{Y}_i^*]
$$
  
= 
$$
E_{v_i, b_i} [H(\lambda_i + v_i) F(b_i) | x_i, \bar{Y}_i^*]
$$
  
= 
$$
\frac{\int \int H(\lambda_i + v_i) F(b_i) f(x_i, \bar{Y}_i^* | v_i, b_i) \phi(v_i, b_i) dv_i db_i}{\int \int f(x_i, \bar{Y}_i^* | v_i, b_i) \phi(v_i, b_i) dv_i db_i}
$$

where

$$
f(x_i,\bar{Y}^*_i|v_i,b_i)=[H(\lambda_i+v_i)]^{x_i}[1-H(\lambda_i+v_i)]^{n_i-x_i}\phi(\bar{Y}^*_i-\mu_i-b_i;0,\sigma_w^2/x_i)
$$

Both nominator and denominator can be approximated by a *two-dimensional Gauss-Hermite integration*. Note that for the log-transform  $F(b_i) = T_i = \exp(\mu_i + b_i + \sigma_w^2)/2$  can be calculated exactly; for the *power transformation* an approximation must be used. It can be expected that the predicted usual intake will not have the correct variance. This can possibly be remedied by equating the mean and variance of  $U_i$  and  $P_i$ . These are however rather involved to calculate.

For individual with zero intakes on all recall days the model assisted usual intake can be set to zer[o, or can be simulated](#page-349-2) as follows

- 1. Calculate the Model assisted frequency  $P_0$  for usual intake (see LNN0)
- 2. Transform  $P_0$  back to the logistic scale, i.e.  $L_0 = logit(P_0)$ . Get the conditional distribution of

$$
(b|v = L_0 - \lambda_i) \sim N\left(\frac{\sigma_b}{\sigma_v}\rho(L_0 - \lambda_i), (1-\rho^2)\sigma_b^2\right)
$$

3. Simulate a draw  $b_0$  from this conditional distribution and obtain the usual intake as  $P_0 \exp(\mu_i + b_0 + \sigma_w^2)$ 

Note that the backtransformation from  $P_0$  to  $L_0$  is according to the median of the distribution rather than the mean.

#### **Dietary exposures settings**

**Calculation settings**

<b>Name</b>	Description
Dietary exposure calculation	A tier is a pre-specified set of model configurations. By selecting a
tier	model tier, MCRA automatically sets all model settings in this
	module according to this tier. Note that currently tier setting may
	need to be performed separately in sub-modules. Use the Custom
	tier when you want to manually set each model setting.
Risk type	The type of exposure considered in the assessment; acute (short
	term) or chronic (long-term).
Total diet study concentration	Specifies whether exposure is based on sampling data from total
data	diet studies.
Multiple substances analysis	Specifies whether the assessment involves multiple substances.
Express results in terms of	Specifies whether the assessment involves multiple substances and
reference substance equivalents	results should be cumulated over all substances.
(cumulative)	
Sample based	Include co-occurrence of substances in samples in simulations. If
	checked, substance residue concentrations are sampled using the
	correlations between values on the same sample. If unchecked,
	any correlation between substances is ignored, substance residue
	concentrations are sampled ignoring the correlations between
	values on the same sample.
Consumptions on the same day	if checked, in procedure of EFSA Guidance 2012, section 4.1.1,
come from the same sample	all consumptions of a raw commodity of an individual on the
	same day are assumed to come from the same sample. If
	unchecked, all consumptions of a raw commodity of an individual
	on the same day are assumed to come from different samples.
Maximize co-occurrence of	Within each pattern of substance presence. If checked, substance
high values in simulated	residue concentrations are sorted within co-occurrence patterns of
samples	substances on the same samples. After sorting, high residue values
	occur more frequently on the same sample. This choice is
	conservative. If unchecked, substance residue concentrations are
	sampled at random, ignoring any co-occurrence patterns of
	substances on the same samples. This choice is less conservative.
Apply processing factors	Specified in table ProcessingFactor. If checked, processing factors
	are applied. Concentrations in the consumed food may be different from concentrations in the food as measured in
	monitoring programs (typically raw food) due to processing, such as peeling, washing, cooking etc. If unchecked, no processing
	information is used. This is in most (though not all) cases a
Use distribution	worst-case assumption
Use processing factors higher	
than one	
Use unit variability	Controls whether to use unit variability.
Unit variability model	Describes variation between single units when concentration data
	are from composite samples.
Estimates nature	Simulated unit concentrations can be higher or lower than
	composite value (realistic) or only equal or higher (conservative).
Unit variability parameter	Use Coefficient of variation or Variability factor, specified in
	VariabilityFactor table.
Mean of LogNormal simulated	Unbiased: correct unit simulations for difference between median
values (biasing)	and mean.
Default variability factor for	Default variability factor 1 (unit weight $\leq$ 25 g, small crops). Still
unit weight $\leq$ 25g	requires specification of unit weight (FoodProperties table) and, in
	case of beta model, also the Number of units in a composite
	sample (UnitVariability table).

Table 2.90: Calculation settings for module Dietary exposures.

continues on next page

Name	Description		
Default variability factor for	Default variability factor 5 (unit weight $> 25$ g, medium/large		
unit weight $> 25g$	crops). Still requires specification of unit weight (FoodProperties		
	table) and, in case of beta model, also the Number of units in a		
	composite sample (UnitVariability table).		
Model type	The parametric model for between-and within-individual		
	variation, and possibly covariates.		
Model-then-add	Specifies whether to create separate exposure models for specific		
	groups of foods-as-measured (model-then-add).		
Covariate modelling	Specifies whether to model exposures as a function of covariates		
	at individual level.		
Amount model covariate model	Specifies whether, and how to model exposures amounts as		
	function of covariates.		
Frequency model covariates	Specifies whether, and how to model exposure frequency as		
model	function of covariates.		
Use occurrence patterns for	When selected, this simulated samples will be based on		
generating simulated samples	occurrence patterns.		
Details level dietary exposures	Level of detail for summarizing dietary exposure/intakes.		
Iterate survey	Instead of (re-)sampling the individual days, loop over the entire		
	survey $(= 1$ iteration). The number of iterations for a survey is		
	calculated as round (number of Monte Carlo iterations /(number		
	of individuals * surveys days)).		
Monte Carlo iterations	The number of iterations for Monte Carlo simulations, e.g.		
	100.000 (maximum is 100.000).		
Impute exposure distributions	Impute exposure distributions for substances with missing		
	concentrations.		
Allow conversion using food	Step 3c: try to find read across codes. If unchecked, read across		
extrapolations	table is ignored, default is 'Use read across info'. E.g. for		
	pineapple no measurements are found but by specifying that		
	pineapple is converted to FruitMix (with a default proportion of		
	100%), the TDS sample concentration value of FruitMix will be		
	used for pineapple (as-eaten or as ingredient). If successful,		
	restart at step 1.		
Non-detects replacement	How to replace non-detects (when not co-modelled, as in		
	censored models).		
Default concentration model	The concentration model type that will be used as default for all		
	food/substance combinations. If this model type cannot be fitted,		
	e.g., due to a lack of data, a simpler model will be chosen		
	automatically as a fall-back.		

Table 2.90 – continued from previous page

# **Output settings**

<b>Name</b>	Description
Include drill-down on 9	Specifies whether drilldown on 9 individuals is to be included in
individuals around specified	the output.
percentile.	
Summarize simulated data	Specifies whether a summary of the simulated consumptions and
	concentrations should be included in the output.
Store simulated individual day	Store the simulated individual day exposures. If unchecked, no
exposures	additional output will be generated. If checked, the output will
	contain an additional section with the simulated individual day
	exposures.
Show percentiles for	Give specific percentiles of exposure distribution $(\%)$ , e.g. 50 90
	95 97.5 99 (space separated).
Percentage for drilldown	Gives detailed output for nine individuals near this percentile of
	the exposure distribution.
Percentage for upper tail	Gives detailed output for this upper percentage of the exposure
	distribution.
Show % of population below	Exposure levels can be generated automatically or by explicit
level(s)	specification (Manual).
<b>Exposure levels</b>	Specify exposure levels for which to give the percentage of
	exposure below these levels, e.g. 1 10 50 100 200 500. Specify
	below whether these levels are absolute or relative to ARfD/ADI.
Exposure levels are	Specify whether exposure levels are absolute or percentages of
	ARfD/ADI.
Number of levels of covariable	Specify the number of levels, e.g. 20. The range of the covariable
to predict exposure	is divided by the number of levels: range $=$ $(max - min)/levels$ .
	For these covariable levels exposures are predicted.
Predict exposure at extra	Specify specific prediction levels in addition to the automatically
covariable levels	generated prediction levels (space separated).
Lower percentage for	The default value of 25% may be overruled.
variability (%)	
<b>Upper percentage for</b>	The default value of 75% may be overruled.
variability (%)	
Report consumptions and	Specifies whether body weights should be ignored and
exposures per individual	consumptions and exposures should be expressed per individual.
instead of per kg body weight	Otherwise, the consumptions and exposures are per kg body
	weight.
Cutoff for ratio total exposure/	For selection of individual(day) exposures specify cutoff for ratio
maximum (MCR plot)	total exposure/ maximum.
Show tail percentiles (MCR	Give specific percentiles of exposure distribution $(\%)$ , e.g. 97.5
plot) for	99 (space separated).
Set minimum percentage	Set minimum percentage contribution per substance to the tail
contribution per substance to	exposure.
the tail exposure (MCR plot)	

Table 2.91: Output settings for module Dietary exposures.

# **Uncertainty settings**

Description
Specifies whether to resample the imputated exposure
distributions.

Table 2.92: Uncertainty settings for module Dietary exposures.

## **Dietary exposures tiers**

# **Overview**

<b>Name</b>	Table 2.33. TIET OVERVIEW TOT HOUGHE DICTALY CAPOSITIES. <b>EFSA</b>	<b>EFSA</b>	<b>EFSA</b>	EC 2018	<b>EC 2018</b>
	2012 Op-	2012	2012	Tier 1	Tier <sub>2</sub>
	timistic	Pes-	Pes-		
		simistic -	simistic -		
		Acute	Chronic		
Risk type		Acute	Chronic		
Total diet study	false		false	false	false
concentration data					
Sample based	true	true	true	true	true
Consumptions on the	false	true	true	false	false
same day come from					
the same sample					
Apply processing	true	true	true	true	true
factors					
Use distribution	false	false	false	false	false
Use processing factors	false	true	true	false	false
higher than one					
Use unit variability	false	true		true	true
Unit variability model	NoUnit-	BetaDis-		BetaDis-	BetaDis-
	Variability	tribution		tribution	tribution
Estimates nature		Realistic		Realistic	Realistic
Unit variability		Variabili-		Variabili-	Variabili-
parameter		tyFactor		tyFactor	tyFactor
Model type	$\overline{\text{OIM}}$		$\overline{\text{OIM}}$	$\overline{\text{OIM}}$	$\overline{\text{OIM}}$
Model-then-add	false		false	false	false
Covariate modelling	false	false	false	false	false
Iterate survey	false	false	false	false	false
Report consumptions	false	false	false	false	false
and exposures per					
individual instead of					
per kg body weight					

Table 2.93: Tier overview for module Dietary exposures.

#### **EFSA 2012 Optimistic**

Use the optimistic model settings according to the EFSA Guidance 2012. Concentration values are sampled using a sample-based empirical distribution. Available processing factors are applied. No unit variability model should be applied.

Name	Setting
Total diet study concentration data	false
Sample based	true
Consumptions on the same day come from the same sample	false
Apply processing factors	true
Use distribution	false
Use processing factors higher than one	false
Use unit variability	false
Unit variability model	NoUnitVariability
Model type	<b>OIM</b>
Model-then-add	false
Covariate modelling	false
Covariate modelling	false
Iterate survey	false
Report consumptions and exposures per individual instead of per	false
kg body weight	

Table 2.94: Tier definition for EFSA 2012 Optimistic.

#### **Input tiers**

Table 2.95: Input tiers for EFSA 2012 Optimistic.

Module	Input tier
Concentration models	EFSA 2012 Optimistic

## **EFSA 2012 Pessimistic - Acute**

Acute probabilistic exposure assessment using the pessimistic model settings according to the EFSA Guidance 2012. Only processing factors > 1 are applied. For unit variability, the Beta distribution is applied.

Table 2.96: Tier definition for EFSA 2012 Pessimistic - Acute.

Name	Setting
Risk type	Acute
Sample based	true
Consumptions on the same day come from the same sample	true
Apply processing factors	true
Use distribution	false
Use processing factors higher than one	true
Use unit variability	true
Unit variability model	<b>BetaDistribution</b>
Estimates nature	Realistic
Unit variability parameter	VariabilityFactor
Covariate modelling	false
Iterate survey	false
Report consumptions and exposures per individual instead of per	false
kg body weight	




### **EFSA 2012 Pessimistic - Chronic**

Chronic probabilistic exposure assessment using the pessimistic model settings according to the EFSA Guidance 2012. Only processing factors > 1 are applied.

Name	Setting
Risk type	Chronic
Total diet study concentration data	false
Sample based	true
Consumptions on the same day come from the same sample	true
Apply processing factors	true
Use distribution	false
Use processing factors higher than one	true
Model type	<b>OIM</b>
Model-then-add	false
Covariate modelling	false
Iterate survey	false
Report consumptions and exposures per individual instead of per	false
kg body weight	

Table 2.98: Tier definition for EFSA 2012 Pessimistic - Chronic.

## **Input tiers**





Name	Setting
Total diet study concentration data	false
Sample based	true
Consumptions on the same day come from the same sample	false
Apply processing factors	true
Use distribution	false
Use processing factors higher than one	false
Use unit variability	true
Unit variability model	<b>BetaDistribution</b>
Estimates nature	Realistic
Unit variability parameter	VariabilityFactor
Model type	<b>OIM</b>
Model-then-add	false
Covariate modelling	false
Iterate survey	false
Report consumptions and exposures per individual instead of per	false
kg body weight	

Table 2.100: Tier definition for EC 2018 Tier 1.

## **Input tiers**

Table 2.101: Input tiers for EC 2018 Tier 1.

Module	Input tier
Concentration models $\mid EC$ 2018 Tier 1	

# **EC 2018 Tier 2**

Name	Setting
Total diet study concentration data	false
Sample based	true
Consumptions on the same day come from the same sample	false
Apply processing factors	true
Use distribution	false
Use processing factors higher than one	false
Use unit variability	true
Unit variability model	<b>BetaDistribution</b>
Estimates nature	Realistic
Unit variability parameter	VariabilityFactor
Model type	<b>OIM</b>
Model-then-add	false
Covariate modelling	false
Iterate survey	false
Report consumptions and exposures per individual instead of per	false
kg body weight	

Table 2.102: Tier definition for EC 2018 Tier 2.



#### **EFSA 2012 Pessimistic**

**Note:** This tier is deprecated and has been replaced by separate acute/chronic tiers.

Probabilistic exposure assessment using the pessimistic model settings according to the EFSA Guidance 2012. Only processing factors > 1 are applied. For unit variability, the Beta distribution is applied.

Name	Setting
Total diet study concentration data	false
Sample based	true
Consumptions on the same day come from the same sample	true
Apply processing factors	true
Use distribution	false
Use processing factors higher than one	true
Use unit variability	true
Unit variability model	<b>BetaDistribution</b>
Estimates nature	Realistic
Unit variability parameter	VariabilityFactor
Model type	<b>OIM</b>
Model-then-add	false
Covariate modelling	false
Iterate survey	false
Report consumptions and exposures per individual instead of per	false
kg body weight	

Table 2.104: Tier definition for EFSA 2012 Pessimistic.

#### **Input tiers**

Table 2.105: Input tiers for EFSA 2012 Pessimistic.



#### **Calculation of dietary exposures**

Dietary exposures are calculated from consumptions per food-as-measured and concentration models. Optionally, also processing factors and unit variability models are applied.

• *Dietary exposures calculation*

Inputs used: *Consumptions by modelled food Concentration models Processing factors Unit variability factors High exposure food-substance combinations Active substances Occurrence patterns Relative potency factors Food conversions Conce[ntration distributions](#page-116-0)*

#### Settings used

• *[Calculation Settings](#page-147-0)*

# **2.4.3 High exposure food substance combinations**

Identification of food-as-eaten/food-as-measured/substance combinations that have the highest expected contribution to exposure based on a simple screening model.

<span id="page-147-0"></span>This module has as primary entities: *Foods Substances Effects*

Output of this module is used by: *Dietary exposures*

## **High exposure food substanc[e com](#page-27-0)[binations](#page-44-0) [calcu](#page-25-0)lation**

<span id="page-147-1"></span>A full Monte Carlo analysis can b[e unwieldy for larg](#page-116-1)e cumulative assessment groups (CAGs) and/or large number of foods or concentration data. An algorithmic approach was developed to handle large CAGs. Two unique features of MCRA are:

- contributions to the exposure results can be seen both in terms of foods as eaten (e.g. white bread) and foods as measured (e.g. wheat), and
- a drill-down can be made into the exact foods and substances contributing for simulated individuals or individual-days in the upper tail.

The number of combinations of simulation, substance, modelled food and food as eaten can be very large. To avoid memory problems with very large datasets, an additional optional modelling step, named *screening*, was added to MCRA. *Screening* should be used if the data dimensions are too large for a direct analysis. Screening identifies risk drivers. A full analysis based on screened risk drivers will still retain all food/substance combinations in the exposure calculation, and will therefore produce exactly the same cumulative exposure distribution, and allow to see contributions of all substances and all foods-as-measured. Details with respect to foods as eaten are however restricted to the risk drivers selected in the screening step. For more details see *screening calculation for large Cumulative Assessment Groups*.

The two-step approach consists of:

- **Step 1: Data screening and selection of risk drivers** Run a [simple analysis for each potential](#page-148-0) [source/](#page-148-0)substance combination (SCC). Here source means the combination of food as eaten and modelled food, for example apple in apple pie. The screening is based on this combination, and not just foods as measured, to avoid problems with potentially multi-modal consumption distributions as much as possible (see van der Voet et al. 2014). SCCs are also referred to as risk driver components. The screening step in MCRA implements a simple model that is applied to each SCC. The model calculates a percentile of interest in a distribution, consisting of a spike of zeroes (non-consumptions), and a mixture of two lognormal distributions for the exposure related to non-detects and positive concentrations, respectively. SCCs (risk driver components) can be combined to measured source/substance combinations (MSCCs, risk drivers). For example APPLE/apple juice/captan and APPLE/apple pie/captan combine to APPLE/captan. MCRA has an interface which identifies the Top- $N$  SCCs (based on a chosen exposure percentile, e.g.  $p95$ ) with an option to select  $N$  based on cumulative importance according to some criterion. Remark: Screening is performed before concentration modelling. Therefore there is no correction for processing at the screening stage. Note, originally SCC stands for Source Compound Combination, MSCC for Measured Source Compound Combination.
- **Step 2: Full MC analysis** Perform the standard MC to all combinations of substances and foods, but restrict the stored information regarding foods as eaten to the SCCs selected in step 1.

The screening method requires to specify:

- Which percentile to consider for each single source/substance combination (SCC, potential risk driver component) (default p95)
- Which percentage of all exposures (according to the screening model) should be covered by the selected set of SCCs (default 95%)
- How to impute non-detect concentrations  $(c < LOR)$  in the screening step. The choice of a factor 0 (default) represents optimistic imputation, the choice of a factor 1 represents a pessimistic imputation. It may be noted that a factor 1 (pessimistic imputation) may select many SCCs (risk driver components) with relatively high

LORs and high RPFs, but with only nondetect measurements. Choosing a lower fraction, e.g. 0.25 can be useful if a more realistic method is sought.

Based on limited experience with the EFSA test data, useful settings of these three screening parameters were found to be (95, 95, 0) in preparation for an EFSA optimistic run, and (50, 95, 0.25) in preparation for an EFSA pessimistic run. See also screening calculation *acute exposure* and *chronic exposure* .

#### **Screening calculation for large Cumulative Assessment Groups**

#### **Statistical model for the screening step (acute exposure)**

<span id="page-148-2"></span><span id="page-148-0"></span>The screening step implements a simple model that is applied to each SCC. Assume independent *NonDetectSpike-LogNormal* (NDS-LN) models for both the consumptions of food-as-measured in source S and the concentrations of substance C in source S. A non-detect consumption is assumed to be a zero consumption. A non-detect concentration will be imputed by a user-specified fraction f of the Limit of Reporting. Then the model for consumption has 3 parameters and the model for concentration has four parameters, as specified in Table 2.106. Note th[at the parameters](#page-62-0) [of the consu](#page-62-0)mption distribution are estimated from the consumption data using sampling weights if these have been provided in the consumption data set.

parameter	consumptions	concentrations
probability of a positive	$\pi_r$	$\pi_{c}$
mean positives (ln scale)	$\mu_x$	$\mu_c$
standard deviation positives (ln scale)	$\sigma_r$	$\sigma_{\alpha}$
value to use for NonDetects (ln scale)		$\cdot$ 1

Table 2.106: Parameters for screening models (per so[urce/substanc](#page-148-1)e)

<span id="page-148-1"></span>Exposure is consumption times concentration, so on logarithmic scale they can be added:

 $e = x + c$ .

The assessment will focus on a chosen percentile of exposure, e.g. p95. The relevant fraction will be denoted by  $p$ , for example  $p = 0.95$  for the 95th percentile. The two NDS-LN models combine to three possibilities, depending on whether there is consumption and if so, whether the concentration is non-detect or positive. In the screening model the two possibilities that lead to potential exposure are modelled with a mixture of two lognormal distribution. For the non-detect case the positive exposure distribution equals the positive consumption distribution modified by the multiplication of a user-chosen factor times an estimate of the average worst-case limit value for concentration (LOR):

$$
\pi_1=\pi_x(1-\pi_c);\mu_1=\mu_x+f\cdot L_c;\sigma_1=\sigma_x
$$

where  $L_c$  is the logarithm of the LOR, or, if there are multiple analytical methods with different LOR, a weighted average of these different LORs.

For the detect case the positive exposure distribution is easily combined from the positive consumption distribution and the positive concentration distribution:

$$
\pi_2 = \pi_x \pi_c; \mu_2 = \mu_x + \mu_c; \sigma_{12} = \sqrt{\sigma_x^2 + \sigma_c^2}
$$

 $p$  can be corrected for the non-consumptions to the appropriate fraction needed in the mixture of the two positive distributions:

$$
p' = \frac{p - (1 - \pi_x)}{\pi_x}
$$

If  $p' \leq 0$  then all positive exposures are beyond the requested fraction, and the estimated exposure is just 0.

If  $p' > 0$  then the relevant log exposure  $e_p$  satisfies

$$
(1-\pi_c)\cdot \Phi\left(\frac{e_p-\mu_1}{\sigma_1}\right)+\pi_c\cdot \Phi\left(\frac{e_p-\mu_{12}}{\sigma_2}\right)=p'
$$

where  $\Phi(\cdot)$  represents the cumulative standard normal distribution function. The value of  $e_p$  can easily be found in a bisection search within the interval

$$
[\mu_{min}-4\sigma_{max},\mu_{max}+max(0,z_{p'}\sigma_{max})].
$$

The final exposure percentile estimate then is  $\exp(e_p)$ .

Denote by  $e_{(p,max)}$  the highest estimate (for the SCC denoted by  $SSC_{highest}$ ). Then evaluate for each SCC the probability to exceed  $e_{(p, max)}$ .

$$
P_i = Pr(e > e_{p,max}) = \pi_x \cdot \left[ (1-\pi_c)*\Phi\left(\frac{e_{p,max}-\mu_1}{\sigma_1}\right) + \pi_c \cdot \Phi\left(\frac{e_{p,max}-\mu_2}{\sigma_1}\right) \right]
$$

 $P_i$  is a tentative measure for the 'probability of a high exposure'. For  $SSC_{highest}$   $P_i = 1 - p$ , for all other SCCs it will be lower. The sum of all these probabilities is not a meaningful probability in itself. However, this sum is used to scale the individual  $P_i$  values to measures of relative importance for the SCCs

$$
Imp_i = P_i / \sum P_i
$$

Rank all SCCs according to *Imp<sub>i</sub>* and calculate cumulative importance. The relative importance of the two mixture components at  $e_p$  can be estimated as

$$
w_{1,2} = \frac{\pi_{1,2} \cdot \phi\left(\frac{e_p - \mu_{1,2}}{\sigma_{1,2}}\right) / \sigma_{1,2}}{\pi_1 \cdot \phi\left(\frac{e_p - \mu_1}{\sigma_1}\right) / \sigma_1 + \pi_2 \cdot \phi\left(\frac{e_p - \mu_2}{\sigma_2}\right) / \sigma_2}
$$

where  $\phi(.)$  represent the standard normal probability density function. The user interface should allow to select the top- $N$  SCCs from the list, based on a chosen percentage (e.g. 95%) of cumulative importance included. The full analysis will calculate exactly the same exposure distribution as a full analysis without screening. However, less information is retained in the output. This concerns tables with information on foods-as-eaten, which is only shown for the selected risk driver components (SCCs). Risk drivers are groupings of SCCs (risk driver components) at the level of measured-source-substance combinations (MSCCs). Note that output for an MSSC (e.g. APPLE/captan) only covers the selected SCCs (e.g. APPLE from apple juice/captan and APPLE from apple pie/captan), but not unselected SCCs (e.g. APPLE from fruit yoghurt/captan).

#### **Statistical model for the screening step (chronic exposure)**

In chronic exposure assessments, the mean concentration of chemicals is calculated first, and combined with the consumption distribution. For this reason a chronic calculation uses less memory, and therefore larger datasets can be handled.

The model described under *acute exposure* can be simplified for a chronic screening. The concentration distribution is only used to estimate a mean exposure, incorporating any effect from the imputation of non-detects. The exposure distribution is therefore only a scaled version of the consumption distribution.

$$
\pi_2=\pi_x\pi_c; \mu_2=\mu_x+\mu_c; \sigma_2=\sigma_x
$$

The parameters of the consumption distribution  $(\pi_x, \mu_x, \sigma_x)$  are calculated from the observed individual means (*OIM*), i.e. the mean daily consumptions over the survey days of each person in the data (allowing for sampling weights). The percentiles are calculated as  $e_p = \mu_2 + z_p$  where z is a percentile of the standard normal distribution. The exceedances of the maximum percentile are calculated as

$$
P_i = Pr(e > e_{p,max}) = \pi_x \cdot \Phi\left(\frac{e_{p,max} - \mu_2}{\sigma_2}\right)
$$

## **Calculation settings**



<span id="page-150-0"></span>

## **Calculation of high exposure food-substance combinations**

Screening results are computed for each combination of source (being a specific combination of food-as-eaten/foodas-measured) and substance by combining simple approximations of the consumption and the concentration distribution.

• *High exposure food-substance combinations calculation*

Inputs used: *Consumptions by modelled food Concentration models Active substances Relative potency factors*

Settings used

• *[Calculation Settings](#page-147-1)*

# **2.4.4 Exposures**

Expos[ures are amounts of](#page-150-0) substances, typically expressed per mass unit and per day, to which individuals in a population are exposed at a chosen target level. This target level may be external exposure (dietary exposure, expressed per unit body weight, or per person) or internal exposure (expressed per unit organ weight). Internal exposures may be aggregated from dietary and non-dietary exposures using either absorption factors or kinetic models to translate the external exposures to internal exposures. Exposures can be short-term/acute exposures and then contain exposures for individual-days, or they can be long-term/chronic exposures, in which case they represent the average exposure per day over an unspecified longer time period.

This module has as primary entities: *Populations Foods Substances*

Output of this module is used by: *Exposure mixtures Human monitoring analysis Risks*

### **Exposures calculation**

Calculation of exposures comprises two main steps:

- <span id="page-151-1"></span>1. Linking *dietary and non-dietary individual/individual-day exposures*.
- 2. Computing *the (aggregated) internal exposures at the specified target compartment*.

Both steps are optional in this module. If none is selected, exposures are external *dietary exposures*, i.e the target level is extern[al/dietary. However, when multiple routes of exposure are c](#page-151-0)onsidered, then the target level should be an internal compartment (organ). In the latter case, *absorption factors* or *kinetic model* are needed to aggregate the exposures from [multiple routes into exposure at the target compartment. It is also poss](#page-155-0)ible to only provide dietary exposures and compute internal exposures at some target compartment.

In cumulative exposure calculations two simple app[roaches are used to](#page-155-1) id[entify and sel](#page-156-0)ect mixtures contributing to the exposure of a target population:

- 1. qualitative approach: *counting of co-exposure*. To which combinations of substances are individuals exposed?
- 2. quantitative approach: *maximum cumulative ratio (MCR)*. To what degree are mixtures more important than single substances?

A quantitative approach isa[vailable in the](#page-158-0) *exposures mixtures module*.

## **Combining dietary and non-dietary exposures**

<span id="page-151-0"></span>If *dietary* and *non-dietary exposures* are available for the same individuals or individual-days, the non-dietary exposures can be matched to specific individuals of the food survey from which the dietary exposures are generated. More commonly, dietary and non-dietary exposures are available from separate surveys, in which case they can be randomly combined. If both dietary and non-dietary information is available for a known population of individuals, th[e user m](#page-116-1)ay select the *[matching opti](#page-189-0)on* such that specific dietary and non-dietary estimates are aggregated for each individual in the food survey population. If matching is enabled, any non-dietary exposures that do not correspond to individuals from the food survey will be ignored (see *Example 2*), unless an individual is specified with *id = General*. In that case, the dietary individual should meet the criteria of the non-dietary survey, specified by the survey properties, to be assigned. Ift[he non-dietary da](#page-164-0)ta relates instead to a population in which individuals have no corresponding records in the food survey (unmatched case), the user may choose to randomly assign the non-dietary exposures to the individuals from the food survey.

When multiple non-dietary surveys are available, the options with or without correlation are important (not relevant when matching is switched on). When correlation is chosen, the exposure contributions of non-dietary individuals with identical ids in different surveys are combined and allocated to a randomly selected dietary individual. When the correlation is not chosen, the non-dietary exposures of randomly selected individuals from different surveys are combined and allocated to a dietary individual.

The user may also define demographic criteria for the assignment (for each source of non-dietary exposure) to indicate that those exposures are relevant only to a defined sub-population. Only those individuals in the food survey who meet the criteria of the non-dietary survey will be assigned non-dietary exposures from that source e.g. only males aged 18 to 65 (see *Example 1*). The simplest assessment consists of a single (deterministic) non-dietary exposure estimate which is assigned to all individuals in the food survey (*idIndividual = General*). This case, and more complex possibilities are illustrated below using hypothetical examples.



Figure 2.22: Aggregate exposure distributions.

## **Example 1**

Deterministic cumulative (multi-substance) non-dietary exposure input, adult male sub-population. Unmatched case.



<span id="page-152-0"></span>



idNonDietary-	Description	Location	Date	NonDietary-
Survey				IntakeUnit
	BROWSE, acute,	York	09/10/2012	$\mu q / day$
	cumulative,			
	operators			

Table 2.110: NonDietarySurveyProperties

<span id="page-152-1"></span>

In this example, there are exposure values for multiple substances in Table 2.108 and the user has provided Table 2.110 which specifies that the non-dietary exposures given in survey number 1 relate to males aged 18 to 65.

When this assessment is performed, only those individuals whose property values fit the criteria in Table 2.110 will receive the non-dietary exposures in survey 1. The use of *idI[ndividual = G](#page-152-0)eneral* indicates that this [is the](#page-152-1) default exposure. All individuals in the dietary survey meeting the criteria receive all exposures given in the 3 rows, corresponding to 3 substances. The following should be noted:

- There should only ever be one *General* entry in the dietary exposures table per substance, survey combination.
- The property names and the values of any text properties must precisely match those given in the **Individual-Properties** and **IndividualPropertyValues** tables for this to work.
- The minimum and maximum values for numeric properties are both inclusive boundaries.

So in this example, all males aged 18 to 65 will receive the given exposures of all three substances and the other members of the population will receive no non-dietary exposure. Note that example 1 describes the unmatched case.

## **Example 2**

<span id="page-153-0"></span>Variability (but no uncertainty) in cumulative non-dietary exposure input (matched to dietary survey individuals).

idIndividual	idNonDietarySurvey	idSubstance	Dermal	Oral
5432		011003001	10	5
5432		011003002	33	22
5433		011003001	12	7
5433		011003002	34	23
5434		011003001	18	9
5434		011003002	35	25
5435		011003001	10	5
5435		011003002	33	21

Table 2.111: NonDietaryExposures





In this example, the non-dietary exposures are being specified explicitly for individuals in the dietary population. Switch 'matching' on to allow exposure variability to be specified at the individual level. For the purposes of illustration, the population is extremely small, consisting of only four individuals. The values in the *idIndividual* column of Table 2.111 match those in the **Individuals** table (dietary population).

It is not mandatory to specify exposures for every individual in the population. Those not included will simply receive a zero non-dietary exposure, unless there is also a default exposure value (*General* row(s) in Table 2.111) and the individual matches the specified demographic criteria for the survey, as specified in Table 2.110. (In this example, a [default expo](#page-153-0)sure would apply to all individuals not listed in Table 2.111 because Table 2.110 has not been used).

There is variability between individuals in this example, but no uncertainty in exposure. Note that these data could also be used with matching switched off. This would be the same as treating the *idIndividual* [values as](#page-153-0) generic individuals, so that each pair of exposure lines would be as[signed at rand](#page-153-0)om toi[ndividuals meeti](#page-152-1)ng the criteria.

## **Example 3**

Variability (no uncertainty) in cumulative non-dietary exposure input (unmatched individuals).

<span id="page-154-0"></span>

idIndividual	idNonDietarySurvey	idSubstance	Dermal	Oral	Inhalation
ND <sub>1</sub>		011003001	10	5	17
ND <sub>1</sub>		011003002	33	22	45
ND2		011003001	12	7	18
ND2		011003002	34	23	47
ND <sub>3</sub>		011003001	18	9	19
ND <sub>3</sub>		011003002	35	25	49
ND <sub>4</sub>		011003001	10	5	17
ND <sub>4</sub>		011003002	33	21	45

Table 2.113: NonDietaryExposures

### Table 2.114: NonDietarySurveys



#### Table 2.115: NonDietarySurveyProperties

<span id="page-154-1"></span>

This example is similar to example 2, except that the values in the *idIndividual* column of Table 2.113 do not match those in the **Individuals** table. In this instance, 'matching' would not be an option, and the non-dietary exposures would be randomly assigned to individuals who meet the criteria in Table 2.115. (In fact for the same result rather than changing the values in the *idIndividual* column in Table 2.111 from the previous example may be used with matching switched off). Exposures in Table 2.113 will be recycled if the number of exp[osure rows is](#page-154-0) less than the number of dietary records with which to aggregate exposures.

Again, there is variability between individuals in this exa[mple, but n](#page-153-0)[o uncertainty](#page-154-1) in exposure.

By allowing generic *idIndividual* valu[es in this wa](#page-154-0)y, correlations between different sources (within individual) can be accounted for even in the unmatched case. If the same *idIndividual* value is used in different surveys, then the corresponding exposure values will be kept together and assigned to an eligible individual as a combined exposure.

So for option matching switched of, it is relevant whether individuals are correlated or not. In the following example, two non-dietary surveys are available, per survey three individuals are specified.

idIndividual	idNonDietarySurvey	idSubstance	Dermal	Oral	<b>Inhalation</b>
ND0		011003001	10		
ND <sub>1</sub>		011003001	23	22	45
N <sub>D</sub> 2		011003001	12		18
ND0		011003001	34	23	47
ND3		011003001	18	Q	19
N <sub>D</sub> 4		011003001	33	16	35

Table 2.116: matching switched of, with correlation or without.

• When a correlation is applied, the non-dietary exposure for individual ND0 from survey 1 and 2 are combined and allocated to a dietary individual. For individual ND1, ND2, ND3 and ND4 just a single non-dietary exposure is found and allocated to a dietary individual.

• When no correlation is applied, the exposure for individual ND0 from survey 1 is combined with one of the exposures of ND0, ND3 or ND4 from survey 2; exposure of ND1 from survey 1 is combined with one of the exposures of ND0, ND3 or ND4 from survey 2, etc.

When the intention is to sample just one exposure for a dietary individual, specify per survey different codes, e.g. ND1, ND2, ND3 for survey 1, ND4, ND5, ND6 for survey 2 and apply correlation, or specify 6 different individual codes and just one *idNonDietarySurvey*. Then, options with or without correlation are irrelevant and sampling results are identical no matter which option is chosen.





Figure 2.23: Contributions by route to aggregate exposure distributions.

See *non-dietary exposure settings*.

### **Internal exposures calculation**

<span id="page-155-0"></span>Computation of internal exposures (internal substance amounts and concentrations) requires a *kinetic model* to translate external doses, possibly from multiple routes, to internal doses at the target compartment/organ of interest.

## **Calculation of internal concentrations using absorption factors**

<span id="page-155-1"></span>In the simplest form, internal concentrations are derived from external exposure concentrations using multiplication factors (or, absorption factors) that can be specified by substance and by route. That is, for a given substance, the internal exposure  $exp_{int}$  is computed as

$$
\exp_{\text{int}} = \sum_{r \in \text{Routers}} f_{\text{abs},r} \cdot \exp_{\text{ext},r}
$$

Here, *Routes* denotes the set external exposure routes,  $exp_{ext,r}$  denotes the external exposure for route r and  $f_{abs,r}$ denotes the absorption factor of route  $r$ . Note that this model assumes that both external and internal exposures refer to amounts or concentrations depending on the *dietary exposures* setting (External exposure: substance amount per individual, or substance amount divided by body weight; internal exposure: substance amount per organ, or substance amount divided by organ weight.) Also, both external and internal exposures are expressed per day.

#### **Calculation of internal concentrations using kinetic models**

<span id="page-156-0"></span>A more detailed alternative to using absorption factors is to use one of the *advanced kinetic models* available in MCRA. In this approach, for each substance independently, the external exposures of an individual (chronic) or individual-day (acute) are presented for a number of simulated day to a PBK model of the individual. This yields a time course of the internal substance amount at the specified target compartment/organ from which a long term average substance amount (chronic) or peak substance amount (acute) can be obtained. [An example of such a tim](#page-246-0)e course is given in Figure 2.24 for acute exposure assessments, and in Figure 2.25 for chronic exposure assessments. By dividing this substance amount by the weight of the compartment, an internal concentration is obtained. Notice that this procedure also changes the unit of the exposures from exposure per day to long term exposure.

<span id="page-156-1"></span>

#### **[Model COS](#page-157-0)MOSv6**

Figure 2.24: Time course of the internal substance amount when applying the same single dose on each day. The acute internal concentration is derived as the peak substance amount (the green line in the figure) divided by the compartment weight. The vertical line at 50 indicates the selected end of an assumed non-stationary period, defining a burn-in period that is to be ignored for computing the peak substance amount.

Mathematically, the calculation of the peak substance amount  $(d_{peak})$  for deriving acute internal exposures is as follows:

$$
d_{\text{peak}} = \max_{i=0,\ldots,n_{\text{stop}}} \left\{ d(t_{\text{start}}+i\Delta t) \right\}.
$$

Here,  $d(t)$  denotes the substance amount at time  $t$ ,  $t_{\text{start}}$  denotes the starting time of the evaluation window (defined by the *non-stationary period*),  $\Delta t$  denotes the time resolution of the kinetic model (e.g., hours or minutes), and  $n_{\text{stop}}$ denotes the total number of time-points, marking the end of the evaluation window (defined by the specified number of simulation days), which is computed as

$$
n_{\rm stop} = \left\lfloor \frac{t_{\rm stop} - t_{\rm start}}{\Delta t} \right\rfloor.
$$

Likewise, chronic long term average substance amounts  $(d_{\text{avg}})$  are computed as:

$$
d_{\text{avg}} = \frac{\sum_{i=0}^{n_{\text{stop}}} d(t_{\text{start}} + i\Delta t)}{n_{\text{stop}}}.
$$

<span id="page-157-0"></span>

Figure 2.25: Time course of the internal substance amount when randomly applying one of the individual-day doses for a number days. The chronic internal concentration is derived as the average substance amount (the blue line in the figure), divided by the compartment weight. The vertical line at 50 indicates the selected end of an assumed non-stationary period, defining a burn-in period that is to be ignored for computing the average substance amount.

### **Dosing patterns**

In MCRA, the dietary and non-dietary exposures are computed at the level of exposures per day. However, when applying advanced PBK models, dosing patterns may be specified at a much finer resolution (e.g., hours or minutes). For this, a method is needed to translate external exposures provided per day to dosing patterns of substance amounts during the day. The simplest, yet not very realistic model is to apply, per route, the full exposure amount in one single dose at the beginning of the day. Alternatively, MCRA offers the possibility to specify, per route, the *number of exposure events per day*. If it is specified to use multiple doses per day, then the total substance amount of each day is divided into equal portions which are applied at regular time-intervals during the day.

#### **[Non-stationary peri](#page-245-0)od**

Especially in the case of chronic exposure assessments, where a long term average exposure is computed based on the simulated time-course, it is important to realise that at time zero, the substance is commonly considered to be completely absent in the simulated system. However, this is not a realistic assumption. It is much more likely that the substance was already present in the system, and that the level is equal to the level obtained from applying the same chronic exposures to the system. For this, a specification of the *number of days skipped* (or burn-in period) is required in order to come to these initial concentration levels. This period is not used for computing the long term average or peak exposures, but just to determine initial (background) concentration levels.

#### **Counting of co-exposure**

<span id="page-158-0"></span>In this qualitative approach, the number of combinations of substances to which an individual is exposed are recorded, see Table 2.117. There is no cut-off level, the only criterion is the presence of a substance in the simulated daily diet or not. For an *acute* or short term exposure assessment, a simulated individual day is the smallest entity to determine co-exposure. For a *chronic* or long term exposure assessment, co-exposures are summarized at the individual level, e.g. co-exposure is determined combining all consumption days of an individual.

Substance	day 1	day 2	day 3	$\cdots$	day n
Tebuconazole	x			$\cdots$	
<b>Bitertanol</b>	x			.	X
Triadimefon	x		.	$\cdots$	X
$\cdots$	.				

Table 2.117: Counting combinations of substances in the exposure matrix: [for exa](#page-122-0)mple, on day 1 there is coexposure to substances Tebuconazole, Bitertanol and Triadimefon

In Table 2.118, the frequency and percentage for the number of substances occurring together are shown.

Number of substances	Frequency	Percentage
	337	3.4
	959	9.6
	1207	12.1
	1275	12.8
.	$\ddot{\phantom{0}}$	.

Table 2.118: Co-exposure of substances

In Table 2.119, the mixtures containing the substance(s) including all other combinations with the specified combination of substance(s), (a maximum number of 15 records is shown).

Table 2.119: Mixtures containing substances

Number of substances	Percentage	<b>Substances</b>
	5.88	Tebuconazole
	3.88	Imazalil (aka enilconazole), Tebuconazole
	3.37	
っ	2.20	Difenoconazole, Imazalil (aka enilconazole), Tebuconazole
	1.78	Imazalil (aka enilconazole)
	1.76	Imazalil (aka enilconazole), Tebuconazole, Triadimenol
$\cdots$	$\cdots$	$\cdots$

## **Maximum Cumulative Ratio**

Price and Han [Price et al., 2011] propose the Maximum Cumulative Ratio (MCR) which is defined as the ratio of the cumulative exposure received by an individual on an intake day to the largest exposure received from a single substance:

### MCR = [Cumulative exposu](#page-356-0)re/ Maximum exposure

This MCR statistic is also picked up as a practical device in a recent JRC report [Bopp et al., 2015] to investigate cumulative exposure. If MCR is large, it is important to consider cumulative effects, if MCR is close to 1, the individual exposure will not be much different from a single-substance assessment. The MCR can therefore be interpreted as the degree to which the risk of being exposed is underestimated by not performing a cumulative risk assessment.

The MCR statistic is implemented in MCRA for both the *acute* risk and the chronic risk cases. In the acute risk case the short-term (single-day) exposures are used, in the *chronic* case the long-term individual exposures (estimated by aggregating over the available survey days of each individual).

Table 2.120 shows an artificial example how the MCR is calculated in the acute risk case. First the cumulative exposure per day is calculated by cumulating the exposure [of eac](#page-117-0)h substance multiplied by the *relative potency factors* (RPF). Then, for each day, the cumulative exposure [\(in equ](#page-122-0)ivalents of the reference substance) is divided by the maximum exposure of a single substance on that day. The last column shows the MCR values within parenthesis [the substance](#page-159-0) with the highest exposure. The MCR has a value of 1 or close to 1 for mixtures where the exposure is dominated by one substance (e.g. day 1, substance B). When all substances have approximat[ely equal exposure \(e.g.](#page-237-0) day 3) the MCR value is equal or close to the number of substances, here 4. Day 2 represents an intermediate case. The MCR suggest that for exposure days (or persons) with MCR values close to 1, the need for a cumulative risk assessment is low.

<span id="page-159-0"></span>

	Substance A	Substance B	Substance C	Substance D	total exposure	ratio
day	0.01	0.99				1.01(B)
day 2	0.1	0.2	0.3	0.4		2.50(D)
day <sub>3</sub>		0.25	0.24	0.26		$3.99$ (D)

Table 2.120: Maximum Cumulative Ratios

In the example, all days have equal values for total exposure. For real data, total exposure will vary. It is obviously of interest to know if the MCR is high or low at those days (or individuals) where the total exposure is highest.

In Figure 2.26, French steatosis data (39 substances, 4079 persons) are used to calculate the chronic exposure matrix. For each individual the MCR is calculated and plotted against the total exposure. The different colours are used to identify the single substances with maximum exposure. From the original 39 substances, 10 different substances have the largest exposures. For the total exposure and MCR, the p5, p50 and p95 percentiles are indicated with the black li[ne segments.](#page-160-0) The red line indicates the ratio with value 5. The dashed green lines indicate the p95 percentiles for the MCR value for different ranges of the total exposure.

The plot shows that MCR values with Imazalil as risk driving substance (purple) are predominantly found in the lower part of the plot for relatively high values of the total exposure. A second finding is that MCR values decline when total exposure increases. This implies that cumulative exposure for most individuals is driven by multiple substances. At the right site of the plot, individuals are found with high exposure. Because MCR values tend to be lower here, higher

<span id="page-160-0"></span>

### **Using MCR to identify substances that drive cumulative exposures**

Figure 2.26: Maximum Cumulative Ratios vs total exposure

exposures are received from one predominant substance and not because many substances are above the average level. For those individuals a cumulative risk assessment has less value.

Because Figure 2.26 can be very dense, in Figure 2.27, 95% confidence regions representing bivariate lognormal distributions of MCR and total exposure are plotted. The latter figure facilitates interpretation of the first figure. Note that substances with just one or two observations cannot be plotted in this display (substances with 2 observations are represented by a line).

<span id="page-161-0"></span>

**[Bivariat](#page-161-0)e distributions**

Figure 2.27: Bivariate distributions MCR vs total exposure

In Figure 2.28 and Figure 2.29 scattered MCR distributions for the total and upper tail (here 37%) that drive the cumulative exposure are shown. The red line indicates the MCR threshold, 1.5. The black lines represent the regression lines MCR vs ln(Cumulative exposure) for each tail. Substances with an exposure contribution less than 15% are not displayed.

In [Table 2.121](#page-162-0) cont[ributions to ta](#page-163-0)il exposures at various percentile are shown. Column MCR = 1 shows the percentage of tail exposure due to individual(day)s with a single substance. Column  $1 < MCR \le 2$  shows the percentage of tail exposure due to individual(day)s with multiple substances, but the MCR  $\leq$  2. Column MCR  $>$  2 shows the percentage of tail exposure due to individual (day)s with multiple substances with  $MCR > 2$ .

<span id="page-162-0"></span>

Figure 2.28: Using MCR to identify substances that drive cumulative exposures, scatter distributions (total).

<span id="page-163-0"></span>

Figure 2.29: Using MCR to identify substances that drive cumulative exposures, scatter distributions (upper tail 37%).

Tail	% with MCR	Sub-	with % $\,<\,$	Substances	$%$ with MCR	Sub-
%	$=1$	stances	$MCR < = 2$		>2	stances
37	20.6	Difeno,	73.7	Difeno, Tebu	5.7	Difeno,
		Tebu				Tebu
50	19.2	Difeno,	75.6	Difeno, Tebu	5.2	Difeno,
		Tebu				Tebu
90	16.3	Difeno,	78.8	Difeno, Tebu	5.0	Difeno,
		Tebu				Tebu
95	15.0	Difeno,	82.5	Difeno, Tebu	2.5	Difeno.
		Tebu				Tebu
99	25.0	Difeno	75.0	Difeno. Tebu	0.0	
				Propi		

Table 2.121: Maximum Cumulative Ratio summary

For MCR settings, see *exposure mixture settings*.

# **Exposures settings**

# <span id="page-164-0"></span>**Calculation setting[s](#page-171-0)**

<span id="page-164-1"></span>

Name	Description
Risk type	The type of exposure considered in the assessment; acute (short)
	term) or chronic (long-term).
Multiple substances analysis	Specifies whether the assessment involves multiple substances.
Express results in terms of	Specifies whether the assessment involves multiple substances and
reference substance equivalents	results should be cumulated over all substances.
(cumulative)	
Include dietary and non-dietary	Specifies whether the assessment involves both dietary and
routes of exposure	non-dietary (oral, inhalatory or dermal) routes of exposure.
Target level	Select to express hazard characterisations at external or internal
	exposure level.
Match non-dietary to dietary	Specifies whether the individuals of one or more non-dietary
survey individuals	surveys should be matched to individuals in the dietary survey
	according to the individual codes (idIndividual). If unchecked,
	nondietary exposures are randomly allocated to dietary survey
	individuals.
Match individuals of multiple	If checked, exposures from identical individuals in non-dietary
non-dietary surveys	surveys are aggregated to obtain the overall nondietary exposures.
	If unchecked, exposures from random individuals in all
	non-dietary surveys are aggregated.
Model-then-add	Specifies whether to create separate exposure models for specific
	groups of foods-as-measured (model-then-add).

Table 2.122: Calculation settings for module Exposures.

# **Output settings**

<b>Name</b>	Description
Include drill-down on 9	Specifies whether drilldown on 9 individuals is to be included in
individuals around specified	the output.
percentile.	
Summarize simulated data	Specifies whether a summary of the simulated consumptions and
	concentrations should be included in the output.
Store simulated individual day	Store the simulated individual day exposures. If unchecked, no
exposures	additional output will be generated. If checked, the output will
	contain an additional section with the simulated individual day
	exposures.
Show percentiles for	Give specific percentiles of exposure distribution $(\%)$ , e.g. 50 90
	95 97.5 99 (space separated).
Percentage for drilldown	Gives detailed output for nine individuals near this percentile of
	the exposure distribution.
Percentage for upper tail	Gives detailed output for this upper percentage of the exposure
	distribution.
Show % of population below	Exposure levels can be generated automatically or by explicit
level(s)	specification (Manual).
<b>Exposure levels</b>	Specify exposure levels for which to give the percentage of
	exposure below these levels, e.g. 1 10 50 100 200 500. Specify
	below whether these levels are absolute or relative to ARfD/ADI.
Exposure levels are	Specify whether exposure levels are absolute or percentages of
	ARfD/ADI.
Number of levels of covariable	Specify the number of levels, e.g. 20. The range of the covariable
to predict exposure	is divided by the number of levels: $range = (max - min)/levels$ .
	For these covariable levels exposures are predicted.
Predict exposure at extra	Specify specific prediction levels in addition to the automatically
covariable levels	generated prediction levels (space separated).
Lower percentage for	The default value of 25% may be overruled.
variability (%)	
Upper percentage for	The default value of 75% may be overruled.
variability (%)	
Report consumptions and	Specifies whether body weights should be ignored and
exposures per individual	consumptions and exposures should be expressed per individual.
instead of per kg body weight	Otherwise, the consumptions and exposures are per kg body
	weight.

Table 2.123: Output settings for module Exposures.

# **Uncertainty settings**

Table 2.124: Uncertainty settings for module Exposures.

Name	Description
Resample kinetic model	Specifies whether kinetic model parameter values are resampled.
parameter values	

### **Calculation of exposures**

Exposures are computed by linking dietary and (if available) non-dietary individual/individual-day exposures and computing the (aggregated) internal exposures at the specified target compartment.

• *Exposures calculation*

Inputs used: *Dietary exposures Non-dietary exposures Active substances Relative potency factors Kinetic models*

Settings used

• *[Calculation Settings](#page-151-1)*

# **2.4.5 Exposure mixtures**

Expos[ure mixtures are m](#page-164-1)ixtures of substances that contribute relatively much to the overall cumulative exposure (potential risk drivers).

This module has as primary entities: *Foods Substances Effects*

## **Exposure mixtures calculation**

<span id="page-166-1"></span>The most common model of cumula[tive ris](#page-27-0)[k assessme](#page-44-0)[nt is to](#page-25-0) focus on substances that belong to the same common assessment groups (CAG). *Substances* in such a group belong to the same chemical family and may or may not have a similar mode of action. In assessing the risk, possible interactions between substances are often ignored and, moreover, little information is available about synergistic effects at low doses. More information is needed about the combined effects of such substances, but it is impractical to investigate all possible mixtures, and therefore instruments are needed to s[elect the mo](#page-44-0)st relevant substances for further investigation. This selection should not only be based on the hazard (highest relative potencies) but also on the exposure of the population to these substances. The potential risk of being exposed to chemicals in a mixture depends on the food *consumption* patterns of *individuals* in a population. A regular diet can contain hundreds of substances, so the number of combinations of substances to which an individual in a population is exposed can be numerous. The exposures mixtures module can be used to identify the most relevant mixtures to which a population is exposed.

Exposure mixtures are identified using a quantitative approach: *sparse non-[negative matr](#page-48-0)ix underap[proximation](#page-50-0) (SNMU)*. What mixtures predominantly determine the underlying patterns in the exposure matrix (substance x person  $(\text{day}))$ ?

### **[Sparse](#page-166-0) nonnegative matrix underapproximation**

<span id="page-166-0"></span>Starting point to identify major mixtures of substances using exposure data was to use Non-negative Matrix Factorization (NMF). Non-negative Matrix Factorization was proposed by Lee & Seung [Lee et al., 1999] and Saul & Lee [Saul et al., 2002] and deals specifically with non-negative data that have excess zeros such as exposure data. Zetlaoui et al. [Zetlaoui et al., 2011], introduced this method in food risk assessment to define diet clusters.

The NMF method was then implemented by Béchaux et al. [Béchaux et al., 2013] in or[der to identify fo](#page-356-1)od consumption patterns and main mixtures of pesticides to which the French population was exposed using *TDS* exposure to 26 priori[ty pesticides.](#page-356-2)

At the start of the Euromix project ideas had evolved: to obtain less components per mixture experiments were made using Sparse Nonnegative Matrix Factorization (S[NMF\) \[Hoyer, 2004\]](#page-354-0). This method was found not to give stable solutions. Better results were obtained with Sparse Nonnegative Matrix Underappr[oxim](#page-129-0)ation (SNMU) [Gillis et al., 2013]. This model also fits better to the problem situation because also the residual matrix after extracting some mixtures should be nonnegative. The SNMU method has been implemented in MCRA.

Indeed, NMF may be used to identify patterns or associations bet[ween substan](#page-355-0)ces in exposure data. NMF can be described as a method that finds a description of the data in a lower dimension. There are standard techniques available [such as principal c](#page-355-1)omponents analysis or factor analysis that do the same, but their lower rank representation is less suited because they contain negative values which makes interpretation hard and because of the modelling with a Gaussian random vectors which do not correctly deal with the excess of 0 and non-negative data. The NMF solution



Figure 2.30: Example of co-exposure distribution (from >1 substance per individual-day, red) super-imposed on the total exposure distribution (blue).

approximates a non-negative input matrix (i.c. exposure data) by two constrained non-negative matrices in a lower dimension such that the product of the two is as close as possible to the original input matrix. In Figure 2.31, the exposure matrix M with dimensions  $m$  (number of substances) and  $n$  (number of intake days or persons) is approximated by matrix U and V with dimensions  $(m \times k)$  and  $(k \times n)$  respectively, where k represents the number of mixtures. Matrix  $U$  contains weights of the substances per mixture, matrix  $V$  contains the coefficients of presence of mixtures in exposure per intake day or person. M is non-negative (zero or positive) and  $U$  and  $V$  [are constraint](#page-168-0) to be non-negative. The minimization criterion is:  $||M-UV||2$  such that  $U \ge 0$  and  $V \ge 0$ .

<span id="page-168-0"></span>

Figure 2.31: NMF approximation of exposure data

The solution found by NMF contains zeros, but mixtures still contain many components which complicates interpretability. Therefore, the Sparse Nonnegative Matrix Underapproximation (SNMU) [Gillis et al., 2013] which also provide sparse results was investigated. The SNMU has also some nice features well adapted to the objective of the Euromix project: the solution is independent of the initialization and the algorithm is recursive. Consequently, the SNMU method which is based on the same decomposition process as the NMF was the one implemented in MCRA.

SNMU is initialized using an optimal nonnegative rank-one approximation usingt[he power method](#page-355-1) (https://en. wikipedia.org/wiki/Power\_iteration). This initialization is based on a singular value decomposition and results in general in a unique solution that is sparse. The SNMU algorithm is called recursive because after identifying the first optimal rank-one underapproximation  $u_1v_1$ , the next rank-one factor is recovered by subtracting  $u_1v_1$  from M and applying the same factorization algorithm to the remainder  $M - u_1v_1$ . The solution  $u_1v_1$  is called a rank-[one under](https://en.wikipedia.org/wiki/Power_iteration)[approximation because an upper bou](https://en.wikipedia.org/wiki/Power_iteration)nd constraint is added to ensure that the remainder  $M - u_1v_1$  is non-negative. For Matlab code see: https://sites.google.com/site/nicolasgillis/code.

For each mixture, a percentage of explained variance is calculated. M is the exposure matrix with  $m$  rows (substances) and *n* columns (individuals or individual days)  $S_t$  is sum of squared elements of M:

$$
S_t = ||M||^2 = \sum_{i,j}^{m,n} e_{i,j}^2
$$

Apply SNMU on  $M$ , then for the first mixture:

- $u$  is  $m \times 1$  vector, contains weights of the mixture.
- $v$  is  $1 \times n$  vector, contains presence of mixture in exposure per individual.

Calculate residual matrix  $R$ :

$$
R = M - uv
$$

Calculate  $S_r$ , residual sum of squared elements of R:

$$
S_r = ||R||^2 = \sum_{i,j}^{m,n} e_{i,j}^2
$$

Percentage explained variance first mixture  $(k = 1)$  is:

$$
V_k=(1-S_r)/S_t)\cdot 100
$$

For the second mixture:

1. continue with residual matrix  $R$  (replace  $M$  by  $R$ ),

- 2. estimate  $u$  and  $v$ ,
- 3. calculate new residual matrix  $R$
- 4. calculate  $S_r$ , residual sum of squared elements of R

Percentage explained variance second mixture  $(k = 2)$  is:

$$
V_k = (1-S_r)/S_t) \cdot 100 - \sum_{l=1}^{k-1} V_l
$$

The last term is de percentage explained variance of the first mixture. Continue with the third mixture etc….

#### **Exposure matrix**

Basically, exposure is calculated as consumption x concentration. By summing the exposures from the different foods for each substance per person day separately, the exposure matrix for mixture selection is estimated:

$$
y_{ijc} = \frac{\sum_{k=1}^{p} x_{ijk} c_{ijkc}}{bw_i}
$$

where  $y_{ijc}$  is the exposure to substance c by individual i on day j (in microgram substance per kg body weight),  $x_{ijk}$  is the consumption by individual i on day j of food k (in g),  $c_{ijkc}$  is the concentration of substance c in food k eaten by individual i on day j (in mg/kg), and  $bw_i$  is the body weight of individual i (in kg). Finally, p is the number of foods accounted for in the model. More precisely, for an *acute* or short term risk assessment, the exposure matrix summarises the  $y_{ijc}$  with columns denoting the individual-days  $(ij)$  and rows denoting the substances  $(c)$ . Each cell represents the sum of the exposures for a substance on that particular day. For a *chronic* or long term risk assessment, the exposure matrix is constructed as the sum of all exposures for a particular substance averaged over the total number of intake days of an individual (substances x per[sons\).](#page-117-0) So each row represents a substance and a column an individual. Each cell represents the observed individual mean exposure for a substance for that individual. Note that in the exposure calculation, the concentration is the average of all residue values [of a sub](#page-122-0)stance.

When *relative potency factors* (RPF) are available then exposures are multiplied by the RPF and thus exposures to the different substances are on the same and comparable scale (toxicological scale). In this case, the selection of the mixture is risk-based. In some cases, RPFs may not be available. In this case exposure to different substances, even in the same unit, may lead to very different effect. To give all substances an equal weight a priori and avoid scalin[g effect, a normalization](#page-237-0) of the data can be applied as done in Béchaux et al. [Béchaux et al., 2013]. In this case, all substances are assigned equal mean and variance, and the selection of the mixtures will work on patterns of correlation only. Then mixture selection is not risk-based anymore but, what could be called, co-exposure-based.

Finally, in the mixture selection module of MCRA, the SNMU expects RPF data for a risk-based selection. If not available, the user should provide alternative RPF data, e.g. values 1 for a purel[y exposure-based sele](#page-354-0)ction, or lower-tier estimates (e.g. a maximum value on RPF thought possible).

#### **Mechanisms to influence sparsity**

Two mechanisms to *influence sparsity* are available. The SNMU algorithm incorporates a sparsity parameter and by increasing the value, the final mixtures will be more sparse (sparsity close to 0: not sparse; sparsity close to 1: sparse). The other approach is by using a subset of the exposure matrix based on a cut-off value for the *MCR*. High ratios correspond to high co-exposure, so it is reasonable to focus on these (person) days and not include days where exposure is received [from a single subst](#page-171-0)ance (ratio close to 1). To illustrate the combined use of MCR and the sparsity parameter, the French steatosis data (39 substances, 4079 persons) are used and person days with a ratio > 5 (see Figure 2.26) are selected yielding a subset of 139 records.

In Figure 2.32, the effect of using a cut-off level is immediately clear. The number of substances of the first mixture is 17 whereas in the unselected case only 4 substances were found The three plots show the influence of increasing the sparsity parameter from 0 to 1 on the number of substances in the mixture. For values close to 0, the mixture [contains 17](#page-160-0) substances. For values > 0.4 the number of substances in the mixture drops to 3.

In [Figure 2.33](#page-171-1) and Figure 2.34 the sparsity parameter is set to 0.0001 (not sparse) and 0.4 (sparse), respectively. This leads to mixtures containing different number of substances.





 $0.4$ 

Specified sparsity

 $0.6$ 

 $0.8$ 

 $\mathbf{1}$ 

**2.4. Exposure modules 163**

 $0.2$ 

 $\mathbf 0$  $\mathbf 0$ 

<span id="page-171-1"></span>

Figure 2.32: Influence of the specified sparsity parameter on the realized sparsity,  $n = 139$ 

As mentioned before, one of the nice features of the SNMU algorithm is its recursive character which results in identical mixtures. In Figure 2.35 and Figure 2.36, two U matrices are visualized. In the upper plot 4 mixtures are estimated, in the lower plot the solution for 5 mixtures is shown. Because of the ordering the plots look different, but a closer inspection of the first 4 mixtures of each solution shows that they are the same. In both figures, mixture 1 contains Imazalil, Thiacloprid, Deltamethrin (cis-deltamethrin) and Deltamethrin including other mixture.

### **Exposure mixtures settings**

#### **Calculation settings**

<span id="page-171-0"></span>

<b>Name</b>	Description	
Sparseness constraints	Sparseness parameter. Should be a value between 0 (not sparse,	
	many substances) and 1 (sparse, few substances).	
Number of mixtures	The number of mixtures.	
Number of iterations	Number of iterations, e.g. 1000.	
Seed for pseudo-random	Random seed for initializing matrix W and H.	
number generator.		
Exposures are	Exposures are risk based (expressed in equivalents of the	
	reference substance) or standardized.	
Convergence criterion	Convergence criterion for factorization algorithm.	
Cutoff for ratio total exposure/	For selection of individual(day) exposures specify cutoff for ratio	
maximum	total exposure/ maximum.	
Cutoff percentage $(\%)$ for total	For selection of individual(day) exposures specify cutoff	
exposure	percentage $(\%)$ for total exposure.	

Table 2.125: Calculation settings for module Exposure mixtures.

<span id="page-172-0"></span>

**Co-exposure of substances**

Figure 2.33: Heatmap for a solutions with 10 mixtures. The sparsity is set to 0 (not sparse). Each mixture contains many substances (see also Figure 2.34).



**Co-exposure of substances**

Figure 2.34: Heatmap for a solutions with 10 mixtures. The sparsity is set to 0.4 (sparse). Mixtures contain less substances compared to Figure 2.33.



**Co-exposure of substances**

Figure 2.35: Heatmap for solution with 4 mixtures. The first 4 mixtures in Figure 2.35 and Figure 2.36 are identical.



**Co-exposure of substances**

Figure 2.36: Heatmap for solution with 5 mixtures. The first 4 mixtures in Figure 2.35 and Figure 2.36 are identical.

#### **Calculation of exposure mixtures**

A multivariate decomposition method, sparse non-negative matrix underestimation (SNMU), is applied to the matrix of exposures per substance and per individual (chronic) or individual-day (acute) to find substance combinations that contribute most to the cumulative exposure.

• *Exposure mixtures calculation*

Inputs used: *Exposures*

Settings used

• *[Calculation Settings](#page-166-1)*

# **2.4.6 Food conversions**

Food [conversions relate fo](#page-171-0)ods-as-eaten, as found in the consumption data, to modelled foods (foods-as-measured), which are the foods for which concentration data are available. A food-as-eaten can be linked to one, or multiple foodas-measured using various conversion steps (e.g., using food recipes to translate a composite food into its ingredients). There are several types of conversion steps, and a conversion path may comprise multiple conversion steps between a food-as-eaten and a food-as-measured.

This module has as primary entities: *Foods Substances*

Output of this module is used by: *Consumptions by modelled food Dietary exposures*

## **Food conversions calculation**

<span id="page-176-0"></span>Food conversions are computed u[sing a recursive search algorithm](#page-115-0) [to link foods-as-ea](#page-116-1)ten to foods-as-measured, possibly through intermediate conversion steps. For instance, if (unpeeled) apple and grapes are the foods-as-measured, the food-as-eaten apple pie contains peeled apple and raisins, peeled apple is linked to unpeeled apple, and raisins are dried grapes. Hence, for this apple pie, there are two conversions, one to apple (with processing type 'peeled') and one to grapes (with processing type 'dried'), each with its own conversion path of intermediate conversion steps.

For each food-as-eaten, the food conversion algorithm recursively builds up the conversion paths using the following procedure:

- 1. **Check food-as-measured:** Check whether the current food is a food-as-measured, i.e., when substance concentration measurements are available, the food is considered a modelled food. If successful, a modelled-food has been found, and the current search stops.
- 2. **Find processing link (deprecated):** Check whether the current food can be considered to be a processed variant (e.g., cooked or peeled) of another food.
- a. **Match processing factor:** try to find the code in the processing factors table.

If successful, try to find the corresponding food translation proportion in the food recipes data to correct for a weight reduction or increase. Then, restart at step 1 with the new code of the unprocessed food.

**Warning**: the 'Find processing link' step is not recommended and is currently maintained for backwards compatibility reasons only. We do not recommended this step because occasionally, a conversion using processing factor data may lead to different conversion paths compared to a conversion where this step is skipped or data are not available. This is undesirable behaviour. In fact, the processing link is not needed, because processed foods are recognized in the 'Food translation link' where the translation proportion to correct for a weight reduction or increase is stored. In the conversion algorithm the processing factor itself is irrelevant and the identification may be postponed.

- 3. **Food translation link:** Check whether the current food translates to one or more foods through composition or read-across. Identify any processing types.
- a. **Food recipe link:** Try to find food translations for the current food (i.e., the ingredients of a composite food). This may result in one or more food codes for ingredients, and the iterative algorithm will proceed with each of the ingredient food codes in turn. Simultaneously check, whether the current food is a processed food or not. If so, determine the processing type or facets.
- b. **TDS food sample composition link:** Try to find the code in the TDSFoodSampleCompositions table (column idFood), a default translation proportion of 100% is assumed. The iterative algorithm will proceed with a TDS food (column idTDSFood) sample.
- c. **Read-across link:** Try to find a food extrapolation rule for the current food, a default translation proportion of 100% for 'idToFood' is assumed.

If successful, restart at step 1 with each of the new codes of the ingredient foods, TDS foods or Read Across foods.

- 4. **Subtype link:** try to find subtype codes, e.g. 'xxx\$\*' in the MarketShares table. In general, marketshares should sum to 100%. Foods with marketshares not summing to 100% are ignored in the analysis unless the checkbox 'Allow marketshares not summing to 100%' is checked. This step is optional, see advanced settings if you want to use this. If successful, restart at step 1 with each of the new codes of the subtype foods.
- 5. **Supertype link:** try to find supertypes, e.g. 'xxx\$yyy' is converted to 'xxx'. This step is optional, see advanced settings if you want to use this. If successful, restart at step 1 with the new code of the supertype food.
- 6. **Default processing factor:** remove processing part (-xxx) of the code. If successful, restart at step 1 with the new code without processing part.
- 7. **Maximum residue limit:** try to find the code in the MaximumResidueLimits table. If successful, the current search stops. If not successful, then stop anyway and the search is marked as failed food conversion.

## <span id="page-177-0"></span>**Food conversion settings**

# **Calculation settings**

Name	Description
Allow conversion using	Warning, the processing step is deprecated and is currently only
processing info	maintained for backwards compatibility reasons. See
	documentation for more details how processed foods are
	converted in the upgraded conversion algorithm. Step 2a: try to
	find the code in the processing table. Try to find the code in the
	FoodTranslation table (step 3a) to account for weight
	reduction/increase (translation proportion). If unchecked
	(default), processing table is ignored. If successful, restart at step
	1.
Allow conversion using food	Step 3a: try to find food translations for the current food (i.e., the
translations	ingredients of a composite food). This may result in one or more
	food codes for ingredients, and the iterative algorithm will
	proceed with each of the ingredient food codes in turn.
Allow conversion using TDS	Step 3b: try to find the code in the TDS food sample compositions
food sample compositions	table (idFood), a default translation proportion of 100% is
	assumed. The iterative algorithm will proceed with a TDS food
	(column idTDSFood) sample.
Allow conversion using food	Step 3c: try to find read across codes. If unchecked, read across
extrapolations	table is ignored, default is 'Use read across info'. E.g. for
	pineapple no measurements are found but by specifying that
	pineapple is converted to FruitMix (with a default proportion of
	100%), the TDS sample concentration value of FruitMix will be
	used for pineapple (as-eaten or as ingredient). If successful,
	restart at step 1.
Allow conversion using market	Step 4: try to find subtype codes, e.g. 'xxx\$*' in the market shares table.
shares Allow marketshares not	
	Specify whether to rescale market share percentages that do not
summing to 100%	sum to 100%. If checked, then foods with marketshares not
	summing to 100% are allowed. If not, then these foods are
	ignored in the analysis. Step 5: try to find supertypes, e.g. 'xxx\$yyy' is converted to 'xxx'
Allow conversion to supertypes	(optional, check box if you want to use this). If checked, allows
	for linkage of consumed foods coded at a lower hierarchical level
	to foods with measured concentrations at a higher hierarchical
	level e.g. consumed is Apple (code PF\$Apple) -> measured is
	Pome Fruit (code PF). Note: food codes are split on '\$'.
	Measurements of substances on food are available at a less
	detailed food coding level than consumption data. MCRA allows
	to use the concentration data of a supertype for all underlying
	food codes. If successful, restart at step 1.
Allow conversion using default	Step 6: remove processing part. If unchecked, no default
processing factors	processing factors are assumed, default is 'Use default processing
	factors'. If successful, restart at step 1.

Table 2.126: Calculation settings for module Food conversions.

## **Calculation of food conversions**

Food conversions are computed recursively, starting with a food-as-eaten and following a path to ingredients (food recipes), super/sup-type foods, etc. until either arriving at a food-as-measured (commonly the raw primary commodities) or concluding that the path does not lead to a food-as-measured.

• *Food conversions calculation*

Inputs used: *Consumptions Modelled foods Processing factors Food recipes Market shares Food extrapolations Total diet study sample compositions Active substances*

#### Settin[gs used](#page-176-0)

• *Calcul[ation Settings](#page-48-0)*

# **2.4.7 Human monitoring analysis**

Huma[n monitoring analysi](#page-177-0)s compares observed human monitoring data with predictions made for the same population of individuals from dietary survey data, concentration data and (optionally) non-dietary exposure data.

This module has as primary entities: *Populations Substances*

## **Human monitoring analysis calculation**

Human monitoring analysis comput[es internal s](#page-41-0)[ubstance co](#page-44-0)ncentration estimates based on provided human monitoring data. These estimates are specified at the level of long term average concentrations for individuals in case of *chronic assessments*, or the average concentrations for individual-days in case of *acute assessments*. The internal concentrations are computed independently for each substance, compartment, and sampling type.

The main steps for computing the human monitoring concentration estimates are:

- 1. [Imputation of no](#page-122-0)n-detects.
- 2. Imputation of missing values.
- 3. Calculation of individual concentrations (chronic) or individual day concentrations (acute).
- 4. Comparison of monitoring versus modelled exposures by substance and compartment (optional).

## **Imputation of non-detects**

Similar to *concentrations measurements in food*, human monitoring measurements can also contain measurements below the limit of reporting and similar to *concentrations modelling in foods*, human monitoring analysis needs to address these non-detects and replace them with imputed concentration values. For this, two approaches are available:

- 1. Rep[lace non-detects by zero.](#page-76-0)
- 2. Replace non-detects by a factor times [LOR, in which the factor is set be](#page-62-1)tween zero and one.

### **Imputation of missing values**

Concentration measurements may be missing. The following imputation methods are available for imputation of missing values:

- 1. Replace missing values by zero.
- 2. For each substance, sampling type, and compartment, replace missing values by a random other sample of this substance, sampling type, and compartment.
**Note:** For the second imputation method, more refined methods could be useful as well. E.g., when for a given day multiple samples are available, of which one is missing, then it may alternatively be sensible to leave this sample out when computing an average exposure. Also, when samples have been taken at different times during the day, it may be better to impute missing records using samples approximately from the same time-slot.

#### **Calculation of acute human monitoring concentrations**

For acute assessments, the monitoring concentrations are computed for each substance, compartment, and sampling type as average individual-day concentrations. That is, for a given substance, compartment, and sampling type, the acute individual-day concentration  $c_{ij}$  for individual i on day j is:

$$
c_{ij} = \frac{\sum_{k=1}^{n_{\text{samples}}} c_{ijk} \cdot s g_{ijk}}{n_{\text{samples}}},
$$

where  $n_{\text{samples}}$  is the number of samples available for individual i on day j, and  $c_{ijk}$  and  $s_{qijk}$  denote the concentration and specific gravity, respectively, of the  $k$ -th sample of the individual day.

**Note:** Note that currently, the acute concentrations are computed as mean concentrations when multiple samples are available for one day. In acute scenarios, one may be more interested in peak concentrations. I.e., the highest concentration of a day.

#### **Calculation of chronic human monitoring concentrations**

**Note:** The implementation for chronic is not yet available. Below is a description of the foreseen implementation.

For chronic assessments, the monitoring concentrations are computed as the average monitoring concentrations of multiple individual-days for each substance, compartment, and sampling type. That is, for a given substance, compartment, and sampling type, the chronic concentration  $c_i$  for individual i is:

$$
c_i = \frac{\sum_{j=1}^{n_{\text{days}}} c_{ij}}{n_{\text{days}}},
$$

where  $n_{\text{days}}$  is the number of days that individual i was monitored, and  $c_{ij}$  denotes the average monitoring concentration of individual  $i$  on day  $j$ .

#### **Compare measured and modelled exposures**

An optional step of the human monitoring analysis is to compare the monitoring concentrations with *modelled exposures* that were obtained from *dietary* (and optionally *non-dietary*) exposure assessments. This comparison may provide insight in the coherence between modelled exposures and the measured reality. A requirement is that both monitoring data and dietary/non-dietary use data is available for the same individuals or individual-days. An example of a graphical output of these comparison is given in Figure 2.37.



**Monitoring versus modelled exposures BPA**



# **Human monitoring analysis settings**

# **Calculation settings**

Name	Description	
Non-detects handling method	Method for dealing with non-detects samples in human	
	monitoring data.	
Fraction of LOR	Factor for replacing non-detects with factor times LOR.	
Missing value imputation	Imputation method for missing values.	
method		
Correlate monitoring with	Correlate monitoring with modelled exposures.	
modelled exposures		

Table 2.127: Calculation settings for module Human monitoring analysis.

#### **Calculation of human monitoring analysis**

Human monitoring analysis calculations comprise two parts. The first part is to compute estimates of the human monitoring concentrations based on the human monitoring data. The second part, which is optional, is to relate the human monitoring concentrations to modelled concentrations from exposure assessments.

• *Human monitoring analysis calculation*

Inputs used: *Human monitoring data Exposures*

Settings used

• *[Calculation Settings](#page-179-0)*

# **2.4.8 Human monitoring data**

Human monitoring data quantify substance concentrations found in humans collected in human monitoring surveys.

This module has as primary entities: *Substances*

Output of this module is used by: *Human monitoring analysis*

# **Human monitoring data dataf[ormats](#page-44-0)**

Data are provided on the survey, th[e individuals in the survey, th](#page-179-1)e samples taken, the analyses performed, the analytical methods used, the properties for substances analysed, and the concentrations found.

#### <span id="page-182-0"></span>**Data are provided in the following relational tables:**

- Human monitoring surveys
- Human monitoring individuals
- Human monitoring samples
- Human monitoring sample analyses
- Sample concentrations
- Analytical methods
- Analytical method properties for substances

#### **Human monitoring samples**

Suggested table definitions for human monitoring data.

# **Human monitoring surveys**

Contains the survey definitions.





Table aliases: HumanMonitoringSurveys, HumanMonitoringSurvey.

# **Human monitoring individuals**

The individuals of a survey are recorded in the individuals table.





Table aliases: HumanMonitoringIndividuals, HumanMonitoringIndividual.

# **Human monitoring samples**

Contains the samples taken during the study.





Table aliases: HumanMonitoringSamples, HumanMonitoringSample.

# **Human monitoring sample analyses**

Contains the measurements of the samples of human monitoring studies.

Name	<b>Type</b>	Description	<b>Aliases</b>	Required
idSample-	AlphaNumeric(50)	Unique identification code of	idSample-	Yes
Analysis		the sample analysis.	Analysis,	
			SampleAnalysis	
idSample	AlphaNumeric(50)	Code of the measured	idSample,	Yes
		monitoring sample.	Sample	
idAnalytical-	AlphaNumeric(50)	The code of method of	idAnalytical-	Yes
Method		analysis.	Method.	
			Analytical-	
			MethodName,	
			Analytical-	
			MethodId	
AnalysisDate	AlphaNumeric(50)	Date of analysis.	AnalysisDate,	N <sub>o</sub>
			DateAnalysis	
Substance	AlphaNumeric(100)	One or more columns with		Yes
concentration(s)		the measured concentrations		
		of the substances in the unit		
		as specified by the analytical		
		method. The column name $(s)$		
		should match the substance		
		codes of the substances		
		measured by the analytical		
		methods. Empty fields for		
		substances that should have		
		been measured by the		
		analytical method are		
		considered to be non-detects		
		with measurement values		
		below LOR.		

Table 2.131: Table definition for HumanMonitoringSampleAnalyses.

Table aliases: HumanMonitoringSampleAnalyses, HumanMonitoringSampleAnalysis.

# **Sample concentrations**

The positive concentration values for substances from analysis in the unit specified in table human monitoring sample analyses. Non-detects (i.e. results 'less than LOR') are not included, their existence can be inferred from the tables AnalysisSamples and AnalyticalMethodSubstances, and the LOR itself from the analytical method.

Table 2.132: Table definition for HumanMonitoringSampleConcentrations.

Name	⊺ype	Description	Aliases	Required
idAnalysis-	AlphaNumeric(50)	The identification number of	idAnalysis-	Yes
Sample		the analysed sample.	Sample,	
			AnalysisSample-	
			Id	
idSubstance	AlphaNumeric(50)	The substance code.	idSubstance,	Yes
			SubstanceId,	
			Substance	
Concentration	Numeric	The measured concentration.	Concentration	Yes

Table aliases: HumanMonitoringSampleConcentrations, HumanMonitoringSampleConcentration.

# **Analytical methods**

The analytical methods used for analyzing the samples are recorded in the analytical methods table. Each analytical method should have a unique identification code (idAnalyticalMethod). The description field may be used for a more detailed description of the analytical method. The records of this table should be linked to one or more analytical method substance properties table, which record the substances that are measured by this method (and their limits of reporting).





Table aliases: AnalyticalMethod, AnalyticalMethods.

# **Analytical method properties for substances**

Name	Type	Description	<b>Aliases</b>	Required
idAnalytical-	AlphaNumeric(50)	The code of method of	idAnalytical-	Yes
Method		analysis.	Method,	
			Analytical-	
			MethodName,	
			Analytical-	
			MethodId	
idSubstance	AlphaNumeric(50)	The substance code.	idSubstance,	Yes
			SubstanceId,	
			Substance	
LOR	Numeric	The limit of reporting (LOR).	LOR	Yes
		In MCRA, LOR just means		
		the limit below which no		
		quantitative result has been		
		reported. Depending on a		
		laboratory's format of		
		reporting, LOR may be a		
		limit of detection (LOD), a		
		limit of quantification (LOQ)		
		or another limit.		
Concentration-	AlphaNumeric(50)	The code of the unit as used	Concentration-	N <sub>o</sub>
Unit		for substance concentration	Unit, Units,	
		data. Allowed code: kg/kg or	Unit	
		kilogram/kilogram; g/kg or		
		gram/kilogram; mg/kg or		
		milligram/kilogram (default);		
		$\mu$ g/kg or		
		microgram/kilogram; ng/kg		
		or nanogram/kilogram; pg/kg		
		or picogram/kilogram.		

Table 2.134: Table definition for AnalyticalMethodCompounds.

Table aliases: AnalyticalMethodSubstances, AnalyticalMethodSubstance, AnalyticalMethodCompounds, AnalyticalMethodCompound.

# **Human monitoring data settings**

# **Selection settings**

Name	Description		
Surveys	The surveys that should be included in the action.		
Sampling methods	The sampling methods that should be included in the action.		

Table 2.135: Selection settings for module Human monitoring data.

#### **Human monitoring data as data**

Data are provided in the form of surveys consisting of individuals from which the human monitoring samples taken. Substance concentration measurements are linked to analyses performed on the human monitoring samples. The data should also include information about the analytical methods that were used.

• *Human monitoring data data formats*

# **2.4.9 Non-dietary exposures**

<span id="page-189-0"></span>Non-[dietary exposures are the amounts of](#page-182-0) substances to which individuals in a population are exposed via any of three non-dietary routes: dermal, inhalation or oral, per day. Non-dietary exposures can be used for *computing aggregate exposure distributions* from both dietary and non-dietary routes of exposure. Depending on the exposure type, non-dietary exposures can be short-term/acute exposures and then contain exposures for individual-days, or they can be long-term/chronic exposures, in which case they represent the average exposure per day over an unspecified longer time period. Examples are presented as case studies in Kennedy et al. ([Kenned[y et al., 2012\],](#page-150-0) [\[Kennedy et al., 2015a\], \[Ken](#page-150-0)nedy et al., 2015b], [Kennedy et al., 2017]) and R code to generate these examples is available for general use.

Datasets are typically generated by external programs, e.g. Browse, Bream2 or PACEM. The Browse and Bream2 models both simulate distributions of potential exposure of residents and bystanders to pesti[cides sprayed on crop](#page-355-0)s. [Probability distribution](#page-355-1)sa[re included to quantify](#page-356-0) va[riations in input param](#page-356-1)eters representing conditions during a spray event. PACEM is a probabilistic exposure model for substances present in consumer products. Browse was an EU FP7 project (https://secure.fera.defra.gov.uk/browse/software), that in addition to bystanders and residents from boomsprayers includes various arable and orchard scenarios. It includes dermal, oral and inhalation routes of exposure and can generate exposure files in the correct format for MCRA non-dietary exposure. The underlying simulation of dermal spray deposits on bystanders and residents was taken from Bream, although Browse includes post-processing to mode[l indirect exposures, multiple routes and long-ter](https://secure.fera.defra.gov.uk/browse/software)m exposure [Kennedy et al., 2017]. Volatilisation is also included through the PEARL-OPS model [van den Berg et al., 2016] to account for inhalation of vapours. Bream2 is an updated version of the original Bream model [Kennedy et al., 2012] and software is available online (http: //www.ssau.co.uk/bream2-calculator). Results from Bream had been used as part of EFSA guidance on bystander and resident exposure. Bream2 was recently shown to produce more rea[listic exposure distrib](#page-356-1)utions, when compared to measured dermal exposure [Butler et al., [2018\].](#page-356-2)

[This module has as primary entities:](http://www.ssau.co.uk/bream2-calculator) *Populations Subst[ances](#page-355-0)*

Output of this module is used by: *Exposures*

# **Non-dietary exposures data f[ormats](#page-41-0)**

Non-dietary exposures may be sp[ecified for](#page-150-0) multiple routes of exposure (dermal, oral and inhalation), for multiple substances, and for multiple exposure sources. Also, they can be provided as single deterministic exposure levels or as probabilistic exposure estimates and it is possible, but not mandatory, to specify uncertainty. The non-dietary exposures may be short term (acute) or longer term averages (chronic), and the user must ensure to supply appropriate non-dietary data for the type of exposure assessment of interest. For chronic assessments this means the non-dietary exposure is averaged over an appropriate time interval.

Non-dietary exposures are defined by non-dietary surveys to which dietary exposures are linked. For these surveys, individual properties can be specified to define non-dietary exposures for particular sub-groups of the population (e.g., specific age groups, or a specific gender). For each non-dietary survey a percentage of the target population that is not exposed from this source can be specified by means of a percentage. Uncertainty about non-dietary exposures can be specified by specifying multiple records for each individual in an additional table.

The use of multiple surveys can be used when multiple sources are relevant. For example, when modelling individuals taking part in various activities involving pesticide use or incidental exposures as a resident. Each non-dietary source is characterised in a particular user-selected or user-supplied non-dietary survey. By default, exposures from separate non-dietary surveys (sources) are considered to be independent events, but as explained below correlations between substances and/or activity types per individual can be represented if generated prior to uploading to MCRA. When including multiple non-dietary surveys it is possible to supply some with uncertainty/variability and others without variability/uncertainty according to the requirements and data availability.

When the user supplies probabilistic non-dietary exposure estimates (i.e., there is a distribution for the non-dietary exposure rather than a single nominal value), then this information will be propagated as part of the *exposure assessment*. Distributions may be included to represent variability, uncertainty or both, and in these cases the aggregate exposure estimates are reported with variability and/or uncertainty as appropriate. Multiple (uncertain) values from the non-dietary exposure distribution may be supplied per individual and per substance.

Exposures within a non-dietary survey may be expressed as correlated or independent for the differen[t substances. For](#page-150-0) [exam](#page-150-0)ple, if the exposures are a mixture of substances in a known ratio (e.g. from a specific tank mix of pesticides), or if exposure to one substance strongly implies that exposure to another is likely, these relationships may be included in the non-dietary data supplied by the user. Inference for the matched-case scenario with uncertainty analysis can use exposure sets. These are specific sets of exposures defined for each individual and in any uncertainty iteration an individual will receive exactly one of the exposure sets for that individual. Alternatively, independence may be represented by generating sets drawn from independent distributions when generating these tables.

#### **Non-dietary exposures**

Non-dietary exposure data is provided per non-dietary surveys. Each non-survey has some general information about the exposed population and the origin of the non-dietary exposure data. Also, a number of properties, such as specific age groups, can be specified for a survey. To each non-dietary survey, non-dietary exposures can be linked. These exposures may originate from dermal, oral and/or inhalatory exposure routes.

# **Non-dietary surveys**

This table provides detail about non-dietary surveys (source of non-dietary exposure): description, location, date and unit of exposure).

Name	Type	Description	Aliases	Required
idNonDietary-	AlphaNumeric(50)	The survey identification	idNonDietary-	<b>Yes</b>
Survey		number.	Survey	
Description	AlphaNumeric $(200)$	Description of non-dietary	Description	N <sub>o</sub>
		survey.		
Location	AlphaNumeric(50)	The location of survey.	Location	N <sub>o</sub>
Date	<b>DateTime</b>	The date of survey.	Date	N <sub>o</sub>
NonDietary-	AlphaNumeric(50)	The unit of the non-dietary	Unit.	<b>Yes</b>
IntakeUnit		exposure.	NonDietary-	
			IntakeUnit,	
			NonDietary-	
			ExposureUnit	
Percentage-	Numeric	The proportion zeros,	PercentageZeros	N <sub>o</sub>
Zeros		specified as a percentage $(\%).$		
idPopulation	AlphaNumeric $(50)$	Unique identification code of	IdPopulation,	N <sub>o</sub>
		the population.	PopulationId	

Table 2.136: Table definition for NonDietarySurveys.

Table aliases: NonDietarySurveys, NonDietarySurvey.

# **Non-dietary survey properties**

This table specifies demographic properties that apply to the individuals in the surveys. These properties could be used to link the individuals of a non-dietary survey with individuals from dietary surveys. That is, if demographic criteria are defined, only those individuals in the dietary survey that meet these criteria will be assigned non-dietary exposures. This table is not relevant when matching is switched on (i.e., when individuals are matched based on individual id).





Table aliases: NonDietarySurveyProperties, NonDietarySurveyProperty.

# **Non-dietary exposures**

This table defines nominal non-dietary exposure values (such as means) for individuals within the non-dietary surveys. It can also be used to specify non-dietary exposures for individuals within the food surveys. Each exposure comprises a non-dietary survey (source of exposure); a string identifying an individual, which may or may not correspond to the ID of an individual in a food survey; a substance; and dermal, oral and inhalation exposure values. Exposures are assumed to be external doses.





Table aliases: NonDietaryExposures, NonDietaryExposure.

# **Non-dietary exposure uncertainty records**

This table may be used to supply uncertainty sets of multiple (uncertain) non-dietary exposure values for individuals within the non-dietary surveys. Multiple non-dietary values are generated by probabilistic exposure calculations i.e. when there is a distribution for the non-dietary exposure rather than a single nominal value. If this table is supplied, aggregate exposure estimates will be reported with uncertainty using the uncertainty set approach. Each exposure set comprises a non-dietary survey (source of exposure); an individual ID; a substance; and dermal, oral and inhalation exposure values. In addition, the id column is used to define the uncertainty set. Summarizing, an uncertainty set is identified by column id and contains all exposure sets defined for each individual. In each uncertainty run (outer loop) an uncertainty set is sampled and in each iteration (inner loop) nondietary individuals are sampled from this set.

Name	Type	Description	Aliases	Required
idIndividual	AlphaNumeric(50)	Non-dietary individual	idIndividual	<b>Yes</b>
		identification number. The		
		idIndividual value may		
		correspond to an id in the		
		Individuals table (dietary		
		exposures matched to food		
		survey individuals), may not		
		correspond to an id in the		
		Individuals table (unmatched		
		individuals), or may contain a		
		default exposure (indicated by		
		$idIndividual = 'General' -$		
		demographic criteria for the		
		assignment of exposures are defined in the		
		NonDietarySurveyProperties		
		table). For matching to occur,		
		the user will need to tick the		
		option to 'match specific		
		dietary survey individuals' in		
		the user-interface. The		
		software will then assign		
		non-dietary exposures to the		
		dietary individuals according		
		to the values in this column.		
		Any idIndividual values that		
		do not correspond to		
		individuals in the food survey		
		will be ignored, unless a value		
		'General' is specified. Then		
		the individual should meet the		
		demographic criteria as		
		defined in the		
		NonDietarySurveyProperties		
		table. If this box is left		
		unticked, the non-dietary exposures will be randomly		
		allocated to the dietary		
		population provided they meet		
		the demographic criteria.		
idNonDietary-	AlphaNumeric(50)	code of survey (must	idNonDietary-	Yes
Survey		correspond to values in id	Survey	
		column of		
		NonDietarySurveys table)		
idCompound	AlphaNumeric(50)	Substance code (must	idSubstance,	Yes
		correspond to values in id	SubstanceId,	
		column of Substances table).	SubstanceCode,	
			Substance	
id	AlphaNumeric(50)	Uncertainty set identification	id	Yes
		number.		
Dermal	Numeric	Dermal non-dietary exposure	Dermal	N <sub>o</sub>
		value.		
Oral	Numeric	Oral non-dietary exposure	Oral	N <sub>o</sub>
		value.		
Inhalation	Numeric	Inhalation (non-dietary)	<b>Inhalation</b>	N <sub>o</sub>
		exposure value.		

Table 2.139: Table definition for NonDietaryExposuresUncertain.

#### **Non-dietary exposures settings**

#### **Uncertainty settings**

Name	Description		
Resample non-dietary	Specifies whether non-dietary exposures are resampled. Note that		
exposures	non-dietary uncertainty is only ignored when individual		
uncertainty is set to false (uncheck box: do NOT resample			
	individuals).		

Table 2.140: Uncertainty settings for module Non-dietary exposures.

#### **Non-dietary exposures uncertainty**

In an aggregate exposure assessment, dietary and nondietary data are combined into an aggregate exposure distribution. The nondietary data are supplied in table NonDietaryExposures. In an uncertainty analysis, MCRA provides two ways to assess the uncertainty:

- 1. the uncertainty set approach
- 2. the bootstrap algorithm.

When table **NonDietaryExposuresUncertain** is not supplied, the nondietary data in table **NonDietaryExposures** is resampled and the bootstrapped sets are used in the uncertainty run. More precisely, in each outer loop of the 2D Monte Carlo, within each nondietary survey (multiple surveys may be supplied), the nondietary individuals are resampled. Each individual represents a nondietary exposure set containing dermal and/or oral and/or inhalation exposure values for multiple substances. Bootstrapping is the default behaviour when the **NonDietaryExposuresUncertain** table is missing. When uncertainty distributions supplied in this table represent sampling uncertainty (individual exposure sets are repeatedly sampled using the same nondietary exposure generator without changing the input parameters), then bootstrapping the data performs equally well and is more efficient.

#### **Non-dietary exposures as data**

Non-dietary exposures are collected in non-dietary surveys. Data may be specified on population level or individual level, and may or may not include variability and uncertainty.

• *Non-dietary exposures data formats*

Inputs used: *Active substances*

See also *Combining dietary and non dietary exposures*.

# **2.4.10 Si[ngle value d](#page-202-0)ietary exposures**

Single v[alue dietary exposures are based on the single v](#page-151-0)alue concentrations of substances, expressed per standard (kg) bodyweight and/or single value amounts of consumed food as measured. Depending on the exposure type, dietary exposures can be short-term/acute exposures.

This module has as primary entities: *Populations Foods Substances*

Output of this module is used by: *Single value risks*

## **Single value dietary exposures data formats**

Single value dietary exposures are IESTI etc.

# **Dietary exposures**

Dietary exposure data is specified through dietary exposure models. To each dietary exposure model, exposure distributions are linked.

## **Dietary exposure models**

High level description of the dietary exposure models, specifying the id, name, description and the (reference) substance and exposure unit used for reporting the exposures. To this models, exposure percentiles and bootstrap values of the percentile may be linked.





Table aliases: DietaryExposureModels, DietaryExposures.

## **Dietary exposure percentiles**

Exposure percentiles linked to a dietary exposure model. The percentiles are reported in the unit specified by the exposure model to which they belong.

Name	Гуре.	Description	Aliases	Required
idDietary-	AlphaNumeric(50)	The code of the dietary	idDietary-	Yes
<b>ExposureModel</b>		exposure model to which this	ExposureModel	
		record belongs.		
Percentage	Numeric	The percentage to which the	Individual-	Yes
		percentile value belongs.	PropertyDouble-	
			ValueMin	
Exposure	Numeric	The percentile value. I.e., the	Exposure	Yes
		exposure value belonging to		
		the specified percentage.		

Table 2.142: Table definition for DietaryExposurePercentiles.

Table aliases: DietaryExposurePercentiles.

# **Dietary exposure percentile bootstrap values**

Uncertainty values, obtained from bootstrap runs, of the dietary exposure percentiles.

Name	Type	Description	Aliases	Required
idDietary-	AlphaNumeric $(50)$	The code of the dietary	idDietary-	<b>Yes</b>
<b>ExposureModel</b>		exposure model to which this	<b>ExposureModel</b>	
		record belongs.		
idUncertainty-	AlphaNumeric $(50)$	The uncertainty set identifier.	idUncertainty-	<b>Yes</b>
<b>Set</b>			Set.	
			UncertaintyId	
Percentage	Numeric	The percentage to which the	Individual-	Yes
		percentile value belongs.	PropertyDouble-	
			ValueMin	
Exposure	Numeric	The percentile value. I.e., the	Exposure	<b>Yes</b>
		exposure value belonging to		
		the specified percentage.		

Table 2.143: Table definition for DietaryExposurePercentilesUncertain.

Table aliases: DietaryExposurePercentilesUncertain, DietaryExposurePercentileUncertains.

# **Single value dietary exposures calculation**

<span id="page-196-0"></span>Either the short-(acute) or long-term (chronic) dietary exposure to a substances via food can be estimated as a single value calculated from single value inputs. This is often referred to as deterministic estimation. MCRA implements the IESTI, TMDI, IEDI and NEDI (Rees-Day) calculation methods that are also available in the EFSA PRIMo (Pesticide Residue Intake Model) tool revision 3, [EFSA, 2018].

The implementation in MCRA allows more choices than EFSA PRIMo by choosing other inputs or input combinations for the calculation formula. Moreover, the calculations can in all cases be adapted for processing factors or occurrence frequencies. For the chronic [estimates, al](#page-355-2)so the contributions per food or processed food are reported.

#### **Acute single value dietary exposure assessment**

The short term (acute) exposure assessment is usually the exposure related to a consumption of food over a single day. MCRA applies in principle the IESTI equations as shown in EFSA PRIMo revision 3 [EFSA, 2018], but the equations are extended with a factor OF to allow adaptation for an occurrence frequency lower than 1. So the inputs to the equations are not necessarily the same as used in PRIMo. For example, the large portion (LP) and body weight (BW) can be computed instead of just being standard values.

#### **IESTI (International Estimated Short-Term Intake)**

The IESTI (International Estimated Short-Term Intake) is calculated according to different equations depending on the unit weight of the raw agricultural commodity (RAC) and the unit weight of the edible portion (EP). The following cases are distinguished.

- **Case 1** refers to commodities with unit weight of the raw agricultural commodity  $U_{RAC} \leq 25$  g (e.g. walnuts, strawberries and peas. It is also used for meat, liver, kidney, edible offal, eggs and for post-harvest uses in cereal grains, oilseeds and pulses).
- **Case 2a** for food product with a  $U_{RAC} > 25$  g, where the meal portion is  $> U_{ep}$  (unit weight edible portion).
- **Case 2b** for food products with a  $U_{RAC} > 25$  g, where the meal portion is  $U_{ep}$ .
- **Case 3** for food products that are usually bulked or blended before they are consumed (e.g. cereals, pulses, oilseeds and milk).

The calculations are as follows.

**Case 1**

$$
\frac{\text{LP}\cdot\text{HR}\cdot\text{PF}\cdot\text{CF}\cdot\text{OF}}{\text{BW}}
$$

**Case 2a**

 ${\tt U}_{\tt ep} \cdot {\tt HR} \cdot {\tt PF} \cdot {\tt CF} \cdot {\tt VF} \cdot {\tt OF} + ({\tt LP} - {\tt U}_{\tt ep}) \cdot {\tt HR} \cdot {\tt PF} \cdot {\tt CF} \cdot {\tt OF}$ BW

**Case 2b**



**Case 3**

$$
\frac{\text{LP}\cdot\text{STMR}\cdot\text{PF}\cdot\text{CF}\cdot\text{OF}}{\text{BW}}
$$

**New Case 1 and 3:**

$$
\frac{\text{LP}\cdot\text{MRL}\cdot\text{PF}\cdot\text{CF}\cdot\text{OF}}{\text{BW}}
$$

**New Case 2a and 2b**

LP ⋅ MRL ⋅ PF ⋅ CF ⋅ VF ⋅ OF BW

Parameters used in the equations

MRL: Maximum residue level for the RAC concerned (in mg/kg);

STMR: Supervised Trials Median Residue for raw agricultural commodity (RAC) concerned (in mg/kg);

CF: Conversion factor residue definition enforcement to residue definition risk assessment (calculated as the ratio of residues according to the residue definition for risk assessment divided by the residue concentration according to the residue definition for enforcement);

OF: Use Frequency of the raw agricultural commodity (RAC),

BW: mean body weight for the subgroup of the population related to the LP or mean consumption (in kg). It is noted that for  $\text{TEST}_{\text{new}}$ , it was recommended to express the LP on individual body weight. This recommendation could not yet be fully implemented since the LP data were used as provided by the Member States. The LP would have to be recalculated on the basis of the individual consumption and individual body weight of the respondent of the survey;

LP: Large portion reported (in kg/day) (97.5th percentile of eaters (or alternative percentile, depending on the number of reported eating occasions);

HR: Highest residue according to residue definition for enforcement in composite sample (in mg/kg);

 $U_{en}$ : Unit weight of edible portion (in kg), provided by the country from which the LP was reported (or mean unit weight calculated from all available unit weight data, if no unit weight is available from the country matching the highest LP;

PF: Processing factor or peeling factor (calculated as the ratio of residues in processed/peeled product, divided by residue concentration in unprocessed/unpeeled product);

VF: variability factor, depending on the unit weight of the whole product ( $U_{RAC}$ ), different default VF are used in the calculations.

 $(U_{\text{PAC}})$  < 25 g, the calculations are performed according to case 1 (VF = 1).

 $(U_{RAC})$  between 25 and 250 g:  $VF = 7$ .

 $(U<sub>RAC</sub>)$  greater than 250: VF = 5.

In  $\text{TEST}_{new}$ , a default VF of 3 is used.

In case the empirically derived variability factors are available, the default VF is to be replaced.

#### **Alternative IESTI-styled assessments**

If consumption survey data for a specific population are available, the LP values in the IESTI equations may be replaced by statistics calculated from these data (at the consumed modelled food level).

If concentration monitoring data (retrospective) or concentration field trial data (prospective) are available, the MRL, HR, STMR values in the IESTI equations may be replaced by statistics calculated from these data (at the consumed modelled food level).

In the current use of IESTI, the occurrence frequency (use frequency)  $\circ$ F is assumed to be 1. In alternative assessments, a more realistic estimate may be used. Such an estimate could be derived for example as the highest occurrence frequency observed in a retrospective assessment for either the same substance or the same food.

#### **IESTI special cases**

For some foods, substances are applied after harvest, i.c. post-harvest use. For those combinations of food and substance, Case 1 should be used in the calculation. However, commodities with post-harvest use like cereal grains, oilseeds and pulses are typically bulked or blended (Case 3). To overrule Case 3, specify in table *IESTISpecialCases* the food and substance combination with 'PostHarvest' as application type. For those food and substance combinations with a unit weight of the raw agricultural commodity  $U_{RAC} \leq 25$  g, Case 1 is applied. When substances are applied before harvest, i.c. pre-harvest use, Case 1 should be overruled by Case 3. Specify in table *IESTISpecialCases* the food and substance combination with 'PreHarvest' as application type. See also *IESTISpecialCases table format*.

#### **Chronic single value dietary exposure assessment**

The long term (chronic) exposure assessment is usually the exposure related to a consumption over a longer period of time. MCRA applies in principle the TMDI, IEDI or NEDI (Rees-Day) equations as shown in EFSA PRIMo revision 3 [EFSA, 2018]. However, the equations are extended with factors  $PF_i$  and  $OF_i$  to allow adaptation for processing factors and occurrence frequencies lower than 1. Also, the inputs to the equations are not necessarily the same as used in PRIMo. For example, the consumption statistics ( $MC$ ,  $p_{97.5}$ ) and body weight ( $BW$ ) can be computed instead of just being standard values. Note that TMDI, IEDI and NEDI (Rees-Day) estimates are summations over foods (r[aw agricultura](#page-355-2)l products). In addition to the summations, MCRA will also report the individual terms (single value dietary exposures per food).

#### **TMDI (Theoretical Maximum Dietary Intake)**

$$
\sum_{X=i}^{n} \frac{MRL_i \cdot CF_i \cdot PF_i \cdot OF_i \cdot MC_i}{BW}
$$

 $i, j, k, \ldots n$ : individual raw agricultural products

#### **IEDI (International Estimated Dietary Intake)**

$$
\sum_{X=i}^{n} \frac{STMR_i \cdot CF_i \cdot PF_i \cdot OF_i \cdot MC_i}{BW}
$$

 $i, j, k, \ldots n$ : individual raw agricultural products

#### **NEDI (National Estimated Dietary Intake): Rees-Day model (I)**

$$
\sum_{X=i}^{j}\frac{\textit{MRL}_{i}\cdot CF_{i}\cdot PF_{i}\cdot OF_{i}\cdot p_{97.5} \textit{consumption}_{i}}{\textit{BW}}+\sum_{X=k}^{n}\frac{\textit{MRL}_{k}\cdot CF_{k}\cdot PF_{i}\cdot OF_{i}\cdot MC_{k}}{\textit{BW}}
$$

 $i, j$ : two raw agricultural products leading to the highest intake;

 $k, l, m, \ldots n$ : remaining raw agricultural commodities consumed

#### **NEDI (National Estimated Dietary Intake): Rees-Day model (II)**

$$
\sum_{X=i}^{j} \frac{STMR_i \cdot CF_i \cdot PF_i \cdot OF_i \cdot p_{97.5} consumption_i}{BW} + \sum_{X=k}^{n} \frac{STMR_k \cdot CF_i \cdot PF_i \cdot OF_i \cdot MC_k}{BW}
$$

 $i, j$ : two raw agricultural products leading to the highest intake;

 $k, l, m, \ldots n$ : remaining raw agricultural commodities consumed

Parameters used in the equations

*MRL* : Maximum residue level for the RAC concerned (in mg/kg);

*STMR* : Supervised Trials Median Residue for raw agricultural commodity (RAC) concerned (in mg/kg);

*CF* : Conversion factor residue definition enforcement to residue definition risk assessment (calculated as the ratio of residues according to the residue definition for risk assessment divided by the residue concentration according to the residue definition for enforcement);

*MC*<sup>2</sup>: mean consumption for a given raw agricultural product (RAC) calculated for the whole survey/subgroup of the survey, including processed products (recalculated to the unprocessed RAC) (in kg/day);

 $p_{97.5}$  *consumption*<sub>i</sub> for a given raw agricultural product (RAC), calculated from the individual consumption reported by the participants of the whole survey/subgroup of the survey, including processed products (recalculated to the unprocessed RAC) (in kg/day);

 $BW$ : mean body weight for the subgroup of the population related to the  $LP$  or mean consumption (in kg). It is noted that for  $IESTI_{new}$ , it was recommended to express the  $LP$  on individual body weight. This recommendation could not yet be fully implemented since the  $LP$  data were used as provided by the Member States. The  $LP$  would have to be recalculated on the basis of the individual consumption and individual body weight of the respondent of the survey;

*OF* : Occurrence Frequency of the substance on the food (typically, a raw agricultural commodity, RAC),

*PF* : Processing factor or peeling factor (calculated as the ratio of residues in processed/peeled product, divided by residue concentration in unprocessed/unpeeled product);

# **Alternative TMDI-, IEDI- or NEDI-styled assessments**

If consumption survey data for a specific population are available, the  $MC$ ,  $p_{97.5}$ *consumption* values in the IESTI equations may be replaced by statistics calculated from these data (at the consumed modelled food level).

If concentration monitoring data (retrospective) or concentration field trial data (prospective) are available, the  $MRL, STMR$  values in the IESTI equations may be replaced by statistics calculated from these data (at the consumed modelled food level).

In the current use of IESTI, the occurrence frequency (use frequency)  $OF$  is assumed to be 1. In alternative assessments, a more realistic estimate may be used. Such an estimate could be derived for example as the highest occurrence frequency observed in a retrospective assessment for either the same substance or the same food.

#### **Selection settings**

Name	Description	
Risk type	The type of exposure considered in the assessment; acute (short)	
	term) or chronic (long-term).	
Dietary exposure calculation	A tier is a pre-specified set of model configurations. By selecting a	
tier	model tier, MCRA automatically sets all model settings in this	
	module according to this tier. Note that currently tier setting may	
	need to be performed separately in sub-modules. Use the Custom	
	tier when you want to manually set each model setting.	

Table 2.144: Selection settings for module Single value dietary exposures.

# **Calculation settings**

Table 2.145: Calculation settings for module Single value dietary exposures.



# **Calculation of single value dietary exposures**

Single value dietary exposures are calculated from single value consumptions per food-as-measured and single value concentrations. Optionally, also processing factors, unit variability models and use frequencies are applied.

• *Single value dietary exposures calculation*

Inputs used: *Single value consumptions Single value concentrations Processing factors Unit variability factors Occurrence frequencies*

#### Settin[gs used](#page-196-0)

• *Calcul[ation Settings](#page-56-0)*

# **2.5 Hazard modules**

Hazard data exist at two levels: at a lower level *dose response data* give *responses* measured in *test systems* from doses of *active substances*. Such data can be modelled with *dose response models*.

At a higher level *responses* can be linked to *effects*, optionally via *AOP networks*, using *effect representations*. If benchmark responses (BMRs) have been specified, *dose response models* can calculate Benchmark Doses (BMDs), which are the preferred Points of departure i[n hazard assessmen](#page-209-0)[ts. I](#page-216-0)[n addition](#page-42-0), or alternati[vely, extern](#page-46-0)al *points of departure* [can be spe](#page-202-0)cified for *active substances* and *effects*.

BMDs from *dose [response m](#page-42-0)odels* and/or other *[points](#page-25-0) of departure* ca[n be converted](#page-206-0) to *haza[rd characterisations](#page-220-0)* at the intended level (external or internal dose, without or [with safety factors\), usi](#page-216-0)ng *kinetic models*, *inter-species co[nversions](#page-234-0)* [and/or](#page-234-0) *intra-species factors*. Finally, *[hazard cha](#page-202-0)ract[erisatio](#page-25-0)ns* can be translated to *relative potency factors*.

# **2.5.1 Ac[tive substance](#page-216-0)s**

<span id="page-202-0"></span>Active [substances are the s](#page-233-0)ubstance[s that may lead with non-](#page-221-0)zero probability (P [\(AG\)>0\) to a specific](#page-237-0) *health effect* (adverse outcome). In the simplest case, all substances in the scope of the action will form one assessment group (AG) of active substances. In more advanced cases, the list of active substances is derived from possibly multiple assessment group memberships, which are scores for substances that determine whether a substance is included (score  $> 0$ ) or excluded (score = 0) in the set of active substances. Substances with membership 0 are exclu[ded from the](#page-25-0) list of active substances. Memberships scores between 0 and 1 are treated as probabilities of being in the set of active substances. Assessment group memberships can be either specified directly as data or derived from *QSAR membership models*, *molecular docking models*, or from availability of *points of departure*.

This module has as primary entities: *Effects Substances*

Output of this module is used by: *Concentrations Single value concentrations Occurrence patterns Occurrence fr[equen](#page-241-0)cies [Substance conv](#page-241-0)ersions [Non-dietary exposu](#page-239-0)res Kinetic models Rel[ative potency factor](#page-234-0)s Hazard characterisations Inter-species conversions Intra specie[s factor](#page-25-0)s [Food con](#page-44-0)versions High exposure food-substance combinations Dietary exposures Exposures*

# **[Active substances](#page-231-0)[d](#page-231-0)[ata formats](#page-233-0)**

<span id="page-202-1"></span>[Active su](#page-116-0)[bstances as](#page-150-0) data have to be specified via assessment group (AG) memberships in an AG membership model. For each effect one or more AG membership models can be available, one of which should be chosen in assessments. The AG memberships can be crisp, i.e. a positive list of active substances (with default memberships 1, although it is also allowed to include the negative memberships with membership 0 explicitly) or probabilistic ( $0 \le P \le 1$ ).

#### **Assessment group membership models**

Assessment group membership models contain substance membership definitions for a given (health) effect. This data is described using two tables: the assessment group membership models table and the assessment group memberships table. The groups for a specified health effect are defined in the assessment group membership models table. The assessment group memberships table describes the substance memberships (or membership probabilities) in each group.

# **Assessment group membership models**

This table contains the definitions of the assessment group membership models. Each model contains a id, name, an optional description, and refers to its related health effect.

Name	Type	Description	Aliases	Required
id	AlphaNumeric(50)	The unique identification code	id, idModel,	Yes
		of the assessment group	Model,	
		membership model.	idAssessment-	
			GroupModel,	
			Assessment-	
			GroupModel,	
			idGroup-	
			Membership-	
			Model,	
			Group-	
			Membership-	
			Model	
Name	AlphaNumeric(100)	The name of the assessment	Name	N <sub>0</sub>
		group membership model.		
Description	AlphaNumeric(200)	Description of the assessment	Description	N <sub>o</sub>
		group membership model.		
idEffect	AlphaNumeric(50)	The effect code.	idEffect,	Yes
			EffectId, Effect	
Accuracy	Numeric	If applicable, the accuracy of	Accuracy	N <sub>o</sub>
		the assessment group		
		membership model		
		memberships.		
Sensitivity	Numeric	If applicable, the sensitivity of	Sensitivity	N <sub>0</sub>
		the assessment group		
		membership model.		
Specificity	Numeric	If applicable, the specificity of	Specificity	No
		the assessment group		
		membership model.		
Reference	AlphaNumeric(200)	External reference(s) to	References	$\overline{No}$
		sources containing more		
		information about the		
		assessment group model.		

Table 2.146: Table definition for AssessmentGroupMembershipModels.

Table aliases: AssessmentGroupMembershipModels, AssessmentGroupMembershipModel.

#### **Assessment group memberships**

Substances belong to an assessment group with certainty (probability 1), or the membership are uncertain. This table allows to specify membership probabilities for assessment group membership models. The probability should be a value between zero and one. For example, set to 1 or 0, or prior probabilities, or probabilities or 0/1 values estimated from QSAR, from Molecular Docking or from expert elicitation. The table can contain prior or posterior memberships. Default membership are specified with an empty idSubstance field.





Table aliases: AssessmentGroupMemberships, AssessmentGroupMembership.

#### **Active substances calculation**

Depending on the *model settings*, the set of active substances for a specified effect can be computed in several ways:

- <span id="page-204-0"></span>1. From the list of substances with available *points of departure (POD) data* for the specified effect. If there is a POD, then the substance is considered an active substance, with membership 1. If not, the membership is 0, and the sub[stance is exclu](#page-205-0)ded from the list of active substances.
- 2. From one or more in-silico (QSAR and/[or molecular docking\) models.](#page-234-0) The results of the in-silico models should be provided as *QSAR membership models data* and/or *molecular docking models data*. Binding energies from molecular docking models are first translated to crips memberships using a threshold value. The results from multiple in-silico models can be combined in any of four membership calculation methods:
	- 1. (crisp, any) the [substance is considered an active](#page-241-0) subst[ance if any in-silico model indic](#page-239-0)ates activity;
- 2. (crisp, majority) the substance is considered an active substance if the majority of in-silico models indicates activity;
- 3. (probabilistic, ratio) the membership probability is the fraction of in-silico models that indicate activity;
- 4. (probabilistic, Bayesian) the membership probability is calculated using a Bayesian model according to Kennedy et al. [Kennedy, 2019] and a specified prior probability (which is by default 0.5).

For substances within the scope of the assessment but without in-silico data, the default is to omit them in the AG. There is an option however to include such substances with a default membership probability.

3. From a combination [of 1 and 2, usin](#page-356-3)g either the union (OR) method or the intersection (AND method) of results.

# **Active substances settings**

# **Calculation settings**

<span id="page-205-0"></span>

Name	Description
Filter by certain assessment	Filter substances by certain assessment group membership.
group membership	
Filter by possible assessment	Filter substances by possible assessment group membership.
group membership	
Derive memberships from	Determine assessment group membership based on
POD presence	presence/absence of points of departure.
Restrict active substances to	Determine assessment group membership based on
substances with available	presence/absence of hazard characterisations.
hazard characterisations	
Derive memberships from	Specifies whether QSAR membership data is used for computing
QSAR membership data	the assessment group memberships.
Derive memberships from	Specifies whether molecular docking data is used for computing
molecular docking data	the assessment group memberhips.
Include substances without	For non-probabilistic methods: specifies whether substances for
membership information	which no membership information is available in the specified
	inputs should be included in the assessment group.
Combination method	Specifies whether to take the intersection or the union of the set of
memberships from available	substances with available PoDs and the set of substances with
PODs and in-silico data	positive/probable (in-silico) membership score.
Membership calculation	Calculation method for computing assessment group
method	memberships: majority/any (crisp methods), ratio/Bayesian
	(probabilistic methods)
Default/prior membership	Default substance membership probability for which no
probability	membership information is available in the specified inputs. Prior
	probability for Bayesian method.
Use probabilistic assessment	Specifies whether substance memberships should be expressed in
group memberships	terms of probabilities (probabilistic). Otherwise, substance
	memberships are expressed as in or out (crisp).

Table 2.148: Calculation settings for module Active substances.

Name	Description
Resample assessment group	Specifies whether assessment group memberships of substances
memberships	should be resampled using the assessment group membership
	probabilities.

Table 2.149: Uncertainty settings for module Active substances.

#### **Active substances as data**

When provided as data, in the form of assessment group memberships, the active substances are derived from the specified memberships.

• *Active substances data formats*

Inputs used: *AOP networks Points of departure Hazard characterisations*

#### **Calc[ulation of active substanc](#page-202-1)es**

Active subst[ances and asses](#page-206-0)[sment group memb](#page-234-0)[erships may be computed](#page-221-0) from PoD presence of in-silico data.

• *Active substances calculation*

Inputs used: *Molecular docking models QSAR membership models*

#### Settings used

• *[Calculation Settings](#page-204-0)*

# **2.5.2 AOP networks**

<span id="page-206-0"></span>Effect[s are related to each](#page-205-0) other using the toxicological concept of adverse outcome pathways (AOPs) and adverse outcome pathway networks (see https://aopwiki.org). Adverse Outcome Pathway (AOP) Networks specify how biological events (effects) can lead to an adverse outcome (AO) in a qualitative way through relations of upstream and downstream key events (KEs), starting from molecular initiating events (MIEs). Using AOPs, the adverse outcome (AO), e.g., liver steatosis, is linked to key events (KEs), e.g., triglyceride accumulation in the liver, and to molecular initiating events (MIEs), e.g., PPAR-alpha receptor antagonism. In general, multiple AOPs may lead to the same AO, and therefore AOP networks can be identified.

This module has as primary entities: *Effects*

Output of this module is used by: *QSAR membership models Molecular docking models Active substances Relative potency factors Hazard characterisations Points of departure Effect representations*

#### **AOP networks data formats**

#### **[AOP networ](#page-237-0)[ks](#page-221-0)**

AOP networks are described using two tables: the AOP networks table, and the effect relations table. The AOP networks table records the ids, names, descriptions, and other metadata of the AOP networks. The effect relations table describes the effects and effect relations (i.e., upstream and downstream key event relations) that are part of the AOP network.

# **AOP networks**

Data format for specification of adverse outcome pathway (AOP) networks.





Table aliases: AOPNetworks, AOPNetwork.

# **Effect relations**

Dataformat for specification of the effect (key event) relationships of adverse outcome pathway (AOP) networks.





Table aliases: EffectRelations, EffectRelation, EffectRelationships, EffectRelationship, KeyEventRelationships, KeyEventRelationship.

# **AOP networks settings**

# **Selection settings**

Name	Description
<b>AOP</b> Network	The AOP networks of interest.
Restrict AOP network by focal	Restrict the AOP network to a specific sub-network, containing
upstream event	only the AOPs that include both the focal key event (KE) defined
	here (which must be upstream of the AO) and the focal effect
	(adverse outcome, AO)
Focal upstream event	The focal key event used for restricting the AOP network to a
	specific sub-network of interest.

Table 2.152: Selection settings for module AOP networks.

#### **AOP networks as data**

AOP networks can only be provided as data in the form of network definitions containing effect relations (key-event relationships) collections.

• *AOP networks data formats*



Figure 2.38: AOP network

# **2.5.3 Dose response data**

Dose response data are data on response values of test systems at specified doses of substances (or mixtures of substances) from dose response experiments.

<span id="page-209-0"></span>This module has as primary entities: *Substances Test systems Responses*

Output of this module is used by: *Dose response models*

# **Dose response data data form[ats](#page-44-0)**

The meta-data of dose response e[xperiments \(such as na](#page-216-0)me, description, etc.) are specified in the DoseResponseExperiments table.

For presenting the data of these experiments to the system, there are two formats: a single table format (DoseResponseData) and a relational data format (three tables DoseResponseExperimentDoses, ExperimentalUnitProperties, DoseResponseExperimentMeasurements). Usually, the single table format will be the easier one. For internal use in MCRA, this single table data is converted to the relational data format.

# **Dose response data**

Dose response data are used to extract assessment group membership or hazard doses . The meta-data of dose response experiments (such as name, description, etc.) are specified in the DoseResponseExperiments table. For presenting the data of these experiments to the system, there are two formats: a single table format (DoseResponse-Data) and a relational data format (three tables). Usually, the single table format will be the easier one. For internal use in MCRA, this single table data is converted to the relational data format.

#### **Dose response experiments**

General information about the dose response experiments, such as the (unique) identifier, name, description, the used test-system, and the dose unit is stored in the table DoseResponseExperiments. If the data of an experiment is provided in a single table format, then the fields Time, Covariates, Substances, and Responses are used to map the column header names of the columns of the single data table to these their respective types.

Table 2.153: Table definition for DoseResponseExperiments.

Name	<b>Type</b>	Description	<b>Aliases</b>	Required
idExperiment	AlphaNumeric(50)	Unique identification code of	idExperiment,	Yes
		the dose effect experiment.	Id, Code	
Name	AlphaNumeric $(100)$	Name of the dose effect	Name <sup></sup>	$\overline{No}$
		experiment.		
Description	AlphaNumeric(200)	Description of the dose effect	Description	N <sub>o</sub>
		experiment.		
Date	<b>DateTime</b>	The starting date of the	Date	N <sub>o</sub>
		experiment.		
Reference	AlphaNumeric(200)	External reference, for	Reference	No
		instance, to the experiment		
		protocol and/or supporting		
		material.		
Experimental-	AlphaNumeric(100)	The name of the experimental	Experimental- Unit	No
Unit		unit of the experiment, e.g.,		
		rat, cage, litter, vial, cup, petridish.		
<b>DoseRoute</b>	AlphaNumeric(100)	For in-vivo test systems, the	<b>DoseRoute</b>	No
		route in which the dose was		
		administered		
Substances	AlphaNumeric	Code or comma separated list	idSubstance,	Yes
		of the codes of the substances	SubstanceId,	
		measured in the experiment.	SubstanceCode,	
		E.g., 'Cyproconazole,	Substance,	
		Thiram'. Required when	idSubstances,	
		presenting the dose-response	SubstanceIds,	
		data in a single table. Make	SubstanceCodes,	
		sure that in table	Substances	
		DoseResponseData the		
		column headers exactly match		
		these names.		
DoseUnit	<b>DoseUnits</b>	Unit of the doses	<b>DoseUnit</b>	Yes
		administered in this		
		experiment.		
Responses	AlphaNumeric	Code or comma separated list	Responses,	Yes
		of codes of the responses	Response,	
		measured in the experiment.	idResponses,	
		E.g., 'AngleM_PQ,	idResponse	
		Mortality'. Required when		
		presenting the dose-response		
		data in a single table. Make		
		sure that in table		
		DoseResponseData the		
		column headers exactly match		
		these names.		
Time	AlphaNumeric(100)	Identifier of the time field of	Time, Times	N <sub>o</sub>
		the experiment. Required		
		when presenting the		
		dose-response data in a single		
		table and responses are		
		measured at multiple times.		
		Make sure that in the table		
		DoseResponseData the		
		column header of the		
		time-column exactly matches		
		this name.		
TimeUnit	TimeUnit	Unit of the time scale used in	TimeUnit	No
		the experiments.		
$204$ Covariates	AlphaNumeric(200)	Comma separated list of the	<b>CovariatesChapter R<sub>o</sub> Modules</b>	
		names/codes of the covariates	Covariate	
		of the experiment. E.g.		
		'Gender, Inhibitor,		

Table aliases: DoseResponseExperiments, DoseResponseExperiment.

# **Dose response data**

Single (two-way) table data format for specifying data of dose response experiments (as alternative for the relational format). The column headers are dynamic and should be defined in the table DoseResponseExperiments through fields Substances and Responses (and, optionally, Covariates and Time). For responses given as aggregated statistics, also SD, CV, N and Uncertainty are specified as [Datatype:Response]. E.g., 'SD:Y', 'CV:Y', 'N:Y'. Uncertainty upper 95%limits are specified as 'UncertaintyUpper:Y'. For each quantal response an additional column 'N:[responsename]'is required with binomial totals (e.g. Mortality = 3, N:Mortality = 10).





Table aliases: TwoWayDoseResponseData, DoseResponseDataTwoWay, DoseResponseData.

# **Relational dose response data**

In the relational data format, dose response experiment data are specified using the triplet of tables: DoseResponseExperimentDoses, DoseResponseExperimentMeasurements, and ExperimentalUnitProperties. These tables describe, respectively, the experiment designs (including the administered substance doses), the response measurements, and additional properties of the experimental units of the experiment.

# **Dose response experiment doses**

The table DoseResponseExperimentDoses describes the experiment design, being a complete specification of which doses of which substances were applied to which experimental unit and if relevant at what time.

<b>Name</b>	Type	Description	<b>Aliases</b>	Required
idExperiment	AlphaNumeric $(50)$	Identification code of the	idExperiment,	<b>Yes</b>
		experiment to which this	Experiment	
		design record belongs.		
idExperimental-	AlphaNumeric $(50)$	Identification code of the	idExperimental-	<b>Yes</b>
Unit		experimental unit to which	Unit.	
		the dose is applied.	Experimental-	
			Unit	
Time	Numeric	The time of administration of	Time	No
		the dose.		
idSubstance	AlphaNumeric $(50)$	Code of the substance that	idSubstance,	<b>Yes</b>
		was administered.	SubstanceId,	
			SubstanceCode,	
			Substance	
Dose	Numeric	The dose that was	Dose	<b>Yes</b>
		administered.		

Table 2.155: Table definition for DoseResponseExperimentDoses.

Table aliases: DoseResponseExperimentDoses, DoseResponseExperimentDose.

# **Experimental unit properties**

The table ExperimentalUnitProperties are used to specify additional properties of the experimental units of the experiment. For instance, the gender of the rat, in case rats are the experimental units.





Table aliases: ExperimentalUnitProperties, ExperimentalUnitProperty.

# **Dose response experiment measurements**

The table DoseResponseMeasurements describes the measurements that were done in the experiments. That is, for each response and experimental unit, at each observation time, one measurement should be recorded. If the response is an aggregated statistic, then this record may also include a standard deviation and number of units over which was aggregated.





Table aliases: DoseResponseExperimentMeasurements, DoseResponseExperimentMeasurement, DoseResponseMeasurements, DoseResponseMeasurement.

# **Dose response data settings**

# **Selection settings**

Name	Description
Experiments	The dose response experiments of interest.
Merge dose response data of	Specifies whether the dose response data of multiple experiments
multiple experiments	should be merged into one large dose response data set.

Table 2.158: Selection settings for module Dose response data.
#### **Dose response data as data**

Dose response data are provided per experiment or study in which several responses (on in-vitro or in-vivo test systems) are measured from several administered substance doses.

• *Dose response data data formats*

# **2.5.4 Dose response models**

<span id="page-216-1"></span>Dose [response models are models fitted](#page-209-0) to dose response data and can be provided as data or calculated using a local or remote version of PROAST. The main results for hazard and risk assessment are benchmark doses (BMDs), related to a specified substance, response, optionally covariate value, and the benchmark response (BMR). Dose response models can be uploaded as data, retrieved from PROASTweb through *linked remote repositories*, or *calculated using an internal version of PROAST*.

This module has as primary entities: *Test systems Responses Substances*

Output of this module is used by: *Hazard char[acterisations](https://proastweb.rivm.nl/user/login)*

### **Dose response models data f[ormats](#page-46-0)**

#### **Dose response models**

<span id="page-216-0"></span>Dose response models are specified using three tables: the dose response models table holds the dose response model definitions (id, name, description) and other information about the dose response models. The dose response model benchmark doses table records the benchmark doses and (optionally) the model parameters for specific substances and covariates. The dose response model benchmark doses uncertainty table records results from bootstrap runs for the benchmark doses per substance/covariate combination.

### **Dose response models**

Each dose response model has a unique id, a name (optional), and description (optional). Also, each dose response model is associated with a specific dose response experiment (idExperiment) from which the data used to create the model is obtained, a response (idResponse), one or more substances, and, optionally, specific covariates considered by the dose response model. The combination of the benchmark response type and the associated value define the benchmark response of the model. The dose unit specifies the unit used for the doses, and if applicable, the model equation can be specified.

<b>Name</b>	Type	Description	<b>Aliases</b>	Required
idDose-	AlphaNumeric $(50)$	The unique identification code	idDose-	<b>Yes</b>
<b>ResponseModel</b>		of the fitted dose response	ResponseModel,	
		model.	idModel	
idExperiment	AlphaNumeric(50)	The identification code of the	experiment-	Yes
		experiment from the dose	Code,	
		response model.	experimentId	
Name	AlphaNumeric(100)	The name of the dose	Name	$\overline{No}$
		response model.		
Description	AlphaNumeric(200)	Description of the dose	Description	$\overline{No}$
		response model.		
Substances	<b>AlphaNumeric</b>	Code or comma separated list	Substances	Yes
		of the codes of the substances		
		in the Dose Response Model.		
		E.g., 'Cyproconazole,		
		Thiram'.		
idResponse	<b>AlphaNumeric</b>	The response of the dose	idResponse,	Yes
		response model.	Response	
Covariates	AlphaNumeric	The covariates considered by	Covariates,	$\overline{No}$
		the dose response model.	Covariate	
Benchmark-	Numeric	The value of the benchmark	Benchmark-	Yes
Response		response or critical effect size.	Response,	
			CriticalEffect-	
			Size,	
			<b>CES</b>	
Benchmark-	Benchmark-	Specifies how the benchmark	Benchmark-	$\overline{No}$
ResponseType	ResponseTypes	response is expressed. E.g.,	ResponseType,	
		using a percent change in	HazardEffect-	
		mean response or, for quantal	SizeType,	
		response types, in terms of	CriticalEffect-	
		extra risk, additional risk, or	SizeType	
		ED50.		
LogLikelihood	Numeric	Loglikelihood of the model	LogLikelihood	$\overline{No}$
		fit.		
<b>DoseUnit</b>	AlphaNumeric(50)	The dose unit (if not specified,	DoseUnit,	$\overline{\text{No}}$
		then mg/kg is assumed).	UnitDose	
ModelEquation	AlphaNumeric(500)	If available, the model	ModelEquation,	$\overline{No}$
		equation of the dose response	DoseResponse-	
		model (R model equation) or	ModelEquation,	
		the identifier of the dose	Equation	
		response model type.		

Table 2.159: Table definition for DoseResponseModels.

Table aliases: DoseResponseModels, DoseResponseModel.

# **Dose response model benchmark doses**

The benchmark responses and benchmark doses belonging to the dose response models are recorded per substance/covariate in the dose response model benchmark doses table. Optionally, if the model equation of the dose response model has been specified in the dose response models table, the model parameter values for this specific substance/covariate can be specified here.





Table aliases: DoseResponseModelBenchmarkDoses.

# **Dose response model benchmark dose bootstraps**

Empirical uncertainty values of the benchmark benchmark doses of dose response models can be recorded in the dose response model benchmark doses bootstraps table. The uncertainty set identifier (idUncertaintySet) can be specified to retain correlations between uncertainty records that originate from the same bootstrap run.



Table 2.161: Table definition for DoseResponseModelBenchmarkDosesUncertain.

Table aliases: DoseResponseModelBenchmarkDosesBootstraps, DoseResponseModelBenchmarkDosesUncertain.

### **Dose response models calculation**

Besides uploading dose response models as data or retrieving them from PROASTweb through *linked remote repositories*, there is also a possibility to compute dose response models using an integrated version of PROAST. When computing dose response models using the integrated version, MCRA will attempt to fit a dose response model for each response of each dose response experiment. Depending on the type of data (e.g., response type, covariates y/n, single or multiple substances) a PROAST run is configured and exec[uted. If](https://proastweb.rivm.nl/user/login) *effect representations* [are provided,](#page-15-0) [then b](#page-15-0)enchmark responses specified by the effect representations data are used, otherwise only the model fits will be computed without benchmark doses.

### **Dose response models as data**

Dose response models as data contain the details of fitted dose response models. The main elements for hazard and risk assessment are the benchmark doses (BMDs) related to specified substances, responses, and optionally covariate values for specified benchmark responses (BMR). These specifications can be provided in data files or can be retrieved/imported from PROAST output files on the PROAST website https://proastweb.rivm.nl/user/login using a PROASTweb user account and an application access key.

• *Dose response models data formats*

Inputs used: *Dose response data*

# **Calc[ulation of dose response mod](#page-216-0)els**

Used as a c[alculator, dose resp](#page-209-1)onse models are fitted to dose response data using an MCRA-internal version of PROAST. Currently, all available models appropriate for the response type will be fitted, and for the Hill and Exponential model families, the best fitting model based on maximum likelihood will be selected. The set of results for the calculation will include BMDs etc. for all fitted models.

• *Dose response models calculation*

Inputs used: *Effect representations*

# **2.5.5 Effect representations**

Effect representations specify the responses that can be used to measure specified effects and which response levels, the benchmark response (BMR), define the hazard limits for the effects.

This module has as primary entities: *Effects Responses*

Output of this module is used by: *Hazard characterisations Dose response models*

### **Effect representations data fo[rmats](#page-25-0)**

### **Effect representations**

<span id="page-220-0"></span>Effect representations specify responses that may represent the effect.

### **Effect representations**

One response can be set as the canonical response (golden standard). For a quantitative or stochastically qualitative canonical response a benchmark response should be defined.





Table aliases: EffectRepresentations, EffectRepresentation.

### **Effect representations as data**

Effect representations are provided as data in the form of specified combinations of effect and response, optionally with a benchmark response that defines a hazard limit for the effect.

• *Effect representations data formats*

Inputs used: *AOP networks*

# **2.5.6 [Hazard characterisatio](#page-220-0)ns**

Hazard char[acterisations a](#page-206-0)re reference exposure values for active substances at the chosen biological target level (external or internal). Hazard characterisations may be specified for specific effects or for the critical effect as defined in hazard characterisation. Hazard characterisations are specified as external values (e.g. human based guidance values, such as ADI or ARfD) or are based on points of departure, such as BMDs from dose-response models or externally specified points of departure (NOAEL, LOAEL, MDS). The computation may involve assessment factors, e.g. for inter-species conversion, intra-species variation or additional sources of uncertainty. The calculation may also use kinetic models or absorption factors to convert external doses to internal doses or vice versa.

This module has as primary entities: *Substances Effects*

Output of this module is used by: *Active substances Relative potency factors Risks Single value risks*

# **Hazard characterisations data [formats](#page-44-0)**

### **Hazard characterisations**

<span id="page-221-0"></span>Hazard characterisations provide reference threshold values associated with the hazard of interest. Examples are health-based guidance values such as ADI or ARfD, and points of departure such as BMD or NOAEL.

### **Hazard characterisations**

Hazard characterisations are specified for combinations of hazard characterisation type, effect, substance, population type, target level, and exposure route (for external) or target organ (for internal). Effects can be specific, but can also be labeled as being the critical effect and used as such if this has been specified in the hazard characterisation settings.

Table 2.163: Table definition for HazardCharacterisations.

Name	$\mathsf{\overline{Type}}$	Description	Aliases	Required
idHazard-	AlphaNumeric(50)	Id of the hazard	id, idHazard-	Yes
Characterisation		characterisation.	Characterisation	
idEffect	AlphaNumeric(50)	Code of the (critical) effect	idEffect,	$\overline{No}$
		linked to this hazard	EffectId, Effect	
		characterisation.		
idSubstance	AlphaNumeric(50)	The code of the substance.	idSubstance	Yes
idPopulation-	AlphaNumeric(50)	The code of the population	idPopulation-	$\overline{N_{0}}$
Type		type for which this reference	Type	
		value is defined. If not		
		specified, PS06A, Consumers		
		is assumed.		
TargetLevel	TargetLevelType	The target level. I.e., internal	TargetLevel	$\overline{No}$
		or external. If omitted,		
		external is assumed		
<b>ExposureRoute</b>	ExposureRouteTypes	The exposure route (only	<b>ExposureRoute</b>	$\overline{No}$
		applicable if target level is		
		external). If not specified,		
		Dietary is assumed.		
<b>TargetOrgan</b>	AlphaNumeric(50)	The target organ (should be		$\overline{No}$
		specified when target level is		
		internal).		
<b>IsCriticalEffect</b>	Boolean	Specifies whether this value is	<b>IsCriticalEffect</b>	$\overline{No}$
		the value associated with the		
		critical effect. If omitted, No		
		is assumed		
<b>ExposureType</b>	<b>ExposureTypes</b>	The exposure type associated	<b>ExposureType</b>	Yes
		with the hazard		
		characterisation (i.e., chronic		
		or acute).		
Hazard-	Hazard-	The type of the hazard	Hazard-	Yes
Characterisation-	Characterisation-	characterisation (e.g., ARfD,	Characterisation-	
Type			Type	
Qualifier	<b>Types</b> QualifierType	ADI, NOAEL, BMD). Qualifier of the hazard	QualifierType	$\overline{No}$
		characterisation value, e.g.		
		equal-to $(=)$ or smaller-than		
		$(<)$ . If omitted, = is assumed.		
Value	Numeric	Reference value that	Value, Hazard-	Yes
		characterises the hazard.	Characterisation-	
			Value	
				Yes
DoseUnit	<b>DoseUnits</b>	Unit of the hazard	DoseUnit, Unit	
		characterisation value.		
idPointOf-	AlphaNumeric $(50)$	The code of the point of	idHazardDose,	$\overline{No}$
Departure		departure from which this	idPod	
		hazard characterisation was		
		derived.		
Combined-	Numeric	Combined assessment factor	Combined-	N <sub>o</sub>
Assessment-		(includes, e.g., safety factor,	Assessment-	
Factor		but also other extrapolation	Factor	
		factors that may be used to		
		derive the hazard		
		characterisation from the		
		underlying PoD).		
<b>PublicationTitle</b>	AlphaNumeric	Title of the publication of the	PublicationTitle,	N <sub>o</sub>
		study in which this hazard	Title	
		characterisation was		
		established.		
Publication- 2.5 Hazard modules Authors		Author(s) of the publication	Publication-	$\overline{\text{No}}$
		of the study in which this	Authors,	215
		hazard characterisation was	Publication-	
		established.	Author, Author,	
			Authors	

# **Hazard characterisations calculation**

<span id="page-223-0"></span>Hazard characterisations are defined as deterministic threshold values (e.g. ADI, ARfD) or as distributions (using probabilistic models). They are linked to an effect of interest or alternatively are defined for the critical effect. Hazard characterisations depend on the *risk type* (acute or chronic) and the biological *target level* of the human body (external via some route of exposure or internal for a specific defined organ or compartment). Hazard characterisations are derived from *points of departure* provided as data and/or from *dose-response models*. The procedure for computing hazard characterisations has two main phases: 1) collection of all available hazard characterisation candidates and alignment with the target syste[m, and 2\)](#page-230-0) aggregation over multiple available [hazard char](#page-230-0)acterisations and imputation of missing hazard characterisations.

Collection of [available hazard cha](#page-234-0)racterisation candidates invol[ves collecting the appr](#page-216-1)opriate points of departure data and/or dose-response models that are used for deriving the hazard characterisations. In MCRA, a distinction is made between three *methods for computing hazard characterisations*:

- 1. Calculation of hazard characterisations from externally specified in-vivo points of departure (PoD, e.g. BMD, NOAEL, LOAEL).
- 2. Calculat[ion of hazard characterisations from PoDs \(in th](#page-230-0)is case BMD) calculated from dose response data.
- 3. Calculation of hazard characterisations based on an *in-vivo PoD for the index substance and in-vitro RPFs from dose-response models for the other substances (IVIVE model)* (cumulative assessments only).

For all three methods, the collected points of departure and benchmark doses should be aligned with the target system. This alignment may involve various conversion steps for [each point of departure and specific substance, and can be](#page-225-0) forma[lly specified as:](#page-225-0)

$$
HC = f_{\text{expression-type}} \cdot f_{\text{kinetic}} \cdot \frac{1}{f_{\text{inter-species}}} \cdot \frac{1}{f_{\text{intra-species}}} \cdot \frac{1}{f_{\text{additional}}} \cdot PoD
$$

where:

- *HC* denotes the hazard characterisation.
- *f*expression−type denotes the *expression type correction factor*, e.g., for extrapolation from LOAEL or NOAEL, or from NOAEL to BMD.
- $f_{\text{kinetic}}$  denotes the kinetic conversion factor for *conversion from internal to external or external to internal hazard characterisations*.
- *f*inter−species denotes the inte[r-species factor for](#page-224-0) *extrapolation from animal to human (inter-species)*.
- *f*intra−species denotes the intra-species factor for *[extrapolation from the average to the sensitive human or](#page-226-0) [probabilistic calculation](#page-226-0) of the distribution of human individuals (intra-species)*.
- $f_{\text{additional}}$  denotes the additional assessment factor for *[extrapolation from the POD to the hazard cha](#page-224-1)racterisation in humans for sources where appropriate data [or information is scarce or missing \(additional\)](#page-224-2)*.
- *PoD* [denotes the point of departure.](#page-224-2)

Note that inter- and intra-species extrapolation and the use ofa[n additional assessment factor are optional. However,](#page-225-1) expre[ssion type correction and the kinetic conversion are always applied \(when relevant\) whatever option i](#page-225-1)s chosen.

Occasionally, for some substances multiple hazard characterisations are available (e.g., obtained from multiple experiments) and for others substance hazard characterisations are still missing. Hence, two final steps remain to come to the final set of hazard characterisation:

- *Aggregation over multiple available hazard characterisations*. Set the *selection method in case of multiple candidate hazard characterisations* from *MostToxic* to *Aggregate*.
- *Imputation of missing hazard characterisations*. Check the option *Imput missing hazard characterisations*, to [select the Imputation method.](#page-229-0)

### **Hazard characterisation type extrapolation**

<span id="page-224-0"></span>Hazard doses, or points of departure can be of *various types*. E.g., BMDs, NOAELs, or LOAELs. When computing hazard characterisations, the type in which the hazard characterisations are expressed (i.e., the *hazard characterisation expression type*) should be specified explicitly. When points of departure from types different from the expression type are provided, these should be translated to the specified expression level. In the current implementation, the simple conversion factors shown in Table 2.164 are us[ed, roughly b](#page-347-0)ased on the WHO guidance document on evaluating and expressing uncertainty in hazard characterization [WHO, 2018].

From	Т٥	Conversion factor
<b>BMD</b>	<b>NOAEL</b>	1/3
<b>BMD</b>	<b>LOAEL</b>	
<b>NOAEL</b>	<b>BMD</b>	3
<b>NOAEL</b>	<b>LOAEL</b>	1/3
<b>LOAEL</b>	<b>BMD</b>	
<b>LOAEL</b>	<b>NOAEL</b>	1/3

Table 2.164: Conversion factors for hazard characterisation types.

### <span id="page-224-3"></span>**Inter-species extrapolation**

<span id="page-224-1"></span>Hazard doses, or points of departure, are commonly only determined for animals, not for humans. In order to derive hazard characterisations for humans, the animal hazard doses need to be converted to toxicologically equivalent doses for humans. This extrapolation is usually expressed as a multiplication factor, and traditionally a factor of 10 is used (which is roughly obtained from the product of a factor of 3.2 for toxicokinetic variability and a factor 3.2 for toxicodynamic variability).

The following methods are available within the toolbox:

- 1. **No inter-species extrapolation:** Assume that for all available points of departure, the animal hazard dose is equal to the human hazard dose. Effectively, this is equivalent to using a conversion factor of 1.
- 2. **Default distribution:** Use a conversion factor drawn from a default, substance and species independent lognormal uncertainty distribution specified (as *model settings*) by a geometric mean (GM) and geometric standard deviation (GSD). In the *nominal run*, the nominal value of this distribution (i.e., the geometric mean) is used as a conversion factor. In the *uncertainty analysis loop*, provided that inter-species extrapolation uncertainty is *included in the uncertainty analysis*, a single factor is drawn from the lognormal distribution.
- 3. **Substance/species specific distributions:** [Use conversio](#page-232-0)n factors drawn from substance/species specific lognormal uncertainty distr[ibutions spec](#page-11-0)ified (as *data*) by a geometric mean (GM) and geometric standard deviation (GSD). In the *nominal run*[, a factor equal to the ge](#page-11-0)ometric mean is used for all combinations of substance [and species. In the](#page-233-0) *uncertainty analysis loop*, provided that inter-species extrapolation uncertainty is *included in the uncertainty analysis*, a uncertainty factor is drawn from the lognormal distribution with  $\mu = 0$  and  $\sigma^2 = 1$ , which is used to obtain correlated draws for [all av](#page-231-0)ailable inter-species conversion factor distributions. If the distribution param[eters are mis](#page-11-0)sing for a specific substance/species, then the default distribution is used as a fallback.

### **Intra-species extrapolation of hazard characterisations**

<span id="page-224-2"></span>There is variation between individuals concerning their individual sensitivities to experience health effects. In some scenarios the aim is to perform assessments for the sensitive individuals instead of the average individuals for which the points of departure are derived. If this is the case, then extrapolation is required to translate hazard characterisations derived for the average individual to hazard characterisations for a sensitive individual. In traditional exposure assessments, a safety of 100 is commonly used as a margin of safety, that is assumed to be composed of a interspecies extrapolation factor (factor 10), and inter-individual extrapolation factor (factor 10). I.e., the hazard characterisation defined for the sensitive individual is defined as

$$
HC_{\text{sens}} = \frac{1}{f_{\text{intra-species}}} \cdot HC_{\text{avg}}
$$

Here  $f<sub>inter-species</sub>$  denotes the intra-species factor. An alternative to using a fixed safety factor, is to define intraspecies variability may be explicitly *a lognormal distribution*, characterised by a geometric mean (GM) equal to 1 and a geometric standard deviation (GSD). For *risks calculations*, this distribution could be used to sample individual hazard characterisations. This effectively converts the description of hazard characterisations to include variability, with an unbiased central value.

### **Additional assessment factors**

<span id="page-225-1"></span>In cases where where appropriate data or information is scarce or missing, an additional assessment factor is used for extrapolation from the POD to the hazard characterisation in humans. This factor can be used as a worst case value (preferably  $AF = 100$ ) for inter- and intra-species extrapolation, but it may equally well serve as an additional extrapolation factor next to inter- and/or intra-species extrapolation. In the latter case the factor merely serves to account for differences in for example route to route, metabolic rates in interspecies, completeness and consistency of available data, reliability of alternative data (e.g. read-across) or quality of data in general.

# **In-vitro in-vivo extrapolation (IVIVE)**

The in-vitro in-vivo extrapolation method implemented in MCRA is based on the following prerequisites:

- <span id="page-225-0"></span>1. For one substance, the index substance, a reliable point of departure is available for the adverse outcome of interest obtained from an in-vivo assay (i.e., external dose).
- 2. There are other substances for which there is a dose-response model available from an in-vitro assay on a response representing an early key event of the adverse outcome for these substances and the index substance.

In IVIVE, these RPFs, in combination with the known hazard characterisation of the index substance, can be used to derive hazard characterisations for the other substances as well. Figure 2.39 shows the conceptual model that forms the basis of the IVIVE methodology of MCRA.



Figure 2.39: Conceptual model IVIVE.

### **IVIVE for calculating internal hazard characterisations**

- 1. Translate the (external) PoD of the index substance substance to an internal hazard characterisation for the human target system/compartment.
- 2. If the RPFs are obtained are obtained using mol-based specification of the doses, then convert the mol-based RPFs to mass-based RPFs. I.e.,

$$
\mathit{RPF}_{\mathtt{mass-based},i} = \mathit{RPF}_{\mathtt{mol-based},i} \cdot \frac{\mathit{MW}_{\mathtt{ref}}}{\mathit{MW}_i}
$$

3. Derive internal hazard characterisations for the human target system for the other substances by multiplying the RPF obtained from dose-response modelling with the hazard characterisation of the index substance. I.e.,

$$
HC_i = HC_{\text{ref}} \cdot RPF_{\text{mass-based},i}
$$

### **IVIVE for calculating external hazard characterisations**

- 1. Translate the PoD of the index substance to an external human hazard characterisation (dietary/oral exposure route).
- 2. Derive an internal hazard characterisation for the index substance, with an target organ/compartment representative for the response of the dose-response model.
- 3. If the RPFs are obtained are obtained using mol-based specification of the doses, then convert the mol-based RPFs to mass-based RPFs.
- 4. Derive internal hazard characterisations for the human target system for the other substances by multiplying the RPF obtained from dose-response modelling with the hazard characterisation of the index substance.
- 5. Convert the internal hazard characterisations of the other substance to external hazard characterisations for the dietary/oral exposure route using.

### **Kinetic conversion of hazard characterisations**

<span id="page-226-0"></span>When the *hazard characerisation level* is internal and points of departure are available for external exposures (e.g., NOAELs from in-vivo animal studies) or when the hazard characterisation level is external and benchmark doses are available at the internal level, then *kinetic conversion models* are needed to *translate the external doses to equivalent internal doses at the target compartment/organ* of interest or *vice-versa*.

In the too[lbox, this alignment from int](#page-230-0)ernal to external or from external to internal is generally termed *kinetic conversion*, associated with a *kinetic conversion factor*. The kinetic conversion factor is a multiplication factor needed to obtain a hazard characterisation [on the target level from](#page-244-0) a hazard char[acterisation of the point of departure or](#page-227-0) [benchmark dose. Depending on the chosen ki](#page-227-0)netic modelli[ng tier, this](#page-228-0) kinetic conversion factor may be 1) assumed to be one, 2) derived from absorption factors, or 3) derived using PBPK models.

An important detail in the use of kinetic conversion factors for computing hazard characterisations is the order between kinetic conversion and inter-species extrapolation. Notice that when points of departure are determined for animals, a choice should be made regarding the order of inter-species extrapolation and kinetic modelling. That is, one may first choose to convert animal external point of departure to an internal hazard characterisation for that animal, using the available animal kinetic model. Alternatively, one may first extrapolate the animal external point of departure to a human external hazard characterisation, and thereafter apply the human kinetic model to obtain internal hazard characterisations. In the toolbox, only the latter approach is currently implemented.

### **Extrapolation from external to internal hazard characterisations**

<span id="page-227-0"></span>The calculation of internal hazard characterisations based on external hazard characterisations is similar to the procedure for *computing internal exposures*. In the simplest tier, equivalence can be assumed between internal and external hazard characterisations, and in higher tiers absorption factors, respectively PBPK models can be used.

### **Calcul[ation of internal doses u](#page-155-0)sing absorption factors**

In the simplest form, internal doses are obtained from external exposure concentrations using multiplication factors (or, absorption factors) that can be specified by substance and by route. That is, for a given substance, the internal hazard characterisation  $HC_{\text{int}}$  can be derived from an external hazard characterisation  $HC_{\text{ext}}$  as

$$
H\!C_{\text{int}} = f_{\text{abs},r} \cdot H\!C_{\text{ext},r}
$$

Here, *r* denotes the route of the external exposure  $HC_{ext}$ , and  $f_{abs,r}$  denotes the absorption factor of route *r* for the specified compartment. Note that this model assumes that the external hazard characterisations are specified as concentrations (i.e., substance amount divided by the body weight).

### **Calculation of internal doses using human PBPK models**

A more detailed alternative to using absorption factors is to use one of the *advanced PBPK models* available in MCRA. In this approach, for each substance independently, an external exposure equivalent to the dose of the external hazard characterisation is presented to a representative simulated individual for a number of simulated days to the PBPK model of the individual. This representative individual should represent the "average" individual of the population, with nominal physiological properties (e.g., an average bodyweight of 7[0kg\). This yields a time](#page-246-0) course of the internal substance amount at the specified target compartment/organ from which a long term average substance amount (chronic) or peak substance amount (acute) can be obtained. By dividing this substance amount by the weight of the compartment, an internal concentration is obtained, which then represents the internal hazard characterisation.

More details on computing internal doses from external doses can be found in the description of the *calculation of internal exposures from external exposures*. For both tasks, the procedure for computing internal exposures/doses is exactly and the same *kinetic model settings*, such as *dosing patterns* and *non-stationary period* period apply for calculation of internal hazard characterisations as well.

### **[Calculation of intern](#page-156-0)[al doses using an](#page-245-0)imal PB[PK models](#page-158-0)**

In the above methods, the assumption is that the external points of departure (often obtained from experiments on animals) are first converted to external hazard characterisations for humans, and a human kinetic model is used for obtaining the internal hazard characterisations. As mentioned, an alternative approach is to use first the animal PBPK models to derive an internal hazard characterisation specific for the tested animal species and thereafter extrapolate to humans. When there are more precise kinetic models available for the animal used in the experiments for obtaining the point of departure, this could be a preferred path.

**Note:** Notice that this procedure is not yet implemented.

### **Extrapolation from internal to external hazard characterisations**

<span id="page-228-0"></span>In some cases, hazard characterisations are available at the internal level whereas the specified *hazard characerisation level* is external. This situation may occur, for instance, in *in-vitro in-vivo extrapolation (IVIVE)*. In this case, conversion is needed from the internal level to the external level, where the external level is implicitly defined as coming from the dietary/oral route of exposure.

When using absorption factors, the external (dietary) hazard characterisation of a substance [is simply computed by](#page-230-0) [divid](#page-230-0)ing the internal hazard characterisation by the dietary [absorption factor. I.e.,](#page-225-0)

$$
HC_{\text{ext},\text{dict}} = \frac{HC_{\text{int}}}{f_{\text{abs},\text{dict}}}
$$

When using PBPK models, reverse dosimetry is needed to find for the available internal hazard characterisation, the corresponding external (dietary) doses that yield the internal concentrations specified by the internal hazard characterisation. In MCRA, this is done using a bisection method, in which external doses are systematically fed to the PBPK model in order to converge to an external dose that yields the specified internal hazard characterisation with some level of precision.

### **Hazard characterisation imputation**

In some cases it may be that there are substances that are known to cause (or may possibly cause) the effect of interest, but for which there are no data available for obtaining hazard characterisations. I.e., for these substances, there are no points of departure or dose response models. Instead of excluding these substances in quantitative analyses, it is also possible to impute hazard characterisations for these substances based on hazard characterisations of other (similar) substances, and use these for calculating, e.g., relative potency factors or for risk assessment.

### **Munro P5 (TTC approach)**

The Threshold of Toxicological Concern (TTC) is an example of a tier for extrapolation of hazard characterisations from other substances that is already in common use (see [Munro et al., 1996]). The *Munro collection of NOELs/LOAELs* is a collection of NOELs/LOAELs for chemicals for the critical (i.e., first occurring) effect. In the TTC approach, the toxicity of an unknown substance is, depending on its Cramer class (see [Cramer et al., 1976]), imputed by the 5th percentile NOAEL of the sub-collection of the corresponding Cramer class.

Two variations of this approach are to use the empirical NOAEL [distribution itself \(ju](#page-356-0)st samp[le from the NOAEL](#page-339-0) [data\), or to fit a](#page-339-0) distribution (e.g. lognormal) to the empirical data and sample from the parametric distribution. MCRA provides an implementation of the TTC approach that uses the empirical distributio[n. In the nominal ru](#page-354-0)n, this implementation imputes the hazard characterisations with a value equivalent to the TTC. In the uncertainty runs, NOAELs are sampled from the empirical distribution.

The TTC is a conservative estimate of the NOAEL for at least two reasons:

- 1. TTCs are calculated from a collection of NOELs for the critical (i.e., first occurring) effect within each study and often the effect of interest will not be the critical effect, and therefore higher NOAELs are expected.
- 2. The TTC is a low percentile and therefore a conservative estimate for a random class member with unknown  $NOAFI$

#### **Munro central value**

To avoid the conservatism of taking the 5th percentile in the Munro P5 approach, alternatively, a nominal (or central) value could be taken from the Munro collection for each Cramer class. For a nominal run without uncertainty, the expected contribution of a substance with missing hazard characterisation to the risk as quantified in the hazard index is obtained from

$$
H\!I = SF \cdot \sum_i^n \frac{\exp_i}{HC_i}
$$

Here SF are all combined safety factors. It follows from this equation that an unbiased estimate for the contribution from a substance with missing hazard characterisations is obtained by taking the harmonic mean from the available NOAELs:

$$
\textit{NOAEL} = \left(\sum_{i=1}^{n} \frac{1}{\textit{NOAEL}_i}\right)^{-1}
$$

This is the value to use in a nominal run without uncertainty for the Munro central value approach. For the uncertainty runs, this approach also uses random sampling from the empirical distribution of the corresponding Cramer class.

#### **Available hazard characterisations distribution P5**

Another conservative aspect of the TTC approach is the fact that the Munro set lists NOELs/LOAELs for critical effects, not for the specific effect under study. Therefore an alternative is to use the effect-specific hazard characterisations of the substances for which these are available. This collection will have on average higher NOAELs than those of the Munro NOEL collection, because for many substances, the effect of interest will not be the critical effect.

#### **Available hazard characterisations distribution central value**

Similar to the Munro central value approach, a central value could also be obtained from the set of effect-specific hazard characterisations distribution for imputation of hazard characterisations. This approach may yield the most realistic, or unbiased imputation value for missing hazard characterisations.

#### **Aggregation over multiple available hazard characterisations**

<span id="page-229-0"></span>In some scenarios, it may be that for a given substance and effect there are multiple available hazard characterisations. This can happen, for instance, if there are two different NOAELs originating from different studies. In such cases, a single hazard characterisation should be derived from the available candidates.

A conservative approach is to choose the lowest hazard characterisation (HC) of the available hazard characterisations. I.e.,

$$
\mathrm{HC} = \min_{i=1,\ldots,n} \mathrm{HC}_i
$$

Alternatively, it is possible to aggregate the candidates into a new "average" hazard characterisation. For this, the harmonic mean, also used for obtaining central value estimates in the *imputation of missing hazard characterisations*, is a suitable approach.

$$
\text{HC} = \left(\sum_i^n \frac{1}{\text{HC}_i}\right)^{-1}
$$

# **Hazard characterisations settings**

# **Selection settings**

Table 2.165: Selection settings for module Hazard characterisations.

Name	Description	
Risk type	The type of exposure considered in the assessment; acute (short)	
	term) or chronic (long-term).	
Target level	Select to express hazard characterisations at external or internal	
	exposure level.	
Consider critical effect	Specifies whether the analysis should look at critical effects such	
	as specified in the Hazard characterisation data source.	

# <span id="page-230-0"></span>**Calculation settings**

Name	Description
Method	Choose method for computing the hazard characterisations: from
	in-vivo or in-vitro points of departure or both.
Expression type	Specifies how hazard characterisations are expressed: as BMD, as
	NOAEL, or the expression type is ignored.
Selection method in case of	Choose either the most toxic (default) or an aggregated hazard
multiple candidate hazard	characterisation when in nominal runs there are multiple available
characterisations	candidates. In uncertainty runs, multiple candidates are
	resampled.
Impute missing hazard	If checked, missing hazard characterisations are imputed based on
characterisations	Munro NOELs or on other available points of departure.
Imputation method	Imputation of Hazard characterisations: use low percentile (P5) or
	unbiased central estimate from either the Munro set or the
	available POD collection.
Use BMDs from dose response	If checked, preferably BMDs from dose response models will be
models	used. If these data are not available, other POD data are used.
Use inter-species conversions	Use inter-species conversion factors (default value, e.g. 10, or
	data).
Use intra-species factors	Use intra-species conversion factors (default value, e.g. 10, or
	data).
Use additional assessment	Use additional assessment factor for extrapolation of PODs to
factor	(human) hazard characterisations.
Include dietary and non-dietary	Specifies whether the assessment involves both dietary and
routes of exposure	non-dietary (oral, inhalatory or dermal) routes of exposure.

Table 2.166: Calculation settings for module Hazard characterisations.

# **Uncertainty settings**





### **Hazard characterisations as data**

Hazard characterisations can be provided as data e.g., in the form of ADI or ARfD.

• *Hazard characterisations data formats*

Inputs used: *AOP networks Active substances Points of departure*

### **Calc[ulation of hazard characterisatio](#page-221-0)ns**

Hazard char[acterisations ca](#page-206-0)[n be computed fr](#page-202-0)[om points of depart](#page-234-0)ure. The computation may involve assessment factors, e.g. for inter-species conversion, intra-species variation or additional sources of uncertainty. The additional assessment factor can be used to bypass inter- and intra species conversion, or as an additional assessment factor to account for extrapolation for sources where appropriate data or information is scarce or missing (e.g. to implement a mixture assessment factor). The hazard characterisation calculation may also use kinetic models or absorption factors to convert external doses to internal doses or vice versa.

• *Hazard characterisations calculation*

Inputs used: *Dose response models Effect representations Inter-species conversions Intra species factors Kinetic models*

Settings used

• *[Calculation Settings](#page-223-0)*

# **2.5.7 Inter-species conversions**

Inter-[species conversions s](#page-230-0)pecify how to convert a hazard characterisation for a given species to a hazard characterisation for humans. In the simplest approach, this specifies a fixed inter-species factor. In a higher tier, this specifies a geometric mean (GM) and geometric standard deviation (GSD) for a lognormal uncertainty distribution of the interspecies factor. Inter-species conversion are specified per effect and can be general or substance-specific.

This module has as primary entities: *Substances Effects*

Output of this module is used by: *Hazard characterisations*

### **Inter-species conversions dat[a formats](#page-44-0)**

### **Inter-species conversions**

<span id="page-231-1"></span><span id="page-231-0"></span>Inter-species conversion models specify how to convert a hazard dose for a given species to a hazard dose for humans.

### **Inter-species model parameters**

Inter-species extrapolation factors are described using a lognormal distribution specified by a geometric mean (GM) and geometric standard deviation (GSD). Inter-species factors are defined for an effect and a species and may optionally be specified specifically for a substance.





Table aliases: InterSpeciesModelParameters, InterSpeciesModelParameter, InterSpeciesFactors, InterSpeciesFactor.

# **Inter-species conversions settings**

# <span id="page-232-0"></span>**Selection settings**

Name	Description
Default interspecies factor	Default interspecies factor geometric mean.
geometric mean	
Default interspecies factor	Default interspecies factor geometric standard deviation.
geometric standard deviation	

Table 2.169: Selection settings for module Inter-species conversions.



<span id="page-233-0"></span>

### **Inter-species conversions as data**

Data are provided in the form of a geometric mean (GM) and geometric standard deviation (GSD)

• *Inter-species conversions data formats*

Inputs used: *Active substances*

# **2.5.8 [Intra species factors](#page-231-1)**

Intra-species [factors specify ho](#page-202-0)w to convert a hazard characterisation from the average to a sensitive human individual.

This module has as primary entities: *Substances Effects*

Output of this module is used by: *Hazard characterisations*

### **Intra-species factors data for[mats](#page-44-0)**

### **Intra-species factors**

<span id="page-233-1"></span>Intra-species factors.

### **Intra-species model parameters**

Intra species factors.

Name	Type	Description	Aliases	Required
idEffect	AlphaNumeric(50)	The effect code.	idEffect.	<b>Yes</b>
			EffectId, Effect	
idSubstance	AlphaNumeric(50)	The code of the substance.	idSubstance,	N <sub>o</sub>
			SubstanceId.	
			SubstanceCode,	
			Substance	
IntraSpecies-	Numeric	The lower variability factor.	IntraSpecies-	N <sub>o</sub>
Lower-		The lower and upper factor	LowerVariation-	
<b>VariationFactor</b>		are used to derive a geometric	Factor	
		standard deviation (gsd) and		
		degrees of freedom (df).		
IntraSpecies-	Numeric	The upper variability factor.	IntraSpecies-	Yes
UpperVariation-		The lower and upper factor	UpperVariation-	
Factor		are used to derive a geometric	Factor	
		standard deviation (gsd) and		
		degrees of freedom (df).		
idPopulation	AlphaNumeric(50)	Unique identification code of	IdPopulation,	N <sub>o</sub>
		the population.	PopulationId	

Table 2.171: Table definition for IntraSpeciesModelParameters.

Table aliases: IntraSpeciesModelParameters, IntraSpeciesModelParameter, IntraSpeciesFactors, IntraSpeciesFactor.

### **Intra species factors settings**

### **Selection settings**





#### **Intra species factors as data**

In the simplest approach, intra-species factors are fixed factors. In a higher tier, lower and upper values for the intraspecies factor are used to derive a variability distribution (log-normal around 1) and an uncertainty distribution for the geometric standard deviation related to human variability in sensitivity.

• *Intra species factors data formats*

Inputs used: *Active substances*

# **2.5.9 [Points of departure](#page-233-1)**

<span id="page-234-0"></span>Externally sp[ecified points of d](#page-202-0)eparture can be used as an alternative to calculated BMDs from dose response models. Points of departure can be of various types, such as NOAEL, LOAEL or BMD. They can be used to construct the list of active substances, to derive relative potency factors, and to perform health impact assessments.

This module has as primary entities: *Substances Effects*

Output of this module is used by: *Active substances Hazard characterisations*

### **Points of departure data form[ats](#page-44-0)**

#### **Points of departure**

<span id="page-234-1"></span>Points of departure, such as NOAELS and BMDs, describe the critical/reference levels of substance dose in relation to the presence or absence of an effect. If available, the uncertainty of externally specified points of departure can be specified with uncertainty sets (empirical distributions representing possible values) in the points of departure uncertainty table.

#### **Points of departure**

Nominal points of departure should be presented in this table. Each point of departure should be linked to an effect using the effect code (idEffect) and to substances using the substance code (idSubstance).

Table 2.173: Table definition for HazardDoses.

<b>Name</b>	Type	Description	<b>Aliases</b>	Required
idModel	AlphaNumeric(50)	The dose response model code.	idDose- ResponseModel, idModel, idPod, idPointOf- Departure, Pod, PointOf- Departure	$\overline{No}$
idEffect	AlphaNumeric(50)	The effect code.	idEffect, EffectId, Effect	Yes
idSubstance	AlphaNumeric $(50)$	The code of the substance.	idSubstance, SubstanceId, SubstanceCode, Substance	Yes
<b>Species</b>	AlphaNumeric(50)	The species used to obtain this point of departure.	Species, System	$\overline{No}$
Point of departure	Numeric	Point of departure, can be of various types, e.g. NOAEL, LOAEL, BMD, CED	PointOf- Departure, LimitDose, HazardDose, Value, CED	Yes
Point of departure type	HazardDoseTypes	The type of the point of departure: e.g. NOAEL, LOAEL, BMD (default).	PODType, HazardDose- Type, LimitDoseType	N <sub>0</sub>
DoseUnit	AlphaNumeric $(50)$	The dose unit (if not specified, then mg/kg is assumed).	DoseUnit, UnitDose	$\overline{No}$
<b>Benchmark</b> response (BMR)	AlphaNumeric(100)	The effect size.	Benchmark- Response, CriticalEffect- Size, HazardEffect- Size	$\overline{No}$
<b>ExposureRoute</b>	AlphaNumeric(100)	The route of dose administration used in the study to obtain this point of departure. If not specified $exposure route = Dietary is$ assumed.	ExposureRoute, RouteExposure	$\overline{No}$
<b>IsCriticalEffect</b>	Boolean	Specifies whether this value is the value associated with the critical effect. If omitted, No is assumed	<b>IsCriticalEffect</b>	N <sub>o</sub>

Table aliases: PointsOfDeparture, PointOfDeparture, HazardDoses, HazardDose.

# **Points of departure uncertainty**

Often, the PODs found for a substance/effect combination are uncertain. This table facilitates in specifying the POD uncertainty in the form of a set of uncertainty values that may additionally be specified for a substance/effect combination.





Table aliases: PointsOfDepartureUncertain, PointOfDepartureUncertain, HazardDosesUncertain, HazardDoseUncertain.

# **Points of departure settings**

# **Uncertainty settings**

Name	Description
Resample hazard	Specifies whether to resample the hazard characterisations or
characterisations or RPFs	relative potency factors. Requires hazard characterisation or RPF
	uncertainty to be quantified in DoseResponseModelsUncertain or
	RelativePotencyFactorsUncertain tables.

Table 2.175: Uncertainty settings for module Points of departure.

### **Points of departure as data**

Points of departure are provided as data for combinations of substance and effect and each is minimally described by a reference value and a type (e.g., NOAEL or LOAEL). In addition, the exposure route, specifies, and references may be specified.

• *Points of departure data formats*

Inputs used: *AOP networks*

# **2.5.10 [Relative potency fac](#page-234-1)tors**

Relative pot[ency factors \(R](#page-206-0)PFs) quantify potencies of substances with respect to a defined effect, relative to the potency of a chosen index substance. RPFs can be used to express combined exposures of multiple substances in terms of a the exposure value of the chosen index substance (i.e., in index substance equivalents). In MCRA, hazard characterisations, and therefore also RPFs are based on mass units (e.g., µg), and not on mol units. RPFs can be different for different levels of the human organism (external, internal, specific compartment). RPFs can be given as data or computed from hazard characterisations. RPFs can be specified with uncertainty. Computation from uncertain hazard characterisations allows to include correlations between uncertain RPFs which originate from using the same index substance.

This module has as primary entities: *Substances Effects*

Output of this module is used by: *Concentrations Concentration models High exposure food-substance combinations Dietary exposures Exposures*

# **Relative potency factors dat[a formats](#page-76-0)**

### **[Relative poten](#page-116-0)[cy factor](#page-150-0)s**

<span id="page-237-0"></span>Relative potency factors quantify relative potencies of substances with respect to an effect and can be used to express combined exposures of multiple substances in terms of the exposure value of the chosen index substance (i.e., as index substance equivalents). Relative potency factors can be provided in case hazard characterisations are missing. If available, the uncertainty of externally specified RPFs can be specified with uncertainty sets (empirical distributions representing possible values) in an additional table.

### **Relative potency factors**

Relative potency factors are linked to an effect using the effect code (idEffect) and to substances using the substance code (idSubstance).





Table aliases: RelativePotencyFactors, RelativePotencyFactor.

### **Relative potency factor uncertainty**

This table contains sets of values representing the uncertainty for relative potency factors.

Name	Type	Description	Aliases	Required
idUncertainty-	AlphaNumeric $(50)$	The uncertainty set	idUncertainty-	<b>Yes</b>
<b>Set</b>		identification number. During	Set.	
		each uncertainty iteration one	UncertaintyId	
		set is used.		
idEffect	AlphaNumeric $(50)$	The effect code (must)	idEffect,	<b>Yes</b>
		correspond to values in id	EffectId, Effect	
		column of Effects table).		
idSubstance	AlphaNumeric $(50)$	The substance code (must)	idSubstance.	Yes
		correspond to values in id	SubstanceId.	
		column of Substances table).	SubstanceCode.	
			Substance	
<b>RPF</b>	Numeric	The relative potency factor.	RPF, Relative-	Yes
			PotencyFactor	

Table 2.177: Table definition for RelativePotencyFactorsUncertain.

Table aliases: RelativePotencyFactorsUncertain, RelativePotencyFactorUncertain.

### **Relative potency factors calculation**

<span id="page-238-0"></span>Relative potency factors (RPFs) describe the potency of substances with respect to a defined effect, relative to the potency of a chosen index substance. RPFs can be given as data or computed from *hazard characterisations*. The RPF for substance i is defined by the ratio of hazard characterisation value for the index substance ( $r \in f$ ) and the hazard characterisation value for substance  $i$ . That is,

$$
\mathsf{RPF}_i = \mathtt{POD}_\mathtt{ref} / \mathtt{POD}_i.
$$

When the hazard characterisations are resampled in the uncertainty runs, RPFs are also recomputed based on the bootstrapped hazard characterisations. In this way, RPF uncertainty can also included in the uncertainty analysis.

# **Relative potency factors settings**

### **Calculation settings**

Table 2.178: Calculation settings for module Relative potency factors.

<span id="page-238-1"></span>

Name	Description
Index substance	The substance of interest or index substance.

# **Uncertainty settings**





### **Relative potency factors as data**

Data are provided in the form of a RPF for a specific substance and effect.

• *Relative potency factors data formats*

Inputs used: *Active substances AOP networks*

### **Calc[ulation of relative potency facto](#page-237-0)rs**

RPFs are co[mputed from haza](#page-202-0)[rd characterisa](#page-206-0)tions.

• *Relative potency factors calculation*

Inputs used: *Hazard characterisations*

Settings used

• *[Calculation Settings](#page-238-0)*

# **2.6 [In-silico mo](#page-238-1)dules**

Two types of in-silico models are available: QSAR models specify assessment group memberships for active substances, as numbers in the interval [0,1]. This allows both crisp (0 or 1) and probabilistic memberships. Molecular docking models specify binding energies and thresholds which can be used to convert binding energies to assessment group memberships for active substances.

# **2.6.1 Molecular docking models**

Molecular docking models specify binding energies for substances in specific molecular docking models related to a specific health effect (adverse outcome).

This module has as primary entities: *Substances Effects*

Output of this module is used by: *Active substances*

# **Molecular docking models dat[a formats](#page-44-0)**

### <span id="page-239-0"></span>**Required data tables:**

- Molecular docking models, to identify models for a specified effect (receptor)
- Molecular docking binding energies, to specify the binding energies per substance for the receptor

# **Molecular docking models**

Contains definitions of molecular docking models for a given effect (molecular initiating event), for example parameters needed in the conversion of binding energies to group memberships or to relative potency factors. Substance specific binding energies are specified in the binding energies table.

# **Molecular docking models**

Each docking model has a unique identifier, and optionally a name and a description. Each model is linked to an effect using the idEffect field and optionally a binding threshold and the number of receptors can be added. A reference to the source of the data can be stored in the reference field.





Table aliases: MolecularDockingModels, MolecularDockingModel, BindingEnergyModels, BindingEnergyModel.

# **Molecular docking binding energies**

Molecular docking model binding energies per substance

Name	Type	Description	<b>Aliases</b>	Required
idMolecular-	AlphaNumeric(50)	The id of the molecular	idMolecular-	N <sub>0</sub>
DockingModel		docking model or source.	Docking,	
			Molecular-	
			DockingModel	
idSubstance	AlphaNumeric $(50)$	The code of the substance.	idSubstance,	Yes
			SubstanceId,	
			SubstanceCode,	
			Substance	
BindingEnergy	Numeric	Molecular Docking binding	Molecular-	<b>Yes</b>
		energy.	Docking-	
			BindingEnergy	

Table 2.181: Table definition for MolecularBindingEnergies.

Table aliases: MolecularBindingEnergies, MolecularBindingEnergy, BindingEnergies, BindingEnergy, MolecularDockingBindingEnergies, MolecularDockingBindingEnergy.

### **Molecular docking models as data**

Binding energies for substances in specific molecular docking models related to a specific health effect (adverse outcome) are provided as data.

• *Molecular docking models data formats*

Inputs used: *AOP networks*

# **2.6.2 [QSAR membership model](#page-239-0)s**

QSAR mem[bership models](#page-206-0) specify assessment group memberships for active substances related to a specific health effect (adverse outcome). Memberships should be derived externally from Quantitative Structure-Activity Relationship (QSAR) models.

This module has as primary entities: *Substances Effects*

Output of this module is used by: *Active substances*

# **QSAR membership models da[ta format](#page-44-0)[s](#page-25-0)**

Required data tables:

- <span id="page-241-0"></span>• QSAR membership models, to identify QSAR models for a specified health effect
- QSAR membership scores, to specify the memberships per substance per QSAR model

Note that only memberships 1 (include) and 0 (exclude) are allowed.

# **QSAR membership models**

Substance membership models obtained from QSAR for a given (health) effect. The models are defined in the membership models table, and substance specific memberships are specified in the QSAR memberships table.

# **QSAR membership models**

This table contains the definitions of the QSAR membership models. Each model contains a id, name, an optional description, and refers to its related health effect.





Table aliases: QSAR, QSARMembershipModels, QSARMembershipModel, QSARModels, QSARModel.

Substance membership score according to the QSAR model.

Name	Type	Description	<b>Aliases</b>	Required
idQSAR-	AlphaNumeric(50)	The id of the QSAR model.	Model,	Yes
Membership-			ModelCode,	
Model			idModel,	
			QSARModel,	
			idQSARModel,	
			<b>QSAR-</b>	
			Membership-	
			Model,	
			idQSAR-	
			Membership-	
			Model,	
			Membership-	
			Model,	
			idMembership-	
			Model	
idSubstance	AlphaNumeric(50)	The code of the substance.	idSubstance,	Yes
			SubstanceId,	
			SubstanceCode,	
			Substance	
Membership-	Numeric	QSAR membership score.	Membership-	Yes
Score		Value should be 1 for positive	Score,	
		membership, or 0 for negative	Membership,	
		membership.	QSARScore,	
			Score	

Table 2.183: Table definition for QSARMembershipScores.

Table aliases: QSARMembershipScores, QSARMembershipScore, QSARMemberships, QSARMembership.

# **QSAR membership models as data**

QSAR memberships models are provided as data, per QSAR model assessment group memberships for active substances related to a specific health effect are specified.

• *QSAR membership models data formats*

Inputs used: *AOP networks*

# **2.7 [Kinetic modules](#page-241-0)**

Kinetic models convert exposures or hazard characterisations from one or more external routes or compartments to an internal (target) compartment. The reverse conversion from internal to external can also be made (reverse dosimetry).

In a simple tier, kinetic models are specified as absorption factors. In a higher tier, physiologically based toxicokinetic (PBTK) models of a specified type (currently available is the EuroMix generic PBTK model) are linked to MCRA. Kinetic model instances for specific substances and test systems (e.g. cypermethrin in the rat) are specified with parameter sets for the chosen kinetic model.

# **2.7.1 Kinetic models**

External exposure can be from on more more exposure routes: oral (dietary or non-dietary), dermal or inhalation. Internal exposure can be systemic or related to a specific compartment in a kinetic model. There are four tiers for relating external to internal exposures (doses):

- <span id="page-244-0"></span>1. Assume 100% absorption: internal exposures are equal to external exposures.
- 2. Assume conservative absorption factors as suggested by EFSA ([EFSA, 2014], [EFSA, 2017]): oral and inhalation 100%, dermal 50%.
- 3. Use externally provided absorption factors (*absorption factors data tables*).
- 4. Use one of the *implemented kinetic models*, with instances for sp[ecific substanc](#page-355-0)es [defined in da](#page-355-1)ta table *kinetic model instances* and model parameters specified in data table *kinetic model instance parameters*.

Given a chosen tier, the calculation will fall backt[o the next lower tier in case of](#page-245-1) missing data.

This module has as pr[imary entities:](#page-246-0) *Substances*

Outpu[t of this module](#page-244-1) is used by: *Exposures Hazard characterisatio[ns](#page-245-2)*

# **Kinetic models data formats**

### <span id="page-244-2"></span>**Data tables:**

- Absorption factors
- Kinetic model instances
- Kinetic model instance parameters

### **Kinetic models**

Kinetic models may be specified as kinetic model instances that contain parameter specifications of built in kinetic models or as simple absorption factors.

### **Kinetic model instances**

<span id="page-244-1"></span>Kinetic model instances.

Name	Type	Description	Aliases	Required
idModel-	AlphaNumeric $(50)$	Unique identification code of	idModel-	Yes
Instance		the kinetic model instance.	Instance, Id.	
			Code	
idModel-	KineticModelType	Identifier of the kinetic model	idModel-	Yes
Definition		definition for which this is an	Definition,	
		instance.	ModelDefinition	
idTestSystem	AlphaNumeric(200)	The species on which the	System,	Yes
		experiment was performed.	<b>TestSystem</b>	
idSubstance	AlphaNumeric $(50)$	Unique identification code of	idSubstance.	N <sub>0</sub>
		substance, Default: valid for	SubstanceId,	
		all substances. Should be	SubstanceCode.	
		omitted for parameters in the	Substance	
		class Physiological		
Reference	AlphaNumeric(100)	Reference or author.	References	N <sub>0</sub>

Table 2.184: Table definition for KineticModelInstances.

Table aliases: KineticModelInstances, KineticModelInstance.

# <span id="page-245-2"></span>**Kinetic model instance parameters**

Kinetic model parameters

Name	Type	Description	Aliases	Required
idModel-	AlphaNumeric $(50)$	Unique identification code of	Id, Code	<b>Yes</b>
Instance		the kinetic model instance to		
		which this parameter belongs		
Parameter	AlphaNumeric(100)	Name of the parameter in the		<b>Yes</b>
		kinetic model.		
Description	AlphaNumeric	Description of or reference		No
		for the parameter values in		
		the kinetic model.		
Value	Numeric	Mean.	MEAN, mean	Yes
Distribution-	AlphaNumeric(20)	Distribution.	Distribution-	N <sub>o</sub>
Type			Type,	
			<b>Distribution</b>	
CvVariability	Numeric	Variability.		N <sub>o</sub>
CvUncertainty	Numeric	Uncertainty.		N <sub>o</sub>

Table 2.185: Table definition for KineticModelInstanceParameters.

Table aliases: KineticModelInstanceParameters, KineticModelInstanceParameter.

# <span id="page-245-1"></span>**Kinetic model absorption factors**

Kinetic absorption factors



# Table 2.186: Table definition for KineticAbsorptionFactors.

Table aliases: KineticAbsorptionFactors, KineticAbsorptionFactor, AbsorptionFactors, AbsorptionFactor.

# <span id="page-245-0"></span>**Kinetic models settings**

Name	Description		
Default oral absorption factor	When there is no kinetic model and absorption factors are not		
for non-dietary exposure	specified in file, non-dietary oral exposures (external doses) are		
	multiplied by this factor to determine the absorbed (internal) dose.		
Default oral absorption factor	When there is no kinetic model and absorption factors are not		
for dietary exposure	specified in file, dietary exposures (external doses) are multiplied		
	by this factor to determine the absorbed (internal) dose.		
Default dermal absorption	When there is no kinetic model and absorption factors are not		
factor for non-dietary exposure	specified in file, dermal exposures (external doses) are multiplied		
	by this factor to determine the absorbed (internal) dose.		
Default inhalation absorption	When there is no kinetic model and absorption factors are not		
factor for non-dietary exposure	specified in file, inhalation exposures (external doses) are		
	multiplied by this factor to determine the absorbed (internal) dose.		
Number of days	The number of days.		
Number of events per day for	The daily dose is administered in equal portions $(dose / number of$		
the ORAL dietary dose	events) per event.		
Number of initial days skipped	This period is skipped in the calculation of the mean internal		
	exposure.		
Kinetic model	Code Kinetic Model.		
Use parameter variability	When specified, use parameter variability.		

Table 2.187: Calculation settings for module Kinetic models.

# **Uncertainty settings**





# **Kinetic models as data**

Specify nondietary absorption factors as data.

• *Kinetic models data formats*

Inputs used: *Active substances*

### **Avail[able kinetic models](#page-244-2)**

<span id="page-246-0"></span>Physiologica[lly based toxicoki](#page-202-0)netic (PBTK) models, or kinetic models for short, are mathematical representations of the animal or human body aimed at describing and predicting the time course distribution of chemicals in tissues and organs. Those internal dose metrics can usefully replace external exposure dose in the derivation of the quantitative dose-response relationships and following risk assessments. PBTK models can simulate both internal doses from exposure scenarios (forward dosimetry) and external dose from biomonitoring data (reverse dosimetry).

The following generic PBTK models are currently implemented in MCRA:

- *EuroMix generic PBTK model* [Cleo et al., 2019].
- *bisphenol model ETHZ* [Karrer et al., 2019].

# **EuroMix Generic PBTK model v6**

Cosmos version 6 (received 3/27/2019)

Table 2.189: Exposure routes (forcings)

	Description	Unit	Order
Dietary	Dietary exposure	mmoles	
Dermal	Dermal exposure	mmoles	
Inhalation	Inhalatory exposure	mmoles	



# Table 2.190: Output

ld	Description	Unit	Type	Order	
BM	<b>Body</b> mass	kg	Physiological	$\overline{0}$	
<b>BSA</b>	Body surface area (internally scaled by	dm2	Physiological	$\mathbf{1}$	
	an allometric scaling				
	factor $s =$				
	70/BM^0.3)				
scVFat	Fat as fraction of		Physiological	$\overline{2}$	
	total body volume				
scVRich	<b>Richly perfused</b>		Physiological	$\overline{3}$	
	tissues (viscera) as				
	fraction of total				
	body volume				
scVLiver	Liver as fraction of		Physiological	$\overline{4}$	
	total body volume				
scVBlood	Blood as fraction of		Physiological	$\overline{5}$	
	total body volume				
Height_sc	Skin thickness	decimeter	Physiological	6	
Height_vs	Viable skin		Physiological	$\overline{7}$	
scFBlood	Total blood flow per	L/h/kg	Physiological	$\overline{8}$	
	unit mass				
scFFat	Fat fraction of total		Physiological	$\overline{9}$	
	blood flow going to				
	compartments				
scFPoor	Poorly perfused		Physiological	10	
	tissues (muscles) fraction of total				
	blood flow going to compartments				
scFLiver	Liver fraction of		Physiological	11	
	total blood flow				
	going to				
	compartments				
scFSkin	Skin fraction of total		Physiological	12	
	blood flow going to				
	compartments				
Falv	Alveolar ventilation	L/h	Physiological	13	
	rate				
mic	Microsomal proteins	mg/gr liver	Physiological	14	
	content				
$Kp\_sc\_vs$	Diffusion rate from	decimeter/h	Metabolic	$\overline{22}$	
	stratum corneum to				
	viable skin				
Ke	Renal excretion rate	L/h	Metabolic	$\overline{23}$	
Michaelis	Flag for		Metabolic	$\overline{24}$	
	Michaelis-Menten vs				
	linear metabolism (0				
	$=$ linear)				
Vmax	Maximum rate of	mmoles/h/L liver	Metabolic	$\overline{25}$	
	metabolism				
$\overline{\text{Km}}$	Michaelis-Menten	mM	Metabolic	$\overline{26}$	
	constant for				
	metabolism				
CLH	Hepatic metabolic		Metabolic	$\overline{27}$	
	clearance				
fub	Unbound fraction in		Metabolic	$\overline{28}$	
	blood				
2.7. Kiriatic modules	Fraction absorbed by		Metabolic	$\overline{29}$	241
	the gut				
$\overline{\text{kGut}}$	Oral 1st order	1/h	Metabolic	$\overline{30}$	
	absorption rate constant				

Table 2.191: Input

Model aliases: Cosmos6, CosmosV6.

# **EuroMix Generic PBTK model v5**

Cosmos version 5 (adapted 9/11/2018)









$\overline{\mathsf{Id}}$	Description	Unit	Type	Order
BM	<b>Body</b> mass	kg	Physiological	$\boldsymbol{0}$
<b>BSA</b>	Body skin surface area	dm2	Physiological	$\mathbf{1}$
scVFat	Fat as fraction of total body volume		Physiological	$\overline{2}$
scVRich	<b>Richly perfused</b> tissues (viscera) as		Physiological	$\overline{3}$
	fraction of total body volume			
scVLiver	Liver as fraction of total body volume		Physiological	$\overline{4}$
scVBlood	Blood as fraction of total body volume		Physiological	5
Height_sc	Skin thickness	decimeter	Physiological	6
Height_vs	Viable skin		Physiological	7
scFBlood	Total blood flow per unit mass	L/h/kg	Physiological	$\overline{8}$
scFFat	Fat fraction of total blood flow going to compartments		Physiological	$\overline{9}$
scFPoor	Poorly perfused tissues (muscles) fraction of total blood flow going to		Physiological	10
scFLiver	compartments Liver fraction of total blood flow going to compartments		Physiological	$\overline{11}$
scFSkin	Skin fraction of total blood flow going to compartments		Physiological	$\overline{12}$
Falv	Alveolar ventilation rate	L/h	Physiological	13
mic	Microsomal proteins content	$mg/gr$ liver	Physiological	$\overline{14}$
PCAir	Partition coefficient: blood over air		Partition coefficient	15
Kp_sc_vs	Diffusion rate from stratum corneum to viable skin	decimeter/h	Metabolic	$\overline{22}$
Ke	Renal excretion rate	L/h	Metabolic	$\overline{23}$
Michaelis	Flag for Michaelis-Menten vs linear metabolism (0 $=$ linear)		Metabolic	$\overline{24}$
<b>V</b> max	Maximum rate of metabolism	mmoles/h/L liver	Metabolic	25
Km	Michaelis-Menten constant	mM	Metabolic	$\overline{26}$
CLH	Hepatic clearance		Metabolic	$\overline{27}$
fup	<b>Unbound</b> fraction in blood		Metabolic	28
Frac	Fraction absorbed by the gut		Metabolic	$\overline{29}$
$\overline{kGut}$	Oral 1st order	1/h	Metabolic	$\overline{30}$
Kinetic modules	absorption rate constant			
fSA_exposed	Fraction of skin surface area actually $r$ nosod		Metabolic	$\overline{35}$

Table 2.194: Input

Model aliases: Cosmos4, CosmosV4, Cosmos5, CosmosV5.

# **Generic Model BPA**

Generic model Cecile Karrer 23 juli 2018









### Table 2.197: Input


Id	Description	shimada nom providas pago Unit	Type	Order
<b>VbrainC</b>	Fractional volume of		Physiological	11
	brain tissue			
<b>VskinC</b>	Fractional volume of		Physiological	$\overline{12}$
	skin tissue			
VgonadC	Fractional volume of		Physiological	$\overline{13}$
	gonads			
VmuscleC	Fractional volume of		Physiological	$\overline{14}$
	muscle tissue			
VrichC	Fractional volume of		Physiological	$\overline{15}$
	richly perfused			
	tissue			
VbodygC	Distribution volume		Physiological	16
	of BPA-g			
<b>MW</b>	Molecular weight	g/mol	Chemical property	18
pliver	Partition coefficient		Partition coefficient	19
	liver to blood			
pfat	Partition coefficient		Partition coefficient	$\overline{20}$
	fat to blood			
pslow	Partition coefficient		Partition coefficient	$\overline{21}$
	slowly perfused			
	tissue to blood			
prich	Partition coefficient		Partition coefficient	$\overline{22}$
	richly perfused			
	tissue to blood			
pgonad	Partition coefficient		Partition coefficient	$\overline{23}$
	gonads to blood Partition coefficient		Partition coefficient	24
pbrain				
	brain to blood Partition coefficient		Partition coefficient	$\overline{25}$
pskin	skin to blood			
			Metabolic	
geC k0C	Gastric emptying Oral uptake from the	$1/h/kg$ bw^-0.25 $1/h/kg$ bw^-0.25	Metabolic	26 $\overline{27}$
	stomach into the			
	liver			
k1C	Oral uptake from the	$1/h/kg$ bw^-0.25	Metabolic	28
	small intestine into			
k4C	the liver Fecal elimination	$1/h/kg$ bw^-0.25	Metabolic	29
	from small intestine			
	after oral			
	administration			
			Metabolic	
kGIingC	<b>Transport</b> of	$1/h/kg$ bw^-0.25		$\overline{30}$
	glucuronide from			
	enterocytes into			
	serum			
kGIinsC	Transport of sulfate	$1/h/kg$ bw^-0.25	Metabolic	$\overline{31}$
	from enterocytes			
	into serum			$\overline{32}$
kmgutg	Km of Glucuronidation in	nM	Metabolic	
	the gut Vmax of		Metabolic	$\overline{33}$
vmaxgutgC	Glucuronidation in	nmol/h/kg bw		
	the gut			

Table 2.197 – continued from previous page

ld	Description	Unit	Type	Order
fgutg	Correction factor of		Metabolic	34
	glucuronidation in			
	the gut			
kmguts	Km of Sulfation in	nM	Metabolic	$\overline{35}$
	the gut			
vmaxgutsC	Vmax of Sulfation in	nmol/h/kg bw	Metabolic	$\overline{36}$
	the gut			
fguts	Correction factor of		Metabolic	$\overline{37}$
	sulfation in the gut			
met1g	Fraction of		Metabolic	$\overline{38}$
	glucuronide in the			
	liver taken up			
	directly into serum			
	(the rest undergoes			
	EHR)			
met1s	Fraction of sulfate in		Metabolic	$\overline{39}$
	the liver taken up			
	directly into serum			
enterocytes	Sum of enterocytes	L	Metabolic	40
	weights in			
	duodenum, jejunum			
	and ileum			
kmliver	Km of	nM	Metabolic	41
	Glucuronidation in			
	the liver			
vmaxliverC	Vmax of	nmol/h/g liver	Metabolic	42
	Glucuronidation in			
	the liver			
fliverg	Correction factor of		Metabolic	43
	glucuronidation in			
	the liver			
kmlivers	Km of Sulfation in	nM	Metabolic	44
	the liver			
vmaxliversC	Vmax of Sulfation in	nmol/h/g liver	Metabolic	$\overline{45}$
	the liver			
flivers	Correction factor of		Metabolic	46
	sulfation in the liver			
<b>EHRtime</b>	Time until EHR	h	Metabolic	47
	occurs			
<b>EHRrateC</b>	EHR of glucuronide	$1/h/kg$ bw^-0.25	Metabolic	$\overline{48}$
k4C_IV	Fecal elimination of	$1/h/kg$ bw^-0.25	Metabolic	49
	glucuronide from the			
	EHR compartment			
kurinebpaC	Clearance, urine	L/h/kg bw $\sqrt{0.75}$	Metabolic	50
	excretion of parent			
	compound			
kurinebpagC	Clearance, urine	L/h/kg bw $^{\wedge}0.75$	Metabolic	$\overline{51}$
	excretion of			
	glucuronide			
kurinebpasC	Clearance, urine	L/h/kg bw $\sqrt{0.75}$	Metabolic	52
	excretion of sulfate			
vreabsorptiong-	Vmax for renal	nmol/h/kg bw^0.75	Metabolic	$\overline{53}$
C	reabsorption of			
	glucuronide			

Table 2.197 – continued from previous page





Model aliases: PBPKModel\_BPA, PBPKModelBPA, ModelBPA, BPA.

**Note:** Additional kinetic models can be implemented, please contact the MCRA administrator.

### **EuroMix generic PBTK model**

Reference: Tebby et al, 2019: [Cleo et al., 2019]

In MCRA updated versions (version 4b, 6) of the PBTK model developed at INERIS in the framework of the COS-MOS project is used. The model describes the distribution of chemicals in venous blood, arterial blood, adipose tissues, poorly perfused tissues (muscles), gut lumen, liver, richly perfused tissues (other viscera), and skin. Each of those is described as a compart[ment \(homogeneo](#page-354-0)us virtual volume) in which distribution is instantaneous and limited only by the incoming blood flow or rate of entry in the compartment. Exposure can occur through the dermal route, ingestion or inhalation. The absorbed molecules can be excreted to urine, exhaled through the lung, or metabolized in liver.



Figure 2.40: Schematic representation of the EuroMix Generic PBTK model.

The EuroMix generic PBTK model is coded as a set of ordinary differential equations. There is one such equation per time-dependent chemical quantity of the model (so-called state variables). There are 13 state variables in the model: the quantity of chemical in venous blood ( $Q_{ven}$ ), in arterial blood ( $Q_{art}$ ), in adipose tissues ( $Q_{fat}$ ), in poorly perfused tissues  $(Q_p)$ , in well perfused tissues  $(Q_r)$ , in liver  $(Q_{liv})$ , in unexposed skin  $(Q_{s,u})$ , in exposed skin  $(Q_{s,e})$ , in the stratum corneum of unexposed skin  $(Q_{sc,u})$ , in exposed stratum corneum  $(Q_{sc,e})$ , in gut lumen  $(Q_{gut})$ , the quantity excreted to urine ( $Q_{ex}$ ), and the quantity metabolized ( $Q_{met}$ ). The model can predict, as a function of time, for given oral, dermal and/or inhalation exposures, all the above quantities and the corresponding concentrations as a function of time. Concentrations are obtained by dividing quantities by compartment volumes (cited: Bois, Tebby & Brochot).

In Figure 2.41 a time course of the internal substance amount  $(\mu g)$  for Clothianidin in the liver is shown. For 50 consecutive days a bolus per day is submitted. The red line shows the substance amount varying over time. The green line displays the average of the peaks representing acute exposure, the blue line displays the steady state representing chronic exposure, all after skipping a nonstationary period of 10 days (the vertical black line).

From the substance amount, a concentration is computed by dividing it by the total compartment weight (i.e., the mass/volume of the compartment/organ).



**Model CosmosV6**

Figure 2.41: Time course of exposure ( $\mu q$ ) for Clothianidin in the liver (EuroMix generic PBTK model version 6).

In Figure 2.42, for a large number of individuals the internal exposure (acute, green dots) in the liver is plotted versus the external exposure ( $\mu g/kgbw$ ). The diagonal represents the 1:1 ratio of internal vs external exposure.

#### **Bi[sphenol m](#page-257-0)odel**

Reference: Karrer et al. 2019: [Karrer et al., 2019]

'Structural analogs such as the bisphenols S, F, and AF (BPS, BPF, BPAF) are used to replace the endocrine disrupting chemical bisphenol A (BPA), but they exert estrogenic effects in the same order of magnitude. In order to investigate the consequences of BPA restrictions, we assessed the cumulative risk from BPA, BPS, BPF, and BPAF in Europe before and after the first [BPA restrictions in](#page-355-0) 2011. We modelled external exposures from food, personal care products (PCPs), thermal paper, and dust, using the models MCRA and PACEM for food and PCPs, respectively. We calculated internal concentrations of unconjugated BPs with substance-specific PBPK models and cumulated concentrations by taking into account relative estrogenic potencies. Average cumulative exposure to unconjugated BPs was 3.8 and 2.1 ng/kg bw/day before and after restrictions, respectively. The decline was mostly caused by the replacement of BPA with BPS in thermal paper. Therefore, the margins of exposure (MOEs) for estrogenic effects were mostly higher after the restrictions. However, in high uncertainty percentiles the MOEs were partly lower than before (e.g. the MOEs for the uncertainty P97.5 of the variability P99 were 2.6 and 1.9 before and after restrictions, respectively), which shows the higher uncertainty around exposures for substitutes compared to BPA.'

Abstract: Linking probabilistic exposure and pharmacokinetic modelling to assess the cumulative risk from the bisphenols BPA, BPS, BPF, and BPAF for Europeans. Authors: Cecile Karrer, Waldo de Boer, Christiaan Delmaar, Yaping Cai, Amélie Crépet, Konrad Hungerbühler, Natalie von Goetz

<span id="page-257-0"></span>

Figure 2.42: Internal versus external exposure for Clothianidin in the liver (EuroMix Generic PBTK model version 6).



Figure 2.43: Graphical abstract 'Linking probabilistic exposure and pharmacokinetic modelling to assess the cumulative risk from the bisphenols BPA, BPS, BPF, and BPAF for Europeans.'

# **2.8 Risk modules**

*Exposures* and *hazard characterisations* are compared in risk metrics. If both exposure and hazard characterisation are characterised by a single value, the risk metric (e.g. a traditional margin of exposure, hazard quotient or hazard index) can be calculated using module Single value risks. Module Risks allows for probabilistic risk calculations. In both cases a threshold can be specified to assist in interpretation. The threshold value should be chosen in relation to [the assessm](#page-150-0)ent [factors used in the hazar](#page-221-0)d characterisation, e.g. a threshold MOE=100 is often used if no assessment factors have been used, but a threshold 1 would be appropriate if assessment factors have already been used to address relevant uncertainties.

# **2.8.1 Risks**

<span id="page-259-1"></span>Risks (health impacts) are defined as a function of exposure and hazard characterisation at a chosen biological level (external or internal). Risk metrics are margins of exposure (MOE) or hazard indices (HI) or more generalised MOE or HI distributions.

This module has as primary entities: *Substances Effects Populations*

Output of this module is used by: *Single value risks*

# **Risks calculation**

<span id="page-259-0"></span>Probabilistic risk is calculated as [a distribution of](#page-263-0) either margin of exposure (MOE) or hazard index (HI), if at least one of the inputs exposure and hazard characterisation is a distribution. The risk distibution is characterised by percentiles. To accommodate for matching results of MOE and HI=1/MOE in the case of percentiles, there is an option to calculate percentiles via the complementary percentile of the inverse distribution in order to handle numerical differences when calculating percentiles for a left or right tail. E.g. p1 of the MOE distribution can optionally be calculated as 1 divided by p99 of the corresponding 1/MOE distribution. For more details about the graphical displays and the calculations see Individual risks.

# **Individual risks**

A (cumulative) risk assessment aims to characterise the health impact due to one or multiple substances present in food causing one or more health effects. The health impact is characterized by a distribution of individual risks: exposures and hazard characterizations are compared at the chosen level (external or internal) via margins of exposure (MOE) or hazard indices (HI = 1/MOE). Hazard characterisations are included as single values or in a probabilistic way.

The aim is to specify the probability that a random individual from a defined (sub)population will have an exposure high enough to cause a particular health effect of a predefined magnitude, the critical effect size. The exposure level that results in exactly that critical effect in a particular person is that person's individual critical hazard dose. Individuals in a population typically show variation, both in their individual exposure and in their hazard characterization. Both the variation in exposure and the variation in hazard characterization are quantified in the form of probability distributions. Assuming independence between both distributions, they are combined by Monte Carlo methods.

The proportion of the MOE distribution below unity is the probability of critical exposure (*PoCE*) in the particular (sub)population. Uncertainties involved in the overall risk assessment (i.e., both regarding exposure and effect assessment) are quantified using Monte Carlo and bootstrap methods. This results in an uncertainty distribution for any statistic of interest, such as the probability of critical exposure (*PoCE*).

In Figure 2.44, margin of exposures for a number of substances are shown. As shown, the distinction between variability (grey bars, 90% probability) and uncertainty (whiskers) is retained. This is discussed in [van der Voet et al., 2007] and [van der Voet et al., 2009].

In [Figure 2.45](#page-260-0), hazards versus exposures are plotted for the same substances.

<span id="page-260-0"></span>

Figure 2.44: Individual margin of exposure (MOE) plot for multiple substances.

### **Risks settings**

# <span id="page-260-1"></span>**Calculation settings**







Figure 2.45: Hazard vs. exposure plot for multiple substances. 95% bivariate confidence areas for target hazard dose distribution and exposure distribution. Inner ellipses express variability, outer ellipses uncertainty.

# **Output settings**

<b>Name</b>	Description
Number of plot labels	Maximum number of labels to plot in hazard vs exposure plot.
Number of substances in	Maximum number of substances to plot in hazard vs exposure
hazard vs. exposure plot	plot.
Left margin safety plot	Left margin of the plot for margins of exposure or hazard indices.
Right margin safety plot	Right margin of the plot for margins of exposure of hazard indices.
Inclusion percentage variability interval	The central percentage of the variability distribution to include in intervals for exposure, hazard and MOE (e.g. 90 means p5-p95).
Include drill-down on 9	Specifies whether drilldown on 9 individuals is to be included in
individuals around specified percentile.	the output.
Summarize simulated data	Specifies whether a summary of the simulated consumptions and
	concentrations should be included in the output.
Store simulated individual day exposures	Store the simulated individual day exposures. If unchecked, no additional output will be generated. If checked, the output will contain an additional section with the simulated individual day exposures.
Show percentiles for	Give specific percentiles of exposure distribution $(\%)$ , e.g. 50 90 95 97.5 99 (space separated).
Percentage for drilldown	Gives detailed output for nine individuals near this percentile of the exposure distribution.
Percentage for upper tail	Gives detailed output for this upper percentage of the exposure distribution.
Show % of population below	Exposure levels can be generated automatically or by explicit
level(s)	specification (Manual).
<b>Exposure levels</b>	Specify exposure levels for which to give the percentage of exposure below these levels, e.g. 1 10 50 100 200 500. Specify below whether these levels are absolute or relative to ARfD/ADI.
Exposure levels are	Specify whether exposure levels are absolute or percentages of ARfD/ADI.
Number of levels of covariable to predict exposure	Specify the number of levels, e.g. 20. The range of the covariable is divided by the number of levels: range $=$ $(max - min)/levels$ . For these covariable levels exposures are predicted.
Predict exposure at extra	Specify specific prediction levels in addition to the automatically
covariable levels	generated prediction levels (space separated).
Lower percentage for variability (%)	The default value of 25% may be overruled.
Upper percentage for variability (%)	The default value of 75% may be overruled.
Report consumptions and exposures per individual instead of per kg body weight	Specifies whether body weights should be ignored and consumptions and exposures should be expressed per individual. Otherwise, the consumptions and exposures are per kg body weight.

Table 2.199: Output settings for module Risks.

### **Calculation of risks**

Risk (health impact) is quantified as exposure relative to hazard characterisation, which in MCRA is called a hazard index (HI) for any type of inputs, or as hazard characterisation relative to exposure, which in MCRA is called a margin of exposure (MOE) for any type of inputs. Exposures or hazards can be single values or distributions, the risk metric is a distribution if at least one of the inputs is a distribution (if both are single values, see the module Single value risks). Risk metrics are valid for a specific biological level (external or internal at a specific organ).

• *Risks calculation*

Inputs used: *Exposures Hazard characterisations*

Settings used

• *[Calculation Settin](#page-259-0)gs*

Risks are ex[pressed as](#page-150-0) [distribution of margin of](#page-221-0) exposure or hazard index. The distribution is summarised by percentiles, and by the probability to exceed the specified threshold value (e.g. 1 or 100). The hazard vs. exposure plot compares the exposures and the hazard characterisation for individuals or individual-days in a population. Exposures, hazar[d characterisations an](#page-260-1)d risks can be acute or chronic. The default unit for exposures and hazard characterisations is  $\mu q/kgBW/day$ , but this can be changed by choosing non-default units for consumptions, concentrations and/or body weight.

By using probabilistic tiers for both exposure and hazard characterisation, the calculated MOE distribution is equal to the Integrated Margin Of Exposure (IMOE) distribution, as described for the Integrated Probabilistic Risk Asssessment (IPRA) approach in [van der Voet et al., 2007] and [van der Voet et al., 2009].

# **2.8.2 Single value risks**

Single value risks are risk [estimates obtained from c](#page-357-0)ombi[ning single value exposure](#page-357-1)s with single value hazard characterisations or as a percentile from a risk distribution.

<span id="page-263-0"></span>This module has as primary entities: *Substances Effects Populations*

# **Single value risks calculation**

Single value risks can be calculatedi[n two ways](#page-44-0)[.](#page-25-0)

- <span id="page-263-1"></span>• If the option 'Single value risk calculation method' is set to 'From single value risks' then *single value exposures* are combined with (single value) *hazard characterisations*.
- If the option 'Single value risk calculation method' is set to 'As percentile from risks distribution' then a percentile can be selected from a *risks* distribution.

See below for a more detailed explanati[on.](#page-221-0)

#### **Combining single value expo[sures](#page-259-1) and hazard characterisations**

Single value risks are computed by combining *single value exposures* by route/source and substance with (single value) *hazard characterisations* by substance. They are computed as margin of exposure (hazard characterisation / exposure), hazard quotient or hazard index (exposure / hazard characterisation), or as a percentage of the reference dose (100 \* exposure / hazard characterisation).

#### **Single value risks from individual risks**

In this option, a percentage point can be specified for the chosen risk metric (margin of exposure or hazard index). The corresponding percentile is calculated from the distribution of individual *risks*. The default percentiles are a margin of exposure at 0.1% or a hazard index at 99.9%, but another value can be chosen. It can also be indicated whether the risk metric is calculated using the inverse distribution or not. This option is provided because percentile calculation in small data sets is asymmetric in both tails. When this option is set, the percentile is calculated as the inverse of the complementary percentage of the inverse distribution. E.g., the  $p_{0.1}$  of th[e MO](#page-259-1)E distribution is calculated as  $1/(p_{99.9}$ of 1/MOE distribution); the  $p_{99.9}$  of the HI distribution is calculated as  $1/(p_{0.1}$  of 1/HI distribution).

#### **Adjustment factors and uncertainty specification**

Many sources of uncertainty that may affect input data, model assumptions and assessment methodology do not enter the assessment. In [EFSA, 2020a] and [EFSA, 2020b], thirty-four sources of uncertainty were identified and the impact of each source on the MOE was quantified. Some uncertainties tend to overestimate the MOE, others tend to underestimate it. Following the guidance of the EFSA Scientific Committee, specific MOE and/or HI percentiles are adjusted using adjustment factors for exposure and hazard, e.g. from expert elicitation. They may be available as fixed values or as para[metric uncertai](#page-355-1)nty di[stributions. In](#page-355-2) the nominal run, the percentile is adjusted with the median of the uncertainty distribution. In each uncertainty run, adjustment factors are sampled from the uncertainty distribution. In the MCRA interface, for both exposure and hazard distribution separately, a fixed value or a parametric uncertainty distribution is specified. The available parametric uncertainty distributions are the same as avaible in the SHELF package that was used by EFSA. The SHeffield ELicitation Framework (SHELF) is a package of documents, templates and software to carry out elicitation of probability distributions for uncertain quantities from a group of experts (http://www.tonyohagan.co.uk/shelf/).

#### **Options for specifying uncertainty distributions are:**

- Lognormal( $\mu$ , s) with offset c. Parameters  $\mu$  and s specify the mean and standard deviation of the un[derlying normal.](http://www.tonyohagan.co.uk/shelf/)
- Log Student t( $\mu$ , s,  $\nu$ ) with offset d. Parameters mu and s specify the mean and standard deviation of the underlying normal,  $\nu$  the degrees of freedom,  $\nu > 0$
- Beta(a, b) scaled to the interval [c, d], with shape parameters a and  $b > 0$ .
- Gamma(a, b) with offset c, with shape and rate parameters a and  $b > 0$ .

#### **Background-only adjustment factor**

When exposures are calculated by *combining focal food/substance concentrations with background concentrations*, it may be appropriate to have a separate adjustment for the foreground and background. A pragmatic solution agreed with EFSA is to estimate the contribution of the foreground in the tail above the selected percentile. Suppose this contribution is  $c$ . Note that  $c$  will vary in uncertainty runs. Then, the adjustment factor should be multiplied by  $(1 - c)$ , i.e. no adjustment for the [focal part.](#page-86-0)

The calculation proceeds as follows:

$$
\begin{array}{rcl} p_{\text{MOE},\text{adjusted}} & = p_{\text{MOE}} \cdot (c + (1-c) \cdot \text{AdjustmentFactor}_{\text{exposure}} \cdot \text{AdjustmentFactor}_{\text{hazard}}) \\ & = \frac{p_{\text{HI}}}{c + (1-c) \cdot \text{AdjustmentFactor}_{\text{exposure}} \cdot \text{AdjustmentFactor}_{\text{hazard}}} \end{array}
$$

Note that when the focal substance measurements are converted to active substances using *substance conversions* or *deterministic substance conversions*, then *c* is the sum of the contributions of the focal food in and all active substances to which the substance translates.

In Figure 2.50, an example is shown where the margin of exposure is adjusted for the exposure and hazard distribution based on expert elicitation. The median adjustment factors for exposure and hazard are res[pectively, 1.77 and 3.0](#page-109-0)1. [The overall adjustment factor is 5.3](#page-91-0)3.



Figure 2.46: Scaled lognormal ( $\mu$  =0.705, s=0.566, offset=1), table 8, EFSA 2020 [EFSA, 2020b].



Figure 2.47: Scaled logstudents t ( $\mu$  =-0.593, s=0.367,  $\nu$  =3, offset=0.5), table 9, EFSA 2020 [EFSA, 2020b].



Figure 2.48: Scaled beta (a=2.37, b=4.26, lowerbound=0.5, upperbound=6), table 7, EFSA 2020 [EFSA, 2020a].



Figure 2.49: Scaled gamma (a=3.26, b=3.56, offset=0.9), table 6, EFSA 2020 [EFSA, 2020a].



Figure 2.50: Margin of exposure (model) and adjusted margin of exposure (model + expert) with uncertainty bounds.

# **Single value risks settings**

# **Calculation settings**

freedom Logstudent-t)

Name	Description	
Multiple substances analysis	Specifies whether the assessment involves multiple substances.	
Express results in terms of	Specifies whether the assessment involves multiple substances and	
reference substance equivalents	results should be cumulated over all substances.	
(cumulative)		
Risk type	The type of exposure considered in the assessment; acute (short term) or chronic (long-term).	
Risk metric type	Report risks in terms of hazard index $(HI = 1/MOE)$ or margin of exposure.	
Single value risk calculation	Calculate single value from exposures and hazard or from an	
method	individual risks distribution.	
Percentage for percentile	Percentage for percentile (default 0.1 for MOE or 99.9 for HI).	
Use inverse distribution to calculate percentile	Calculate percentile via the complementary percentage of the inverse distribution (default: no). Description: E.g., P0.1 of MOE distribution is calculated via P99.9 of 1/MOE distribution. Note:	
	This option is provided because percentile calculation in small data sets is asymmetric in both tails.	
Apply adjustment factors to the specified MOE percentile	Specify adjustment factors, e.g. based on expert knowledge elicitation, to a specified MOE percentile (default 0.1%). If the selected risk metric is HI, the adjustment factors should still be specified for the complementary percentile of MOE (e.g. P0.1 of	
	MOE if P99.9 of HI is selected).	
Adjustment type related to exposure	Specify the factor and/or distribution of the adjustment factor for the MOE percentile. Default is no adjustment. Alternatives are a fixed factor or an uncertainty distribution. If distributions are selected, default values are set based on EFSA cumulative risk	
Parameter A (Fixed factor,	reports 2020. This parameter can be: 1) the fixed adjustment factor; 2) for	
mean Lognormal or LogStudent-t, or shape	Lognormal or LogStudent-t, the mean of the underlying normal distribution; 3) For Beta or Gamma. the shape parameter.	
parameter Beta or Gamma)		
Parameter B (standard deviation Lognormal or LogStudent-t or second shape parameter Beta or rate parameter Gamma)	This parameter can be: 1) for Lognormal or LogStudent-t, the standard deviation of the underlying normal distribution; 2) For Beta, the second shape parameter; 3) for Gamma, the rate parameter.	
Parameter C (Lower bound	This parameter can be: 1) for Beta, the lower bound value; 2) for	
Beta, offset Gamma or Lognormal or degrees of freedom Logstudent-t)	Gamma or Lognormal, the offset; 3) for LogStudent-t, the degrees of freedom.	
Parameter D (Upper bound Beta or offset LogStudent-t)	This parameter can be: 1) for Beta, the upper bound value; 2) for LogStudent-t, the offset.	
Adjustment type related to hazard	Specify the factor and/or distribution of the adjustment factor for the MOE percentile. Default is no adjustment. Alternatives are a fixed factor or an uncertainty distribution. If distributions are selected, default values are set based on EFSA cumulative risk reports 2020.	
Parameter A (Fixed factor, mean Lognormal or LogStudent-t, or shape parameter Beta or Gamma)	This parameter can be: 1) the fixed adjustment factor; 2) for Lognormal or LogStudent-t, the mean of the underlying normal distribution; 3) For Beta or Gamma. the shape parameter.	
Parameter B (standard	This parameter can be: 1) for Lognormal or LogStudent-t, the	
deviation Lognormal or	standard deviation of the underlying normal distribution; 2) For Beta, the second shape parameter; 3) for Gamma, the rate	
LogStudent-t or second shape parameter Beta or rate	parameter.	
<b>RPSKameter (Fegama)</b>		
Parameter C (Lower bound	This parameter can be: 1) for Beta, the lower bound value; 2) for	
Beta, offset Gamma or Lognormal or degrees of	Gamma or Lognormal, the offset; 3) for LogStudent-t, the degrees of freedom.	

Table 2.200: Calculation settings for module Single value risks.

# **Calculation of single value risks**

Single value risk can be computed by route and substance in the form of hazard quotients or margins of exposure. Single value risks are risk estimates obtained from combining single value exposures with single value hazard characterisations or as a percentile from a risk distribution. Optionally, the exposure and hazard behind the percentile calculation can be adjusted with fixed adjustment factors or factors drawn fromn parametric uncertainty distributions.

• *Single value risks calculation*

Inputs used: *Single value dietary exposures Hazard characterisations Risks*

#### Settings used

• *[Calculation Settings](#page-263-1)*



















Category	Module	Inputs	Used by	Description
	Single value	Active	Modelled	Single value concentrations
	concentra-	substances,	foods, Single	data are the single value
	<i>tions</i>	Concentra-	value dietary	estimates (High Residue,
		tions.	exposures.	Maximum Residue Limit,
		Concentra-		<b>Supervised Trials Median</b>
		tion limits,		Residue) of residue
		Deterministic		concentrations on foods as
		substance		measured.
		conversion		
		<i>factors.</i>		
	Processing	Foods,	Food	Processing factors are
	factors	Substances.	conversions,	multiplication factors to derive
			<b>Dietary</b>	the concentration in a
			exposures,	processed food from the
			Single value	concentration in an
			dietary	unprocessed food and can be
			exposures.	specified for identified
				processing types (e.g., cooking,
				washing, drying). Processing
				factors are primarily used in
				dietary exposure assessments
				to correct for the effect of
				processing on substance
				concentrations in dietary
				exposure calculations.
	Unit	Foods,	<b>Dietary</b>	Unit variability factors specify
	variability	Substances.	exposures,	the variation in concentrations
	factors		Single value	between single units of the
			dietary	same food, which have been
			exposures.	put together in a mixture
				sample on which the
				concentration measurements
				have been made. Unit
				variability factors are used to
				account for the fact that
				concentration data often relate
				to composite samples, whereas
				an acute risk may result from
				single food units.

Table 2.201 – continued from previous page





Category	Module	Inputs	Used by	Description
	Deterministic	Substances,	Concentra-	Deterministic substance
	substance	Foods.	tions, Single	conversion factors.
	conversion		value con-	
	factors		centrations.	
	Concentra-	Foods,	Concentra-	Concentration limits specify
	tion	Substances.	tions, Single	(legal) limit values for
	limits		value con-	substance concentrations on
			centrations.	foods and are sometimes used
			Concentra-	as conservative values for
			tion models,	concentration data. In the
			Modelled	framework of pesticides the
			foods.	legal Maximum Residue Limit
				(MRL) is the best known
				example.
	Concentra- $tion$	$\overline{\mathit{Concentra}}$ -	High	Concentration models are
		tions.	exposure	distributional models of
	models	Concentra-	food-	substance concentrations on
		tion limits,	substance	foods. They describe both the
		<b>Modelled</b>	combina-	substance presence (yes/no,
		foods,	tions,	with no representing an
		Substance	<b>Dietary</b>	absolute zero concentration)
		authorisa-	exposures.	and the substance
		tions.		concentrations. Concentration
		Occurrence		models are specified per
		frequencies,		food/substance combination.
		Relative		
		potency		
		factors, Con-		
		centration		
		distributions,		
		Total diet		
		study sample		
		composi-		
		tions.		
	Modelled	Concentra-	Concentra-	Modelled foods are foods
	foods	tions, Single	tion models,	within the foods scope for
		value con-	Food	which concentration data or
		centrations,	conversions.	MRLs of substances are
		Concentra-		
		tion		available (or expected).
		limits.		
	Focal food	Foods,	Concentra-	In some cases the attention in
	concentra-	Substances.	tions.	an assessment is on a specific
	tions			food (focal food), against the
				background of other foods.
				Focal food concentrations are
				separate concentration data for
				one or more focal food
				commodities, that will take the
				place of any other
				concentration data for the focal
				food in the ordinary
				concentration data.

Table 2.201 – continued from previous page

Category	Module	Inputs	Used by	Description
	Total diet	Foods.	$\overline{Concentra}$ -	Total diet study sample
	study sample		tion models,	compositions specify the
	compositions		Food	composition of mixed food
			conversions.	samples, such as used in a total
				diet study (TDS), in terms of
				their constituting foods.
	Food extrap-	Foods.	Concentra-	Food extrapolations data
	olations		tions, Food	specify which foods (data rich
			conversions.	foods) can be used to impute
				concentration data for other
				foods with insufficient data
	Food			(data poor foods). Food conversions relate
Exposure	conversions	Consump- tions,	Consump- tions by	foods-as-eaten, as found in the
		Modelled	modelled	consumption data, to modelled
		foods,	food,	foods (foods-as-measured),
		Processing	Dietary	which are the foods for which
		factors,	exposures.	concentration data are
		Food recipes,		available. A food-as-eaten can
		<b>Market</b>		be linked to one, or multiple
		shares, Food		food-as-measured using
		extrapola-		various conversion steps (e.g.,
		tions, Total		using food recipes to translate a
		diet study		composite food into its
		sample com-		ingredients). There are several
		positions,		types of conversion steps, and a
		Active		conversion path may comprise
		substances.		multiple conversion steps
				between a food-as-eaten and a
				food-as-measured.
	Consump-	Consump-	Single value	Consumptions by modelled
	tions by	tions, Food	consump-	food are consumptions of
	modelled	conversions.	tions, High	individuals expressed on the
	food		exposure	level of the foods for which
			food-	concentration data are available
			substance	(i.e., the modelled-foods).
			combina-	These are calculated from
			tions,	consumptions of foods-as-eaten and food
			<b>Dietary</b> exposures.	conversions that link the
				foods-as-eaten amounts to
				modelled-foods amounts.
	High	Consump-	<b>Dietary</b>	Identification of
	exposure	tions by	exposures.	food-as-eaten/food-as-
	food-	modelled		measured/substance
	substance	food, Con-		combinations that have the
	combina-	centration		highest expected contribution
	tions	models,		to exposure based on a simple
		Active		screening model.
		substances,		
		Relative		
		potency		
		<i>factors.</i>		

Table 2.201 – continued from previous page

Category	Module	Inputs	Used by	Description
	<b>Dietary</b>	Consump-	Exposures.	Dietary exposures are the
	exposures	tions by		amounts of substances,
		modelled		expressed per kg bodyweight
		food, Con-		or per individual, to which
		centration		individuals in a population are
		models,		exposed from their diet per
		Processing		day. Depending on the
		factors, Unit		exposure type, dietary
		variability		exposures can be
		factors, High		short-term/acute exposures and
		exposure		then contain exposures for
		food-		individual-days, or they can be
		substance		long-term/chronic exposures,
		combina-		in which case they represent
		tions, Active		the average exposure per day
		substances,		over an unspecified longer time
		<i>Occurrence</i>		period.
		patterns,		
		Relative		
		potency		
		factors,		
		Food		
		conversions,		
		Concentra-		
		tion		
		distributions.		
	Single value	Single value	Single value	Single value dietary exposures
	dietary	consump-	risks.	are based on the single value
	exposures	tions, Single		concentrations of substances,
		value con-		expressed per standard (kg)
		centrations,		bodyweight and/or single value
		Processing		amounts of consumed food as
		factors, Unit		measured. Depending on the
		variability		exposure type, dietary
		factors,		exposures can be
		Occurrence		short-term/acute exposures.
		frequencies.		
	Non-dietary	Populations,	Exposures.	Non-dietary exposures are the
	exposures	Substances,		amounts of substances to which
		Active		individuals in a population are
		substances.		exposed via any of three
				non-dietary routes: dermal,
				inhalation or oral, per day.

Table 2.201 – continued from previous page

Category	Module	Inputs	Used by	Description
	<b>Exposures</b>	<i>Dietary</i>	<b>Exposure</b>	Exposures are amounts of
		exposures,	mixtures,	substances, typically expressed
		Non-dietary	Human	per mass unit and per day, to
		exposures,	monitoring	which individuals in a
		Active	analysis,	population are exposed at a
		substances,	Risks.	chosen target level. This target
		Relative		level may be external exposure
		potency		(dietary exposure, expressed
		factors,		per unit body weight, or per
		Kinetic		person) or internal exposure
		models.		(expressed per unit organ
				weight). Internal exposures
				may be aggregated from
				dietary and non-dietary
				exposures using either
				absorption factors or kinetic
				models to translate the external
				exposures to internal
				exposures. Exposures can be
				short-term/acute exposures and
				then contain exposures for
				individual-days, or they can be
				long-term/chronic exposures,
				in which case they represent
				the average exposure per day
				over an unspecified longer time
				period.
	Exposure	<i>Exposures.</i>		Exposure mixtures are
	mixtures			mixtures of substances that
				contribute relatively much to
				the overall cumulative exposure
			Human	(potential risk drivers).
	Human	Substances.		Human monitoring data
	monitoring		monitoring	quantify substance concentrations found in
	data		analysis.	
				humans collected in human
				monitoring surveys.
	Human	Human		Human monitoring analysis
	monitoring	monitoring		compares observed human
	analysis	data,		monitoring data with
		Exposures.		predictions made for the same
				population of individuals from
				dietary survey data,
				concentration data and
				(optionally) non-dietary
				exposure data.

Table 2.201 – continued from previous page

Category	Module	Inputs	Used by	Description
<i>In-silico</i>	<b>QSAR</b>	Substances.	Active	QSAR membership models
	membership	Effects, AOP	substances.	specify assessment group
	models	networks.		memberships for active
				substances related to a specific
				health effect (adverse
				outcome). Memberships
				should be derived externally
				from Quantitative
				<b>Structure-Activity Relationship</b>
				(QSAR) models.
	Molecular	Substances,	Active	Molecular docking models
	docking	Effects, AOP	substances.	specify binding energies for
	models	<i>networks.</i>		substances in specific
				molecular docking models
				related to a specific health
				effect (adverse outcome).
<b>Kinetic</b>	<b>Kinetic</b>	Substances,	Exposures,	Kinetic models convert
	models	Active	Hazard	exposures or hazard
		substances.	<i>characteri-</i>	characterisations from one or
			sations.	more external routes or
				compartments to an internal
				(target) compartment. The
				reverse conversion from
				internal to external can also be
				made (reverse dosimetry).

Table 2.201 – continued from previous page

Category	Module	Inputs	Used by	Description
Hazard	Active	$\overline{AOP}$	Concentra-	Active substances are
	substances	networks,	tions, Single	substances that may lead (P>0)
		Points of	value con-	to a specific health effect
		departure,	centrations,	(adverse outcome). Active
		Hazard	<b>Occurrence</b>	substances are specified
		characteri-	patterns,	directly as data or calculated
		sations,	<i>Occurrence</i>	from POD presence, QSAR
		Molecular	frequencies,	models or Molecular docking
		docking	<b>Substance</b>	models. Active substances can
		models,	conversions,	have an assessment group
		<b>QSAR</b>	Non-dietary	membership 1 (crisp), or
		membership	exposures,	values in the range $(0,1]$
		models.	Kinetic	(probabilistic).
			models,	
			Relative	
			potency	
			factors,	
			Hazard	
			<i>characteri-</i>	
			sations.	
			Inter-species	
			conversions,	
			Intra species	
			factors,	
			Food	
			conversions,	
			High	
			exposure	
			food-	
			substance	
			combina-	
			tions,	
			Dietary	
			exposures,	
			Exposures.	

Table 2.201 – continued from previous page

Category	Module	Inputs	Used by	Description
	Relative	Active	Concentra-	Relative potency factors
	potency	substances,	tions.	(RPFs) quantify potencies of
	factors	AOP	Concentra-	substances with respect to a
		networks.	tion models.	defined effect, relative to the
		Hazard	High	potency of a chosen index
		characteri-	exposure	substance. RPFs can be used to
		sations.	food-	express combined exposures of
			substance	multiple substances in terms of
			combina-	a the exposure value of the
			tions.	chosen index substance (i.e., in
			Dietary	index substance equivalents).
			exposures,	In MCRA, hazard
			Exposures.	characterisations, and therefore
				also RPFs are based on mass
				units (e.g., $\mu$ g), and not on mol
				units. RPFs can be different for
				different levels of the human
				organism (external, internal,
				specific compartment). RPFs
				can be given as data or
				computed from hazard
				characterisations. RPFs can be
				specified with uncertainty.
				Computation from uncertain
				hazard characterisations allows
				to include correlations between
				uncertain RPFs which originate
				from using the same index
				substance.

Table 2.201 – continued from previous page





Category	Module	Inputs	Used by	Description
	<b>Dose</b>	Substances,	<b>Dose</b>	Dose response data are data on
	response	Test systems,	response	response values of test systems
	data	Responses.	models.	at specified doses of substances
				(or mixtures of substances)
				from dose response
				experiments.
	Effect repre-	Effects,	Hazard	Effect representations specify
	sentations	Responses,	characteri-	the responses that can be used
		AOP	sations, Dose	to measure specified effects
		networks.	response	and which response levels, the
			models.	benchmark response (BMR),
				define the hazard limits for the
				effects.
	Inter-species	Substances,	Hazard	Inter-species conversions
	conversions	Effects,	characteri-	specify how to convert a hazard
		Active	sations.	characterisation for a given
		substances.		species to a hazard
				characterisation for humans. In
				the simplest approach, this
				specifies a fixed inter-species
				factor. In a higher tier, this
				specifies a geometric mean
				(GM) and geometric standard
				deviation (GSD) for a
				lognormal uncertainty
				distribution of the interspecies
				factor. Inter-species conversion
				are specified per effect and can
				be general or
				substance-specific.
	Intra species	Substances,	Hazard	Intra-species factors specify
	factors	Effects,	characteri-	how to convert a hazard
		<b>Active</b>	sations.	characterisation from the
		substances.		average to a sensitive human
				individual.

Table 2.201 – continued from previous page

Category	Module	Inputs	Used by	Description
	<b>AOP</b>	Effects.	<b>QSAR</b>	Effects are related to each
	networks		membership	other using the toxicological
			models,	concept of adverse outcome
			Molecular	pathways (AOPs) and adverse
			docking	outcome pathway networks
			models,	(see https://aopwiki.org).
			Active	Adverse Outcome Pathway
			substances,	(AOP) Networks specify how
			Relative	biological events (effects) can
			potency	lead to an adverse outcome
			factors,	(AO) in a qualitative way
			Hazard	through relations of upstream
			characteri-	and downstream key events
			sations,	(KEs), starting from molecular
			Points of	initiating events (MIEs). Using
			departure,	AOPs, the adverse outcome
			Effect repre-	(AO), e.g., liver steatosis, is
			sentations.	linked to key events (KEs),
				e.g., triglyceride accumulation
				in the liver, and to molecular
				initiating events (MIEs), e.g.,
				PPAR-alpha receptor
				antagonism. In general,
				multiple AOPs may lead to the
				same AO, and therefore AOP
				networks can be identified.
<b>Risks</b>	<b>Risks</b>	Exposures,	Single value	Risks (health impacts) are
		Hazard	risks.	defined as a function of
		characteri-		exposure and hazard
		sations.		characterisation at a chosen
				biological level (external or
				internal). Risk metrics are
				margins of exposure (MOE) or
				hazard indices (HI) or more
				generalised MOE or HI
				distributions.
	Single value	Single value		Single value risks are risk
	risks	dietary		estimates obtained from
		exposures,		combining single value
		Hazard		exposures with single value
		characteri-		hazard characterisations or as a
		sations,		percentile from a risk
		Risks.		distribution.

Table 2.201 – continued from previous page
# **STANDARD ACTIONS**

A standard action is a user friendly way to perform a complex probabilistic calculation. By using a standard action predefined settings are used and the user can set only a limited number of selections. All settings (pre-defined and set by the user) are visible in the output. As a result a short output is presented. More detailed output is still available.

# **3.1 EU acute cumulative exposure assessment (2018) Tier I and Tier II**

#### This standard action is of type: *Dietary exposures*

This standard action is based on work done in 2018 [van Klaveren et al., 2019a]. In the context of the second framework partnership agreement between the National Institute for Public Health and the Environment of the Netherlands (RIVM) and the European Food Safety Authority (EFSA) acute cumulative dietary exposure assessments were performed for two cumulative asse[ssment groups \(CA](#page-116-0)Gs) of pesticides that affect the nervous system: pesticides causing brain and/or erythrocyte AChE inhibition (CAG-N[AN, 47 pesticides\) and pes](#page-357-0)ticides causing functional alterations of the motor division (CAG-NAM, 100 pesticides). The exposure assessments used monitoring data collected by the Netherlands under their official monitoring programmes in 2014, 2015 and 2016 and individual Dutch food consumption data. Exposure estimates were obtained for each group of pesticides using the MCRA software. The Standing Committee on Plants, Animals, Food and Feed (SC PAFF) discussed the scope of the assessment in 2018 and agreed on the parameters to be used for the cumulative exposure assessment. Based on that discussion, a very conservative tier I modelling approach and a refined, but still conservative tier II modelling approach were used. In these assessments, common risk assessment practice was followed and the cumulative exposure was expressed as the total margin of exposure (MOET) at the 50th, 90th, 95th, 99th and 99.9th percentile of the exposure distribution.

Table 3.1: Datasources for EU acute cumulative exposure assessment (2018) Tier I and Tier II.

<b>Table Group</b>	<b>Name</b>	Repository	Type
AssessmentGroup-	SecondaryInput-	Standard Actions/Cumulative Exposure Assessment/EU	Fixed
Memberships	Data.mdb	Acute Cumulative Exposure Assessment	
AuthorisedUses	SecondaryInput-	Standard Actions/Cumulative Exposure Assessment/EU	Fixed
	Data.mdb	Acute Cumulative Exposure Assessment	
Compounds	SecondaryInput-	Standard Actions/Cumulative Exposure Assessment/EU	Fixed
	Data.mdb	Acute Cumulative Exposure Assessment	
Concentrations	a_ConcentrationsSSD	NtandbActions/Cumulative Exposure Assessment/EU	Vari-
		Acute Cumulative Exposure Assessment	able
Concentrations	a_ConcentrationsSSD	NtandhandbActions/Cumulative Exposure Assessment/EU	Vari-
		Acute Cumulative Exposure Assessment	able
Effects	SecondaryInput-	Standard Actions/Cumulative Exposure Assessment/EU	Fixed
	Data.mdb	<b>Acute Cumulative Exposure Assessment</b>	
FoodExtrapola-	SecondaryInput-	Standard Actions/Cumulative Exposure Assessment/EU	Fixed
tions	Data.mdb	Acute Cumulative Exposure Assessment	
FoodTranslations	SecondaryInput-	Standard Actions/Cumulative Exposure Assessment/EU	Fixed
	Data.mdb	Acute Cumulative Exposure Assessment	
Foods	SecondaryInput-	Standard Actions/Cumulative Exposure Assessment/EU	Fixed
	Data.mdb	Acute Cumulative Exposure Assessment	
<b>HazardDoses</b>	SecondaryInput-	Standard Actions/Cumulative Exposure Assessment/EU	Fixed
	Data.mdb	Acute Cumulative Exposure Assessment	
MaximumResidu-	SecondaryInput-	Standard Actions/Cumulative Exposure Assessment/EU	Fixed
eLimits	Data.mdb	<b>Acute Cumulative Exposure Assessment</b>	
Processing	SecondaryInput-	Standard Actions/Cumulative Exposure Assessment/EU	Fixed
	Data.mdb	<b>Acute Cumulative Exposure Assessment</b>	
<b>ResidueDefinitions</b>	SecondaryInput-	Standard Actions/Cumulative Exposure Assessment/EU	Fixed
	Data.mdb	Acute Cumulative Exposure Assessment	
Survey		a_ConsumptionsNL2.m&tandard Actions/Cumulative Exposure Assessment/EU	Fixed
		Acute Cumulative Exposure Assessment	
Survey		a_ConsumptionsNL3_6Staalldard Actions/Cumulative Exposure Assessment/EU	Fixed
		<b>Acute Cumulative Exposure Assessment</b>	
UnitVariability	UnitVarPrimo.mdb	Standard Actions/Cumulative Exposure Assessment/EU	Vari-
		Acute Cumulative Exposure Assessment	able
UnitVariability	UnitVar36.mdb	Standard Actions/Cumulative Exposure Assessment/EU	Vari-
		Acute Cumulative Exposure Assessment	able

# **3.2 EU chronic cumulative exposure assessment (2018) Tier I and Tier II**

This standard action is of type: *Dietary exposures*

This standard action is based on research done in 2018 [van Klaveren et al., 2019b].

This standard action will enable you to reproduce the exposure assessment of chronic cumulative effects of pesticide residues in food affecting the thy[roid. These are ret](#page-116-0)rospective exposure assessments of the cumulative exposure for the thyroid using monitoring data from 2014, 2015 and 2016. In this standard action Dutch monitoring and consumption data are used. The results, data used and methodology [are reported in a scientific r](#page-357-1)eport following published on the EFSA website in September 2019. The methodology fulfils the requirements set by the European Commission.

Table 3.2: Datasources for EU chronic cumulative exposure assessment (2018) Tier I and Tier II.

<b>Table Group</b>	Name	Repository	$\overline{\mathsf{Type}}$ $\mid$
Assessment-	SecondaryInput-	Standard Actions/Cumulative Exposure Assessment/EU	Fixed
GroupMember-	Data.mdb	<b>Chronic Cumulative Exposure Assessment</b>	
ships			
AuthorisedUses	SecondaryInput-	Standard Actions/Cumulative Exposure Assessment/EU	Fixed
	Data.mdb	<b>Chronic Cumulative Exposure Assessment</b>	
Compounds	SecondaryInput-	Standard Actions/Cumulative Exposure Assessment/EU	Fixed
	Data.mdb	<b>Chronic Cumulative Exposure Assessment</b>	
Concentrations		c_ConcentrationsSSD_StGRlandb Actions/Cumulative Exposure Assessment/EU	Vari-
		<b>Chronic Cumulative Exposure Assessment</b>	able
Concentrations		c_ConcentrationsSSD_StoHdandb Actions/Cumulative Exposure Assessment/EU	Vari-
		<b>Chronic Cumulative Exposure Assessment</b>	able
Effects	SecondaryInput-	Standard Actions/Cumulative Exposure Assessment/EU	Fixed
	Data.mdb	<b>Chronic Cumulative Exposure Assessment</b>	
FoodExtrapola-	SecondaryInput-	Standard Actions/Cumulative Exposure Assessment/EU	Fixed
tions	Data.mdb	<b>Chronic Cumulative Exposure Assessment</b>	
FoodTranslations	SecondaryInput-	Standard Actions/Cumulative Exposure Assessment/EU	Fixed
	Data.mdb	Chronic Cumulative Exposure Assessment	
Foods	SecondaryInput-	Standard Actions/Cumulative Exposure Assessment/EU	Fixed
	Data.mdb	<b>Chronic Cumulative Exposure Assessment</b>	
<b>HazardDoses</b>	SecondaryInput-	Standard Actions/Cumulative Exposure Assessment/EU	Fixed
	Data.mdb	<b>Chronic Cumulative Exposure Assessment</b>	
MaximumResidu-	SecondaryInput-	Standard Actions/Cumulative Exposure Assessment/EU	Fixed
eLimits	Data.mdb	<b>Chronic Cumulative Exposure Assessment</b>	
Processing	SecondaryInput-	Standard Actions/Cumulative Exposure Assessment/EU	Fixed
	Data.mdb	Chronic Cumulative Exposure Assessment	
<b>ResidueDefinitions</b>	SecondaryInput-	Standard Actions/Cumulative Exposure Assessment/EU	Fixed
	Data.mdb	<b>Chronic Cumulative Exposure Assessment</b>	
Survey		c_ConsumptionsNL2.rfiddindard Actions/Cumulative Exposure Assessment/EU	Fixed
		Chronic Cumulative Exposure Assessment	
Survey		c_ConsumptionsNL3_6tandbard Actions/Cumulative Exposure Assessment/EU	Fixed
		<b>Chronic Cumulative Exposure Assessment</b>	

• *EU acute cumulative exposure assessment (2018) Tier I and Tier II*

• *EU chronic cumulative exposure assessment (2018) Tier I and Tier II*

**FOUR**

# **EXAMPLES**

#### **Note:** This section is under construction. Please contribute!

Training materials used in EuroMix training sessions:

- EuroMix dietary exposure
- RPF-exercise 1-for training-draft

There are a few exercises prepared that you could follow to get started.

# **4.1 [Cumulative dietary](https://mcra-test.rivm.nl/Mcra91/WebApp/manual/_static/RPF-exercise 1-for training-draft.pdf) exposure assessment**

## **4.1.1 Introduction**

The goal of this exercise is to perform a probabilistic cumulative dietary exposure assessment, illustrating all data needed. In Example 1 we will upload and use nine different files containing the data. In Example 2 we will upload and use a single data file for the same purpose. In the example the exposure will be characterised by upper tail percentiles, and the risk driving substances and foods can be examined. In Example 3 an uncertainty analysis is added.

## **4.1.2 Preparation**

In the workspace browser ( $\Box$  icon), create a new workspace *Examples*, using the  $+$  button in the bottom right corner.

## **4.1.3 Example 1**

Calculate a cumulative chronic dietary exposure for Dutch young adults in 2003 regarding a group of eight triazole substances according to the basic optimistic model of the EFSA 2012 guidance document. Use liver steatosis as a focal effect and Cyproconazole as an index substance. The data files are already available in the data folder *Documentation-Examples / Exercise Dietary Exposure Assessment*.

Detailed steps are as follows.

- In the *Examples* workspace, create a new action using the <sup>+</sup> button in the bottom right corner.
	- Select action type *Dietary exposures*
		- Name it, e.g. *Triazoles exposures*
		- (Optional) You can also add tags (e.g. triazoles, NL, steatosis) as labels that can be used later to find similar actions
		- (Optional) You can add a description for further information
		- Click Next
- Specify Dietary exposures settings
	- Tier: *EFSA 2012 Optimistic*
	- Risk type *Chronic*
	- Click *Create*

You are now directed to the main page of the new action. You can always return to this main page by clicking Action settings ͊ or the action type name (*Dietary exposures*) in the green bar.

The main page contains at least three blocks of information: Scope, Inputs and Settings. We will now first link all nine data files needed for this cumulative assessment. For most settings we will use default values in accordance with the chosen tier (*EFSA 2012 Optimistic*).

Scope of the assessment:

- Click *Effects* (path in the green bar changes Total *Dietary exposures / Effects*)
	- At *Effects data source*, click  $\bullet$  and browse to the file *Effect Steatosis.xlsx*, then click *Select*
	- At *Effect Settings* for *focal effect select Steatosis-liver* and click **a** *Save Changes*
	- In the green navigation bar, click *Dietary exposures* to go up one level.
- Click *Foods* (path: *Dietary exposures / Foods*)
	- At *Foods data source*, click **/** and browse to the file *Foods.xlsx*, then click *Select*
	- In the green navigation bar, click *Dietary exposures* to go up one level
- Click *Populations (optional)* (path: *Dietary exposures / Populations*)
	- At *Populations data source*, click  $\bullet$  and browse to the file *Populations.xlsx*, then click *Select*
	- $\bullet$  This file contains two populations, only one is allowed. Click  $\bullet$  under Populations selection, this opens a pop-up window. Deselect  $NL_2006$ , then click *Save*. The red warning signs  $\triangle$  should now be gone. (Note: green warning signs  $\triangle$  point at details and can usually be ignored)
	- In the green navigation bar, click *Dietary exposures* to go up one level.
- Click *Substances* (path: *Dietary exposures / Substances*)
	- At *Substances data source*, click  $\triangle$  and browse to the file *Substances Triazoles.xlsx*, then click *Select*
	- At *Substance settings* for *Index substance* select *Cyproconazole* and click **a** *Save Changes*
	- In the green navigation bar, click *Dietary exposures* to go up one level

Next we choose the other input data:

- Click *Consumptions by modelled foodd* (path: *Dietary exposures / Consumptions by modelled food*)
	- Click *Consumptions* (path: *Dietary exposures / Consumptions by modelled food / Consumptions*)
		- At *Consumptions data source*, click  $\bullet$  and browse to the file *FoodConsumptions.xlsx* and *Select*
		- At *Consumptions data selection*, with  $\triangle$  open the food consumption surveys selection.
			- The file contains two surveys, but only one is allowed. Click  $\rightarrow$  under Consumptions data selection, this opens a pop-up window. Deselect *VCP-kids*, then click *Save* (the red warning  $\triangle$ should now be gone)
		- In the green navigation bar, click *Consumptions by modelled food* to go up one level
	- Click *Food conversions* (path: *Dietary exposures / Consumptions by modelled food / Food conversions*)
		- Click *Foods as measured* (path: *Dietary exposures / Consumptions by modelled food / Food conversions / Foods as measured*)
			- Click *Concentrations* (path: *Dietary exposures / Consumptions by modelled food / Food conversions / Foods as measured / Concentrations*)
- At *Concentrations data source*, click  $\triangle$  and browse to the file *ConcentrationData.xlsx*, then click *Select*
- In the green navigation bar, click *Food conversions* to go up two levels
- Click *Food recipes* (path: *Dietary exposures / Consumptions by modelled food / Food conversions / Food recipes*)
	- At *Food recipes data source*, click **a** and browse to the file *FoodTranslations.xlsx*. then click *Select*
	- In the green navigation bar, click *Dietary exposures* to go up three levels
- Click *Concentration models* (path: *Dietary exposures / Concentration models*)
	- Click *Relative potency factors* (path: *Dietary exposures / Concentration models / Relative potency factors*)
		- At *Relative potency data source*, click  $\bullet$  and browse to the file *RPFs.xlsx*, then click *Select*
		- In the green navigation bar, click *Dietary exposures* to go up two levels
- Click *Processing factors* (path: *Dietary exposures / Processing factors*)
	- At *Processing factors data source*, click **or** and browse to the file *ProcessingFactors.xlsx*, then click *Select*
	- In the green navigation bar, click *Dietary exposures* to go up one level
- Click *Active substances (optional)* (path: *Dietary exposures / Active substances*)
	- In this example we have a fixed list of relative potency factors for the eight substances, and don't need point of departure (POD) data to decide which substances are active with respect to the health effect and therefore belong to the cumulative assessment group. Deselect the setting "Derive memberships from POD presence", then click **a** *Save Changes*
	- In the green navigation bar, click *Dietary exposures* to go up one level

Now run the model, either by clicking the  $\blacktriangleright$  run icon in the grey bar, or by clicking the  $\blacktriangleright$  run icon in the green bar (Note:  $\triangleright$  in the green bar can also be used to run subactions on their own).

The  $\triangleright$  icon is replaced by the text "Running". When the run has finished, the interface automatically changes to the Results screen. You can also click the Results icon  $\bullet$  to go there.

As an exercise, try find the following results:

- 1. The 99th percentile of cumulative exposure
- 2. The substance(s) with highest contribution to the total exposure
- 3. The food(s)-as-measured with the highest contribution to the upper tail of the exposure distribution

Answers:

- In the grey bar, browse to the results panel by clicking the  $\bullet$  icon and click on the latest output (path: *Results / Dietary exposures*)
	- In the *Dietary exposures* tab, browse in the tree (unfold by clicking  $\rightarrow$  where necessary) to  $\rightarrow$  *Dietary exposures* ţ *Distribution (OIM)* ţ *Percentiles*
		- In the table it states that the 99% exposure percentile is at an exposure of 0.02127 µg/kg bw/day.
	- In the *Dietary exposures* tab, browse in the tree (unfold by clicking  $\rightarrow$  where necessary) to  $\rightarrow$  *Dietary exposures* ţ *Details* ţ *Exposures by substance* ţ *Total distribution*
		- From the pie chart it is clear that Tebuconazole contributes the most to the total exposure distribution with 32.7%. In the table below the graph more details can be found.
	- In the *Dietary exposures* tab, browse in the tree (unfold by clicking  $\rightarrow$  where necessary) to  $\rightarrow$  *Dietary* exposures  $\checkmark$  *Details*  $\checkmark$  *Exposures by food and substance*  $\checkmark$  *Risk drivers upper tail* 
		- From the pie chart it is clear that Flusilazole in grapefruit contributes the most (16.7%) to the upper tail exposure distribution

## **4.1.4 Example 2**

We will create a new action to demonstrate uploading all the data at once. All data is now contained within one file, *MCRA-Documentation Example Dietary exposures.xlsx*.

Detailed steps are as follows.

- In the *Examples* workspace, create a new action (using  $+$ )
	- Select action type *Dietary exposures*
	- Name it, e.g. *Triazoles exposures from one data file*
	- Click Next
- Specify Dietary exposures settings
	- Tier: *EFSA 2012 Optimistic*
	- Risk type *Chronic*
	- Click *Create*
- Then go to the actions settings  $\bullet$  of this action (path: *Dietary exposures*)
	- Click *Effects* (path: *Dietary exposures / Effects*)
		- At *Effects data source*, click  $\triangle$  and browse to the file MCRA-Documentation Example Dietary expo*sures.xlsx*. Click *Toggle all*, then *Select*. This will load all available data tables for all subactions of *Dietary exposures*.

You still need to specify the focal effect (under *Effects*), index substance (under *Substances*), and food surveys (under *Consumptions by modelled food / Consumptions*). You also need to deselect the "Derive memberships from POD presence" setting under *Active substances*. Navigate to the subaction where these changes have to be made using the green bar.

You now have achieved the same as in Example 1, only with the upload of one single file. You can now run the model, and inspect the results, which should be the same as for Example 1.

# **4.1.5 Example 3**

Repeat the run of the previous task, but in addition to the nominal run, perform an uncertainty analysis as well.

- Click on the  $\vec{f}$  icon (in the grey bar) to open the uncertainty settings panel
	- At *Uncertainty settings, check*  $\checkmark$  *Perform uncertainty analysis* 
		- For *Monte Carlo iterations per uncertainty run* choose 100, and press **a** *Save Changes*
- Now run the model, by pressing the  $\blacktriangleright$  run icon in the grey bar. Note that the run will take much more time.

Compare with the previous results, to find:

- 1. 95% uncertainty bounds for the 99% exposure percentile
- 2. 95% uncertainty bounds for the highest contribution from a substance to the total exposure distribution
- 3. 95% uncertainty bounds for the highest contribution from a food to the total exposure distribution

# **4.2 Aggregate exposure assessment**

# **4.2.1 Introduction**

The goal of this exercise is to assess aggregate exposure assessment.

# **4.2.2 Preparation**

If you haven't done so, in the workspace browser (use the  $\Box$  icon), create a new workspace named *Examples*, using the  $+$ .

The data files used in the example(s) in this section, are located in the data folder *Documentation-Examples / Exercise Aggregate Exposure Assessment*.

# **4.2.3 Example 1**

- In the *Examples* workspace, create a new action (using  $+$ )
	- Then select ɮ *Show all action types*, select *Exposures*
	- Name it *exposures*
	- At *Exposure settings* choose:
		- As *Risk type Chronic*
		- Check  $\checkmark$  *Include dietary and non-dietary routes of exposure*
	- Press *Create*
- Then go to the Actions settings  $\clubsuit$  of this action (path: *Exposures*)
	- At *Scope*, click *Effects* (path: *Exposures / Effects*)
		- At *Effects data source* with **/** browse to the file *Effect Steatosis.xlsx* and *Select*
		- At *Effect settings* for *Focal effect select Steatosis-liver* and press **B** *Save Changes*
		- In the green navigation bar, click *Exposures* to go up one level
	- At *Scope*, click *Foods* (path: *Exposures / Foods*)
		- At *Foods data source* with **b** browse to the file *Foods.xlsx* and *Select*
		- In the green navigation bar, click *Exposures* to go up one level
	- At *Scope*, click *Substances* (path: *Exposures / Substances*)
		- At *Substances data source* with **/** browse to the file *Substances.xlsx* and *Select*
		- At *Substance settings*
			- for *Index substance* select *Cyproconazole* and press **B** *Save Changes*
		- In the green navigation bar, click *Exposures* to go up one level
	- At *Inputs*, click *Dietary exposures* (path: *Exposures / Dietary exposures*)
		- At *Inputs*, click *Consumptions by modelled food* (path: *Exposures / Dietary exposures / Consumptions by modelled food*)
			- At *Inputs*, click *Consumptions* (path: *Exposures / Dietary exposures / Consumptions by modelled food / Consumptions*)
				- At *Consumptions data source* with **b** browse to the file *Consumptions.xlsx* and *Select*
				- At *Consumptions data selection* with  $\triangle$  open the food consumption surveys selection.
- The file contains two surveys, but only one is allowed. So deselect everything by clicking ɮ on the first line, next to the word *Code*
- Now select *DNFCS\_2003* and press *Save* (the red warning  $\triangle$  should now be gone)
- In the green navigation bar, click *Consumptions by modelled food* to go up one level
- At *Inputs*, click *Food conversions* (path: *Exposures / Dietary exposures / Consumptions by modelled food / Food conversions*)
	- At *Inputs*, click *Foods as measured* (path: *Exposures / Dietary exposures / Consumptions by modelled food / Food conversions / Foods as measured*)
		- At *Inputs*, click *Concentrations* (path: *Exposures / Dietary exposures / Consumptions by modelled food / Food conversions / Foods as measured Concentrations*)
			- At *Concentration data source* with  $\bullet$  browse to the file *ConcentrationData.xlsx* and *Select*
			- In the green navigation bar, click *Food conversions* to go up two levels
	- At *Inputs*, click *Food recipes* (path: *Exposures / Dietary exposures / Consumptions by modelled food / Food conversions / Food recipes*)
		- At *Food recipes data source*, with **b** browse to the file *FoodRecipes.xlsx* and *Select*
		- In the green navigation bar, click *Dietary exposures* to go up three levels
- At *Inputs*, click *Concentration models* (path: *Exposures / Dietary exposures / Concentration models*)
	- At *Inputs*, click *Relative potency factors* (path: *Exposures / Dietary exposures / Concentration models / Relative potency factors*)
		- At *Relative potency factors data source* with  $\bullet$  browse to the file *RelativePotencyFactors.xlsx* and *Select*
		- In the green navigation bar, click *Dietary exposures* to go up two levels
- At *Inputs*, click *Processing factors* (path: *Exposures / Dietary exposures / Processing factors*)
	- At *Processing factors data source* with  $\bullet$  browse to the file *ProcessingFactors.xlsx* and *Select*
	- In the green navigation bar, click *Dietary exposures* to go up one level
- At *Inputs*, click *Active substances (optional)* (path: *Exposures / Dietary exposures / Active substances*)
	- At *Inputs*, click *Points of departure* (path: *Exposures / Dietary exposures / Active substances / Points of departure*)
		- At *Points of departure data source* with **b** browse to the file *HazardDoses* Triazoles.xlsx
		- In the green navigation bar, click *Dietary exposures* to go up two levels
- At *Dietary exposure settings*, for *Dietary exposure calculation tier* select *EFSA 2012 Optimistic*, and press Ê *Save Changes*
- In the green navigation bar, click *Exposures* to go up one level
- At *Inputs*, click *Non-dietary exposures* (path: *Exposures / Non-dietary exposures*)
	- At *Non-dietary exposures data source* with **V** browse to the file *NonDietaryExposures.xlsx* and *Select*
- Now run the model, by pressing the  $\triangleright$  run icon in the grey bar.

Try to find the following results:

- 1. Exposure percentiles daily intakes with uncertainty bounds
- 2. Substance with highest contribution to the total exposure distribution
- 3. The food-as-measured with the highest contribution to the upper tail of the exposure distribution

## **4.2.4 Example 2**

In this example we will elaborate on the previous one with kinetic models.

- Go to the Actions settings  $\bullet$  of this action (path: *Exposures*)
	- At *Inputs*, click *Kinetic models (default)* (path: *Exposures / Kinetic models*)
		- At *Kinetic models data source* with  $\triangle$  browse to the file *UserGroupDemo-KineticModelsArtificial.xlsx* and *Select*
		- At *Kinetic model settings* for *Kinetic model* select *Cosmos Version 5*
- Now run the model, by pressing the  $\triangleright$  run icon in the green bar.

# **4.3 Hazard characterisations from PoDs**

## **4.3.1 Introduction**

The goal of this exercise is to try to establish hazard characterisations from PoDs (NOAELs).

# **4.3.2 Preparation**

If you haven't done so, in the workspace browser (use the  $\Box$  icon), create a new workspace named *Examples*, using the  $+$ .

The data files used in the example(s) in this section, are located in the data folder *Documentation-Examples / Exercise Hazard characterisations*.

## **4.3.3 Example 1**

In this example, Imazalil target dose from NOAEL will be calculated.

- In the *Examples* workspace, create a new action (using  $+$ )
	- Then select  $\checkmark$  *Show all action types, and select Hazard characterisations*
	- Name it *TargetDoseImazalil*
	- Use as Hazard characterization settings
		- Risk type: *Chronic*
		- Target level: *External*
	- Press Create
- Then go to the Actions settings  $\clubsuit$  of this action.
	- At *Scope*, click *Effects* (path: *Hazard characterisations / Effects*)
		- At *Effects data source* with **b** browse to the file *Effects and AOP Network Steatosis.xlsx* and *Select*
		- At *Effects selection* with  $\triangle$ 
			- Deselect everything by clicking  $\checkmark$  on the first line, next to the word *Code*
			- On the second page, select only *Steatosis-liver*, and **a** Save
		- At *Effect Settings* for *focal effect select Steatosis-liver* and press **B** Save Changes.
		- In the green navigation bar, click *Hazard characterisations* to go up one level
	- At *Scope*, click *Substances* (path: *Hazard characterisations / Substances*)
- At *Substances data source* with **b** browse to the file *TargetDosescalculation-Substances.xlsx* and *Select*
- At *Substances selection* with  $\bigtriangledown$ 
	- Deselect everything, by clicking the  $\checkmark$  on the first line, next to the word *Code*
	- Select only *Imazalil*, and  $\blacksquare$  Save
	- In the green navigation bar, click *Hazard characterisations* to go up one level
- At *Inputs*, click *Points of departure* (path: *Hazard characterisations / Points of departure*)
	- At *Points of departure data source* with **i** browse to the file *TargetDosesCalculation*-*HazardDoses.xlsx* and *Select*
	- In the green navigation bar, click *Hazard characterisations* to go up one level
- At *Hazard characterisations settings*, for *Expression type* select *NOAEL (convert all hazard characterisations as NOAELs)*
- At *Hazard characterisations settings*, Select  $\checkmark$  Use inter-species conversions
- At *Hazard characterisations settings*, Select  $\checkmark$  Use intra-species factors, and press  $\blacksquare$  Save Changes
- Now run the model, by pressing the  $\blacktriangleright$  run icon in the grey bar.

Try to find the following results:

- 1. The NOAEL for Imazalil used as point of departure.
- 2. The target hazard dose based on the default assessment factors 1/10 and 1/10 for inter-species and withinspecies conversion.

Answers:

- In the grey bar, browse to the results panel by clicking the  $\bullet$  icon and click on the latest output (path: *Results / TargetDoseImazalil*)
	- In the *Hazard characterisations* tab, browse in the tree (unfold by clicking  $\rightarrow$  where necessary) to  $\rightarrow$ *Available hazard characterisations*
		- The NOAEL for Imazalil is 40 µg/kg bw/day.

# **4.4 Health impact estimates**

# **4.4.1 Introduction**

The goal of this exercise is to assess a health impact estimate.

# **4.4.2 Preparation**

If you haven't done so, in the workspace browser (use the  $\Box$  icon), create a new workspace named *Examples*, using the  $+$ .

The data files used in the example(s) in this section, are located in the data folder *Documentation-Examples / Exercise Health Impact*.

## **4.4.3 Example 1**

- In the *Examples* workspace, create a new action (using  $+$ )
	- Then select  $\checkmark$  *Show all action types*, select *Risks*
	- Name it *Risks*
	- Press *Create*
- Then go to the Actions settings  $\bullet$  of this action (path: *Risks*)
	- At *Scope*, click *Effects* (path: *Risks / Effects*)
		- At *Effects data source* with  $\triangle$  browse to the file *Effects and AOP Network Steatosis.xlsx* and *Select*
		- At \* Effect settings<sup>\*</sup>, for *focal effect* select *Steatosis-liver* and press **a** *Save Changes*
		- In the green navigation bar, click *Risks* to go up one level
	- At *Scope*, click *Foods* (path: *Risks / Foods*)
		- At *Foods data source* with **\*** browse to the file *Foods.xlsx* and *Select*
		- In the green navigation bar, click *Risks* to go up one level
	- At *Scope*, click *Substances* (path: *Risks / Substances*)
		- At *Substances data source* with  $\triangle$  browse to the file *Substances.xlsx* and *Select*
		- At *Substance settings*, for *index substance* select *Cyproconazole*, and press **B** *Save Changes*
		- In the green navigation bar, click *Risks* to go up one level
	- At *Inputs*, click *Exposures* (path: *Risks / Exposures*)
		- At *Inputs*, click *Dietary exposures* (path: *Risks /Exposures / Dietary exposures*)
			- At *Inputs*, click *Consumptions by food measured* (path: *Risks /Exposures / Dietary exposures / Consumptions by modelled food*)
				- At *Inputs*, click *Consumptions* (path: *Risks /Exposures / Dietary exposures / Consumptions by modelled food / Consumptions*)
					- At *Consumptions data source* with  $\mathcal{\mathcal{E}}$  browse to the file *Consumptions.xlsx* and *Select*
					- At *Consumptions data selection* with  $\triangle$  open the food consumption surveys selection.
						- The file contains two surveys, but only one is allowed. So deselect everything by clicking  $\checkmark$  on the first line, next to the word *Code*
						- Now select *DNFCS* 2003 and press *Save* (the red warning  $\triangle$  should now be gone)
					- At *Consumptions settings* for *Food survey* select *DNFCS\_2003* and press **a** *Save Changes*
					- In the green navigation bar, click *Consumptions by modelled food* to go up one level
				- At *Inputs*, click *Food conversions* (path: *Risks /Exposures / Dietary exposures / Consumptions by modelled food / Food conversions*)
					- At *Inputs* click *Food as measured* (path: *Risks /Exposures / Dietary exposures / Consumptions by modelled food / Food conversions / Food as measured*)
						- At *Inputs*, click *Concentrations* (path: *Risks /Exposures / Dietary exposures / Consumptions by modelled food / Food conversions / Food as measured / Concentrations*)
							- At *Concentrations data source* with **i** browse to the file *UserGroupDemo-ConcentrationData.xlsx* and *Select*
							- In the green navigation bar, click *Food conversions* to go up two levels
- At *Inputs*, click *Food recipes* (path: *Risks /Exposures / Dietary exposures / Consumptions by modelled food / Food conversions / Food recipes*)
	- At *Food recipes data source* with **b** browse to the file *UserGroupDemo-FoodRecipes.xlsx* and *Select*
	- In the green navigation bar, click *Dietary exposures* to go up three levels
- At *Inputs*, click *Processing factors* (path: *Risks /Exposures / Dietary exposures / Processing factors*)
	- At *Processing factors data source* with **i** browse to the file *UserGroupDemo-ProcessingFactors.xlsx* and *Select*
	- In the green navigation bar, click *Risks* to go up three levels
- At *Inputs*, click *Hazard chararcterisations* (path: *Risks / Hazard chararcterisations*)
	- At *Inputs*, click *Active substances* (path: *Risks / Hazard chararcterisations / Active substances*)
		- At *Inputs*, click *Points of departure* (path: *Risks / Hazard chararcterisations / Active substances / Points of departure*)
			- At *Points of departure data source* with **b** browse to the file *UserGroupDemo-HazardDoses.xlsx* and *Select*
			- In the green navigation bar, click *Active substances* to go up one level
		- At 'Active substances' click  $\clubsuit$  Compute

# **4.5 Assessment group membership probabilities**

## **4.5.1 Introduction**

The goal of this exercise is to assess group membership probabilities.

# **4.5.2 Preparation**

If you haven't done so, in the workspace browser (use the  $\Box$  icon), create a new workspace named *Examples*, using the  $+$ .

The data files used in the example(s) in this section, are located in the data folder *Documentation-Examples / Exercise Dietary Exposure Assessment*.

## **4.5.3 Example 1**

- In the *Examples* workspace, create a new action (using  $+$ )
	- Then select *Dietary exposures*
	- Name it *Dietary exposures*
	- Use as Dietary exposures settings
		- Tier: *EFSA Guidance Optimistic*
		- Risk type *Chronic*
		- Select  $\checkmark$  *Cumulative*
	- Press *Create*
- Then go to the actions settings  $\bullet$  of this action (path: *Dietary exposures*)
- At *Scope*, click *Foods* (path: *Dietary exposures / Foods*)
	- At *Foods data source* with **b** browse to the file *UserGroupDemo-Foods.xlsx* and *Select*
	- In the green navigation bar, click *Dietary exposures* to go up one level
- At *Scope*, click *Substances* (path: *Dietary exposures / Substances*)
	- At *Substances data source* with **i** browse to the file *UserGroupDemo-Substances.xlsx* and *Select*
	- At *Substance settings* for *Index substance* select *Cyproconazole* and press  $\blacksquare$  *Save Changes*
	- In the green navigation bar, click *Dietary exposures* to go up one level
- At *Scope*, click *Effects* (path: *Dietary exposures / Effects*)
	- At *Effects data source* with **b** browse to the file *Effect Steatosis.xlsx* and *Select*
	- At *Effect Settings* for *focal effect select Steatosis-liver* and press **B** *Save Changes*
	- In the green navigation bar, click *Dietary exposures* to go up one level.
- At *Inputs*, click *Consumptions by modelled food* (path: *Dietary exposures / Consumptions by modelled food*)
	- At *Inputs*, click *Consumptions* (path: *Dietary exposures / Consumptions by modelled food / Consumptions*)
		- At *Consumptions data source* with **b** browse to the file *UserGroupDemo-Consumptions.xlsx* and *Select*
		- At *Consumption settings* for *Food survey* select *DNFCS\_2003* and press **D** *Save Changes*
		- In the green navigation bar, click *Consumptions by modelled food* to go up one level
	- At *Inputs*, click *Food conversions* (path: *Dietary exposures / Consumptions by modelled food / Food conversions*)
		- At *Inputs*, click *Foods as measured* (path: *Dietary exposures / Consumptions by modelled food / Food conversions / Foods as measured*)
			- At *Inputs*, click *Concentrations* (path: *Dietary exposures / Consumptions by modelled food / Food conversions / Foods as measured / Concentrations*)
				- At *Concentrations data source* with **b** browse to the file *UserGroupDemo-ConcentrationData.xlsx* and *Select*
				- In the green navigation bar, click *Food conversions* to go up two levels
		- At *Inputs*, click *Food recipes* (path: *Dietary exposures / Consumptions by modelled food / Food Food recipes*)
			- At *Food recipes data source*, with **b** browse to the file *UserGroupDemo-FoodRecipes.xlsx*
			- In the green navigation bar, click *Dietary exposures* to go up three levels
- At *Inputs*, click *Concentration models* (path: *Dietary exposures / Concentration models*)
	- At *Inputs*, click *Relative potency factors* (path: *Dietary exposures / Concentration models / Relative potency factors*)
		- At *Relative potency data source* with **b** browse to the file *UserGroupDemo-RelativePotencyFactors.xlsx* and *Select*
		- In the green navigation bar, click *Dietary exposures* to go up two levels
- At *Inputs*, click *Processing factors* (path: *Dietary exposures / Processing factors*)
	- At *Processing factors data source* with  $\bullet$  browse to the file *UserGroupDemo-ProcessingFactors.xlsx* and *Select*
	- In the green navigation bar, click *Dietary exposures* to go up one level
- At *Inputs*, click *Active substances (optional)* (path: *Dietary exposures / Active substances*)
- At *Inputs*, click *Points of departure* (path: *Dietary exposures / Active substances / Points of departure*)
	- At *Points of departure data source*, with  $\bullet$  browse to the file *HazardDoses Triazoles.xlsx*
	- In the green navigation bar, click *Dietary exposures* to go up two levels
- Now run the model, by pressing the  $\triangleright$  run icon in the grey bar.

Try to find the following results:

- 1. Exposure percentiles daily intakes
- 2. Substance with highest contribution to the total exposure distribution
- 3. The food-as-measured with the highest contribution to the upper tail of the exposure distribution

# **4.5.4 Example 2**

Repeat the run of the previous task, but instead of the nominal run, now do an uncertainty analysis loop.

- Click on the  $\pm \infty$  icon (in the grey bar) to open the uncertainty settings panel, and check  $\checkmark$  *Perform uncertainty analysis*
	- For *Monte Carlo iterations per uncertainty run* choose 100, and press **a** *Save Changes*
- Now run the model, by pressing the  $\blacktriangleright$  run icon in the grey bar.

Compare with the previous results, to find:

- 1. Exposure percentiles daily intakes with uncertainty bounds
- 2. Substance with highest contribution to the total exposure distribution
- 3. The food-as-measured with the highest contribution to the upper tail of the exposure distribution

# **FIVE**

# **PUBLICATIONS USING MCRA**

- European Food Safety Authority (EFSA), P.S. Craig, B. Dujardin, A. Hart, A.F. Hernandez-Jerez, S. Hougaard Bennekou, C. Kneuer, B. Ossendorp, R. Pedersen, G. Wolterink, and L. Mohimont. Cumulative dietary risk characterisation of pesticides that have chronic effects on the thyroid. *EFSA Journal*, 18(4):e06088, 2020. URL: https://efsa.onlinelibrary.wiley.com/doi/abs/10.2903/j.efsa.2020.6088.
- European Food Safety Authority (EFSA), P.S. Craig, B. Dujardin, A. Hart, A.F. Hernández-Jerez, S. Hougaard Bennekou, C. Kneuer, B. Ossendorp, R. Pedersen, G. Wolterink, and L. Mohimont. Cumulative dietary risk characterisation of pesticides that have acute effects on the nervous system. *EFSA Journal*, 18(4):e06087, 2020. URL: [https://efsa.onlinelibrary.wiley.com/doi/abs/10.2903/j.e](https://efsa.onlinelibrary.wiley.com/doi/abs/10.2903/j.efsa.2020.6088)fsa.2020.6087.
- A. Beronius, J. Zilliacus, A. Hanberg, M. Luijten, van der Voet, H, and J. van Klaveren. Methodology for health risk assessment of combined exposures to multiple chemicals. *Food and Chemical Toxicology*, pages 111520, July 2020. URL: h[ttps://doi.org/10.1016/j.fct.2020.111520.](https://efsa.onlinelibrary.wiley.com/doi/abs/10.2903/j.efsa.2020.6087)
- J. Cotterill, N. Price, E. Rorije, and A. Peijnenburg. Development of a QSAR model to predict hepatic steatosis using freely available machine learning tools. *Food and Chemical Toxicology*, 142:111494, August 2020. URL: https://doi.org/10.1016/j.f[ct.2020.111494.](https://doi.org/10.1016/j.fct.2020.111520)
- B.C. Fischer, S. Rotter, J. Schubert, P. Marx-Stoelting, and R. Solecki. Recommendations for international harmonisation, implementation and further development of suitable scientific approaches regarding the assessment of mixture effects. *Food and Chemical Toxicology*, 141:111388, July 2020. URL: https: [//doi.org/10.1016/j.fct.2020.111388.](https://doi.org/10.1016/j.fct.2020.111494)
- C. Karrer, M. Andreassen, N. von Goetz, F. Sonnet, A.K. Sakhi, K. Hungerbühler, H. Dirven, and T. Husøy. The EuroMix human biomonitoring study: source-to-dose modeling of cumulative and aggregate exposure for the bisphenols BPA, BPS, and BPF and comparison with measured urinary levels. *Environment Internat[ional](https://doi.org/10.1016/j.fct.2020.111388)*, [136:105397, March 2020. URL:](https://doi.org/10.1016/j.fct.2020.111388) https://doi.org/10.1016/j.envint.2019.105397.
- M.C. Kennedy, A.D.M. Hart, J.W. Kruisselbrink, M. van Lenthe, W.J. de Boer, H. van der Voet, E. Rorije, C. Sprong, and J. van Klaveren. A retain and refine approach to cumulative risk assessment. *Food and Chemical Toxicology*, 138:111223, April 2020. URL: [https://doi.org/10.1016/j.fct.2020.1](https://doi.org/10.1016/j.envint.2019.105397)11223.
- C. Sprong, A. Crépet, F. Metruccio, U. Blaznik, C. Anagnostopoulos, D.L. Christodoulou, B.H. Jensen, M. Kennedy, N. González, I. Rehurkova, J. Ruprich, J.D. te Biesebeek, M. Vanacker, A. Moretto, and J. van Klaveren. Cumulative dietary risk assessment overarching different regulatory silos using a margin of exposure approach: a case study with three chemical silos. *[Food and Chemical Toxicology](https://doi.org/10.1016/j.fct.2020.111223)*, 142:111416, August 2020. URL: https://doi.org/10.1016/j.fct.2020.111416.
- C. Tebby, H. van der Voet, G. de Sousa, E. Rorije, V. Kumar, W. de Boer, J.W. Kruisselbrink, F.Y. Bois, M. Faniband, A. Moretto, and C. Brochot. A generic PBTK model implemented in the MCRA platform: predictive performance and uses in risk assessment of chemicals. *Food and Chemical Toxicology*, 142:111440, August 2020. URL: [https://doi.org/10.1016/j.fct.2020.](https://doi.org/10.1016/j.fct.2020.111416)111440.
- A.D. van den Brand, M. Beukers, M. Niekerk, G. van Donkersgoed, M. van der Aa, B. van de Ven, A. Bulder, H. van der Voet, and C.R. Sprong. Assessment of the combined nitrate and nitrite exposure from food and drinking water: application of uncertainty around the nitrate to nitrite conversion factor. *Food Additives & Contaminants: Part A*, 3[7\(4\):568–582, January 2020. URL:](https://doi.org/10.1016/j.fct.2020.111440) https://doi.org/10.1080/19440049.2019.1707294.
- H. van der Voet, J.W. Kruisselbrink, W.J. de Boer, M.S. van Lenthe, J.J.B. van den Heuvel, A. Crépet, M.C. Kennedy, J. Zilliacus, A. Beronius, C. Tebby, C. Brochot, C. Luckert, A. Lampen, E. Rorije, C. Sprong, and J.D. van Klaveren. The MCRA toolbox of models and data to support chemical mixture risk assessment. *Food and Chemical Toxicology*, 138:111185, April 2020. URL: https://doi.org/10.1016/j.fct.2020.111185.
- M. Vanacker, P. Quindroit, K. Angeli, C. Mandin, P. Glorennec, C. Brochot, and A. Crépet. Aggregate and cumulative chronic risk assessment for pyrethroids in the French adult population. *Food and Chemical Toxicology*, 143:111519, September 2020. URL: https://d[oi.org/10.1016/j.fct.2020.111519.](https://doi.org/10.1016/j.fct.2020.111185)
- C. Vlachou, D. Hofstädter, E. Rauscher-Gabernig, A. Griesbacher, K. Fuchs, and J. König. Risk assessment of nitrites for the Austrian adult population with probabilistic modelling of the dietary exposure. *Food and Chemical Toxicology*, 143:111480, September 2020. URL: [https://doi.org/10.1016/j.fct.202](https://doi.org/10.1016/j.fct.2020.111519)0.111480.

- F.Y. Bois, C. Tebby, and C. Brochot. EuroMix PBPK m[odel for combined exposures. 2019. URL](https://doi.org/10.1016/j.fct.2020.111480): https: //zenodo.org/record/2532334.
- A. Boobis. Report of EuroMix workshops on international harmonisation on the risk assessment of combined exposure to multiple chemicals. 2019. URL: https://zenodo.org/record/3479150.
- [P.E. Boon, M. Van Der Aa, A](https://zenodo.org/record/2532334). Dusseldorp, P. Janssen, M.J. Zeilmaker, and S. Schulpen. Loodinna[me via](https://zenodo.org/record/2532334) kraanwater: blootstellingsschatting en risicobeoordeling voor diverse risicogroepen. RIVM Letter report 2019- 0090, 2019. URL: https://rivm.openreposito[ry.com/handle/10029/623516.](https://zenodo.org/record/3479150)
- P.E. Boon, G. Van Donkersgoed, W. Van Der Vossen, M. Sam, M.Y. Noordam, and H. Van Der Schee. Tussenevaluatie van de nota 'gezonde groei, duurzame oogst'. RIVM Letter report 2018-0127, 2019. URL: https://rivm.openr[epository.com/handle/10029/623125.](https://rivm.openrepository.com/handle/10029/623516)
- P.E. Boon, M.J. Zeilmaker, and M.J.B. Mengelers. Risicobeoordeling van GenX en PFOA in moestuingewassen in helmond. RIVM Letter report 2019-0024, 2019. URL: https://rivm.openrepository.com/handle/ [10029/622988.](https://rivm.openrepository.com/handle/10029/623125)
- A. Crépet, M. Vanacker, C. Sprong, W. de Boer, U. Blaznik, M. Kennedy, C. Anagnostopoulos, D.L. Christodoulou, J. Ruprich, I. Rehurkova, J.L. Domingo, B.H. Jensen, F. Metruccio, A. Moretto, L. Jacxsens, P. Spanoghe, D. Senaeve, H. van der Voet, and J. van Klaveren. [Selecting mixtures on the basis of di](https://rivm.openrepository.com/handle/10029/622988)[etary exposure](https://rivm.openrepository.com/handle/10029/622988) and hazard data: application to pesticide exposure in the European population in relation to steatosis. *International Journal of Hygiene and Environmental Health*, 222(2):291–306, March 2019. URL: https://doi.org/10.1016/j.ijheh.2018.12.002.
- J. de Rop, D. Senaeve, L. Jacxsens, M. Houbraken, J. van Klaveren, and P. Spanoghe. Cumulative probabilistic risk assessment of triazole pesticides in Belgium from 2011-2014. *Food Additives & Contaminants: Part A*, [36\(6\):911–921, April 2019. URL:](https://doi.org/10.1016/j.ijheh.2018.12.002) https://doi.org/10.1080/19440049.2019.1606943.
- B. Fischer, J. Schubert, S. Rotter, and R. Solecki. Specific recommendations regarding implementation of mechanism-based test strategy for harmonised cumulative risk assessment according oecd, who, efsa and EuroMix guidance. 2019. URL: https[://zenodo.org/record/3490547.](https://doi.org/10.1080/19440049.2019.1606943)
- G. Heinemeyer, M. Jantunen, and P. Hakkinen. *The Practice of Consumer Exposure Assessment*. Springer International Publishing, 2019. URL: https://doi.org/10.1007/978-3-319-96148-4.
- C. Karrer, W. de Boer, C. De[lmaar, Y. Cai, A. Crépet, K. Hunge](https://zenodo.org/record/3490547)rbühler, and N. von Goetz. Linking probabilistic exposure and pharmacokinetic modeling to assess the cumulative risk from the bisphenols BPA, BPS, BPF, and BPAF for Europeans. *Envi[ronmental Science & Technology](https://doi.org/10.1007/978-3-319-96148-4)*, 53(15):9181–9191, July 2019. URL: https://doi.org/10.1021/acs.est.9b01749.
- M. Kennedy, A. Hart, J.W. Kruisselbrink, M. van Lenthe, W. de Boer, H. van der Voet, E. Rorije, C. Sprong, and J. van Klaveren. Methodology and results of the retain and refine approach. 2019. URL: https://zenodo. [org/record/3465690.](https://doi.org/10.1021/acs.est.9b01749)
- M.C. Kennedy, D.G. Garthwaite, W.J. de Boer, and J.W. Kruisselbrink. Modelling aggregate exposure to pesticides from dietary and crop spray sources in UK residents. *Environmental Science and Poll[ution Research](https://zenodo.org/record/3465690)*, [26\(10\):9892–9907, F](https://zenodo.org/record/3465690)ebruary 2019. URL: https://doi.org/10.1007/s11356-019-04440-7.
- A.E. Kolbaum, K. Berg, F. Müller, O. Kappenstein, and O. Lindtner. Dietary exposure to elements from the German pilot total diet study (TDS). *Food Additives & Contaminants: Part A*, 36(12):1822–1836, October 2019. URL: https://doi.org/10.1080/19440049.2019.1668967.
- B. Sachse, A.E. Kolbaum, R. Ziegenhagen, S. Andres, K. Berg, B. Dusemund, K.I. Hirsch-Ernst, O. Kappenstein, F. Müller, C. Röhl, O. Lindtner, A. Lampen, and B. Schäfer. Dietary manganese exposure in the adult population in Germany—what does it mean in relation to health risks? *Molecular Nutrition & Food Research*, 63(16):1900065, July 2019. URL: [https://doi.org/10.1002/mn](https://doi.org/10.1080/19440049.2019.1668967)fr.201900065.
- T. Tietz, A. Lenzner, A.E. Kolbaum, S. Zellmer, C. Riebeling, R. Gürtler, C. Jung, O. Kappenstein, J. Tentschert, M. Giulbudagian, S. Merkel, R. Pirow, O. Lindtner, T. Tralau, B. Schäfer, P. Laux, M. Greiner, A. Lampen, A. Luch, R. Wittkowski, and A. Hensel. Aggregated aluminium exposure: risk assessment for the general population. *Archives of Toxicology*[, 93\(12\):3503–3521, October](https://doi.org/10.1002/mnfr.201900065) 2019. URL: https://doi.org/10. 1007/s00204-019-02599-z.
- H. van der Voet, J.W. Kruisselbrink, W.J. de Boer, M.S. van Lenthe, J.J.B. van den Heuvel, A. Crépet, M.C. Kennedy, J. Zilliacus, A. Beronius, E. Rorije, C. Sprong, and J.D. van Klaveren. The Euro[Mix model toolbox](https://doi.org/10.1007/s00204-019-02599-z) [MCRA 9. 2019. URL:](https://doi.org/10.1007/s00204-019-02599-z) https://zenodo.org/record/3462181.
- H. van der Voet, J.W. Kruisselbrink, W.J. de Boer, M.S. van Lenthe, J.J.B. van den Heuvel, A. Crépet, M.C. Kennedy, J. Zilliacus, A. Beronius, C. Tebby, C. Brochot, E. Rorije, C. Sprong, and J.D. van Klaveren. Draft paper on the EuroMixt[oolbox of models and data to suppo](https://zenodo.org/record/3462181)rt chemical mixture risk assessment. 2019. URL: https://zenodo.org/record/3474943.
- J.D. van Klaveren, J.W. Kruisselbrink, W.J. de Boer, G. van Donkersgoed, J.D. te Biesebeek, M. Sam, and H. van der Voet. Cumulative dietary exposure assessment of pesticides that have acute effects on the nervous system using MCRA software. *EFSA Supporting Publications*, 16(9):1708E, 2019. URL: https: [//efsa.onlinelibrary.wiley.com/doi/a](https://zenodo.org/record/3474943)bs/10.2903/sp.efsa.2019.EN-1708.
- J.D. van Klaveren, J.W. Kruisselbrink, W.J. de Boer, G. van Donkersgoed, J.D. te Biesebeek, M. Sam, and H. van der Voet. Cumulative dietary exposure assessment of pesticides that have chronic effects on the thyroid using MCRA software. *EFSA Supporting Publications*, 16(9):1707E, 2019. URL: https://efsa.onlineli[brary.](https://efsa.onlinelibrary.wiley.com/doi/abs/10.2903/sp.efsa.2019.EN-1708) [wiley.com/doi/abs/10.2903/sp.efsa.2019.EN-1707.](https://efsa.onlinelibrary.wiley.com/doi/abs/10.2903/sp.efsa.2019.EN-1708)
- M.S. van Lenthe, W.J. de Boer, J.W. Kruisselbrink, H. van der Voet, A. Crépet, M. Vanacker, and L. Trocellier. Validation of the EuroMix model toolbox and comparison with us software. 2019. URL: [https://zenodo.org/](https://efsa.onlinelibrary.wiley.com/doi/abs/10.2903/sp.efsa.2019.EN-1707) [record/3467409.](https://efsa.onlinelibrary.wiley.com/doi/abs/10.2903/sp.efsa.2019.EN-1707)
- J. Zilliacus, A. Beronius, A. Hanberg, M. Luijten, J. van Klaveren, and H. van der Voet. EuroMix handbook for mixture risk assessment. 2019. URL: https://zenodo.org/record/3560719.
- [J. Zilliacus, E. R](https://zenodo.org/record/3467409)orije, M. Kennedy, and J. van Klaveren. Proceedings and training m[aterial from second](https://zenodo.org/record/3467409) training session for stakeholders. 2019. URL: https://zenodo.org/record/3560731.

- P.E. Boon, J.D. Te Biesebeek, H. Brants, M.C[. Bouwmeester, and E.V.S. Hessel. D](https://zenodo.org/record/3560731)ietary sources of exposure to bisphenol A in the Netherlands. RIVM Letter report 2017-0187, 2018. URL: http://rivm.openrepository. com/rivm/handle/10029/621792.
- P.E. Boon, G. Van Donkersgoed, J.D. Te Biesebeek, G. Wolterink, and A.G. Rietveld. Cumulative exposure to residues of plant protection products via food in the Netherlands. RIVM Lett[er report 2017-0018, 2018.](http://rivm.openrepository.com/rivm/handle/10029/621792) URL: [http://rivm.openrepository](http://rivm.openrepository.com/rivm/handle/10029/621792).com/rivm/handle/10029/622169.
- Jardim, A.N.O, D.C. Mello, A.P. Brito, H. van der Voet, P.E. Boon, and E.D. Caldas. Probabilistic dietary risk assessment of triazole and dithiocarbamate fungicides for the Brazilian population. *Food and Chemical Toxicology*[, 118:317–327, August 2018. URL:](http://rivm.openrepository.com/rivm/handle/10029/622169) https://doi.org/10.1016/j.fct.2018.05.002.
- Jardim, A.N.O, D.C. Mello, A.P. Brito, G. van Donkersgoed, P.E. Boon, and E.D. Caldas. Dietary cumulative acute risk assessment of organophosphorus, carbamates and pyrethroids insecticides for the Brazilian population. *Food and Chemical Toxicology*, 112:108–[117, February 2018. URL:](https://doi.org/10.1016/j.fct.2018.05.002) https://doi.org/10.1016/j.fct.2017. 12.010.
- M. Mengelers, J.D. Te Biesebeek, M. Schipper, W. Slob, and P.E. Boon. Risicobeoordeling van GenX en PFOA in moestuingewassen in Dordrecht, Papendrecht en Sliedrecht. RIVM Letter report 2017-0017, 2018. URL: http://rivm.openrepository.com/rivm/handle/10029/621785.
- S. Rotter, A. Beronius, A.R. Boobis, A. Hanberg, J. van Klaveren, M. Luijten, K. Machera, D. Nikolopoulou, H. van der Voet, J. Zilliacus, and R. Solecki. Overview on legislation and scientific approaches for risk assessment of combined exposure to multiple chemicals: the potential EuroMix contribution. *Critical Reviews in Toxicology*[, 48\(9\):796–814, October 2018. URL:](http://rivm.openrepository.com/rivm/handle/10029/621785) https://doi.org/10.1080/10408444.2018.1541964.
- J. Suomi, P. Tuominen, S. Niinistö, S.M. Virtanen, and K. Savela. Dietary heavy metal exposure of Finnish children of 3 to 6 years. *Food Additives & Contaminants: Part A*, 35(7):1305–1315, June 2018. URL: https: //doi.org/10.1080/19440049.2018.1480065.
- B.M. van De Ven, S. Fragki, J.D. te Biesebeek, A.[G. Rietveld, and P.E. Boon. Mineral oils in food; a](https://doi.org/10.1080/10408444.2018.1541964) review of toxicological data and an assessment of the dietary exposure in the Netherlands. RIVM Letter report [2017-](https://doi.org/10.1080/19440049.2018.1480065) 0018, 2018. URL: [http://rivm.openrepositor](https://doi.org/10.1080/19440049.2018.1480065)y.com/rivm/handle/10029/622044.

- P.E. Boon, J.D. te [Biesebeek, and G. van Donkersgoed. Dietary exposure to lea](http://rivm.openrepository.com/rivm/handle/10029/622044)d in the Netherlands. RIVM Letter report 2016-0206, 2017. URL: https://www.rivm.nl/bibliotheek/rapporten/2016-0206.pdf.
- K. Presser, C. Zoom, J. Szymanek, and G. Zappa. Development of a pilot service for the electronic infrastructure of METROFOOD-RI. In *Proceedings of 3rd IMEKOFOODS Conference: Metrology Promoting Harmonization and Standardization in Food and Nutrition*. International Measurement Confederation, 2017. URL: https://imeko.org/publications/tc23-2[017/IMEKO-TC23-2017-045.pdf.](https://www.rivm.nl/bibliotheek/rapporten/2016-0206.pdf)
- C. Sieke, B. Michalski, and T. Kuhl. Probabilistic dietary risk assessment of pesticide residues in foods for the German population based on food monitoring data from 2009 to 2014. *Journal of Exposure Science & Environmental Epidemiology*[, 28\(1\):46–54, July 2017. URL:](https://imeko.org/publications/tc23-2017/IMEKO-TC23-2017-045.pdf) https://doi.org/10.1038/jes.2017.7.
- R.C. Sprong, E.M. Niekerk, and M.H. Beukers. Intake assessment of the food additives nitrites (e 249 and e 250) and nitrates (e 251 and e 252). RIVM Letter report 2016-0208, 2017. URL: https://www.rivm.nl/ bibliotheek/rapporten/2016-0208.pdf.

- [P.E. Boon and J.D. te Biesebeek. Preli](https://www.rivm.nl/bibliotheek/rapporten/2016-0208.pdf)minary assessment of dietary exposure to 3-MCPD in the Netherlands. RIVM Letter report 2015-0199, 2016. URL: https://www.rivm.nl/bibliotheek/rapporten/2015-0199.pdf.
- P.E. Boon, J.D. te Biesebeek, S.P.J. van Leeuwen, M.J. Zeilmaker, and L.A.P. Hoogenboom. Dietary exposure to polybrominated diphenyl ethers in the Netherlands. RIVM Letter report 2016-0037, 2016. URL: https: //www.rivm.nl/bibliotheek/rapporten/2016-0[037.pdf.](https://www.rivm.nl/bibliotheek/rapporten/2015-0199.pdf)
- C Rompelberg, M.B. Heringa, G. van Donkersgoed, J. Drijvers, A. Roos, S. Westenbrink, R. Peters, G. van Bemmel, W. Brand, and A.G. Oomen. Oral intake of added titanium dioxide and its nanofraction from food products, food supplements and toothpaste by the Dutch population. *Nanotoxicology*, 10(10):1404–[1414,](https://www.rivm.nl/bibliotheek/rapporten/2016-0037.pdf) September 2016. URL: [https://doi.org/10.1080/17435](https://www.rivm.nl/bibliotheek/rapporten/2016-0037.pdf)390.2016.1222457.
- R.C. Sprong, L. de Wit-Bos, J.D. te Biesebeek, M. Alewijn, P. Lopez, and M.J.B. Mengelers. A mycotoxindedicated total diet study in the Netherlands in 2013: part III – exposure and risk assessment. *World Mycotoxin Journal*, 9(1):109–128, February 2016. URL: [https://doi.org/10.3920/wm](https://doi.org/10.1080/17435390.2016.1222457)j2015.1905.
- C.L. Stephenson and C.A. Harris. An assessment of dietary exposure to glyphosate using refined deterministic and probabilistic methods. *Food and Chemical Toxicology*, 95:28–41, September 2016. URL: https://doi.org/ 10.1016/j.fct.2016.06.026.
- H. van der Voet, W.J. de Boer, J.W. Kruisselbr[ink, G. van Donkersgoed, and J.D. van K](https://doi.org/10.3920/wmj2015.1905)laveren. MCRA made scalable for large cumulative assessment groups. *EFSA Supporting Publications*, 13(1):910[E, 2016. URL:](https://doi.org/10.1016/j.fct.2016.06.026) [https://efsa.onlinelibrary.w](https://doi.org/10.1016/j.fct.2016.06.026)iley.com/doi/abs/10.2903/sp.efsa.2016.EN-910.
- **2015**
	- Y. Akhandaf, J. van Klaveren, S. de Henauw, G. van Donkersgoed, T. van Gorcum, A. Papadopoulos, V. Sirot, M. Kennedy, H. Pinchen, J. Ruprich, I. Rehurkova, G. Perelló, and I. Sioen. Exposure assessment within a total diet study: a comparison of the use of the pan-European classification system FoodEx-1 with national food classification systems. *Food and Chemical Toxicology*, 78:221–229, April 2015. URL: https://doi.org/ 10.1016/j.fct.2015.01.019.
	- U. Blaznik, A. Yngve, I. Eržen, and C.H. Ribič. Consumption of fruits and vegetables and probabilistic assessment of the cumulative acute exposure to organophosphorus and carbamate pesticides of schoolchildren in Slovenia. *Public Health Nutrition*, 19(3):557–563, May 2015. URL: https://d[oi.org/10.1017/](https://doi.org/10.1016/j.fct.2015.01.019) [s1368980015001494.](https://doi.org/10.1016/j.fct.2015.01.019)
	- P.E. Boon and H. Van der Voet. Probabilistic dietary exposure models relevant for acute and chronic exposure assessment of adverse chemicals via food. RIVM Letter report 2015-0191, 2015. URL: [https://www.rivm.nl/](https://doi.org/10.1017/s1368980015001494) [bibliotheek/rapporten](https://doi.org/10.1017/s1368980015001494)/2015-0191.pdf.
	- P.E. Boon, G. van Donkersgoed, D. Christodoulou, A. Crépet, L. D'Addezio, V. Desvignes, B. Ericsson, F. Galimberti, E. Ioannou-Kakouri, B.H. Jensen, I. Rehurkova, J. Rety, J. Ruprich, S. Sand, C. Stephenson, A. Strömberg, A. Turrini, H. van der Voet, P. Ziegler, P. Hamey, and J.D. van Klavere[n. Cumulative dietary](https://www.rivm.nl/bibliotheek/rapporten/2015-0191.pdf) [exposure to a selected group of pestic](https://www.rivm.nl/bibliotheek/rapporten/2015-0191.pdf)ides of the triazole group in different European countries according to the EFSA guidance on probabilistic modelling. *Food and Chemical Toxicology*, 79:13–31, May 2015. URL: https://doi.org/10.1016/j.fct.2014.08.004.
	- D. He, X. Ye, Y. Xiao, N. Zhao, J. Long, P. Zhang, Y. Fan, S. Ding, X. Jin, C. Tian, S. Xu, and C. Ying. Dietary exposure to endocrine disrupting chemicals in metropolitan population from China: a risk assessment based on probabilistic approach. *Chemosphere*, 139:2–8, November 2015. URL: https://doi.org/10.1016/j. [chemosphere.2015.05.036.](https://doi.org/10.1016/j.fct.2014.08.004)
	- R. Jacobs, H. van der Voet, and C.J.F. ter Braak. Integrated probabilistic risk assessment for nanoparticles: the case of nanosilica in food. *Journal of Nanoparticle Research*, June 2015. URL: [https://doi.org/10.1007/](https://doi.org/10.1016/j.chemosphere.2015.05.036) [s11051-015-2911-y.](https://doi.org/10.1016/j.chemosphere.2015.05.036)
	- M.C. Kennedy, C.R. Glass, B. Bokkers, A.D.M. Hart, P.Y. Hamey, J.W. Kruisselbrink, W.J. de Boer, H. van der Voet, D.G. Garthwaite, and J.D. van Klaveren. A European model and case studies for aggregate exposure assessment of pesticides. *Food and Chemical Toxicology*, 79:32[–44, May 2015. URL:](https://doi.org/10.1007/s11051-015-2911-y) [https://doi.org/10.10](https://doi.org/10.1007/s11051-015-2911-y)16/j.fct.2014.09.009.
	- M.C. Kennedy, C.R. Glass, S. Fustinoni, A. Moretto, S. Mandic-Rajcevic, P. Riso, A. Turrini, H. van der Voet, M.T. Hetmanski, R.J. Fussell, and J.D. van Klaveren. Testing a cumulative and aggregate exposure model using biomonitoring studies and dietary records for Italian vineyard spray operators. *Food and Chemical Toxicology*, [79:45–53, May 2015. URL:](https://doi.org/10.1016/j.fct.2014.09.009) https://doi.org/10.1016/j.fct.2014.12.012.
	- M.C. Kennedy, H. van der Voet, V.J. Roelofs, W. Roelofs, C.R. Glass, W.J. de Boer, J.W. Kruisselbrink, and A.D.M. Hart. New approaches to uncertainty analysis for use in aggregate and cumulative risk assessment of pesticides. *Food and Chemical Toxicology*[, 79:54–64, May 2015. UR](https://doi.org/10.1016/j.fct.2014.12.012)L: https://doi.org/10.1016/j.fct.2015. 02.008.
	- F.R. Mancini, V. Sirot, L. Busani, J.L. Volatier, and M. Hulin. Use and impact of usual intake models on dietary exposure estimate and risk assessment of chemical substances: a practical example for cadmium, acrylamide and sulphites. *Food Additives & Contaminants: Part A*, 32(7):1065–1074[, May 2015. URL:](https://doi.org/10.1016/j.fct.2015.02.008) https://doi.org/ [10.108](https://doi.org/10.1016/j.fct.2015.02.008)0/19440049.2015.1041428.
	- R.C. Sprong and P.E. Boon. Dietary exposure to cadmium in the Netherlands. RIVM Letter report 2015-0085, 2015. URL: https://www.rivm.nl/bibliotheek/rapporten/2015-0085.pdf.
	- [J. Suomi, J. Ranta, P. Tuominen, T](https://doi.org/10.1080/19440049.2015.1041428). Putkonen, C. Bäckman, M.L. Ovaskainen, S.M. Virtanen, [and K. Savela.](https://doi.org/10.1080/19440049.2015.1041428) Quantitative risk assessment on the dietary exposure of Finnish children and adults to nitrite. *Food Additives & Contaminants: Part A*[, 33\(1\):41–53, November 2015. URL:](https://www.rivm.nl/bibliotheek/rapporten/2015-0085.pdf) https://doi.org/10.1080/19440049.2015. 1117145.
	- H. van der Voet, W.J. de Boer, J.W. Kruisselbrink, P.W. Goedhart, G.W.A.M. van der Heijden, M.C. Kennedy, P.E. Boon, and J.D. van Klaveren. The MCRA mode[l for probabilistic single-compound and](https://doi.org/10.1080/19440049.2015.1117145)

cumulative risk assessment of pesticides. *Food and Chemical Toxicology*, 79:5–12, May 2015. URL: https://doi.org/10.1016/j.fct.2014.10.014.

• J.D. van Klaveren, M.C. Kennedy, A. Moretto, W. Verbeke, H. van der Voet, and P.E. Boon. The ACROP-OLIS project: its aims, achievements, and way forward. *Food and Chemical Toxicology*, 79:1–4, May 2015. URL: [https://doi.org/10.1016/j.fct.2015.0](https://doi.org/10.1016/j.fct.2014.10.014)3.006.

## **2014**

- P.E. [Boon. Estimation of the acute dietary exp](https://doi.org/10.1016/j.fct.2015.03.006)osure to pesticides using the probabilistic approach and the point estimate methodology. *European Journal of Nutrition & Food Safety*, 4(1):1–3, January 2014. URL: https://doi.org/10.9734/ejnfs/2014/6899.
- P.E. Boon, J.D. te Biesebeek, L. de Wit, and G. van Donkersgoed. Dietary exposure to dioxins in the Netherlands. RIVM Letter report 2014-0001, 2014. URL: https://www.rivm.nl/bibliotheek/rapporten/2014-0001. [pdf.](https://doi.org/10.9734/ejnfs/2014/6899)
- P.E. Boon, H. van der Voet, J. Ruprich, A. Turrini, S. Sand, and J.D. van Klaveren. Computational tool for usual intake modelling workable at the European level. *[Food and Chemical Toxicology](https://www.rivm.nl/bibliotheek/rapporten/2014-0001.pdf)*, 74:279–288, December [201](https://www.rivm.nl/bibliotheek/rapporten/2014-0001.pdf)4. URL: https://doi.org/10.1016/j.fct.2014.10.019.
- H. van der Voet, J.W. Kruisselbrink, W.J. Boer, and P.E. Boon. Model-then-add: usual intake modelling of multimodal intake distributions. RIVM Letter report 090133001/2014, 2014. URL: http://hdl.handle.net/ 10029/3143[61.](https://doi.org/10.1016/j.fct.2014.10.019)

## **2013**

- [A.J.C. Roodenb](http://hdl.handle.net/10029/314361)urg, A.J. van Ballegooijen, M. Dötsch-Klerk, H. van der Voet, and J.C. Seidell. Modelling of usual nutrient intakes: potential impact of the Choices programme on nutrient intakes in young Dutch adults. *PLoS ONE*, 8(8):e72378, August 2013. URL: https://doi.org/10.1371/journal.pone.0072378.
- E.H.M. Temme, H. van der Voet, J.T.N.M. Thissen, J. Verkaik-Kloosterman, G. van Donkersgoed, and S. Nonhebel. Replacement of meat and dairy by plant-derived foods: estimated effects on land use, iron and SFA intakes in young Dutch adult females. *Public Health Nutrition*[, 16\(10\):1900–1907, Februa](https://doi.org/10.1371/journal.pone.0072378)ry 2013. URL: https://doi.org/10.1017/s1368980013000232.

- [P.E. Boon, J.D. te Biesebeek, I. Sioen, I. Hu](https://doi.org/10.1017/s1368980013000232)ybrechts, J. Moschandreas, J. Ruprich, A. Turrini, M. Azpiri, L. Busk, T. Christensen, M. Kersting, L. Lafay, K.-H. Liukkonen, S. Papoutsou, L. Serra-Majem, I. Traczyk, S. de Henauw, and J.D. van Klaveren. Long-term dietary exposure to lead in young European children: comparing a pan-European approach with a national exposure assessment. *Food Additives & Contaminants: Part A*, 29(11):1701–1715, November 2012. URL: https://doi.org/10.1080/19440049.2012.709544.
- P.W. Goedhart, H. van der Voet, S. Knüppel, A.L.M. Dekkers, K.W. Dodd, H. Boeing, and J. van Klaveren. A comparison by simulation of different methods to estimate the usual intake distribution for episodically consumed foods. *EFSA Supporting Publications*[, 9\(6\):299E, 2012. URL:](https://doi.org/10.1080/19440049.2012.709544) https://efsa.onlinelibrary.wiley.com/ doi/abs/10.2903/sp.efsa.2012.EN-299.
- I. Sioen, T. Fierens, M. van Holderbeke, L. Geerts, M. Bellemans, M. de Maeyer, K. Servaes, G. Vanermen, P.E. Boon, and S. de Henauw. Phthalates dietary exposure and food sources for Belgian preschool children and adults. *Environment International*, 48:102–108, November 2012. URL: [https://doi.org/10.1016/j.envint.](https://efsa.onlinelibrary.wiley.com/doi/abs/10.2903/sp.efsa.2012.EN-299) [2012.07.004.](https://efsa.onlinelibrary.wiley.com/doi/abs/10.2903/sp.efsa.2012.EN-299)
- J.D. van Klaveren, P.W. Goedhart, D. Wapperom, and H. van der Voet. A European tool for usual intake distribution estimation in relation to data collection by EFSA. *EFSA Supporting Publications*[, 9\(6\):300E, 2012.](https://doi.org/10.1016/j.envint.2012.07.004) URL: [https://](https://doi.org/10.1016/j.envint.2012.07.004)efsa.onlinelibrary.wiley.com/doi/abs/10.2903/sp.efsa.2012.EN-300.

• EFSA Panel on Plant Protection Products and their Residues (PPR). Guidance on the use of probabilistic methodology for modelling dietary exposure to pesticide residues. *EFSA Journal*, 10(10):2839, 2012. URL: https://efsa.onlinelibrary.wiley.com/doi/abs/10.2903/j.efsa.2012.2839.

## **2011**

- [P.E. Boon, M. Bonthuis, H. van der Voet, and J.D. van Klaveren. Com](https://efsa.onlinelibrary.wiley.com/doi/abs/10.2903/j.efsa.2012.2839)parison of different exposure assessment methods to estimate the long-term dietary exposure to dioxins and ochratoxin A. *Food and Chemical Toxicology*, 49(9):1979–1988, September 2011. URL: https://doi.org/10.1016/j.fct.2011.05.009.
- C.W. Noorlander, S.P.J. van Leeuwen, J.D. te Biesebeek, M.J.B. Mengelers, and M.J. Zeilmaker. Levels of perfluorinated compounds in food and dietary intake of PFOS and PFOA in the Netherlands. *Journal of Agricultural and Food Chemistry*, 59(13):7496–7505, July 2011. URL: [https://doi.org/10.1021/jf](https://doi.org/10.1016/j.fct.2011.05.009)104943p.
- O.W. Souverein, W.J. de Boer, A. Geelen, H. van der Voet, J.H. de Vries, M. Feinberg, and P. van 't Veer. Uncertainty in intake due to portion size estimation in 24-hour recalls varies between food groups. *The Journal of Nutrition*, 141(7):1396–1401, May 2011. URL: https://doi.org/10.3[945/jn.111.139220.](https://doi.org/10.1021/jf104943p)

- P.E. Boon, I. Sioen, H. van der Voet, I. Huybrecht[s, M. de Neve, P. Amiano, M. Azpiri,](https://doi.org/10.3945/jn.111.139220) L. Busk, T. Christensen, A. Hilbig, T. Hirvonen, S. Koulouridaki, L. Lafay, K.-H. Liukkonen, J. Moschandreas, S. Papoutsou, L. Ribas-Barba, J. Ruprich, L. Serra-Majem, M. Tornaritis, A. Turrini, M. Urtizberea, E. Verger, A. Westerlund, M. Kersting, S. de Henauw, and J.D. van Klaveren. Long-term dietary exposure to lead in young children living in different European countries. *EFSA Supporting Publications*, 7(5):51E, 2010. URL: https://efsa.onlinelibrary.wiley.com/doi/abs/10.2903/sp.efsa.2010.EN-51.
- P.E. Boon, J.D. te Biesebeek, I. Sioen, I. Huybrechts, M. de Neve, P. Amiano, C. Arganini, M. Azpiri, L. Busk, T. Christensen, A. Hilbig, T. Hirvonen, S. Koulouridaki, L. Lafay, K.-H. Liukkonen, J. Moschandreas, S. Papoutsouk, L. Ribas-Barba, J. Ruprich, L. Serra-Majem, M. Tornaritis, A. Turrini, M. Urtizberea, E. Verger, [A. Westerlund, M. Kersting, S. de Henauw, and J.D. van Klaveren. Long-t](https://efsa.onlinelibrary.wiley.com/doi/abs/10.2903/sp.efsa.2010.EN-51)erm dietary exposure to chromium in young children living in different European countries. *EFSA Supporting Publications*, 7(5):54E, 2010. URL: https://efsa.onlinelibrary.wiley.com/doi/abs/10.2903/sp.efsa.2010.EN-54.
- I. Huybrechts, I. Sioen, P.E. Boon, M. de Neve, P. Amiano, C. Arganini, E. Bower, L. Busk, T. Christensen, A. Hilbig, T. Hirvonen, A. Kafatos, S. Koulouridaki, L. Lafay, K.-H. Liukkonen, S. Papoutsou, L. Ribas-Barba, J. Ruprich, I. Rehurkova, M. Kersting, L. Serra-Majem, A. Turrini, E. Verger, A. Westerlund, M. Tornaritis, [J.D. van Klaveren, and S. de Henauw. Long-term dietary exposure to diffe](https://efsa.onlinelibrary.wiley.com/doi/abs/10.2903/sp.efsa.2010.EN-54)rent food colours in young children living in different European countries. *EFSA Supporting Publications*, 7(5):53E, 2010. URL: https://efsa. onlinelibrary.wiley.com/doi/abs/10.2903/sp.efsa.2010.EN-53.
- A. König, A.H. Kuiper, H.J.P. Marvin, P.E. Boon, L. Busk, F. Cnudde, S. Cope, H.V. Davies, M. Dreyer, L.J. Frewer, M. Kaiser, G.A. Kleter, I. Knudsen, G. Pascal, A. Prandini, O. Renn, M.R. Smith, B.W. Traill, [H. van der Voet, H. van Trijp, E. Vos, and M.T.A. Wentholt.](https://efsa.onlinelibrary.wiley.com/doi/abs/10.2903/sp.efsa.2010.EN-53) The SAFE FOODS frameworkf[or improved](https://efsa.onlinelibrary.wiley.com/doi/abs/10.2903/sp.efsa.2010.EN-53) risk analysis of foods. *Food Control*, 21(12):1566–1587, December 2010. URL: https://doi.org/10.1016/j. foodcont.2010.02.012.
- EFSA Panel on Contaminants in the Food Chain (CONTAM). Scientific opinion on lead in food. *EFSA Journal*, 8(4):1570, 2010. URL: https://efsa.onlinelibrary.wiley.com/doi/abs/10.2903/j.efs[a.2010.1570.](https://doi.org/10.1016/j.foodcont.2010.02.012)
- [I. Sioen, P.E. Boon, I](https://doi.org/10.1016/j.foodcont.2010.02.012). Huybrechts, M. de Neve, P. Amiano, C. Arganini, L. Busk, C. Chadjigeorgiou, T. Christensen, A. Hilbig, T. Hirvonen, S. Koulouridaki, L. Lafay, K.-H. Liukkonen, J. Moschandreas, S. Papoutsou, L. Ribas-Barba, J. Ruprich, L. Serra-Majem, A. Turrini, M. Urtizberea, M. Kersting, E. Verger, A. Westerlund, J.D. va[n Klaveren, and S. de Henauw. Long-term dietary exposure to selen](https://efsa.onlinelibrary.wiley.com/doi/abs/10.2903/j.efsa.2010.1570)ium in young children living in different European countries. *EFSA Supporting Publications*, 7(5):56E, 2010. URL: https://efsa.onlinelibrary.wiley.com/doi/abs/10.2903/sp.efsa.2010.EN-56.
- W. Slob, W.J. de Boer, and H. van der Voet. Can current dietary exposure models handle aggregated intake from different foods? a simulation study for the case of two foods. *Food and Chemical Toxicology*, 48(1):178– 186, January 2010. URL: [https://doi.org/10.1016/j.fct.2009.09.035.](https://efsa.onlinelibrary.wiley.com/doi/abs/10.2903/sp.efsa.2010.EN-56)
- E.H.M. Temme, H. van der Voet, A.J.C. Roodenburg, A. Bulder, G. van Donkersgoed, and J. van Klaveren. Impact of foods with health logo on saturated fat, sodium and sugar intake of young Dutch adults. *Public Health Nutrition*, 14(4):635–644, September 2010. URL: https://doi.org/10.1017/s1368980010002089.
- J.D. van Klaveren, G. van Donkersgoed, H. van der Voet, C. Stephenson, and P.E. Boon. Cumulative exposure assessment of triazole pesticides. *EFSA Supporting Publications*, 7(2):40E, 2010. URL: https: //efsa.onlinelibrary.wiley.com/doi/abs/10.2903/sp[.efsa.2010.EN-40.](https://doi.org/10.1017/s1368980010002089)

- [B.G.H. Bokkers, M.I. Bakker, P.E. Boon, S. Bosgra, G.W.A.M. v](https://efsa.onlinelibrary.wiley.com/doi/abs/10.2903/sp.efsa.2010.EN-40)an der Heijden, G. Janer, W. Slob, and H. van der Voet. The practicability of the integrated probabilistic risk assessment (IPRA) approach for substances in food. RIVM Report 320121001/2009, 2009. URL: http://hdl.handle.net/10029/260367.
- P.E. Boon, M.I. Bakker, J.D. van Klaveren, and C.T.M. van Rossum. Risk assessment of the dietary exposure to contaminants and pesticide residues in young children in the Netherlands. RIVM report 35007000, 2009. URL: http://www.rivm.nl/bibliotheek/rapporten/350070002.pdf.
- P.E. Boon, K. Svensson, S. Moussavian, H. van der Voet, [A. Petersen, J. Ruprich, F. Debegn](http://hdl.handle.net/10029/260367)ach, W.J. de Boer, G. van Donkersgoed, C. Brera, J.D. van Klaveren, and L. Busk. Probabilistic acute dietary exposure assessments to captan and tolylfluanid using several European food consumption and pesticide concentration databases. *[Food and Chemical Toxicology](http://www.rivm.nl/bibliotheek/rapporten/350070002.pdf)*, 47(12):2890–2898, December 2009. URL: https: //doi.org/10.1016/j.fct.2009.01.040.
- P.E. Boon, E.D. van Asselt, M.I. Bakker, A.G. Kruizinga, and M.C.J.F. Jansen. Trends in diet and exposure to chemicals in Dutch children. Report 2009.002, RIKILT, Wageningen, 2009. URL: http://edepot.wur.nl/[7507.](https://doi.org/10.1016/j.fct.2009.01.040)
- [P.M.J. Bos, P.E. Boon, H. van der Vo](https://doi.org/10.1016/j.fct.2009.01.040)et, G. Janer, A.H. Piersma, B.J. Brüschweiler, E. Nielsen, and W. Slob. A semi-quantitative model for risk appreciation and risk weighing. *Food and Chemical Toxicology*, 47(12):2941– 2950, December 2009. URL: https://doi.org/10.1016/j.fct.2009.03.009.
- S. Bosgra, H. van der Voet, P.E. Boon, and W. Slob. An integrated probabilistic framework for cumulative risk assessment of common mechanism chemicals in food: an example with organophosphorus pesticides. *Regulatory Toxicology and Pharmacology*[, 54\(2\):124–133, July 2009. U](https://doi.org/10.1016/j.fct.2009.03.009)RL: https://doi.org/10.1016/j.yrtph. 2009.03.004.
- W.J. de Boer, H. van der Voet, B.G.H. Bokkers, M.I. Bakker, and P.E. Boon. Comparison of two models for the estimation of usual intake addressing zero consumption and non-normality. *Food Additives & Contaminants: [Part A](https://doi.org/10.1016/j.yrtph.2009.03.004)*, 26(11):1433–1449, November 2009. URL: https://doi.org/10.1080/0[2652030903161606.](https://doi.org/10.1016/j.yrtph.2009.03.004)
- B.H. Jensen, A. Petersen, and T. Christensen. Probabilistic assessment of the cumulative dietary acute exposure of the population of Denmark to organophosphorus and carbamate pesticides. *Food Additives & Contaminants: Part A*, 26(7):1038–1048, July 2009. URL: https://[doi.org/10.1080/02652030902859754.](https://doi.org/10.1080/02652030903161606)
- A.K. Müller, S. Bosgra, P.E. Boon, H. van der Voet, E. Nielsen, and O. Ladefoged. Probabilistic cumulative risk assessment of anti-androgenic pesticides in food. *Food and Chemical Toxicology*, 47(12):2951–2962, December 2009. URL: https://doi.org/10.1[016/j.fct.2009.07.039.](https://doi.org/10.1080/02652030902859754)
- S.D. Muri, H. van der Voet, P.E. Boon, J.D. van Klaveren, and B.J. Brüschweiler. Comparison of human health risks resulting from exposure to fungicides and mycotoxins via food. *Food and Chemical Toxicology*, 47(12):2963–2974, December 2009. URL: [https://doi.org/10.101](https://doi.org/10.1016/j.fct.2009.07.039)6/j.fct.2009.03.035.
- EFSA Panel on Contaminants in the Food Chain (CONTAM). Scientific opinion on arsenic in food. *EFSA Journal*, 7(10):1351, 2009. URL: https://efsa.onlinelibrary.wiley.com/doi/abs/10.2903/j.efsa.2009.1351.
- A. J. C. Roodenburg, E. H. M. Temme, O. [Howell Davies, and J. C. Seidell. Potential](https://doi.org/10.1016/j.fct.2009.03.035) impact of the Choices programme on nutrient intakes in the Dutch population. *Nutrition Bulletin*, 34(3):318–323, September 2009. URL: https://doi.org/10.1111/j.1[467-3010.2009.01767.x.](https://efsa.onlinelibrary.wiley.com/doi/abs/10.2903/j.efsa.2009.1351)
- J. Ruprich, I. Rehurkova, P.E. Boon, K. Svensson, S. Moussavian, H. van der Voet, S. Bosgra, J.D. van Klaveren, and L. Busk. Probabilistic modelling of exposure doses and implications for health risk characterization: glycoalkaloids from potatoes. *Food and Chemical Toxicology*, 47(12):2899–2905, December 2009. URL: https:/[/doi.org/10.1016/j.fct.2009.03.008.](https://doi.org/10.1111/j.1467-3010.2009.01767.x)
- H. van der Voet, G.W.A.M. van der Heijden, P.M.J. Bos, S. Bosgra, P.E. Boon, S.D. Muri, and B.J. Brüschweiler. A model for probabilistic health impact assessment of exposure to food chemicals. *Food and Chemical Toxicology*, 47(12):2926–2940, December 2009. URL: https://doi.org/10.1016/j.fct.2008.12.027.
- H.J. van Ooijen, H. van der Voet, and M.I. Bakker. Identification and handling of uncertainties in dietary exposure assessment. RIVM Report 320103004, 2009. URL: http://hdl.handle.net/10029/261706.
- EFSA Panel on Plant Protection Products and their Residues (PPR [Panel\). Scientific opinion on risk assessmen](https://doi.org/10.1016/j.fct.2008.12.027)t for a selected group of pesticides from the triazole group to test possible methodologies to assess cumulative effects from exposure through food from these pesticides on human health. *EFSA Journal*, 7(9):1167, 2009. URL: https://efsa.onlinelibrary.wiley.com/doi/abs/10.2903/j.[efsa.2009.1167.](http://hdl.handle.net/10029/261706)

- P.E. [Boon, H. Van der Voet, M.T.M. Van Raaij, and J.D. Van Klaveren. Cu](https://efsa.onlinelibrary.wiley.com/doi/abs/10.2903/j.efsa.2009.1167)mulative risk assessment of the exposure to organophosphorus and carbamate insecticides in the Dutch diet. *Food and Chemical Toxicology*, 46(9):3090–3098, September 2008. URL: https://doi.org/10.1016/j.fct.2008.06.083.
- A.L. Brantsæter, M. Haugen, A. de Mul, T. Bjellaas, G. Becher, J. van Klaveren, J. Alexander, and H.M. Meltzer. Exploration of different methods to assess dietary acrylamide exposure in pregnant women participating in the Norwegian mother and child cohort study (MoBa). *Food and Chemical Toxicology*, 46(8):2808– 2814, August 2008. URL: https://doi.org/[10.1016/j.fct.2008.05.020.](https://doi.org/10.1016/j.fct.2008.06.083)
- A. de Mul, M.I. Bakker, M.J. Zeilmaker, W.A. Traag, S.P.J. van Leeuwen, R.L.A.P. Hoogenboom, P.E. Boon, and J.D. van Klaveren. Dietary exposure to dioxins and dioxin-like PCBs in the Netherlands anno 2004. *Regulatory Toxicology and Pharmacology*[, 51\(3\):278–287, Aug](https://doi.org/10.1016/j.fct.2008.05.020)ust 2008. URL: https://doi.org/10.1016/ j.yrtph.2008.04.010.
- B.H. Jensen, J.H. Andersen, A. Petersen, and T. Christensen. Dietary exposure assessment of Danish consumers to dithiocarbamate residues in food: a comparison of the deterministic and probabilistic approach. *Food Additives & Contaminants: Part A*, 25(6):714–721, June 2008. URL: [https://doi.org/10.1080/](https://doi.org/10.1016/j.yrtph.2008.04.010) [0265203070185826](https://doi.org/10.1016/j.yrtph.2008.04.010)2.
- C.J. Seal, A. de Mul, G. Eisenbrand, A.J. Haverkort, K. Franke, S.P.D. Lalljie, H. Mykkänen, E. Reimerdes, G. Scholz, V. Somoza, S. Tuijtelaars, M. van Boekel, J. van Klaveren, S.J. Wilcockson, and L. Wilms. Riskbenefit considerations of mitigation measures on acrylamide content of foods – a ca[se study on potatoes, ce](https://doi.org/10.1080/02652030701858262)[reals and coffee.](https://doi.org/10.1080/02652030701858262) *British Journal of Nutrition*, 99(S2):S1–S46, April 2008. URL: https://doi.org/10.1017/ s0007114508965314.

- [European Food Safet](https://doi.org/10.1017/s0007114508965314)y Authority (EFSA). Opinion of the scientific panel on plant protection products and their residues on acute dietary intake assessment of pesticide residues in fruit and vegetables. *EFSA Journal*, 5(8):538, 2007. URL: https://efsa.onlinelibrary.wiley.com/doi/abs/10.2903/j.efsa.2007.538.
- M.I. Bakker, R. de Winter-Sorkina, A. de Mul, P.E. Boon, G. van Donkersgoed, J.D. van Klaveren, B.A. Baumann, W.C. Hijman, S.P.J. van Leeuwen, W. de Boer, and M.J. Zeilmaker. Dietary intake and risk evaluation of polybrominated diphenyl ethers in the Netherlands. *Molecular Nutrition & Food Research*, 52(2):204–216, December 2007. URL: [https://doi.org/10.1002/mnfr.200700112.](https://efsa.onlinelibrary.wiley.com/doi/abs/10.2903/j.efsa.2007.538)
- P.E. Boon, A.M.J. Ragas., and J.D. van Klaveren. Exploration of aggregate exposure to compounds present in food. Report 2007.016, RIKILT, Wageningen, 2007. URL: http://www.rikilt.wur.nl/NL/publicaties/ Rapporten.
- R. de Winter-Sorkina, [M.I. Bakker, G. Wolterink, and M.J. Zei](https://doi.org/10.1002/mnfr.200700112)lmaker. Brominated flame retardants: occurrence, dietary intake and risk assessment. RIVM report 320[100002/2006, 2007. URL:](http://www.rikilt.wur.nl/NL/publicaties/Rapporten) http://rivm. [openreposi](http://www.rikilt.wur.nl/NL/publicaties/Rapporten)tory.com/rivm/handle/10029/7303.
- H. van der Voet and W. Slob. Integration of probabilistic exposure assessment and probabilistic hazard characterization. *Risk Analysis*, 27(2):351–371, April 2007. URL: https://doi.org/10.1111/j.1539-[6924.2007.](http://rivm.openrepository.com/rivm/handle/10029/7303) [00887.x.](http://rivm.openrepository.com/rivm/handle/10029/7303)
- 
- E.D. Caldas, P.E. Boon, and J. Tressou. Probabilistic assessment of the cumulative acute exposure to organophosphorus and carbamate insecticides in the Brazilian diet. *Toxicology*, 222(1-2):132–142, May 2006. URL: https://doi.org/10.1016/j.tox.2006.02.006.
- E.D. Caldas, J. Tressou, and P.E. Boon. Dietary exposure of Brazilian consumers to dithiocarbamate pesticides—a probabilistic approach. *Food and Chemical Toxicology*, 44(9):1562–1571, September 2006. URL: [https://doi.org/10.1016/j.fct.2006.04.014.](https://doi.org/10.1016/j.tox.2006.02.006)
- J.D. van Klaveren, M.Y. Noordam, P.E. Boon, G. van Donkersgoed, B.C. Ossendorp, M.T.M. van Raaij, and J.G. van der Roest. Trends in normoverschrijdigen, overschrijdingen van de acute referentiewaarde en gesommeerde blootstelling - tussenevaluatie nota duurzame gewasbescherming - deelrapport voedselveiligheid. Repor[t 2006.011, RIKILT, Wageningen, 2006. U](https://doi.org/10.1016/j.fct.2006.04.014)RL: http://edepot.wur.nl/24544.

**2006**

- P.E. Boon, A. de Mul, H. van der Voet, G. van Donke[rsgoed, M. Brette, and J.D.](http://edepot.wur.nl/24544) van Klaveren. Calculations of dietary exposure to acrylamide. *Mutation Research/Genetic Toxicology and Environmental Mutagenesis*, 580(1-2):143–155, February 2005. URL: https://doi.org/10.1016/j.mrgentox.2004.10.014.
- A. de Mul, R. de Winter-Sorkina, P.E. Boon, G. van Donkersgoed, M.I. Bakker, and J.D. van Klaveren. Dietary intake of brominated diphenyl ether congeners by the Dutch population. Report 2005.006, RIKILT, Wageningen, 2005. URL: http://edepot.w[ur.nl/26982.](https://doi.org/10.1016/j.mrgentox.2004.10.014)
- M.J. Paulo, H. van der Voet, M.J.W. Jansen, C.J.F. ter Braak, and J.D. van Klaveren. Risk assessment of dietary exposure to pesticides using a Bayesian method. *Pest Management Science*, 61(8):759–766, 2005. URL: https://doi.org/10.1[002/ps.1060.](http://edepot.wur.nl/26982)
- R.C. Schothorst, H.P. van Egmond, A. de Mul, P.E. Boon, J.D. van Klaveren, and G.J.A. Speijers. Trichothecenes in baby food. RIVM Report 310301002, 2005. URL: http://www.rivm.nl/bibliotheek/rapporten/ 31030[1002.pdf.](https://doi.org/10.1002/ps.1060)

- [P.E. Boon, S. L](http://www.rivm.nl/bibliotheek/rapporten/310301002.pdf)ignell, J.D. van Klaveren, and E.I.M. Tjoe Nij. Estimation of the acute dietary exposure to pesticides using the probabilistic approach and the point estimate methodology - the generation of work examples using food consumption data from the Netherlands and Sweden. Report 2004.008, RIKILT, Wageningen, 2004. URL: http://edepot.wur.nl/28647.
- P.E. Boon, E.I.M. Tjoe Nij, N. Koopman, and J.D. van Klaveren. Dietary habits and exposure to pesticides in Dutch infants. Report 2004.017, RIKILT, Wageningen, 2004. URL: http://edepot.wur.nl/44408.
- P.E. Boon, [E.I.M. Tjoe Nij, G. van Do](http://edepot.wur.nl/28647)nkersgoed, and J.D. van Klaveren. Probabilistic intake calculations performed for the codex committee on pesticide residues. Report 2004.005, RIKILT, Wageningen, 2004. URL: http://edepot.wur.nl/36066.
- H. van der Voet and M.J. Paulo. Some explorations into Bayesian m[odelling of risks due to pes](http://edepot.wur.nl/44408)ticide intake from food. In M.A.J.S. van Boekel, A. Stein, and A.H.C. van Bruggen, editors, *Bayesian statistics and quality modelling in the agro-food production chain*, pages 145–162. Kluwer, Dordrecht, 2004. URL: http://library. wur.nl[/frontis/bayes/13\\_van\\_der\\_](http://edepot.wur.nl/36066)voet.pdf.
- P.E. Boon, H. van der Voet, and J.D. van Klaveren. Validation of a probabilistic model of dietary exposure to selected pesticides in Dutch infants. *Food Additives and Contaminants*, 20(sup001):S36–S49, October 2003. URL: https://doi.org/10.1080/0265203031000134956.
- P.E. Boon and J.D. van Klaveren. Cumulative exposure to acetylcholineterase inhibiting compounds in the Dutch population and young childeren. Report 2003.003, RIKILT, Wageningen, 2003. URL: http://edepot. wur.nl/30057.
- P.E. [Boon and J.D. van Klaveren. Dietary exposure to](https://doi.org/10.1080/0265203031000134956) pesticides relevant variables and probabilistic modelling. Report 2003.008, RIKILT, Wageningen, 2003. URL: http://edepot.wur.nl/23045.
- [R. de Winter-](http://edepot.wur.nl/30057)Sorkina, M.I. Bakker, G. van Donkersgoed, and J.D. van Klaveren. Dietary intake [of brominated](http://edepot.wur.nl/30057) flame retardants by the Dutch population. Report 2003.019, RIKILT, Wageningen, 2003. URL: http://hdl. handle.net/10029/7303.
- R. de Winter-Sorkina, G. van Donkersgoed, M.I. Bakker, and J.D. van Klaveren. Dietary intake of heavy metals (cadmium, lead and mercury) by the Dutch population. Report 2003.016, RIKILT, Wagenin[gen, 2003.](http://hdl.handle.net/10029/7303) URL: [http://edepot.wur](http://hdl.handle.net/10029/7303).nl/41597.
- M.J. Gibney and H. van der Voet. Introduction to the Monte Carlo project and the approach to the validation of probabilistic models of dietary exposure to selected food chemicals. *Food Additives and Contaminants*, 20(su[p001\):S1–S7, October 2003](http://edepot.wur.nl/41597). URL: https://doi.org/10.1080/0265203031000134947.
- H. van der Voet, P.E. Boon, and J.D. van Klaveren. Validation of Monte Carlo models for estimating pesticide intake of Dutch infants. Report 2003.002, RIKILT, Wageningen, 2003. URL: http://edepot.wur.nl/39363.

• P.E. Boon, G. van Donkersgoed, and J.D. van Klaveren. Human acute expos[ure assessment of pesticides](http://edepot.wur.nl/39363) in fruits and vegetables. Report 2002.002, RIKILT, Wageningen, 2002. URL: https://library.wur.nl/WebQuery/ wurpubs/reports/320297.

# **CHAPTER**

# **SIX**

# **LIST OF SYMBOLS**



# **SEVEN**

# **APPENDICES**

# **7.1 Api Documentation**

**Note:** This section is under construction.

### **GET /Api/Actions/GetAll/{idWorkspace} Gets all actions of the workspace with the specified id.**

#### **Parameters**

• **idWorkspace** (*integer:int32, required*) – The id of the workspace.

**Status Codes**

- 200 OK The actions of the workspace with the specified id.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

## **GET /Api/Acti[ons/Get](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)/{id}**

**Gets the acti[on with the specifi](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2)ed id.**

#### **Parame[ters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

• **id** (*integer:int32, required*) – The id of the action.

## **Status Codes**

- 200 OK The action with the specified id.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

## **POST /Api/Act[ions/Cr](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)eate**

## **Creates a ne[w action action.](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2)**

#### **Status [Codes](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

- 200 OK A record of the created action.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

### **POST /Api/Act[ions/Cl](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)one**

**Duplicates a[n action and all of](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2) its settings and output.**

## **Status [Codes](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

- 200 OK A record of the cloned action.
- 401 Unauthorized Authorization error.

• 500 Internal Server Error – Internal server error.

#### **POST /Api/Actions/UploadActionZipFile/{idWorkspace}**

## **Creates an upload task that creates an action from an action+data zip file. Returns the task id for monitoring t[he progress.](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

### **Parameters**

• **idWorkspace** (*integer:int32, required*) – Id of the action that should be cloned.

## **Status Codes**

- 200 OK Task id of the upload task.
- 401 Unauthorized Authorization error.
- 400 Bad Request Bad request.
- [500 Inte](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)rnal Server Error Internal server error.

## **DELETE /Api/A[ctions/Delete](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2)/{id}**

## **Deletes the a[ction with the spe](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.1)cified id.**

#### **Parame[ters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

• **id** (*integer:int32, required*) – The id of the action that should be deleted.

## **Status Codes**

- 200 OK  $-$
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

## **PUT /Api/Acti[ons/Upd](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)ateMetaData/{id}**

# **Updates the [action meta data.](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2)**

## **Parame[ters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

• **id** (*integer:int32, required*) – The id of the action.

### **Status Codes**

- 200 OK A record of the updated action meta data.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

#### **PUT /Api/Acti[ons/Con](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)vertToCustomAction/{id} Converts the [\(standard\)action](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2) to a custom action.**

#### **Parame[ters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

• **id** (*integer:int32, required*) – The id of the action.

## **Status Codes**

- 200 OK A record of the updated action.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

#### **PUT /Api/Acti[ons/Upd](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)ateStandardActionVersion/{id} Updates the([standard\)action t](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2)o use the latest standard action version.**

#### **Parame[ters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

• **id** (*integer:int32, required*) – The id of the action.

## **Status Codes**

- 200 OK A record of the updated action.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

#### **PUT /Api/Acti[ons/Set](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)IsCompute/{id}**

## **Specifies whe[ther the data of a](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2) (sub)module of the action should be computed or obtained from data.**

#### **Parame[ters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

• **id** (*integer:int32, required*) – The id of the action.

### **Status Codes**

- 200 OK A record of the updated action.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

#### **PUT /Api/Acti[ons/Set](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)ActionDataSource/{id}**

#### **Adds a datas[ource to the data](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2) source configuration of the action.**

#### **Parame[ters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

• **id** (*integer:int32, required*) – The id of the action.

### **Status Codes**

- 200 OK A record of the updated action.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

### **PUT /Api/Acti[ons/Rep](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)laceActionDataSource/{id} Replaces an [action data source](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2) with another data source.**

#### **Parame[ters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

• **id** (*integer:int32, required*) – The id of the action.

### **Status Codes**

- 200 OK A record of the updated action.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

## **GET /Api/Acti[ons/Get](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)Summary/{id}**

**Returns thes[ummary of the ac](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2)tion with the specified id.**

### **Parame[ters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

• **id** (*integer:int32, required*) – The id of the action.

#### **Status Codes**

- 200 OK Summary section of the specified action.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

## **GET /Api/Acti[ons/Get](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)DataReadingSummary/{id}**

### **Returns the [data reading summ](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2)ary of the specified (sub-)module for the action with the specified id.**

**Parame[ters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

• **id** (*integer:int32, required*) – The id of the action.

## **Query Parameters**

• **actionType** (*required*) – The id of the (sub-)module

### **Status Codes**

- 200 OK The data reading summary of the specified (sub-)module for the action with the specified id.
- 401 Unauthorized Authorization error.
- [500 Inte](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)rnal Server Error Internal server error.

## **GET /Api/Actions/GetDataSelection/{id} Retrieves a d[ata entities page f](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2)or the provided table.**

#### **Parame[ters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

• **id** (*integer:int32, required*) –

#### **Query Parameters**

• **entityType** (*required*) –

## **Status Codes**

- 200 OK A record of the updated action.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

## **GET /Api/Acti[ons/Get](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)DataReadingEntityRecords/{id} Retrieves a d[ata entities page f](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2)or the provided table.**

#### **Parame[ters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

• **id** (*integer:int32, required*) –

## **Query Parameters**

- **tableGroupId** (*required*) –
- **scopingType** (*required*) –
- **filteredStatusTypes** (*array*) –
- **page** (*integer:int32*) –
- **pageSize** (*integer:int32*) –
- **sort** (*string*) –
- **order** (*string*) –

## **Status Codes**

- 200 OK A collection of data reading summary records.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

## **GET /Api/Acti[ons/Get](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)DataLinkingEntityRecords/{id} Get action da[ta linking records](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2).**

### **Parame[ters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

• **id** (*integer:int32, required*) –

#### **Query Parameters**

- **tableGroupId** (*required*) –
- **scopingType** (*required*) –
- **referencedScopingType** (*required*) –
- **filteredStatusTypes** (*array*) –
- **page** (*integer:int32*) –
- **pageSize** (*integer:int32*) –
- **sort** (*string*) –
- **order** (*string*) –

#### **Status Codes**

- 200 OK A collection of data linking summary records.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

#### **PUT /Api/Acti[ons/Set](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)Scope/{id}**

**Sets the actio[n scope. I.e., spec](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2)ify which data entities (specified by their codes) should be considered in the action.**

**Parame[ters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

• **id** (*integer:int32, required*) – The id of the action.

**Status Codes**

- 200 OK A record of the updated action.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

## **PUT /Api/Acti[ons/Cle](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)arScope/{id} Clears the ac[tion scope/data se](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2)lection.**

#### **Parame[ters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

• **id** (*integer:int32, required*) – The id of the action.

#### **Status Codes**

- 200 OK A record of the updated action.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

## **PUT /Api/Acti[ons/Ext](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)endScope/{id}**

**Extends the [action scope.](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2)**

## **Parame[ters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

• **id** (*integer:int32, required*) – The id of the action.

#### **Status Codes**

- 200 OK A record of the updated action.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

#### **PUT /Api/Acti[ons/Red](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)uceScope/{id} Reduces the [action scope.](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2)**

#### **Parame[ters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

```
• id (integer:int32, required) – The id of the action.
```
#### **Status Codes**

- 200 OK A record of the updated action.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

## **GET /Api/Actions/ExportActionSettings/{id} Returns the settings xml of the action with the specified id.**

#### **Parameters**

• **id** (*integer:int32, required*) – The id of the action.

### **Status Codes**

- 200 OK File response.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.
- **GET /Api/Acti[ons/Exp](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)ortActionDataSources/{id} Returns the [data sources xml o](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2)f the action with the specified id.**

### **Parame[ters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

• **id** (*integer:int32, required*) – The id of the action.

#### **Status Codes**

- 200 OK File response.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

#### **GET /Api/Acti[ons/Dow](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)nloadActionZip/{id}**

**Returns all s[ettings and data o](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2)f the action as a zip-file. Project data is included in the form of csv files (zipped csv f[ormat\).](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

**Parameters**

• **id** (*integer:int32, required*) – The id of the action.

## **Status Codes**

- 200 OK File response.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.
- **GET /Api/Acti[ons/Dow](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)nloadActionZipOriginalData/{id} Returns all s[ettings and data o](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2)f the action as a zip-file. Project data source files are included in their original form[s.](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

#### **Parameters**

• **id** (*integer:int32, required*) – The id of the action.

**Status Codes**

- 200 OK File response.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

#### **GET /Api/Acti[ons/Dow](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)nloadActionZipNoData/{id} Returns all s[ettings and data so](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2)urce configuration of the action as a zip-file.**

#### **Parame[ters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

• **id** (*integer:int32, required*) – The id of the action.

**Status Codes**

- 200 OK File response.
- 401 Unauthorized Authorization error.
• 500 Internal Server Error – Internal server error.

#### **GET /api/Actions/FileUploadProgress**

**Retrieves the task progress of the upload task with the specified task id.**

## **Query [Parameters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

• **taskId** (*string, required*) – The task id of the upload task.

#### **Status Codes**

- 200 OK Task report of the specified task.
- 400 Bad Request Bad request.
- 500 Internal Server Error Internal server error.

### **GET /Api/Data[Sources](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)/Get/{id}**

#### <span id="page-324-0"></span>**Returns a da[ta source descrip](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.1)tion record of the data source with the specified id.**

#### **Parame[ters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

• **id** (*integer:int32, required*) – The id of the data source.

#### **Status Codes**

- 200 OK Data source description records.
- 401 Unauthorized Authorization error.
- 400 Bad Request Bad request.
- [404 Not](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1) Found Record not found.
- [500 Internal Serve](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2)r Error Internal server error.

## **GET /Api/Data[Sources/GetA](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.1)ll**

## <span id="page-324-1"></span>**Gets all data [sources availab](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.5)le to the user.**

#### **Status [Codes](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

- 200 OK Collection of data source description records.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

## <span id="page-324-3"></span>**GET /Api/Data[Sources](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)/GetRepositoryDataSources/{id} Gets the data [sources of the rep](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2)ository with the specified id.**

#### **Parame[ters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

• **id** (*integer:int32, required*) –

#### **Status Codes**

- 200 OK Collection of data source description records.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

## <span id="page-324-2"></span>**GET /Api/Data[Sources](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)/GetRemoteRepositoryDataSources/{id} Gets the data [sources available](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2) in the remote repository with the specified id.**

#### **Parame[ters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

```
• id (integer:int32, required) –
```
- 200 OK Collection of data source description records.
- 401 Unauthorized Authorization error.
- <span id="page-324-4"></span>• 500 Internal Server Error – Internal server error.

## **POST /Api/DataSources/ImportRemoteDataSource Imports the remote data source specified by the import settings.**

## **Status Codes**

- 200 OK Data source description record of the imported data source.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

## **PUT /Api/Data[Sources](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)/Move/{id}**

## **Moves the da[ta source with the](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2) specified id to another repository.**

## **Parame[ters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

• **id** (*integer:int32, required*) – The id of the data source that should be moved.

## **Status Codes**

- 200 OK Data source description record of the moved data source.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

# **DELETE /Api/D[ataSour](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)ces/Delete/{id}**

## **Deletes the d[ata source with th](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2)e specified id.**

## **Parame[ters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

• **id** (*integer:int32, required*) – The id of the data source that should be deleted.

## **Status Codes**

- 200 OK  $-$
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

## <span id="page-325-2"></span>**PUT /Api/Data[Sources](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)/Rename/{id} Renames the [data source with](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2) the specified id.**

## **Parame[ters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

• **id** (*integer:int32, required*) – The id of the data source that is to be renamed.

## **Status Codes**

- 200 OK Data source description record of the moved data source.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

## **POST /Api/Dat[aSource](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)s/UploadNewDataSource/{idRepository}**

<span id="page-325-0"></span>**Creates an u[pload task that add](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2)s a new data source to the repository with the specified id. Returns the task id for m[onitoring the progress.](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

## **Parameters**

• **idRepository** (*integer:int32, required*) – The id of repository in which the data source should be created.

- 200 OK Task id of the upload task.
- 401 Unauthorized Authorization error.
- 400 Bad Request Bad request.
- <span id="page-325-1"></span>• [500 Inte](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)rnal Server Error – Internal server error.

## **POST /Api/DataSources/UploadNewDataSourceVersion/{idDataSource}**

**Creates an upload task that adds a new version to the specified data source. Returns the task id for monitoring the progress.**

#### **Parameters**

• **idDataSource** (*integer:int32, required*) – The id of data source for which this is a new version.

#### **Status Codes**

- 200 OK Task id of the upload task.
- 401 Unauthorized Authorization error.
- 400 Bad Request Bad request.
- [500 Inte](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)rnal Server Error Internal server error.

## <span id="page-326-1"></span>**GET /Api/Data[Sources/GetVe](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2)rsion/{idVersion} Gets the data [source version w](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.1)ith the specified id.**

#### **Parame[ters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

• **idVersion** (*integer:int32, required*) – The id of the data source version.

#### **Status Codes**

- 200 OK Data source version record.
- 401 Unauthorized Authorization error.
- 400 Bad Request Bad request.
- [500 Inte](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)rnal Server Error Internal server error.

## **GET /Api/Data[Sources/GetVe](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2)rsions/{id}**

<span id="page-326-2"></span>**Gets all versi[ons of the data so](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.1)urce with the specified id.**

#### **Parame[ters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

• **id** (*integer:int32, required*) – The id of the data source.

## **Status Codes**

- 200 OK Data source version records.
- 401 Unauthorized Authorization error.
- 400 Bad Request Bad request.
- [500 Inte](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)rnal Server Error Internal server error.

## **GET /Api/Data[Sources/GetWo](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2)rkspaceDataSourceVersions/{idWorkspace} Returns all d[ata source versio](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.1)ns used in the workspace with the specified id.**

#### **Parame[ters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

• **idWorkspace** (*integer:int32, required*) – The id of the workspace.

#### **Status Codes**

- 200 OK Data source version records.
- 401 Unauthorized Authorization error.
- 400 Bad Request Bad request.
- [500 Inte](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)rnal Server Error Internal server error.
- <span id="page-326-0"></span>**GET /Api/Data[Sources/GetDa](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2)taSourceVersionUsage/{idVersion} Returns the [data source versio](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.1)n's usage in actions.**

## **Parame[ters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

• **idVersion** (*integer:int32, required*) – The id of the version.

## **Status Codes**

- 200 OK Data source version usage.
- 401 Unauthorized Authorization error.
- 400 Bad Request Bad request.
- [500 Inte](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)rnal Server Error Internal server error.

## <span id="page-327-2"></span>**GET /Api/Data[Sources/GetDa](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2)taSourceUsage/{idDataSource} Returns the [data source's usa](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.1)ge in actions.**

## **Parame[ters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

• **idDataSource** (*integer:int32, required*) – The id of the data source.

## **Status Codes**

- 200 OK Data source usage.
- 401 Unauthorized Authorization error.
- 400 Bad Request Bad request.
- [500 Inte](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)rnal Server Error Internal server error.

## <span id="page-327-0"></span>**GET /Api/Data[Sources/Downl](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2)oadVersion/{idVersion} Downloads t[he data source ve](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.1)rsion dataset file of the version with the specified id.**

## **Parame[ters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

• **idVersion** (*integer:int32, required*) –

## **Status Codes**

- 200 OK File response.
- 401 Unauthorized Authorization error.
- 400 Bad Request Bad request.
- [404 Not](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1) Found File not found.
- [500 Internal Serve](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2)r Error Internal server error.

#### <span id="page-327-1"></span>**GET /Api/Data[Sources/Down](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.1)loadVersionCsv/{idVersion} Downloads t[he data source v](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.5)ersion raw data (as imported by MCRA) of the version as zipped csv file collection.**

## **Parameters**

• **idVersion** (*integer:int32, required*) –

## **Status Codes**

- 200 OK File response.
- 401 Unauthorized Authorization error.
- 400 Bad Request Bad request.
- [404 Not](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1) Found File not found.
- [500 Internal Serve](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2)r Error Internal server error.

## **POST /Api/Dat[aSources/Upl](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.1)oadActionZipFile**

## <span id="page-327-3"></span>**Creates an u[pload task that](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.5) creates an action from an action+data zip file. Returns the task id for monitoring t[he progress.](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

## **Status Codes**

• 200 OK – Task id of the upload task.

- 401 Unauthorized Authorization error.
- 400 Bad Request Bad request.
- 500 Internal Server Error Internal server error.

## **GET /api/Data[Sources/FileU](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2)ploadProgress**

## <span id="page-328-1"></span>**Retrieves the [task progress of](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.1) the upload task with the specified task id.**

#### **Query [Parameters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

• **taskId** (*string, required*) – The task id of the upload task.

## **Status Codes**

- 200 OK Task report of the specified task.
- 400 Bad Request Bad request.
- 401 Unauthorized Authorization error.
- [500 Inte](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)rnal Server Error Internal server error.

## **GET /api/Repo[sitories/Get](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.1)All**

#### **Status [Codes](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2)**

• [200 OK](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1) –

## <span id="page-328-2"></span>**GET /Api/Repositories/Get/{id}**

#### **Parameters**

• **id** (*[int](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)eger:int32, required*) –

## **Status Codes**

- 200 OK  $-$
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

## <span id="page-328-0"></span>**GET /Api/Repo[sitorie](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)s/GetDetails/{id}**

## **Parame[ters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2)**

• **id** (*[integer:int32,](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1) required*) –

#### **Status Codes**

- 200 OK  $-$
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

## **POST /api/Rep[ositori](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)es/Create**

## **Status [Codes](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2)**

• [200 OK](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1) –

## <span id="page-328-3"></span>**POST /api/Repositories/Update**

**Status Codes**

• [200 OK](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)  $-$ 

## **POST /Api/Repositories/Delete/{id}**

**Parameters**

```
• id (integer:int32, required) –
```
- 200 OK –
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

## <span id="page-329-1"></span>**POST /Api/Rep[ositori](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)es/ForceDelete/{id}**

## **Parame[ters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2)**

```
• id (integer:int32, required) –
```
#### **Status Codes**

- 200 OK  $-$
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

#### **POST /api/Rep[ositori](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)es/Move**

#### **Status [Codes](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2)**

- [200 OK](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)  $-$
- **POST /api/Repositories/ChangeOwner**

**Status Codes**

```
200 OK -
```
## <span id="page-329-4"></span>**POST /api/Repositories/AddUserShare**

**Status Codes**

## • [200 OK](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1) –

#### <span id="page-329-3"></span>**POST /api/Repositories/UpdateUserShare**

**Status Codes**

```
200 OK -
```
**POST /api/Repositories/RemoveUserShare**

**Status Codes**

```
200 OK –
```
#### <span id="page-329-6"></span>**POST /api/Repositories/AddGroupShare**

**Status Codes**

• [200 OK](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1) –

#### <span id="page-329-2"></span>**POST /api/Repositories/UpdateGroupShare**

**Status Codes**

• [200 OK](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1) –

**POST /api/Repositories/RemoveGroupShare**

**Status Codes**

• [200 OK](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1) –

<span id="page-329-5"></span><span id="page-329-0"></span>**GET /Api/Repositories/GetRepositoryDataSourcesInUse/{id} Gets the data sources that are in use in the repository with the specified id.**

**Parame[ters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)**

• **id** (*integer:int32, required*) –

- 200 OK –
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

#### **GET /Api/Repo[sitorie](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)s/GetRepositoryDataSourceUsage/{id}**

<span id="page-330-0"></span>**Gets all proje[ct information sti](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2)ll associated with any datasources in the repository or any subrepositories**

#### **Parame[ters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

• **id** (*integer:int32, required*) – Id of the main repository to search

#### **Status Codes**

- 200 OK  $-$
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

## **GET /Api/Work[space/G](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)etAll**

## **Gets all work[spaces available t](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2)o the user.**

## **Status [Codes](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

- 200 OK Collection of workspace description records.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

## **GET /Api/Work[space/G](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)et/{id}**

## **Gets the wor[kspace with the sp](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2)ecified id.**

#### **Parame[ters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

• **id** (*integer:int32, required*) – The id of the workspace.

#### **Status Codes**

- 200 OK The workspace with the specified id.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

#### **POST /Api/Wor[kspace/](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)Create**

<span id="page-330-1"></span>**Creates a wo[rkspace based on](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2) the provided form data.**

#### **Status [Codes](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

- 200 OK A description record of the created workspace.
- 500 Internal Server Error Internal server error.

## **POST /Api/Workspace/Update/{id}**

<span id="page-330-2"></span>**Updates the [workspac](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)e meta data.**

#### **Parame[ters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

• **id** (*integer:int32, required*) – The id of the workspace that is to be updated.

## **Status Codes**

• 200 OK – A description record of the created workspace.

## **DELETE /Api/Workspace/Delete/{id}**

## **Deletes the workspace with the specified id.**

## **Parame[ters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)**

• **id** (*integer:int32, required*) – The id of the workspace that should be deleted.

**Status Codes**

• 200 OK –

## **GET /Api/Workspace/GetDataSources/{id}**

## <span id="page-331-1"></span>**Returns all data sources used in the workspace with the specified id.**

## **Parame[ters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)**

• **id** (*integer:int32, required*) – The id of the workspace.

#### **Status Codes**

- 200 OK A description record of the created workspace.
- 401 Unauthorized Authorization error.
- 400 Bad Request Bad request.
- [500 Inte](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)rnal Server Error Internal server error.

# **POST /Api/Wor[kspace/AddDat](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2)aSource/{id}**

## <span id="page-331-2"></span>**Adds a datas[ource version to](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.1) the workspace.**

## **Parame[ters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

• **id** (*integer:int32, required*) – The id of the workspace.

## **Status Codes**

- 200 OK On success.
- 401 Unauthorized Authorization error.
- 400 Bad Request Bad request.
- [500 Inte](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)rnal Server Error Internal server error.

## <span id="page-331-3"></span>**POST /Api/Wor[kspace/Remove](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2)DataSource/{id} Removes a d[ata source version](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.1) from the workspace.**

#### **Parame[ters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

• **id** (*integer:int32, required*) – The id of the workspace.

## **Status Codes**

- 200 OK On success.
- 401 Unauthorized Authorization error.
- 400 Bad Request Bad request.
- [500 Inte](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)rnal Server Error Internal server error.

## **GET /Api/Abou[t/VersionInfo](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2)**

## **Returns a ve[rsion info record](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.1) containing version number and build date.**

## **Status [Codes](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

• 200 OK – About info record.

## <span id="page-331-0"></span>**GET /api/ActionTypes/Get**

## **Returns a collection of the action type definitions available in the toolbox.**

## **Status [Codes](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)**

- 200 OK Collection of action type definition records.
- 500 Internal Server Error Internal server error.

## **GET /api/DataFormats/Get**

## <span id="page-331-4"></span>**Returns a col[lection o](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)f the data format definitions available in the toolbox.**

- 200 OK Collection of data format definition records.
- 500 Internal Server Error Internal server error.

## **GET /Api/Outputs/Get/{id}**

## <span id="page-332-0"></span>**Gets the outp[ut with](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1) the specified id.**

## **Parame[ters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

• **id** (*integer:int32, required*) – The id of the output.

#### **Status Codes**

- 200 OK Output record.
- 400 Bad Request Bad request.
- 401 Unauthorized Authorization error.
- [500 Inte](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)rnal Server Error Internal server error.

## **GET /Api/Outp[uts/GetFromA](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.1)ction/{idAction}**

#### <span id="page-332-1"></span>**Gets the outp[uts of the action w](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2)ith the specified id. Returns a list of OutputViewModel records.**

#### **Parame[ters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

• **idAction** (*integer:int32, required*) – Id of the action.

## **Status Codes**

- 200 OK List of output records.
- 400 Bad Request Bad request.
- 500 Internal Server Error Internal server error.

## **GET /Api/Outp[uts/Get](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)OutputReportToc/{id}**

#### <span id="page-332-3"></span>**Get the outp[ut summary dese](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.1)rialized into the object hierarchy.**

#### **Parame[ters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

• **id** (*integer:int32, required*) – Id of the output.

## **Status Codes**

- 200 OK Output hierarchy object.
- 400 Bad Request Bad request.
- 401 Unauthorized Authorization error.
- [500 Inte](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)rnal Server Error Internal server error.

## **GET /Api/Outp[uts/GetOutpu](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.1)tReportTocs**

<span id="page-332-4"></span>**Gets the hier[archical output re](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2)port summary toc objects of the outputs with the specified ids.**

#### **Query [Parameters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

• **ids** (*array*) – Ids of the outputs.

## **Status Codes**

- 200 OK Array of summary hierarchical toc objects.
- 400 Bad Request Bad request.
- 401 Unauthorized Authorization error.
- [500 Inte](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)rnal Server Error Internal server error.

## **GET /Api/Outp[uts/GetOutpu](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.1)tReportSection**

## <span id="page-332-2"></span>**Gets outputr[eport section with](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2) the specified id of the output with the specified id.**

## **Query [Parameters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

• **idOutput** (*integer:int32, required*) – Id of the output

• **idSection** (*string:guid, required*) – Id of the section.

## **Status Codes**

- 200 OK Html string of the section content.
- 400 Bad Request Bad request.
- 401 Unauthorized Authorization error.
- [500 Inte](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)rnal Server Error Internal server error.

## <span id="page-333-2"></span>**GET /Api/Outp[uts/GetShort](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.1)OutputSummary Gets short ou[tput report summ](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2)ary with the specified id of the output**

## **Query [Parameters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

• **idOutput** (*integer:int32, required*) – Id of the output

## **Status Codes**

- 200 OK Html string of the short output summary content.
- 400 Bad Request Bad request.
- 401 Unauthorized Authorization error.
- [500 Inte](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)rnal Server Error Internal server error.

## **GET /Api/Outp[uts/GetOutpu](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.1)tReportTableData**

## <span id="page-333-1"></span>**Gets outputt[able data section](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2) with the specified id of the output with the specified id.**

## **Query [Parameters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

- **idOutput** (*integer:int32, required*) Id of the output.
- **idSection** (*string:guid, required*) Id of the section.
- **maxRecords** (*integer:int32, required*) Maximum number of records.
- **isTree** (*boolean, required*) if true, the hierarchy defining data should be incorporated

## **Status Codes**

- 200 OK Html string of the table data.
- 400 Bad Request **Bad request.**
- 401 Unauthorized Authorization error.
- [500 Inte](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)rnal Server Error Internal server error.

## **GET /Api/Outp[uts/GetOutpu](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.1)tReportTableContent**

## <span id="page-333-0"></span>**Gets outputt[able data section](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2) with the specified id of the output with the specified id.**

## **Query [Parameters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

- **idOutput** (*integer:int32, required*) Id of the output.
- **idSection** (*string:guid, required*) Id of the section.
- **caption** (*string, required*) Table caption
- **maxRecords** (*integer:int32, required*) Maximum number of records.
- **columnOrder** (*string, required*) Comma separated list of column indices determining the shown columns
- **isTree** (*boolean, required*) if true, the hierarchy defining data should be incorporated

## **Status Codes**

• 200 OK – Html string of the full table data.

- 400 Bad Request Bad request.
- 401 Unauthorized Authorization error.
- 500 Internal Server Error Internal server error.

## **GET /Api/Outp[uts/GetOutpu](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.1)tReportChart**

## <span id="page-334-2"></span>**Gets outputt[able data section](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2) with the specified id of the output with the specified id.**

#### **Query [Parameters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

- **idOutput** (*integer:int32, required*) Id of the output.
- **idSection** (*string:guid, required*) Id of the section.

#### **Status Codes**

- 200 OK Html string of the table data.
- 400 Bad Request Bad request.
- 401 Unauthorized Authorization error.
- [500 Inte](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)rnal Server Error Internal server error.

#### **GET /Api/Outp[uts/GetTaskS](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.1)ettingsSection**

#### <span id="page-334-3"></span>**Gets the setti[ngs section conten](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2)t of the task with the specified id.**

## **Query [Parameters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

• **idTask** (*integer:int32, required*) – Id of the task.

#### **Status Codes**

- 200 OK Html string of the section.
- 400 Bad Request Bad request.
- 401 Unauthorized Authorization error.
- [500 Inte](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)rnal Server Error Internal server error.

## **GET /Api/Outp[uts/Download](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.1)FullPdf Returns pdf [file of the output r](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2)eport with the specified id.**

## **Query [Parameters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

• **idOutput** (*integer:int32, required*) – Id of the output.

#### **Status Codes**

- 200 OK Report pdf file.
- 400 Bad Request Bad request.
- 401 Unauthorized Authorization error.
- [500 Inte](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)rnal Server Error Internal server error.

## <span id="page-334-1"></span>**GET /Api/Outp[uts/Download](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.1)ShortReportPdf Returns pdf [file of the output r](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2)eport with the specified id.**

#### **Query [Parameters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

• **idOutput** (*integer:int32, required*) – Id of the output.

- 200 OK Report pdf file.
- 400 Bad Request Bad request.
- 401 Unauthorized Authorization error.
- <span id="page-334-0"></span>• [500 Inte](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)rnal Server Error – Internal server error.

## **GET /Api/Outputs/DownloadSectionPdf**

**Returns pdf file of the report section with the specified id of the output with the specified id.**

#### **Query Parameters**

- **idOutput** (*integer:int32, required*) Id of the output.
- **idSection** (*string:guid, required*) Id of the section.

#### **Status Codes**

- 200 OK Report pdf file.
- 400 Bad Request Bad request.
- 401 Unauthorized Authorization error.
- [500 Inte](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)rnal Server Error Internal server error.

## <span id="page-335-0"></span>**GET /Api/Outp[uts/Download](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.1)ReportCsvZip Returns zip [file with csv files of](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2) the tables of the output report of the output with the specified id.**

#### **Query [Parameters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

• **idOutput** (*integer:int32, required*) – Id of the output.

**Status Codes**

- 200 OK Zip file.
- 400 Bad Request Bad request.
- 401 Unauthorized Authorization error.
- [500 Inte](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)rnal Server Error Internal server error.

## **GET /Api/Outp[uts/Download](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.1)TableCsv**

#### <span id="page-335-2"></span>**Returns csv [file of the report ta](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2)ble with the specified id of the output with the specified id.**

#### **Query [Parameters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

- **idOutput** (*integer:int32, required*) Id of the output.
- **idSection** (*string:guid, required*) Id of the section.

## **Status Codes**

- 200 OK Csv file.
- 400 Bad Request Bad request.
- 401 Unauthorized Authorization error.
- [500 Inte](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)rnal Server Error Internal server error.

## **GET /Api/Outp[uts/Download](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.1)SectionCsvZip**

<span id="page-335-1"></span>**Returns zip [file with csv files o](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2)f the tables of the section with the specified id of the output with the specified id.**

#### **Query Parameters**

- **idOutput** (*integer:int32, required*) Id of the output.
- **idSection** (*string:guid, required*) Id of the section.

- 200 OK Zip file.
- 400 Bad Request Bad request.
- 401 Unauthorized Authorization error.
- [500 Inte](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)rnal Server Error Internal server error.

## **GET /Api/Outputs/DownloadRawData**

**Returns a zip file of the generated raw data of the output with the specified id.**

#### **Query Parameters**

• **idOutput** (*integer:int32, required*) – Id of the output.

#### **Status Codes**

- 200 OK Zip file.
- 400 Bad Request Bad request.
- 401 Unauthorized Authorization error.
- [500 Inte](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)rnal Server Error Internal server error.

#### **GET /api/Scop[ingTypes/Get](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.1)**

#### <span id="page-336-2"></span>**Returns a col[lection of the scop](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2)ing type definitions available in the toolbox.**

## **Status [Codes](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

- 200 OK Collection of workspace description records.
- 500 Internal Server Error Internal server error.

## **GET /api/UnitDefinitions/Get**

<span id="page-336-3"></span>**Returns a col[lection o](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)f the unit definitions available in the toolbox.**

#### **Status [Codes](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

- 200 OK Collection of unit definition records.
- 500 Internal Server Error Internal server error.

# **GET /Api/Tasks/Get/{idWorkspace}**

<span id="page-336-0"></span>**Gets all tasks [of a wo](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)rkspace.**

## **Parame[ters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

• **idWorkspace** (*integer:int32, required*) – The id of the workspace.

## **Status Codes**

• 200 OK – Task record.

#### <span id="page-336-1"></span>**PUT /Api/Tasks/UpdateDescription/{id} Updates the task description.**

#### **Parame[ters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)**

• **id** (*integer:int32, required*) – The id of the task.

## **Status Codes**

- 200 OK Task record.
- 401 Unauthorized Authorization error.
- 400 Bad Request Bad request.
- [500 Inte](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)rnal Server Error Internal server error.

## **GET /Api/Task[s/GetTaskLog/](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2){id}**

**Retrieves the [latest tasklog of](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.1) the task with the specified id.**

#### **Parame[ters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

• **id** (*integer:int32, required*) – The id of the task.

- 200 OK Task log string.
- 401 Unauthorized Authorization error.
- 400 Bad Request Bad request.
- 500 Internal Server Error Internal server error.

## **POST /Api/Tasks/ScheduleTask/{idAction}**

## <span id="page-337-3"></span>**Starts a calcu[lation task for th](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.1)e given action and returns the id of this task.**

#### **Parame[ters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

• **idAction** (*integer:int32, required*) – The id of the action from which the task should be spawned.

## **Status Codes**

- 200 OK Task record.
- 401 Unauthorized Authorization error.
- 400 Bad Request Bad request.
- [500 Inte](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)rnal Server Error Internal server error.

#### **POST /Api/Tas[ks/ScheduleSu](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2)bTask/{idAction}**

<span id="page-337-2"></span>**Starts a conc[entration model c](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.1)alculation task for the given project and returns the id of this task.**

#### **Parame[ters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

• **idAction** (*integer:int32, required*) – The id of the action.

#### **Query Parameters**

• **actionType** (*required*) – The type of the sub-module of the action.

#### **Status Codes**

- 200 OK Task record.
- 401 Unauthorized Authorization error.
- 400 Bad Request Bad request.
- [500 Inte](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)rnal Server Error Internal server error.

## <span id="page-337-1"></span>**GET /Api/Task[s/GetProgress](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2)/{taskId} Returns the [progress of the sp](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.1)ecified task.**

#### **Parame[ters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

• **taskId** (*integer:int32, required*) – Id of the task.

## **Status Codes**

- 200 OK Task progress record.
- 401 Unauthorized Authorization error.
- 400 Bad Request Bad request.
- [500 Inte](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)rnal Server Error Internal server error.

#### <span id="page-337-0"></span>**GET /Api/Task[s/GetActiveTa](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2)skStatuses/{idWorkspace} Gets the stat[uses of all active](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.1) tasks of a workspace.**

#### **Parame[ters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

• **idWorkspace** (*integer:int32, required*) – Id of the workspace.

- 200 OK Task progress record.
- 401 Unauthorized Authorization error.
- 400 Bad Request Bad request.
- <span id="page-337-4"></span>• [500 Inte](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)rnal Server Error – Internal server error.

## **PUT /Api/Tasks/Abort/{id}**

**Aborts the execution of the task with the specified id.**

#### **Parameters**

• **id** (*integer:int32, required*) – The id of the task.

#### **Status Codes**

- 200 OK  $-$
- 401 Unauthorized Authorization error.
- 400 Bad Request Bad request.
- [500 Inte](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)rnal Server Error Internal server error.

# **DELETE /Api/T[asks/Delete/{](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2)id}**

## **Deletes the ta[sk with the speci](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.1)fied id.**

## **Parame[ters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

• **id** (*integer:int32, required*) –

#### **Status Codes**

- 200 OK –
- 401 Unauthorized Authorization error.
- 400 Bad Request Bad request.
- [500 Inte](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)rnal Server Error Internal server error.

## **DELETE /Api/T[asks/BatchDel](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2)ete**

**Deletes the ta[sks with the spec](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.1)ified ids.**

#### **Query [Parameters](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

• **ids** (*array*) – The ids of the tasks that should be deleted.

#### **Status Codes**

- 200 OK  $-$
- 401 Unauthorized Authorization error.
- 400 Bad Request Bad request.
- [500 Inte](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)rnal Server Error Internal server error.

## **GET /Api/Task[s/DownloadTas](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2)kZip/{id}**

<span id="page-338-0"></span>**Download az[ip file containing](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.1) the task's settings and data at the time of the task's creation. The task's settings and [data sources are deserial](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)ized from the ProjectSettings and DataSourceSettings fields**

#### **Parameters**

• **id** (*integer:int32, required*) – The ID of the task

#### **Status Codes**

- 200 OK File result (zip file).
- 401 Unauthorized Authorization error.
- 400 Bad Request Bad request.
- [500 Inte](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)rnal Server Error Internal server error.

## **GET /api/Stan[dardActionTyp](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.2)es/Get**

## <span id="page-338-1"></span>**Returns a col[lection of the sta](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.4.1)ndard action type definitions available for the user.**

## **Status [Codes](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

• 200 OK – Collection of standard action type definition records.

• 500 Internal Server Error – Internal server error.

## **GET /api/ActionClasses/Get**

**Returns a collection of the action class definitions available in the toolbox.**

## **Status [Codes](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

- 200 OK Collection of action class definition records.
- 500 Internal Server Error Internal server error.

# **7.2 Munro [collec](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.2.1)[tion](http://www.w3.org/Protocols/rfc2616/rfc2616-sec10.html#sec10.5.1)**

This collection can be downloaded here.

# **7.3 Unit definitions**

# **7.3.1 Benchmark response types**

Accepted benchmark response types.

Name	Short name	<b>Aliases</b>	Description
Fraction	Fraction	Fraction-	The benchmark response is defined as a
change	change	Change,	fraction change of the background response
		<b>FactorChange</b>	(i.e., defined for both increase and decrease).
			E.g., for a factor of 0.1, the benchmark
			response is at $+/- 10\%$ of background
			response.
Percentage	Percentage	Percent-	The benchmark response is defined as a
change	change	ageChange	percentage change of the background
			response (i.e., defined for both increase and
			decrease). E.g., for a percentage of 10, the
			benchmark response is at $+/-10\%$ of
			background response.
Fraction of	Fraction of	Factor, Facto-	
background	background	rOfBackground	The benchmark response is defined as a fraction of the background response. E.g.,
			for a factor of 0.9, the benchmark response is
response			
			at 0.9 times the background response (i.e., a decrease).
Percentage of	Percentage of	Percentage,	The benchmark response is defined as a
background	background	PercentageOf-	percentage of the background response. E.g.,
response		Background	
			for a percentage of 90, the benchmark
			response is at 90% of the background
			response (i.e., a decrease).
Extra risk	ER	<b>ExtraRisk</b>	For quantal response types. The benchmark
			dose is defined as the dose that
			corresponding with an extra risk of a factor
			times the background risk. A factor of 0.05
			corresponds with 5% extra risk.
<b>Additional risk</b>	<b>AR</b>	<b>AdditionalRisk</b>	For quantal response types. The benchmark
			dose is defined as the dose that
			corresponding with an additional risk of a
			factor times the background risk. A factor of
			0.05 corresponds with $5\%$ additional risk.
ED50	ED50	ED <sub>50</sub>	For quantal response types. The benchmark
			dose is defined as the dose that corresponds
			with an estimated risk of 50% (ED50).
Absolute	Threshold	Absolute	The benchmark dose is defined as an
threshold value	value		absolute threshold value.
Absolute	Absolute	<b>Difference</b>	The benchmark dose is defined an absolute
difference	difference		difference with the background risk.

Table 7.1: Unit definition for Benchmark response types.

# **7.3.2 Body weight units**

Units for describing person body weights.





# **7.3.3 Concentration units**

Units for describing substance concentrations.





## **7.3.4 Concentration value types**

Concentration value types.

Name	Short name	Aliases	Description
Mean	MC	MeanConcen-	Mean value from the residue trials.
concentration		tration.	
		Concentration-	
		Mean,	
		MC.	
Median	<b>MR</b>	MedianConcen-	Median concentration / residue value of the
concentration		tration, MR,	positive measurements of the residue trials.
		STMR, Super-	
		visedTrialMedi-	
		anResidue	
Highest	<b>HR</b>	HighestConcen-	Highest measured residue / concentration
concentration		tration,	value.
		HighestResidue,	
		<b>HR</b>	
Concentration	CP	Percentile	
percentile			
Limit of	LOQ	LOQ	
quantification			
Maximum	MRL	<b>MRL</b>	
residue limit			

Table 7.4: Unit definition for Concentration value types.

# **7.3.5 Consumption intake units**

Units for consumption intakes amounts.

Table 7.5: Unit definition for Consumption intake units.

Name	Short name	Aliases
gram/kilogram	g/kg bw/day	g/kg bw, gram/kg bw, g/kg bw/day, gram/kg bw/day, gr/kg
bodyweight/day		bw/day, G212A
gram/day	g/day	gram, grams, g/day, g/day, gram/day, gr/day

# **7.3.6 Consumption units**

Units for consumption amounts.

Table 7.6: Unit definition for Consumption units.

Name	Short name	<b>Aliases</b>
kilogram	kg	kg, kilograms, kilogr, 3, G167A
Gram		g, grams, gr, 0, G148A

# **7.3.7 Consumption value types**

Consumption value types.

Table 7.7: Unit definition for Consumption value types.

Name	Short name	Aliases
Large portion	LP.	LP, LargePortion
Mean consumption	МC	MC, MeanConsumption
Percentile	Percentile	Percentile, P

## **7.3.8 Dose response model types**

Known dose response model types.





## **7.3.9 Dose units**

Units for describing substance doses.

Name	Short name	Aliases
gram/kilogram	g/kg bw/day	g/kg bw/day, gram/kg bw/day, gr/kg bw/day
bodyweight/day		
milligram/kilogram	mg/kg bw/day	mg/kg bw/day, milligram/kg bw/day, milligr/kg bw/day,
bodyweight/day		G211A

Table 7.9: Unit definition for Dose units.

continues on next page

Name	Short name	<b>Aliases</b>
micro-	µg/kg bw/day	µg/kg bw/day, microgram/kg bw/day, microgr/kg bw/day
gram/kilogram		
bodyweight/day		
nanogram/kilogram	ng/kg bw/day	ng/kg bw/day, nanogram/kg bw/day, nanogr/kg bw/day
bodyweight/day		
picogram/kilogram	pg/kg bw/day	pg/kg bw/day, picogram/kg bw/day, picogr/kg bw/day
bodyweight/day		
fem-	fg/kg bw/day	fg/kg bw/day, femtogram/kg bw/day, femtogr/kg bw/day
togram/kilogram		
bodyweight/day		
gram/gram	g/g bw/day	g/g bw/day, gram/g bw/day, gr/g bw/day
bodyweight/day		
milligram/gram	mg/g bw/day	mg/g bw/day, milligram/g bw/day, milligr/g bw/day
bodyweight/day		
microgram/gram	µg/g bw/day	µg/g bw/day, microgram/g bw/day, microgr/g bw/day
bodyweight/day		
nanogram/gram	ng/g bw/day	ng/g bw/day, nanogram/g bw/day, nanogr/g bw/day
bodyweight/day		
picogram/gram	pg/g bw/day	pg/g bw/day, picogram/g bw/day, picogr/g bw/day
bodyweight/day		
femtogram/gram	fg/g bw/day	fg/g bw/day, femtogram/g bw/day, femtogr/g bw/day
bodyweight/day		
kilogram/day	kg/day	kg/day, kilogram/day, kilogr/day
gram/day	g/day	g/day, gram/day, gr/day
milligram/day	mg/day	mg/day, milligram/day, milligr/day
microgram/day	µg/day	ug/day, microgram/day, microgr/day
nanogram/day	ng/day	ng/day, nanogram/day, nanogr/day
picogram/day	pg/day	pg/day, picogram/day, picogr/day
femtogram/day	fg/day	fg/day, femtogram/day, femtogr/day
kilogram/kilogram	kg/kg	kg/kg, kilogram/kilogram, kilogram/kg, kg/kg bw
gram/kilogram	g/kg	g/kg, gram/kilogram, gram/kg, gr/kg, g/kg bw
milligram/kilogram	mg/kg	mg/kg, milligram/kilogram, milligram/kg, milligr/kg,
		mg/kg bw, G225A
micro-	µg/kg	µg/kg, microgram/kilogram, microgram/kg, microgr/kg,
gram/kilogram		$\mu$ g/kg bw
nanogram/kilogram	ng/kg	ng/kg, nanogram/kilogram, nanogram/kg, nanogr/kg,
		ng/kg bw
picogram/kilogram	pg/kg	pg/kg, picogram/kilogram, picogram/kg, picogr/kg, pg/kg
		bw
Molar	M	$M$ , mol/L
millimolar	mM	mM, mmol/L
micromolar	μM	uM, µM, umol/L
nanomolar	nM	$nM$ , $nmol/L$
moles	moles	moles, Moles
millimoles	mmoles	mmoles, mMoles
micromoles	µmoles	umoles, uMoles
nanomoles	nmoles	nmoles, nMoles

Table 7.9 – continued from previous page

# **7.3.10 Exposure route types**

The different routes in which an individual is exposed to substance concentrations.

Name	Short name	Aliases	Description
Dietary exposure	Dietary	Dietary	Dietary exposure.
Non-dietary oral exposure	Oral	Oral	Non-dietary oral exposure.
Non-dietary dermal exposure	Dermal	Dermal	Non-dietary dermal exposure.
Non-dietary inhalation exposure	Inhalation	Inhalation	Non-dietary inhalation exposure.
At target	At target	<b>AtTarget</b>	Exposures directly at the target (organ).

Table 7.10: Unit definition for Exposure route types.

# **7.3.11 Exposure types**

The different types of exposure. I.e., acute or chronic.

Table 7.11: Unit definition for Exposure types.

Name	Short name	Aliases	Description
Acute	Acute	Acute	Acute exposure.
Chronic	Chronic	Chronic	Chronic exposure.

# **7.3.12 Exposure units**

Units for describing substance exposures.

Table 7.12: Unit definition for Exposure units.

Name	Short name	Aliases
gram/kilogram	g/kg bw/day	g/kg bw/day, g/kg/day, gram/kg bw/day, gr/kg bw/day,
bodyweight/day		G212A
milligram/kilogram	mg/kg bw/day	mg/kg bw/day, mg/kg/day, milligram/kg bw/day,
bodyweight/day		milligr/kg bw/day, G211A
micro-	µg/kg bw/day	µg/kg bw/day, µg/kg/day, microgram/kg bw/day,
gram/kilogram		microgr/kg bw/day, G210A
bodyweight/day		
nanogram/kilogram	ng/kg bw/day	ng/kg bw/day, ng/kg/day, nanogram/kg bw/day, nanogr/kg
bodyweight/day		bw/day, G214A
picogram/kilogram	pg/kg bw/day	pg/kg bw/day, picogram/kg bw/day, picogr/kg bw/day
bodyweight/day		
fem-	fg/kg bw/day	fg/kg bw/day, fg/kg/day, femtogram/kg bw/day,
togram/kilogram		femtogr/kg bw/day
bodyweight/day		
gram/gram	$g/g$ bw/day	g/g bw/day, g/g/day, gram/g bw/day, gr/g bw/day
bodyweight/day		
milligram/gram	mg/g bw/day	mg/g bw/day, mg/g/day, milligram/g bw/day, milligr/g
bodyweight/day		bw/day
microgram/gram	µg/g bw/day	μg/g bw/day, μg/g/day, microgram/g bw/day, microgr/g
bodyweight/day		bw/day

continues on next page

Name	Short name	<b>Aliases</b>
nanogram/gram	ng/g bw/day	ng/g bw/day, nanogram/g bw/day, nanogr/g bw/day
bodyweight/day		
picogram/gram	pg/g bw/day	pg/g bw/day, pg/g/day, picogram/g bw/day, picogr/g
bodyweight/day		bw/day
femtogram/gram	fg/g bw/day	fg/g bw/day, fg/g/day, femtogram/g bw/day, femtogr/g
bodyweight/day		bw/day
kilogram/day	kg/day	kg/day, kilogram/day, kilogr/day
gram/day	g/day	g/day, gram/day, gr/day
milligram/day	mg/day	mg/day, milligram/day, milligr/day
microgram/day	µg/day	µg/day, microgram/day, microgr/day
nanogram/day	ng/day	ng/day, nanogram/day, nanogr/day
picogram/day	pg/day	pg/day, picogram/day, picogr/day
femtogram/day	fg/day	fg/day, femtogram/day, femtogr/day
gram/kilogram	g/kg	g/kg, gram/kg, gr/kg, G015A
milligram/kilogram	mg/kg	mg/kg, milligram/kg, milligr/kg, G061A
micro-	µg/kg	µg/kg, microgram/kg, microgr/kg, G050A
gram/kilogram		
nanogram/kilogram	ng/kg	ng/kg, nanogram/kg, nanogr/kg, G077A
picogram/kilogram	pg/kg	pg/kg, picogram/kg, picogr/kg, G081A
fem-	fg/kg	fg/kg, femtogram/kg, femtogr/kg
togram/kilogram		
gram	g	g, gram, gr, G148A
milligram	mg	mg, milligram, milligr, G155A
microgram	μg	µg, microgram, microgr
nanogram	ng	ng, nanogram, nanogr, G120A
picogram	<u>pg</u>	pg, picogram, picogr, G125A
femtogram	fg	fg, femtogram, femtogr

Table 7.12 – continued from previous page

# **7.3.13 Harvest application types**

Available harvest application types.





# **7.3.14 Hazard characterisation types**

Known hazard characterisation types.





# **7.3.15 Point of departure types**

Known point of departure types.

Name	Short name	<b>Aliases</b>	Description
Benchmark	<b>BMD</b>	<b>BMD</b>	
dose			
No observed	<b>NOAEL</b>	<b>NOAEL</b>	
adverse effect			
level			
Lowest	<b>LOAEL</b>	<b>LOAEL</b>	
observed			
adverse effect			
level			
No observed	<b>NOEL</b>	<b>NOEL</b>	
effect level			
LD50	LD50	LD50	Median lethal dose.

Table 7.15: Unit definition for Point of departure types.

# **7.3.16 Response types**

Available response types.

Name	Short name	Aliases	Description
Continuous	CM.	Continuous-	Response values are positive real numbers,
multiplicative		Multiplicative	e.g., weight, size.
Continuous	CA.	ContinuousAd-	Response values are real numbers, e.g.,
additive		ditive	weight change, temperature.
Binary	B	Binary	Response values have binary outcomes
			(yes/no, true/false, success/failure, 0/1, etc.).
<b>Ouantal</b>	Q	Ouantal,	Response is measured in terms of number of
		<b>Binomial</b>	successes out of N possible.
Quantal group	OG	QuantalGroup	Individual responses are measured as binary
			values, which may be grouped to form a
			quantal response.
Count	$\mathcal{C}$	Count	Number of items (cells, molecules, deaths,
			etc.) in given interval/area/volume.
Ordinal	O	Ordinal	Relative scores (or graded scores) useable
			only for ranking.

Table 7.16: Unit definition for Response types.

## **7.3.17 Target dose level types**

This unit specifies whether a dose is assumed to be an internal or external dose.





## **7.3.18 Test system types**

Available test system types.

Table 7.18: Unit definition for Test system types.

Name	Short name	<b>Aliases</b>	Description
In vivo	In vivo	InVivo	In vivo
Cell line	Cell line	CellLine	CellLine
Primary cells	Primary cells	PrimaryCells	PrimaryCells
<b>Tissue</b>	Tissue	Tissue	<b>Tissue</b>
Organ	Organ	Organ	Organ

# **7.4 Transformations**

## **7.4.1 Box Cox power transformation**

The Box-Cox power transformation is a data transformation to achieve a better normality and to stabilize the variance. In MCRA, the transformation parameter  $p$  in  $(\psi^p - 1)/p$  is determined by maximizing the log-likelihood function

$$
l(p) = -\frac{n}{s} \log \left[ \frac{1}{n} \sum_{i=1}^n (y_i^{(p)} - \overline{y^{(p)}})^2 \right] + (p-1) \sum_{i=1}^n \log y_i
$$

where  $i$  indexes the  $n$  observations and

$$
\overline{y^{(p)}} = \frac{1}{n} \sum_{i=1}^{n} y_i^{(p)}
$$

is the average of the  $y_i^{(p)}$  (Box & Cox, 1964) [Box et al., 1964].

# **7.5 Gauss-Hermite**

## **7.5.1 Gauss-Hermite integration**

## **7.5.2 One-dimensional Gauss-Hermite integration**

Gauss-Hermite integration approximates a specific integral as follows

$$
\int\limits_{-\infty}^{\infty}f(x)\exp(-x^2){\rm d}x\approx\sum\limits_{j=1}^Nw_jf(x_j)
$$

in which  $w_j$  and  $x_j$  are weights and abscissas for N-point Gauss-Hermite integration, see Abramowitz and Stegun (1972) [Abramowitz, 1972]. N-point integration is exact for all polynomials  $f(x)$  of degree 2N-1, see Dahlquist and Björck (1974) [Dahlquist et al., 1974]. This can for instance be used to approximate the mean of a function  $F(Y)$ of a normally distributed random variable Y with mean  $\mu$  and variance  $\sigma^2$ :

$$
\int_{-\infty}^{\infty} F(x) \frac{1}{\sqrt{2\pi\sigma}} \exp\left(-\frac{(y-\mu)^2}{2\sigma^2}\right) dy
$$

$$
= \int_{-\infty}^{\infty} F(\mu + \sqrt{2}\sigma x) \frac{1}{\sqrt{\pi}} \exp(-x^2) dx
$$

$$
= \frac{1}{\sqrt{\pi}} \sum_{j=1}^{N} w_j F(\mu + \sqrt{2}\sigma x_j)
$$

## **7.5.3 Two-dimensional Gauss-Hermite integration**

One-dimensional Gauss-Hermite integration can readily be extended to two dimensions. The following principal result in two dimensions is more or less given in Jäckel (2005) [Jäckel, 2005] for the standard bivariate normal distribution  $\phi(x, y; \rho)$  with correlation parameter  $\rho$ :

$$
\int\limits_{-\infty}^{\infty}\int\limits_{-\infty}^{\infty}F(x,y)\phi(x,y;\rho)\mathrm{d}x\mathrm{d}y \approx \frac{1}{\pi}\sum\limits_{i=1}^{N}\sum\limits_{j=1}^{N}w_{i}w_{j}F(\sqrt{2}[ax_{i}+bx_{j}],\sqrt{2}[bx_{i}+ax_{j}])
$$

in which

$$
a=\frac{\sqrt{1+\rho}+\sqrt{1-\rho}}{2}
$$

and

$$
b = \frac{\sqrt{1+\rho} - \sqrt{1-\rho}}{2}
$$

as given in Jäckel (2005) [Jäckel, 2005] .

Jäckel (2005) discusses other Gauss-Hermite approximations to the two-dimensional integral, but found that the approximation given above generally gives the most accurate results. For the general bivariate normal distribution with means  $(\mu_x, \mu_y)$  and [variances](#page-355-0)  $(\sigma_x^2, \sigma_y^2)$  the integral can be approximated by means of

$$
\frac{1}{\pi}\sum_{i=1}^N\sum_{j=1}^N w_iw_jF(\mu_x+\sigma_x\sqrt{2}[ax_i+bx_j],\mu_y+\sigma_y\sqrt{2}[bx_i+ax_j])
$$

The product  $w_i w_j$  can be very small, especially when many quadrature points are used, thus wasting possibly precious calculation time. This can be remedied by pruning, i.e. by dropping combinations of  $(i, j)$  with very small values of the product  $w_i w_j$ .

## **7.5.4 Maximum likelihood for the LNN model with two-dimensional Gauss-Hermite integration**

Denote non-consumption on day j for individual i as  $Y_{ij} = 0$ . The conditional likelihood, i.e. given random effects  $b_i$  and  $v_i$ , of a non-consumption on day j equals, with  $H()$  the inverse of the logit function

$$
P(Y_{ij} = 0 | b_i, v_i) = 1 - H(\lambda + v_i).
$$

The conditional likelihood of a positive intake  $Y_{ij} > 0$  equals, with  $\phi$  the density of the normal distribution

$$
f(Y_{ij} = y_{ij} | y_{ij} > 0, b_i, v_i) = H(\lambda + v_i) \phi(y_{ij} - \mu - b_i; 0, \sigma_w^2)
$$

The conditional likelihood contribution for individual  $i$  is the product of the individual contributions for each day. The marginal likelihood contribution for individual  $i$  is obtained by integrating over the possible values of  $b_i$  and  $v_i$ . Since the pair  $(b_i, v_i)$  follows a bivariate normal distribution, the likelihood contribution for individual i can be approximated by means of two-dimensional Gauss-Hermite integration. Individually based covariables, such as sex or age, imply that  $\mu_i$  and  $\lambda_i$  must be used instead of  $\mu$  and  $\lambda$ . The likelihood must be optimized by means of some general optimization routine.

# **CHAPTER EIGHT**

# **COLOPHON**



**WUR/Biometris, Wageningen University & Research FERA, Food and Environmental Research Agency RIVM, National Institute for Public Health and the Environment**

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# **8.1 Contributors to MCRA**

MCRA development team: Waldo de Boer, Johannes Kruisselbrink, Marco van Lenthe, Hans van den Heuvel, Hilko van der Voet

Many people contributed to the MCRA code over the years:

Frits van Evert, Jack van Galen, Paul Goedhart, Gerie van der Heijden, Paul Keizer, Marcel Koenders, Jaap Kokorian, Sanne Korzec, Helen Owen, Gerrit Polder, Pim Reijersen, Willem Roelofs, Gert-Jan Swinkels, Jac Thissen.

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Contributors to the MCRA Reference Manual: Waldo de Boer, Paul Goedhart, Andy Hart, Marc Kennedy, Johannes Kruisselbrink, Helen Owen, Willem Roelofs, Hans van den Heuvel, Hilko van der Voet.

## **WUR/Biometris** is the unit for **Mathematical and Statistical Methods of Wageningen University & Research** P.O. Box 16, 6700 AA Wageningen, Netherlands

WUR Campus, Building 107 (Radix), Droevendaalsesteeg 1, 6708 PB Wageningen Telephone: +31 (0)317 476925 https://wur.nl http://www.biometris.nl

## **[Fera Food an](https://wur.nl)d Environmental Research Agency**

[Sand Hutton, York, YO4](http://www.biometris.nl)1 1LZ, United Kingdom Telephone: +44 (0)1904 462000 https://www.fera.co.uk

## **RIVM National Institute for Public Health and the Environment**

[P.O. Box 1, 3729 BA B](https://www.fera.co.uk)ilthoven, Netherlands Antonie van Leeuwenhoeklaan 9, 3721 MA Bilthoven Telephone: +31 30 2749111 https://rivm.nl

## **BIBLIOGRAPHY**

- [Abramowitz, 1972] Milton Abramowitz and Irene A. Stegun. Handbook of mathematical functions. *National Bureau of Standards Applied Mathematics Series*, 55:589–626, 1972.
- [Bopp et al., 2015] Stephanie Bopp and BERGGREN ELISABET; KIENZLER AUDE; VAN DER LINDEN SANDER; WORTH Andrew. Scientific methodologies for the assessment of combined effects of chemicals - a survey and literature review. *EUR - Scientific and Technical Research Reports*, 2015. doi:10.2788/093511.
- [Box et al., 1964] George EP Box and David R Cox. An analysis of transformations. *Journal of the Royal Statistical Society: Series B (Methodological)*, 26(2):211–243, 1964.
- [Butler et al., 2018] [M.C. Butle](https://doi.org/10.2788/093511)r Ellis, Marc C. Kennedy, C.J. Kuster, R. Alanis, and C.R. Tuck. Improvements in modelling bystander and resident exposure to pesticide spray drift: investigations into new approaches for characterizing the 'collection efficiency'of the human body. *Annals of work exposures and health*, 62(5):622–632, 2018. doi:10.1093/annweh/wxy017.
- [Béchaux et al., 2013] Camille Béchaux, Mélanie Zetlaoui, Jessica Tressou, Jean-Charles Leblanc, Fanny Héraud, and Amélie Crépet. Identification of pesticide mixtures and connection between combined exposure and diet. *Food and chemical toxicology*[, 59:191–198, 201](https://doi.org/10.1093/annweh/wxy017)3.
- [Cleo et al., 2019] Tebby Cleo, Hilko van der Voet, Georges de Sousa, Emiel Rorije, Vikas Kumar, Waldo de Boer, Johannes H. Kruiselbrink, Frédéric Y. Bois, Moosa Faniband, Angelo Moretto, and Céline Brochot. In prep: integration of in silico and in vitro data in PBPK modeling for risk assessment of food- and nonfood-borne chemicals using the Euromix toolbox. *xxx*, 2019.
- [Cramer et al., 1976] G.M. Cramer, R.A. Ford, and R.L. Hall. Estimation of toxic hazard—a decision tree approach. *Food and cosmetics toxicology*, 16(3):255–276, 1976. doi:10.1016/S0015-6264(76)80522-6.
- [Dahlquist et al., 1974] G Dahlquist and A Bjorck. Numerical methods (transl. by n. anderson). 1974.
- [de Boer et al., 2009] Waldo J de Boer, Hilko van der Voet, Bas GH Bokkers, Martine I Bakker, and Polly E Boon. Comparison of two models for the estimation of us[ual intake addressing zero consumptio](https://doi.org/10.1016/S0015-6264(76)80522-6)n and nonnormality. *Food Additives and Contaminants*, 26(11):1433–1449, 2009.
- [de Boer et al., 2011] Waldo J de Boer and van der Voet. Mcra 7. a web-based program for monte carlo risk assessment. reference manual 2011-12-19, documenting mcra release 7.1. Technical Report, Biometris, Wageningen UR and National Institute for Public Health and the Environment (RIVM), Bilthoven, Wageningen., 2011. URL: https//mcra.rivm.nl.
- [Dodd, 1996] KW Dodd. A technical guide to c-side. Ames, *Iowa: Department of Statistics and Center for Agricultural and Rural Development, Iowa State University*, 1996.
- [EC, 2018] European Commissio[n Standing Committe](https//mcra.rivm.nl)e on Plants Animals Food and Feed. European commission working document sante-2015-10216 rev. 7. 2018.
- [Efron, 1979] B Efron. Bootstrap methods: another look at the jackknife annals of statistics 7: 1–26. *View Article PubMed/NCBI Google Scholar*, 1979.
- [Efron et al., 1993] Bradley Efron and Robert J Tibshirani. An introduction to the bootstrap chapman & hall. *New York*, 1993.
- [EFSA, 2011a] European Food Safety Authority (EFSA). Report on the development of a food classification and description system for exposure assessment and guidance on its implementation and use. *EFSA Journal*, 9(12):84, 2011. doi:doi:10.2903/j.efsa.2011.2489.
- [EFSA, 2011b] European Food Safety Authority (EFSA). The food classification and description system foodex 2 (draft-revision 1). *EFSA Journal*, pages 438, 2011.
- [EFSA, 2012] European [Food Safety Authority \(EFSA\).](https://doi.org/doi:10.2903/j.efsa.2011.2489) Guidance on the use of probabilistic methodology for modelling dietary exposure to pesticide residues. *EFSA Journal*, 10(10):2839, 2012. doi:10.2903/j.efsa.2012.2839.
- [EFSA, 2014] European Food Safety Authority (EFSA). Guidance on the assessment of exposure of operators, workers, residents and bystanders in risk assessment for plant protection products. *EFSA Journal*, [12\(10\):3874, 2014.](https://doi.org/10.2903/j.efsa.2012.2839) doi:10.2903/j.efsa.2014.3874.
- [EFSA, 2017] European Food Safety Authority (EFSA). Guidance on dermal absorption. *EFSA Journal*, 6 2017. doi:10.2903/j.efsa.2017.4873.
- [EFSA, 2018] European Food [Safety Authority \(EFSA\), Al](https://doi.org/10.2903/j.efsa.2014.3874)ba Brancato, Daniela Brocca, Lucien Ferreira, Luna Greco, Samira Jarrah, Renata Leuschner, Paula Medina, Ileana Miron, Alexandre Nougadere, Ragnor Pedersen, Hermine Reich, Miguel Santos, Alois Stanek, Jose Tarazona, Anne Theobald, and Laura [Villamar-Bouza. Use of efsa](https://doi.org/10.2903/j.efsa.2017.4873) pesticide residue intake model (efsa primo revision 3). *EFSA Journal*, 16(1):e05147, 2018. URL: https://efsa.onlinelibrary.wiley.com/doi/abs/10.2903/j.efsa.2018.5147.
- [EFSA, 2020a] European Food Safety Authority (EFSA), Peter S Craig, Bruno Dujardin, Andy Hart, Antonio F Hernández-Jerez, Susanne Hougaard Bennekou, Carsten Kneuer, Bernadette Ossendorp, Ragnor Pedersen, Gerrit Wolterink, and [Luc Mohimont. Cumulative dietary risk characterisation of pesticides that](https://efsa.onlinelibrary.wiley.com/doi/abs/10.2903/j.efsa.2018.5147) have acute effects on the nervous system. *EFSA Journal*, 18(4):e06087, 2020. URL: https://efsa.onlinelibrary. wiley.com/doi/abs/10.2903/j.efsa.2020.6087.
- [EFSA, 2020b] European Food Safety Authority (EFSA), Peter S Craig, Bruno Dujardin, Andy Hart, Antonio F Hernandez-Jerez, Susanne Hougaard Bennekou, Carsten Kneuer, Bernadette Ossendorp, Ragnor Pedersen, Gerrit Wolterink, and Luc Mohimont. Cumulative dietary risk characte[risation of pesticides that](https://efsa.onlinelibrary.wiley.com/doi/abs/10.2903/j.efsa.2020.6087) [have chronic effects on the thyroid.](https://efsa.onlinelibrary.wiley.com/doi/abs/10.2903/j.efsa.2020.6087) *EFSA Journal*, 18(4):e06088, 2020. URL: https://efsa.onlinelibrary. wiley.com/doi/abs/10.2903/j.efsa.2020.6088.
- [Gillis et al., 2013] Nicolas Gillis and Robert J Plemmons. Sparse nonnegative matrix underapproximation and its application to hyperspectral image analysis. *Linear Algebra and its Applications*[, 438\(10\):3991–4007,](https://efsa.onlinelibrary.wiley.com/doi/abs/10.2903/j.efsa.2020.6088) 2013.
- [Goedhart et al., 2012] [Paul W. Goedhart, Hilko. van der](https://efsa.onlinelibrary.wiley.com/doi/abs/10.2903/j.efsa.2020.6088) Voet, S. Knüppel, Arnold L.M. Dekkers, Kevin W. Dodd, Hermann. Boeing, and Jacob D. van Klaveren. A comparison by simulation of different methods to estimate the usual intake distribution for episodically consumed foods. Technical Report, Report: Supporting Publications 2012:EN-299, 2012. URL: http//www.efsa.europa.eu/publications.
- [Goodhardt et al., 1984] Gerald Joseph Goodhardt, Andrew SC Ehrenberg, and Christopher Chatfield. The dirichlet: a comprehensive model of buying behaviour. *Journal of the Royal Statistical Society. Series A (General)*, pages 621–655, 1984.
- [Hoyer, 2004] Patrik O Hoyer. Non-negative matrix factorization with sparseness constraints. *Journal of machine learning research*, 5(Nov):1457–1469, 2004.
- [Jäckel, 2005] Peter Jäckel. A note on multivariate gauss-hermite quadrature. *London: ABN-Amro. Re*, 2005.
- [Karrer et al., 2019] Karrer, Cecile, Waldo de Boer, Christiaan Delmaar, Yaping Cai, Amélie Crépet, Konrad Hungerbühler, and Natalie van Goetz. In prep: linking probabilistic exposure and pharmacokinetic modeling to assess the cumulative risk from the bisphenols BPA, BPS, BPF, and BPAF for europeans. *xxx*, 2019.
- <span id="page-355-0"></span>[Kennedy et al., 2012] Marc C. Kennedy, Clare M.J. Butler Ellis, and Paul C.H. Miller. Bream: a probabilistic bystander and resident exposure assessment model of spray drift from an agricultural boom sprayer. *Computers and electronics in agriculture*, 88:63–71, 2012. doi:10.1016/j.compag.2012.07.004.
- [Kennedy et al., 2015a] Marc C Kennedy, C Richard Glass, Bas Bokkers, Andy DM Hart, Paul Y Hamey, Johannes W Kruisselbrink, Waldo J de Boer, Hilko van der Voet, David G Garthwaite, and Jacob D van

Klaveren. A european model and case studies for aggregate exposure assessment of pesticides. *Food and Chemical Toxicology*, 79:32–44, 2015.

- [Kennedy et al., 2015b] Marc C Kennedy, Hilko van der Voet, Victoria J. Roelofs, Willem Roelofs, C. Richard Glass, Waldo J de Boer, Johannes W. Kruisselbrink, and Andy D.M. Hart. New approaches to uncertainty analysis for use in aggregate and cumulative risk assessment of pesticides. *Food and Chemical Toxicology*, 79:54–64, 2015.
- [Kennedy et al., 2017] Marc C. Kennedy and M.C. Butler Ellis. Probabilistic modelling for bystander and resident exposure to pesticides using the browse software. *Biosystems engineering*, 154:105–121, 2017. doi:10.1016/j.biosystemseng.2016.08.012.
- [Kennedy, 2019] Marc C. et al. Kennedy. In prep.: a retain and refine approach for grouping chemicals into cumulative exposure assessment. *xxx*, 2019.
- [Kipnis et al., 2009] [Victor Kipnis, Douglas Midthune](https://doi.org/10.1016/j.biosystemseng.2016.08.012), Dennis W Buckman, Kevin W Dodd, Patricia M Guenther, Susan M Krebs-Smith, Amy F Subar, Janet A Tooze, Raymond J Carroll, and Laurence S Freedman. Modeling data with excess zeros and measurement error: application to evaluating relationships between episodically consumed foods and health outcomes. *Biometrics*, 65(4):1003–1010, 2009.
- [Lee et al., 1999] Daniel D Lee and H Sebastian Seung. Learning the parts of objects by non-negative matrix factorization. *Nature*, 401(6755):788, 1999.
- [Mood et al., 1974] Alexander McFarlane Mood, Franklin A Graybill, and Duane C Boes. *Introduction to the Theory of Statistics 1974*. McGraw-Hill Kogakusha, 1974.
- [Munro et al., 1996] Ian C. Munro, Richard A. Ford, Elke Kennepohl, and James G. Sprenger. Correlation of structural class with no-observed-effect levels: a proposal for establishing a threshold of concern. *Food and Chemical Toxicology*, 34(9):829–867, 1996. doi:10.1016/S0278-6915(96)00049-X.
- [Nusser et al., 1996] Sarah M Nusser, Alicia L Carriquiry, Kevin W Dodd, and Wayen A Fuller. A semiparametric transformation approach to estimating usual daily intake distributions. *Journal of the American Statistical Association*, 91(436):1440–1449, 1996.
- [Nusser et al., 1997] Sarah M Nusser, Wayne A Fuller[, Patricia M Guenther, and others. Esti](https://doi.org/10.1016/S0278-6915(96)00049-X)mating usual dietary intake distributions: adjusting for measurement error and nonnormality in 24-hour food intake data. Technical Report, Center for Agricultural and Rural Development (CARD) at Iowa State University, 1997.
- [Price et al., 2011] Paul S Price and Xianglu Han. Maximum cumulative ratio (mcr) as a tool for assessing the value of performing a cumulative risk assessment. *International journal of environmental research and public health*, 8(6):2212–2225, 2011.
- [Saul et al., 2002] Lawrence K Saul and Daniel D Lee. Multiplicative updates for classification by mixture models. In *Advances in Neural Information Processing Systems*, 897–904. 2002.
- [Slob, 2006] Wout Slob. Probabilistic dietary exposure assessment taking into account variability in both amount and frequency of consumption. *Food and Chemical Toxicology*, 44(7):933–951, 2006.
- [Slob et al., 2010] Wout Slob, Waldo J de Boer, and Hilko van der Voet. Can current dietary exposure models handle aggregated intake from different foods? a simulation study for the case of two foods. *Food and chemical toxicology*, 48(1):178–186, 2010.
- [Souverein et al., 2011] Olga W. Souverein, Waldo J. de Boer, Anouk Geelen, Hilko van der Voet, Jeanne H. de Vries, Max Feinberg, and Pieter van't Veer. Uncertainty in intake due to portion size estimation in 24-hour recalls varies between food groups. *The Journal of nutrition*, 141(7):1396–1401, 2011. doi:10.3945/jn.111.139220.
- [Tooze et al., 2006] Janet A Tooze, Douglas Midthune, Kevin W Dodd, Laurence S Freedman, Susan M Krebs-Smith, Amy F Subar, Patricia M Guenther, Raymond J Carroll, and Victor Kipnis. A new statistical method for estimating the usual intake of episodically consumed foods with application to their distribution. *[Journal of the America](https://doi.org/10.3945/jn.111.139220)n Dietetic Association*, 106(10):1575–1587, 2006.
- [van den Berg et al., 2016] F. van den Berg, C.M.J. Jacobs, M.C. Butler Ellis, P. Spanoghe, K. Doan Ngoc, and G. Fragkoulis. Modelling exposure of workers, residents and bystanders to vapour of plant pro-

tection products after application to crops. *Science of the Total Environment*, 573:1010–1020, 2016. doi:10.1016/j.scitotenv.2016.08.180.

- [van der Voet et al., 2007] Hilko van der Voet and Wout Slob. Integration of probabilistic exposure assessment and probabilistic hazard characterization. *Risk Analysis: An International Journal*, 27(2):351–371, 2007. [doi:10.1111/j.1539-6924.2007.0088](https://doi.org/10.1016/j.scitotenv.2016.08.180)7.x.
- [van der Voet et al., 2009] Hilko van der Voet, Gerie W.A.M. van der Heijden, Peter M.J. Bos, Sieto Bosgra, Polly E. Boon, Stefan D. Muri, and Beat J. Brüschweiler. A model for probabilistic health impact assessment of exposure to food chemicals. *Food and Chemical Toxicology*, 47(12):2926–2940, 2009. [doi:10.1016/j.fct.2008.12.027.](https://doi.org/10.1111/j.1539-6924.2007.00887.x)
- [van Klaveren et al., 2019a] J.D. van Klaveren, J.W. Kruisselbrink, W.J. de Boer, G. van Donkersgoed, J.D. te Biesebeek, M. Sam, and H. van der Voet. Cumulative dietary exposure assessment of pesticides that have acute effects on the nervous system using mcra software. *EFSA Supporting Publications*, 16(9):1708E, 2019. URL: [https://efsa.online](https://doi.org/10.1016/j.fct.2008.12.027)library.wiley.com/doi/abs/10.2903/sp.efsa.2019.EN-1708.
- [van Klaveren et al., 2019b] J.D. van Klaveren, J.W. Kruisselbrink, W.J. de Boer, G. van Donkersgoed, J.D. te Biesebeek, M. Sam, and H. van der Voet. Cumulative dietary exposure assessment of pesticides that have chronic effects on the thyroid using mcra software. *EFSA Supporting Publications*, 16(9):1707E, 2019. URL: [https://efsa.onlinelibrary.wiley.com/doi/abs/10.2903/sp.efsa.2019.EN-1707](https://efsa.onlinelibrary.wiley.com/doi/abs/10.2903/sp.efsa.2019.EN-1708).
- [Verkaik-Kloosterman et al., 2011] Janneke Verkaik-Kloosterman, Kevin W Dodd, Arnold LM Dekkers, Pieter van't Veer, and Marga C Ocké. A three-part, mixed-effects model to estimate the habitual total vitamin d intak[e distribution from food and dietary supplements in dutch young children.](https://efsa.onlinelibrary.wiley.com/doi/abs/10.2903/sp.efsa.2019.EN-1707) *The Journal of nutrition*, 141(11):2055–2063, 2011.
- [WHO, 2018] World Health Organization (WHO). *Guidance document on evaluating and expressing uncertainty in hazard characterization*. World Health Organization, 2018.
- [Zetlaoui et al., 2011] Mélanie Zetlaoui, Max Feinberg, Philippe Verger, and Stephan Clémençon. Extraction of food consumption systems by nonnegative matrix factorization (nmf) for the assessment of food choices. *Biometrics*, 67(4):1647–1658, 2011.

# **HTTP ROUTING TABLE**

/Api GET /Api/About/VersionInfo, 324 GET /Api/Actions/DownloadActionZip/{id}, GET /Api/Outputs/DownloadSectionCsvZip, 316 GET /Api/Actions/DownloadActionZipNoDat& $\frac{1}{2}$ {id), Api/Outputs/DownloadSectionPdf, 316 GET /Api/Ac[tio](#page-331-0)ns/DownloadActionZipOriginalData/(Qutputs/DownloadShortReportPdf, [316](#page-323-0) GET /Api/Actions/ExportActionDataSourceG<sup>E</sup>\{id}. 0utputs/DownloadTableCsv, 328 [316](#page-323-1) GET /Api/Actions/ExportActionSettings/{ $\frac{1}{2}$ , Api/Outputs/GetFromAction/{idAction}, [315](#page-323-2) GET /Api/Actions/Get/{id}, 311 GET /Ap[i/A](#page-323-3)ctions/GetAll/{idWorkspace}, 311 GET /Ap[i/A](#page-322-0)ctions/GetDataLinkingEntityRecords/ $\{^{325}_{13}\}$ , 314 GET /Api/Actions/GetDataReadingEntityRecords/ $\{^{326}_{13}\}$ , [314](#page-318-0) GET /Api/Actions/GetDataReadingSummary/{id}, [313](#page-321-0) GET /Api/Actions/GetDataSelection/{id}, [314](#page-321-1) GET /Api/Actions/GetSummary/{id}, 313 GET /Ap[i/D](#page-320-0)ataSources/DownloadVersion/{id\efsion/\Qutputs/GetShortOutputSummary, 320 GET /Ap[i/D](#page-321-2)ataSources/DownloadVersionCsv9\{Id\Api/:Outputs/GetTaskSettingsSection, 320 GET /Api/DataSources/Get/{id}, 317 GET /A[pi/D](#page-327-0)ataSources/GetAll, 317 GET /Api/DataSources/GetDataSourceUsage/{idDat<sup>321</sup>ource}, [320](#page-327-1) GET /Api/DataSo[urc](#page-324-0)es/GetDataSourceVersionUsage<sup>322</sup>idVersion}, 319 GET /Api/DataSources/GetRemo[teR](#page-324-1)epositoryDataSources/{id}, [317](#page-327-2) GET /Api/DataSources/GetRepositoryDataSources $\beta^{\mathfrak{Z} 1}_{{\bf{1d}}}\},$ [317](#page-326-0) GET /Api/DataSources/GetVersion/{idVerstET}/A[pi/T](#page-329-0)asks/GetActiveTaskStatuses/{idWorkspace}, [319](#page-324-2) GET /Api/DataSources/GetVersions/{id}, [319](#page-324-3) GET /Api/DataSources/GetWorkspaceDataSo&\Pte=\\&\Pi{G&\{\&\\\$\{\&\\$\&QQ\{id},32[9](#page-336-0) [319](#page-326-1) GET /Api/Outputs/DownloadFullPdf, 327 GET /A[pi/O](#page-326-2)utputs/DownloadRawData, 328 GET /Api/Outputs/DownloadReportCsvZip, 328 328 [327](#page-335-0) [327](#page-335-1) GET /A[pi/O](#page-334-0)utputs/Get/{id}, 325 [325](#page-334-1) GET /Api/Outputs/GetOutputReportCh[art](#page-335-2), 327 GET /Api/Outputs/GetOutput[Rep](#page-332-0)ortSection, GET /Api/Outputs/GetOutputReportTableContent, GET /Api/Outputs/GetOutputReportTableData, [326](#page-332-2) GET /Api/Outputs/GetOutputReportToc/{id}, [325](#page-333-0) GET /Api/Outputs/GetOutputReportTocs, [325](#page-333-1) [326](#page-332-3) [327](#page-332-4) GET /Api/Repositories/Get/{id}, 321 GET /A[pi/R](#page-333-2)epositories/GetDetails/{id}, GET /A[pi/R](#page-334-3)epositories/GetRepositoryDataSourceUsage/{ GET /Api/Repositories/GetRepositoryDataSourcesInUse/{id}, GET /Api/Tasks/DownloadTaskZip/{id}, GET /Api/Tasks/Get/{idWorkspace}, 329 330 /[Api](#page-338-0)/Tasks/GetProgress/{taskId}, 330 GET /A[pi/W](#page-337-0)orkspace/Get/{id}, 323 GET /Api/Workspace/GetAll, 323 GET /A[pi/W](#page-337-1)orkspace/GetDataSources/{id},

324 POST /Api/Actions/Clone, 311 POST /Api/Actions/Create, 311 POST /Api/Actions/UploadActionZipFile/{GEWorksppaRepositories/GetAll, 321 [312](#page-331-1) POST /Api/DataSources/Im[por](#page-318-1)tRemoteDataSGETceppi/StandardActionTypes/Get,331 317 POST /Api/DataSources/Upl[oad](#page-318-2)ActionZipFil@\$T /api/Repositories/AddGr[oup](#page-328-2)Share, [320](#page-319-0) POST /Api/DataSources/UploadNewDataSour**P@ST**idReppositories/A[ddU](#page-336-2)serS[har](#page-338-1)e, [318](#page-324-4) POST /Api/DataSources/UploadNewDataSour**P@V**er*/sa*ipon//Report backsoniers/C}hange[Own](#page-336-3)er, 322 [318](#page-327-3) POST /Api/Repositories/Delete/{id}, 321 POST /api/Repositories/Move, 322 POST /A[pi/](#page-325-0)Repositories/ForceDelete/{id}, POST /[api/](#page-329-3)Repositories/RemoveGroupShare, 322 POST /A[pi/](#page-325-1)Tasks/ScheduleSubTask/{idActiB@\$T /api/Repositories/RemoveU[ser](#page-328-3)S[hare](#page-329-4), 330 POST /Api/Tasks/ScheduleTask/{idAction}, POST /api/Repositories/Update, 321 [330](#page-329-1) POST /Api/Workspace/AddDataSource/{id}, [324](#page-337-2) POST /Api/Workspace/Create, 323 POST /A[pi/](#page-337-3)Workspace/RemoveDataSource/{id}, 324 POST /[Api/](#page-331-2)Workspace/Update/{id}, 323 PUT /Api/Actions/ClearScop[e/{i](#page-330-1)d}, 315 PUT /Api/Actions/ConvertToCustomAction/{id}, [312](#page-331-3) PUT /Api/Actions/ExtendScope/{i[d}](#page-330-2), 315 PUT /Api/Actions/ReduceScope/{id}, [31](#page-322-1)5 PUT /Api/Actions/ReplaceActionDataSource/{id}, [313](#page-319-1) PUT /Api/Actions/SetActionDataSour[ce/](#page-322-2){id}, 313 PUT /Api/Actions/SetIsCompute/{id}, [313](#page-322-3) PUT /A[pi/A](#page-320-1)ctions/SetScope/{id}, 315 PUT /Api/Actions/UpdateMetaData/{id}, [312](#page-320-2) PUT /Api/Actions/UpdateStandardAct[ionV](#page-320-3)ersion/{id}, 312 PUT /Api/DataSources/Move/{id}, 318 PUT /A[pi/D](#page-319-2)ataSources/Rename/{id}, 318 PUT /Api/Tasks/Abort/{id}, 330 PUT /Ap[i/T](#page-319-3)asks/UpdateDescription/{id}, 329 DELETE /Api/Actions/Delete/{id}, 3[12](#page-325-2) DELETE /Api/DataSources/[Del](#page-337-4)ete/{id}, 318 DELETE [/Ap](#page-336-1)i/Tasks/BatchDelete, 331 DELETE /Api/Tasks/Delete/{id}, 33[1](#page-319-4) DELETE /Api/Workspace/Delete/{id}, 323 /api GET /api/ActionClasses/Get, 332 GET /api/ActionTypes/Get, 324 GET /api/DataFormats/Get, 324 GET /api/DataSources/FileUploadProgress, 321 GET /api/ScopingTypes/Get, [32](#page-331-4)9 GET /a[pi/U](#page-328-1)nitDefinitions/Get, 329 322 322 POST /[api/](#page-329-2)Repositories/Create, 321 322 322 POST /[api/](#page-329-5)Repositories/UpdateGroupShare, 322 POST /[api/](#page-329-6)Repositories/UpdateUserShare, 322